



Root and Rhizosphere Bacterial Phosphatase Activity Varies with Tree Species and Soil Phosphorus Availability in Puerto Rico Tropical Forest

Kristine G. Cabugao^{1,2,3*}, Collin M. Timm⁴, Alyssa A. Carrell⁴, Joanne Childs^{2,3}, Tse-Yuan S. Lu⁴, Dale A. Pelletier⁴, David J. Weston^{3,4} and Richard J. Norby^{2,3}

¹ Bredesen Center for Interdisciplinary Research and Graduate Education, University of Tennessee, Knoxville, TN, United States, ² Environmental Sciences Division, Oak Ridge National Laboratory, Oak Ridge, TN, United States, ³ Climate Change Science Institute, Oak Ridge National Laboratory, Oak Ridge, TN, United States, ⁴ Biosciences Division, Oak Ridge National Laboratory, Oak Ridge, TN, United States

OPEN ACCESS

Edited by:

Pierre-Emmanuel Courty, Institut National de la Recherche Agronomique (INRA), France

Reviewed by:

Megan K. Nasto, Utah State University, United States Luis Sampedro, Consejo Superior de Investigaciones Científicas (CSIC), Spain

> *Correspondence: Kristine G. Cabugao cabugaokm@ornl.gov

Specialty section:

This article was submitted to Plant Microbe Interactions, a section of the journal Frontiers in Plant Science

Received: 07 July 2017 Accepted: 10 October 2017 Published: 30 October 2017

Citation:

Cabugao KG, Timm CM, Carrell AA, Childs J, Lu T-YS, Pelletier DA, Weston DJ and Norby RJ (2017) Root and Rhizosphere Bacterial Phosphatase Activity Varies with Tree Species and Soil Phosphorus Availability in Puerto Rico Tropical Forest. Front. Plant Sci. 8:1834. doi: 10.3389/fpls.2017.01834 Tropical forests generally occur on highly weathered soils that, in combination with the immobility of phosphorus (P), often result in soils lacking orthophosphate, the form of P most easily metabolized by plants and microbes. In these soils, mineralization of organic P can be the major source for orthophosphate. Both plants and microbes encode for phosphatases capable of mineralizing a range of organic P compounds. However, the activity of these enzymes depends on several edaphic factors including P availability, tree species, and microbial communities. Thus, phosphatase activity in both roots and the root microbial community constitute an important role in P mineralization and P nutrient dynamics that are not well studied in tropical forests. To relate phosphatase activity of roots and bacteria in tropical forests, we measured phosphatase activity in roots and bacterial isolates as well as bacterial community composition from the rhizosphere. Three forests in the Luquillo Mountains of Puerto Rico were selected to represent a range of soil P availability as measured using the resin P method. Within each site, a minimum of three tree species were chosen to sample. Root and bacterial phosphatase activity were both measured using a colorimetric assay with para-nitrophenyl phosphate as a substrate for the phosphomonoesterase enzyme. Both root and bacterial phosphatase were chiefly influenced by tree species. Though tree species was the only significant factor in root phosphatase activity, there was a negative trend between soil P availability and phosphatase activity in linear regressions of average root phosphatase and resin P. Permutational multivariate analysis of variance of bacterial community composition based on 16S amplicon sequencing indicated that bacterial composition was strongly controlled by soil P availability (p-value < 0.05). These results indicate that although root and bacterial phosphatase activity were influenced by tree species; bacterial community composition was chiefly influenced by P availability. Although the sample size is limited

1

given the tremendous diversity of tropical forests, our study indicates the importance of roots and bacterial function to understanding phosphatase activity. Future work will broaden the diversity of tree species and microbial members sampled to provide insight into P mineralization and model representation of tropical forests.

Keywords: nutrient acquisition, phosphatase, phosphorus, plant growth promoting rhizobacteria, plant microbiome, root functional traits

INTRODUCTION

Tropical forests annually accumulate the highest amount of biomass among terrestrial ecosystems, constituting a significant global carbon sink. In 2011, tropical forests were estimated to have stored 262.1 Pg C in live above- and belowground biomass. In contrast, boreal forests contained 53.9 Pg C and temperate forests, 46.6 Pg C (Pan et al., 2013). Paradoxically, many tropical forests occur on highly weathered soils with lower amounts of total P and a higher fraction of occluded (and unavailable) P (Yang et al., 2013). Although both nitrogen (N) and P are likely to co-limit natural ecosystems, P is of particular interest because soil P cannot be replenished biologically in the way that N fixation, prevalent in the tropics, can enhance available N (Cleveland et al., 2011; Townsend et al., 2011). Orthophosphate $(H_2PO_4^{-} \text{ and } HPO_4^{2-})$ is the form of P most readily available for plant and microbial uptake, vital for many compounds involved in energy transfer, metabolism, membrane transport, signaling, and in the formation of nucleic acids (Raven, 2015). However, prolonged warm and moist climate of the tropics encourage high weathering rates, which deplete sources of P. In addition, P forms insoluble complexes with aluminum and iron ions, both of which are considered geochemical sinks for phosphate due to their specific surface area (Chacon et al., 2006). High weathering rates and the strong reaction between soil iron and aluminum to phosphate result in tropical soils typically devoid of orthophosphate (Vitousek and Sanford, 1984; Chacon et al., 2006; Dalling et al., 2016). Therefore, it is critical to understand factors that regulate P availability given that P limitation could modulate important tropical ecosystem properties such as photosynthetic activity (Bloomfield et al., 2008), interactions with the N cycle, belowground C cycling, litter decomposition, soil organic matter turnover (Cleveland et al., 2011), growth rate of young trees and seedling survival (Alvarez-Clare et al., 2013), respiration (White and Hammond, 2008), tree distribution (Condit et al., 2013), and microbial biomass and composition (Liu et al., 2012). The immense diversity and vitality of tree species in tropical forests suggests the existence of multiple mechanisms that can alleviate limitations in soil P supply (Lynch and Brown, 2001; Lambers et al., 2006). Although interactions between roots and microbes are important to regulating P acquisition, the correlation between root and microbial traits and their influence on soil P are not well understood (Marschner et al., 2011; Mommer and Weemstra, 2012; Iversen, 2014).

The plant microbiome plays an active role in shaping root and soil characteristics that modulate plant P acquisition by altering existing pathways and providing new biochemical capabilities (Bulgarelli et al., 2013). The direct pathway for P uptake is through root hairs or the root epidermis, restricting P uptake to the soil volume closest to the root surface. However, both arbuscular mycorrhizal fungi (AMF) and plant growth promoting bacteria influence this process in a variety of ways. While there has been much attention in relating mycorrhizae to root traits with regards to P acquisition (Valverde-Barrantes et al., 2013; Comas et al., 2014; Kong et al., 2014), less attention has focused on bacterial members of the microbial community and the ways in which they also shape the root system. Many members in the plant growth promoting bacteria group can alter root architecture and morphology by synthesizing major plant hormones such as auxin and ethylene, stimulating the formation of lateral roots, decreasing primary root length, and increasing root hairs - indirectly influencing the effectiveness of soil foraging (Bulgarelli et al., 2013; Niu et al., 2013; Vacheron et al., 2013). In addition, bacteria also directly enhance P availability in the immediate vicinity of the root. P occurs in both inorganic and organic forms. Orthophosphate availability from minerals (inorganic) is highly dependent on soil pH and by soil processes such as sorption-desorption and dissolutionprecipitation. Bacteria are capable of solubilizing P from these minerals to enhance P availability in the rhizosphere directly. Bacterial release of protons and secondary metabolites such as organic acid anions and amino acids solubilize P from minerals, influencing sorption-desorption and dissolution-precipitation equilibria in the soil (Bunemann et al., 2004; Vacheron et al., 2013). However, in tropical regions, mineralization of organic P is the major source for orthophosphate (Vitousek and Sanford, 1984).

Plants and microbes release phosphatase enzymes to mineralize organic P compounds. Phosphatase activity refers to the actions of two complementary, but distinct enzymes: phosphodiesterase (PDE) and phosphomonoesterase (PME). PDE hydrolyses complex organic P compounds such as nucleic acids and phospholipids into phosphomonoesters (mononucleotides and inositol phosphates). PME further mineralizes these compounds into orthophosphate which can be directly absorbed by plants and microbes (Rejmánková et al., 2011; Stone and Plante, 2014). These extracellular enzymes are key agents in organic P mineralization and play important roles in plant response to limited P availability or increasing P demand (Dakora and Phillips, 2002; Burns et al., 2013; Dalling et al., 2016). More recently, phosphatase activity has been determined to be a crucial uncertainty in modeling P cycling in ecosystem models because P mineralization is among the least understood aspects of P dynamics (Reed et al., 2015). Several modeling studies suggest that the magnitude of phosphatase activity could significantly influence CO₂ uptake in tropical forests (Goll et al., 2012; Yang et al., 2016). Determining the interaction between roots and the bacterial community in regulating phosphatase activity and the factors that influence those interactions is central to understanding P mineralization. However, much of what we currently understand about the role of bacteria in P acquisition is derived from agricultural studies. Although these studies have provided valuable insight, there is a need to extend understanding of those systems to tropical forests, especially with respect to root traits and function (Richardson and Simpson, 2011; Pii et al., 2015). The improved understanding of variations among different tree species of root and bacterial function may provide a means to improve representation of tropical forests in ecosystem models and help us gain a more thorough understanding of how roots and their bacterial community enable tropical forests to thrive in severely limited P soils.

To understand how root and bacterial function differed by tree species and soil P availability, we assessed phosphatase activity in roots and bacterial isolates. Because bacterial community composition may influence this function, we also characterized the bacterial component of the microbial community by sequencing the 16S amplicon. The overarching question was how soil P availability and tree species influenced root and associated bacterial phosphatase activity, as well as root associated bacterial community composition. We hypothesized that (1) phosphatase activity would be higher in areas with low P availability, (2) within a site, tree species would differ in phosphatase activity, and (3) bacterial community composition would primarily depend on tree species. The numerous interactions between roots and the bacterial community determine, in part, P mineralization of tropical ecosystems. Therefore, understanding factors which influence those interactions provides a means to unravel how rhizosphere processes belowground relate to the forest above.

MATERIALS AND METHODS

Study Sites and Tree Species

To examine the influence of P availability on root and bacterial traits, three study sites in the Luquillo mountains of northeastern Puerto Rico (18°30'N, 65°80'W) were chosen to represent a spectrum of P availability. Previous measurements of Rio Icacos (Icacos), El Verde Ridge (Ridge), and El Verde Valley (Valley) indicated lower total P in Icacos and higher total P in Ridge and Valley reflecting differences in parent material. Icacos, formed on quartz-diorite parent material had the lowest total P content $(170 \pm 25 \text{ ppm P})$ followed by Ridge $(290 \pm 15 \text{ ppm P})$ and finally, Valley (410 \pm 43 ppm P), both of which were on volcaniclastic parent material known to contain higher amounts of soil P (Mage and Porder, 2013). In Icacos, Ridge, and Valley, the most common tree species were sampled for root phosphatase. In Icacos, the tree species sampled were Cecropia schreberiana Loefl., Micropholis garcinifolia Grisseb, Cyrilla racemiflora L., and Prestoea montana. In Ridge and Valley, the trees sampled were as follows: Prestoea montana Hook, Dacryodes excelsa Vahl, and Manilkara bidentata Chevalier (Supplementary Table 5).

Sample Collection

Soil Collection and Resin P Assay

At each site, nine cores (2.5 cm diameter) were collected from 0 to 10 cm soil depths to measure site level P availability using the resin P method (Calvert et al., 1993). The resin P method estimates the amount of labile P available for plant and microbial uptake by imitating root removal of phosphate from the soil solution. Charged resin membrane strips in sample soil suspensions attract phosphate ions. Then, the resin strips are washed to remove adhering phosphate ions, resulting in an extract that serves as a proxy for orthophosphate availability in the soil (Van Raij et al., 2009). From each core, 8 g of soil were mixed with 80 mL of distilled water and five resin strips charged with sodium bicarbonate. Samples were shaken for 24 h after which the resin strips were washed of the adsorbed phosphate ions in 50 mL of sulfuric acid. Orthophosphate in the resulting extract was quantified using the Lachat QuikChem 8500. Site level P availability is presented as an average of the nine cores taken at each site.

Root Clusters

The first three most distal ends of the root were collected for the root phosphatase assay according to recent studies that indicate functional differences in the root system. The first three distal ends (orders) are associated with resource acquisition in contrast with higher orders which are involved in resource transport (McCormack et al., 2015). Three root clusters (first three orders) from the most common tree species in each site were sampled during November 2015. Root clusters were collected by tracing roots from the base of each individual tree and gently excavating to the terminal three root orders. In total, 12 root clusters were collected from Icacos, 9 root clusters from Ridge, and 9 root clusters from Valley (n = 3 per tree species). Collected roots were kept cold during transport using cold packs and shipped overnight to Oak Ridge National Laboratory for further analyses.

Bacterial Strain Isolation

To characterize Ridge and Valley bacterial diversity, bacterial isolates and community composition were collected during August 2015. Adhering soil from roots of D. excelsa and P. montana collected at Ridge and Valley were washed off using 10 mM magnesium sulfate. The resulting slurry was pooled for each tree species at each site before taking three serial dilutions. Each dilution was plated onto solid R2A media (18.2 g DifcoTM R2A l⁻¹) (Reasoner and Geldreich, 1985). Throughout 1 week, colonies were selected and re-streaked for isolation on fresh R2A solid agar plates. Single colonies of the resulting plates were re-streaked on fresh R2A solid agar plates three times to ensure pure colonies of each isolate. Bacterial stocks of each isolate were made by growing individual colonies overnight in R2A liquid medium at 25°C and mixing the ensuing bacterial culture with an equal volume of 50% glycerol. These solutions were frozen in -80°C and re-streaked onto fresh plates for downstream functional assays.

Microbial Community

Culturing bacteria enables insight into bacterial function of members of the bacterial community. In addition, a broader survey of the bacterial community via 16S amplicon sequencing provides the opportunity to understand which phyla are present in the bacterial community, placing into context observed bacterial function. Therefore, rhizosphere soil from Ridge and Valley was collected from D. excelsa and P. montana roots for bacterial 16S rRNA gene sequencing. Tweezers were used to separate adhering soil from terminal roots. Approximately 50 mg of those roots were washed in 50 ml falcon tubes with milliQ H₂O before adding rhizosphere soil and pelleting the samples from each tree in Ridge and Valley. A total of 12 samples for DNA extraction from D. excelsa and P. montana roots were collected (three individuals surveyed per species per site). DNA was extracted from the soil pellet using MoBio PowerSoil kit as per manufacturer's instructions (Mo Bio Laboratories) and purified using the Zymo clean-up kit (Zymo Research). Samples were then stored in -20° C. Prior to amplification, DNA concentration was normalized to 10 ng μ l⁻¹.

Root and Bacterial Function

Root Phosphatase

To estimate quantitative phosphorous transformation, PME and PDE activities were measured for the most common tree species at Icacos, Ridge, and Valley. A modified version of the colorimetric soil phosphatase test by Tabatabai and Bremner (1969) was used for root samples para-nitrophenylphosphate (pNPP) and bis-parawith nitrophenylphosphate (bis-pNPP) as a substrate for PME and PDE activities, respectively (Dr. Benjamin L. Turner, personal communication). Enzyme activity for tree species was calculated as the average of all technical replicates for each biological replicate of each species (n = 9). Approximately 0.3 ± 0.1 g of roots were weighed with minimal cutting to maintain natural root surface area. Weighed roots were placed between moistened paper towels to prevent drying during the weighing process for other samples. Each weighed root was then placed into a glass vial with 9 mL of 50 mM sodium acetate (pH = 5.0). Following a 5-min equilibration time in a 27° C shaker, 1 mL of 50 mM pNPP was added for the PME assay, or 1 mL of 50 mM bis-pNPP for the PDE assay. Samples were incubated in the shaker for 1 h, and then 0.5 mL of the sample solution was added to 4.5 mL of 0.11 M NaOH to terminate the reaction. Blank solutions were pure 10 mL of 50 mM sodium acetate incubated along with the samples, 0.5 mL of which was added to 4.5 mL of 0.11 M NaOH. Absorbance at 410 nm was read on a Thermo Spectronic 20D and compared to a standard curve generated from para-nitrophenol (pNP), the yellow end-product released from PME or PDE hydrolysis of pNPP.

Culture of Bacterial Isolates and Bacterial Phosphomonoesterase Activity

Bacterial PME activity was assessed in isolated strains to estimate potential contribution of bacteria to P mineralization. Each bacterial strain was grown in 1.5 mL R2A overnight before centrifugation at 10,000 rpm for 20 min. A total of 1 mL of the resulting supernatant was added to 1 mL of 50 mM pNPP and 4 mL of modified universal buffer (MUB). MUB consisted of (for 1 L): 12.1 g Tris(hydroxymethyl)aminomethane (THAM or Tris base); 11.6 g maleic acid; 14.0 g citric acid; 6.3 g boric acid; and 1 M sodium hydroxide (pH = 6.5). The solution was then incubated at 27°C for 1 h before terminating the reaction by adding 5 mL of 0.5 M NaOH. Blank solutions consisted of 1 mL of pure R2A culture, 1 mL of 50 mM pNPP, 4 mL of MUB, and 5 mL of 0.5 NaOH. Absorbance at 410 nm was read on a Thermo Spectronic 20D and compared to a standard curve generated from pNP (Satta et al., 1979).

Bacterial P Solubilization

Phosphorus (P) solubilization is a key trait in bacteria that is routinely measured to assess bacterial strains for plant growth promoting properties. To induce P solubilization activity of bacterial isolates, each isolate was grown in National Botanical Research Institute's phosphate growth medium (NBRIP), which requires bacteria to solubilize calcium phosphate to acquire P for growth (Panhwar, 2012). Briefly, NBRIP consisted of (for 1 L): 10 g glucose, 5 g calcium phosphate, 5 g magnesium chloride hexahydrate, 0.25 g magnesium sulfate heptahydrate, 0.2 g potassium chloride, 0.1 g ammonium sulfate, and 15 g agar (Nautiyal, 1999). P solubilization was assessed in 96-well plates via a scaled-down version of the United States Geological Survey (USGS) acid-persulfate digestion method (Patton and Kryskalla, 2003). Samples were plated in triplicate, and the average value of those three wells were used as P solubilization for that isolate. Samples were compared to standard curves between 0.005 and 0.1 mg L^{-1} potassium phosphate and read on a Thermo Spectronic 20D at a wavelength of 880 nm.

DNA Analysis

Preparation for Sequencing

The bacterial V4 16S rRNA region was selectively amplified and barcoded using a two-step PCR approach and using established protocols with PNA blockers to prevent plastid and mitochondrial 16S amplification (Lundberg et al., 2013). Barcoded DNA was sequenced on a single lane of an Illumina MiSeq. Forward primers for the first PCR amplification consisted of three 515F universal primers combined with one 515F Crenarchaeota primer and one 515F TM7 primer. Reverse primers were a mixture of three universal 806R primers (Supplementary Table 1). Thermal cycler conditions for the primary PCRs for soils were 5 cycles of 95°C for 1 min, 50°C for 2 min, and 72°C for 1 min. A total of 30 µL of the primary PCR products were cleaned with 21 µL of Agencourt AMPure beads and eluted in 21 µL of nuclease-free water. For secondary PCRs, the same reverse and forward primers were used in the 50 µL reaction. Thermal cycler conditions for secondary soil PCRs consisted of denaturation at 95°C for 45 s followed by 32 cycles of 94°C for 15 s, annealing at 60°C for 30 s, 72°C for 30 s, and final extension at 72°C for 30 s.

Sequence Analysis

Downstream analysis of sequenced bacterial 16S sequences were conducted in MacQIIME (Caporaso et al., 2010). First, bacterial

16S sequences were assigned to respective samples using barcode sequences designated to each sample prior to amplification and sequencing. Sequences were then clustered into OTUs using an open reference OTU picking protocol using UCLUST at a threshold of 97% similarity (Edgar, 2010). Representative sequences were chosen for each OTU and were identified using the Greengenes database. The lowest amount of sequences per sample was 4,363 and the highest was 25,234 with a mean of 12,593. The sampling depth of 7,497 was chosen to include as many samples as possible without compromising diversity analyses with low sampling depths (Supplementary Figures 1, 2). Therefore, samples containing less than 7,497 sequences were removed, which eliminated one P. montana sample from Ridge (4,363), one P. montana sample from Valley (5,250), and one D. excelsa from Valley (6,895). Finally, both mitochondrial and chloroplast sequences were removed. The relative abundances of all the phyla (Figure 3A) are based on the percentage of 16S rRNA sequences assigned to each phylum.

Statistical Analysis

Root phosphatase data were rank transformed prior to performing two-way repeated measures multivariate analysis of variance (MANOVA). Two-way MANOVA avoids errors associated with running multiple ANOVAs because this method can test both enzymes with respect to tree species and site at the same time. The dependent variables were PME and PDE activities, while the independent variables were tree species and site. For each enzyme, three technical replicates of fine root clusters were assayed for three individuals of each tree species at each site. An $\alpha = 0.05$ was chosen to denote statistical significance of either tree species or site on phosphatase activity. Tukey's HSD was used to further examine significant differences of phosphatase activity among samples.

To understand how bacterial community composition responded to tree species and site, a weighted UniFrac distance matrix and principal coordinate analysis plot were constructed using the packages 'phyloseq' and 'vegan' in R (McMurdie and Holmes, 2013; Oksanen et al., 2015). A relative abundance plot was created to show the abundance of phyla in each sample. Principal coordinated analysis (PCoA) was used to visualize the differences between bacterial communities of *P. montana* and *D. excelsa*, where 95% confidence ellipses were drawn around sample points to indicate similarity. Permutational MANOVA (PERMANOVA) was used to statistically test whether bacterial community composition was significantly different between Ridge and Valley and between tree species. An $\alpha = 0.05$ was also chosen to indicate community composition.

RESULTS

Resin P Availability Is Lowest in Icacos and Highest in Valley

Resin P availability increased from Icacos to Ridge to Valley (Figure 1) and varied significantly by site (p-value < 0.05) (Table 1). Post hoc analysis by Tukey's HSD indicated that



Icacos and Valley differed significantly in resin P availability (Supplementary Table 2). Ridge, which was the intermediate site, did not statistically differ from either Icacos or Valley.

Root Phosphatase Activity Varies with Tree Species and P Availability

Both PME and PDE were significantly different among tree species (Pillai's P = 0.59; p-value < 0.05), though marginally not by site (*p*-value = 0.70) at an $\alpha = 0.05$ (Figures 2A,B and Table 2). Furthermore, there was no significant interaction between factors site and tree species. Tukey's HSD of the two-way repeated measures MANOVA indicated that PME and PDE of P. montana differed from all other tree species, except for M. garcinifolia (Supplementary Tables 3A,B). The average PME in Icacos was $60.03 \pm 15.17 \ \mu mol \ pNP \ g_{root}^{-1} hr^{-1}$, $36.54 \pm 7.56 \ \mu mol \ pNP$ $g_{root}^{-1}hr^{-1}$ in Ridge, and 24.98 \pm 7.88 μ mol pNP $g_{root}^{-1}hr^{-1}$ in Valley. For all tree species, PDE activity was roughly one-tenth of PME. Similar to PME, PDE is highest in Icacos $(3.63 \pm 6.78 \,\mu \text{mol})$ bis-pNP $g_{root}^{-1}hr^{-1}$), lower in Ridge (1.31 \pm 1.76 μ mol bispNP $g_{root}^{-1}hr^{-1}$), with the lowest PDE occurring in Valley $(1.06 \pm 1.37 \ \mu\text{mol} \text{ bis-pNP } \text{g}_{\text{root}}^{-1}\text{hr}^{-1})$. Differences in tree species means, as denoted by the letters in Figures 2A,B for PME

TABLE 1 | ANOVA of soil resin P across all three sites.

Resin P ANOVA table					
	Degrees of freedom	Sum of squares	Mean squares	F-value	Pr (> <i>F</i>)
Site Residuals	2 24	0.7643 1.8421	0.3822 0.0768	4.9789	0.0155*

*Denotes p-value < 0.05.



bars are mean \pm SE with n = 3 individual trees per species at each site.

and PDE indicated that *P. montana* was significantly different between *C. racemiflora*, *D. excelsa*, and *M. bidentata*. In addition, PDE activity of *P. montana* differed from *C. schreberiana* (p-value < 0.05).

The initial MANOVA of root phosphatase as related to site and tree species indicated the significant effect of tree species but not of site in determining enzyme activity. However, a second MANOVA performed by replacing site with resin P availability showed that resin P availability also influenced enzyme activity (Pillai's P = 0.0901; *p*-value = 0.0384) (**Table 3**), though this is due to a change in degrees of freedom. Resin P availability within each site was significant in determining root phosphatase in contrast to site itself which was not. Indeed, average root PME and PDE with respect to average resin P availability exhibited negative correlations (PME $r^2 = 0.971$; PDE $r^2 = 0.747$) though neither was significant (PME *p*-value = 0.109; PDE *p*-value = 0.335) (**Figures 2C,D**). In summary, root phosphatase activity is significantly controlled by tree species, not site, though analysis of resin P suggests the importance of soil P availability in co-influencing enzyme activity in the rhizosphere.

Bacterial Community Composition in Ridge and Valley Is Influenced by Site

Our analysis of the rhizosphere bacterial community in *P. montana* and *D. excelsa* in Ridge and Valley identified 415 OTUs representing 36 bacterial phyla. The sequence data are stored in NCBI BioProject Accession No. PRJNA412374. Over 50% of the community in both Ridge and Valley came from two phyla: *Proteobacteria* and *Acidobacteria* with no large differences in composition between Ridge and Valley at the phyla level (**Figure 3A**). The proportion of *Proteobacteria* ranged from 29 to 43% in Valley, whereas it was 37 to 45% in Ridge. The proportion of *Acidobacteria* varied more widely from 13 to 45% in Valley

TABLE 2 | Site and tree species two-way repeated measures MANOVA of root PME and PDE activities.

	Site x tree species root phosphatase						
	Degrees of freedom	Pillai	Approximate F	Num. degrees of freedom	Den. degrees of freedom	Pr (> <i>F</i>)	
Site	2	0.1208	2.2195	4	138	0.0700	
Species	5	0.5798	5.6349	10	138	< 0.001*	
Site:Species	2	0.0414	0.7306	4	138	0.5725	
Residuals	69						

*Denotes p-value < 0.05.

TABLE 3 | Resin P and tree species two-way repeated measures MANOVA of root PME and PDE.

	Resin P x tree species root phosphatase							
	Degrees of freedom	Pillai	Approximate F	Num. degrees of freedom	Den. degrees of freedom	Pr (> <i>F</i>)		
Resin P	2	0.0901	3.4180	2	69	0.0384*		
Species	5	0.5946	5.9232	10	140	<0.001*		
Resin P:Species	2	0.0357	0.6369	4	140	0.6370		
Residuals	69							

*Denotes p-value < 0.05.



and 19 to 34% in the Ridge. The relative abundance plot of bacterial phyla associated with the fine roots of D. *excelsa* and *P. montana* in Ridge and Valley shows no visible differences in relative abundance except for the presence of the *Nitrospirae* phylum found exclusively in Valley.

A weighted UniFrac distance matrix was used to create a PCoA plot with 95% confidence intervals to visually determine similarities among *P. montana* and *D. excelsa* bacterial communities. Points in the PCoA represent each bacterial community, with distances between points indicating

TABLE 4 PERMANOVA for analysis of similarity/dissimilary among bacterial
community from D. excelsa and P. montana.

Bacterial PME and PDE				
	Degrees of freedom	F-value	Pr (> <i>F</i>)	
Site	1	2.5658	0.014*	
Species	1	0.7345	0.642	
Permutations	999			

*Denotes p-value < 0.05.

similarity or dissimilarity. The resulting plot suggests that bacterial communities are more similar between sites rather than tree species (**Figure 3B**). PERMANOVA using bacterial community composition was used to statistically test whether tree species or site influenced bacterial community composition. At an $\alpha = 0.05$, PERMANOVA results verify that bacterial community composition is influenced by site, not tree species (*p*-value < 0.05) (**Table 4**).

Bacterial Isolates from *D. excelsa* Have Higher PME and P Solubilization Activity than Isolates from *P. montana*

A total of 95 isolates from both Ridge and Valley were tested for PME and P solubilization activity after growing the isolates overnight in a P-deficient media (NBRIP). PME activity was found in 36 isolates with a mean of 0.106 \pm 0.021 μ mol pNP ml⁻¹ in isolates from the rhizosphere of *D. excelsa* and a mean of 0.086 \pm 0.005 µmol pNP ml⁻¹ from *P. montana*. P solubilization occurred in 28 isolates with higher activity in D. excelsa (mean = 6.85 \pm 21.58 mg P μ L⁻¹) relative to *P. montana* (mean = 1.31 \pm 1.34 mg P μ L⁻¹). A majority (67%) of the isolate library was capable of directly enhancing P availability through either PME or P solubilization activity. However, only 32% of the isolate library showed both PME and P solubilization given our assay conditions (Figure 4). Of the isolates with both PME and P solubilization, 13 isolates were from Ridge, and nine isolates from Valley. Two-way repeated measures MANOVA conducted on PME and P solubilization activity indicated that tree species was a significant factor in determining both in bacterial isolates (*p*-value < 0.05) but site was not (p-value > 0.05) (Table 5). Tukey's HSD indicated a significant difference between P. montana and D. excelsa in PME (p-value = 0.0006), but not P solubilization (p-value = 0.66); Supplementary Table 4).

DISCUSSION

A major challenge of integrating root traits and function is capturing their complex interactions with microbial soil communities. However, it is precisely these interactions between roots and root-associated microbes that direct multiple root functions. Despite their importance, the knowledge of these microbial communities relative to specific root traits is still limited (Bardgett et al., 2014). Roots and root-associated microbes actively release phosphatase to enhance P availability, suggesting that both root and microbial functions must be studied in tandem to fully understand these important mechanisms that contribute to tree growth in P limited tropical forests. Recent simulations indicate that phosphatase activity could be an important factor in regulating the size of the carbon sink given that P acquisition may be a limiting factor in continued growth under changing climates (Yang et al., 2016).

Decades of work have established the importance of the soil environment on determining root and bacterial function (Vitousek and Sanford, 1984; Treseder and Vitousek, 2001; Nannipieri et al., 2011; Spohn et al., 2015), which influenced our choice to study root and bacterial phosphatase activity and bacterial community composition at a collection of sites that increased in P availability from Icacos to Ridge to Valley (Mage and Porder, 2013). Our results showed the same pattern and confirmed the strong influence of site on soil P availability, though the only significant difference in resin P availability was between Icacos and Valley. Our first hypothesis was that phosphatase activity would differ based on soil P availability. Two-way repeated measures MANOVA of root and bacterial phosphatase indicated a significant influence of tree species, though not of site. Replacing site with resin P in a MANOVA, however, did suggest that both tree species and resin P availability influenced phosphatase though this is attributable to a change in the degrees of freedom. Linear regressions of average root PME and root PDE at each site with respect to average resin P availability indicated a negatively correlated trend between root phosphatase and P availability. Lastly, average bacterial PME was higher in isolates from the Ridge than those in the Valley suggesting that site and resin P availability are modulating phosphatase in roots and bacteria. As expected root phosphatase was negatively correlated with resin P availability, supporting the view that root phosphatase may be an important functional trait that forms a critical aspect of nutrient acquisition in low P environments.

Our second hypothesis that phosphatase activity would be influenced by tree species was supported by both root and bacterial two-way repeated measures MANOVAs albeit our limited sample size. Root PME and PDE activities were significantly different between tree species, most notably between P. montana samples and four of the five tree species. Similarly, bacterial PME and P solubilization were influenced by host tree species. Post hoc analysis placed P. montana samples as a group separate from the other tree species perhaps because P. montana was the only monocot. Monocots were shown to differ from dicots in attributes such as more fibrous root systems (Smith and De Smet, 2012), differences in N:P ratios (Han et al., 2004), and differences in mycorrhizal relationships (Cornwell et al., 2001), which may strongly influence functions in the ecosystem. These results are certainly consistent with previous findings that point to the importance of tree species and of plant functional groups in determining phosphatase activity (Hodge, 2004; White and Hammond, 2008; Lambers et al., 2009; Weintraub, 2011; Keller et al., 2013).

The importance of tree species on root phosphatase can also be tied to the combination of conservative and acquisitive traits that regulate tissue P demand and therefore the necessity of acquiring



P via root phosphatase. Plants can respond to P limitation by altering functional traits to either increase P efficiency or P acquisition, though it is likely that different tree species use varying combinations of both to adjust to local nutrient supplies (White and Hammond, 2008). Within the tropics, multiple lines of evidence confirm that trees are capable of efficiently using P through high P resorption, P recycling, and reduction of P concentrations in metabolic nucleic acid compounds (Vitousek and Sanford, 1984; Hidaka and Kitayama, 2011). These different mechanisms result in a variety of tissue P requirements for different tree species and therefore different responses to P availability and need for root phosphatase (Ushio et al., 2010; Keller et al., 2013). The lack of an interaction term between site and tree species suggests that tree species is driving differences in root phosphatase activity independently. However, it is important to note that site is likely to be a confounding factor with tree species given that the soil environment can modify these

functional traits (Lambers et al., 2006; Richardson et al., 2009; Niu et al., 2013).

Although our alpha level of 0.05 excluded site as a significant determinant of root and bacterial phosphatase activity, we suspect that they are likely influenced by feedbacks between site and tree species. P availability and tree species have consistently been found to regulate root and bacterial function (Treseder and Vitousek, 2001; Costa et al., 2006; Lambers et al., 2006; Haichar et al., 2008; Bardgett et al., 2014; Hinsinger et al., 2015). For example, differences in phosphatase activity between root clusters of the same tree support the notion that heterogeneous distribution of P within each site can cause variation in measured activity. Since roots must respond to local conditions, microsite variations in nutrient content, microbes, and competing roots likely modify phosphatase activity between root clusters of the same individuals and of tree species could be due to variation in

Bacterial PME and P solubilization MANOVA							
	Degrees of freedom	Pillai	Approximate F	Num. degrees of freedom	Den. degrees of freedom	Pr (> <i>F</i>)	
Site	1	0.2057	2.9792	2	23	0.0707	
Species	1	0.3868	7.2540	2	23	0.0036*	
Site:Species	1	0.1416	1.8977	2	23	0.5725	
Residuals	24						

TABLE 5 | Two-way repeated measures MANOVA of bacterial PME and P solubilization activity.

*Denotes p-value < 0.05.

the root orders collected and ensuing differences in the microbes attached to root tissue.

Root order was shown to correspond to different roles in the root system where the first and second orders, or the most distal roots, are associated with higher rates of exudation and uptake than those of higher orders typically associated with transport (McCormack et al., 2015). As such, samples with a greater proportion of higher order roots may have shown much less phosphatase activity than samples of the same individual made up of predominantly first- and second-order, absorptive roots. Preliminary observations of D. excelsa and P. montana roots show that the branching intensity (number of first-order roots per centimeter of second-order roots) of D. excelsa (5.58) was higher compared to P. montana (2.70) (Daniela Yaffar, personal communication). The difference in branching intensity between D. excelsa and P. montana may in part explain why phosphatase activity is higher in *D. excelsa* as it has a much higher amount of first-order (resource acquisitive) roots. Similarly, a comparison between lianas and trees demonstrated that roots of lianas seem to favor high P acquisition, correlated with root traits such as greater root branching intensity, higher N and P concentrations, and higher specific root lengths (Collins et al., 2016). Future work will involve examining how root morphology may impact phosphatase activity and whether the release of phosphatase changes depending on the effectiveness of root architectural and morphological traits to find patches of orthophosphate. Indeed, the role of phosphatase activity in determining P mineralization is likely to depend on merging how phosphatase correlates with other root functional traits and the microbial community.

Finally, our third hypothesis that bacterial communities would be largely influenced by tree species was contradicted by results that indicated the significance of site, but not of tree species in determining bacterial community composition. In both Ridge and Valley bacterial communities, the major phyla representing over half of the microbial communities were *Proteobacteria* and *Acidobacteria*, both of which tend to be common in soils (Bulgarelli et al., 2013). However, the phylum *Nitrospirae*, known to be the most diverse group of nitrite-oxidizing bacteria (NOB), only occurred in the Valley (Daims et al., 2015). Conditions in Valley likely favor *Nitrospirae* because of much wetter conditions that create an anoxic soil environment. While there is no question that soil conditions impact the composition of bacterial communities in significant ways through soil pH (Fierer and Jackson, 2006) and microscale heterogeneity (Vos et al., 2013), the intimate association between plants and their root bacterial symbionts certainly affirms the importance of tree species as key factor (Carney and Matson, 2006; Ushio et al., 2008, 2010; Bulgarelli et al., 2013; Philippot et al., 2013). One theory of how rhizosphere microbial communities are formed suggests that soil conditions, such as nutrient availability, initially determine soil microbial community composition. Growing plants then release root exudates which further alter the soil environment, fostering a unique microbial community distinct from bulk soil (Bulgarelli et al., 2013). This two-step model may explain why microbial community composition is more strongly influenced by soil conditions, though the limited sampling size may have obscured stronger effects of tree species on bacterial composition. Interestingly, although bacterial community composition was influenced by site, a two-way repeated measures MANOVA of bacterial PME and P solubilization indicated that these two functions were only influenced by tree species and not site.

PME activity and P solubilization of bacterial isolates were higher in isolates from D. excelsa than those from P. montana. However, PME was more prevalent in the bacterial isolates collected, despite the importance of P solubilization in enhancing P availability. This might be due to the limitations of the P solubilization assay and the differences in target compounds. First, P solubilization is typically measured using calcium phosphate as a P source, requiring bacterial colonies to solubilize calcium phosphate to survive. However, this favors only those bacterial strains capable of solubilizing calcium phosphate, limiting extrapolation of this function to the broader microbial community. However, P solubilization is a commonly used and relied upon criteria to screen for plant growth promoting activities, thus why we measured it here. Second, phosphatase and P solubilization act on different compounds in the soil. Phosphatase enzymes mineralize organic compounds, which form an abundant portion of total P in tropical forests (Turner and Engelbrecht, 2011). In contrast, P solubilization targets phosphate bound to inorganic compounds, which is a much smaller and more recalcitrant pool (Saghir et al., 2014). It is possible that substrate abundance for PME encourages more members of the bacterial community to express this function. Furthermore, among the 95 isolates tested, only a few isolates were capable of both PME and P solubilization activities, suggesting that it may be uncommon to possess both functions in the larger bacterial community. Given the paradigm that less than 1% of the microbial community can be cultured,

our bacterial isolates likely do not capture the full functional potential of the broader microbial community. However, our bacterial isolates provide an opportunity to test bacteria and root interactions in controlled greenhouse settings. Furthermore, our results indicated that tree species played an important role in determining root and bacterial PME activity, supporting efforts to build more species-specific trait understanding of P mineralization.

The root system is highly plastic, responding to variations in P availability through changes in root morphology, architecture, exudation, and interactions with soil microbes (Lambers et al., 2006; Niu et al., 2013). In addition to root and bacterial phosphatase and P solubilization, there are multiple mechanisms critical to understanding how the root system adjusts to P limitation. For example, P deficiency can induce an increase in lateral roots near the soil surface and morphological changes such as increasing root length, root turnover, and higher biomass allocation to roots (Lynch and Brown, 2001). Furthermore, mycorrhizal hyphae are a critical pathway for P uptake because fungal hyphae are much thinner than the smallest roots, enabling a much larger surface area with which to absorb orthophosphate and enhance foraging of soil P sources (Plassard and Dell, 2010; Smith and Smith, 2011). Numerous reviews describe how this intimate association between trees and fungi influence root branching, fine roots, root: shoot ratio, specific root length, and responses to elevated CO2 as well as the finer details of the rootfungal interface and the exchange of nutrients (Bucher, 2007; Plassard and Dell, 2010; Smith et al., 2011; Bardgett et al., 2014; Liu et al., 2015). However, our results address the (much less known) bacterial component of the plant microbiome.

Root and bacterial phosphatase activity are central components in plant P acquisition and a source of major uncertainty in understanding P mineralization, forest productivity, and modeling tropical forest growth (Cernusak et al., 2013; Hofhansl et al., 2016). The variation of root phosphatase with tree species is encouraging as it may be an important trait for predicting how different tree species respond in altered climatic regimes. Although tropical forests are immensely diverse, our results with our tree species begin pairing root function with bacterial function and community composition to improve understanding of one critical aspect of P dynamics in tropical forests - phosphatase activity. Our main result was that root and bacterial phosphatase, though not bacterial community composition and P solubilization, vary with tree species and soil P availability. Average values of root phosphatase activity show a promising negative correlation with resin P availability that encourages future empirical and modeling research aimed at building trait-based understanding of P dynamics in tropical forests.

CONCLUSION

Production of phosphatase enzymes by both roots and microbes is a crucial function because these enzymes enhance the availability of orthophosphate needed for plant and microbial growth (Marschner et al., 2011). As such, phosphatase activity is an important part of acquiring P. The variation in phosphatase activity among tree species and bacterial communities could be central to understanding how trees adapt to limitations in P supply and how bacterial communities mediate that response. Thus far, roots and microbial communities (if included at all) are modeled as the result of aboveground processes rather than equal determinants of how an ecosystem functions and responds to change (Warren et al., 2015). However, root functional traits and the microbial community are just as central to nutrient cycling, vegetation dynamics, and ecosystem response (Mommer and Weemstra, 2012; Iversen, 2014). Results indicate a negative correlation of root PME activity with increasing P availability, though inference to the broader tropical forest community is limited by the sample size. However, tree species was found to be important in modulating root and bacterial PME activity given local P conditions. These results suggest that efforts to pair root function with microbial community function and composition can indeed provide a platform to unravel complicated rhizosphere interactions that can support the inclusion of roots and microbes in ecosystem models and improve our understanding of key rhizosphere interactions that regulate P mineralization in tropical forests.

AUTHOR CONTRIBUTIONS

KC, DW, and RN designed the study. KC, JC, and RN conducted the field studies. KC, CT, AC, T-YL, DP, and DW isolated bacteria from root samples, extracted rhizobacterial DNA, and prepared it for sequencing. KC, CT, AC, and DW analyzed microbial community sequences. KC, CT, and AC conducted root and microbial enzyme analysis. JC performed the P analysis. KC drafted the manuscript. All authors contributed to writing and finalizing the paper, and also read and approved the final manuscript.

FUNDING

This research was supported as part of the Next Generation Ecosystem Experiments-Tropics, funded by the United States Department of Energy, Office of Science, Office of Biological and Environmental Research. Oak Ridge National Laboratory is managed by UT-Battelle, LLC, for the United States Department of Energy under contract DE-AC05–00OR22725.

ACKNOWLEDGMENTS

We thank Whendee Silver, Jess Zimmerman, and Ariel Lugo for help in identifying research sites in Puerto Rico. We thank Benjamin L. Turner for providing the root phosphatase protocols. Melissa Cregger, Xiaojuan Yang, and Sasha Reed provided helpful reviews of the manuscript. We also thank Xiaocun Sun for assisting with statistical analysis of root phosphatase data. Finally, we thank two reviewers for their thorough critiques and recommendations. This manuscript has been authored by UT-Battelle, LLC under Contract no. DE-AC05-00OR22725 with the United States Department of Energy. The United States Government retains and the publisher, by accepting the article for publication, acknowledges that the United States Government retains a non-exclusive, paid-up, irrevocable, worldwide license to publish or reproduce the published form of this manuscript, or allow others to do so, for United States Government purposes. The Department of Energy will provide public access to these results of federally sponsored research in accordance with

REFERENCES

- Alvarez-Clare, S., Mack, M. C., and Brooks, M. (2013). A direct test of nitrogen and phosphorus limitation to net primary productivity in a lowland tropical wet forest. *Ecology* 94, 1540–1551. doi: 10.1890/12-2128.1
- Baldrian, P. (2014). Distribution of extracellular enzymes in soils: spatial heterogeneity and determining factors at various scales. Soil Sci. Soc. Am. J. 78, 11–18. doi: 10.2136/sssaj2013.04.0155dgs
- Bardgett, R. D., Mommer, L., and De Vries, F. T. (2014). Going underground: root traits as drivers of ecosystem processes. *Trends Ecol. Evol.* 29, 692–699. doi: 10.1016/j.tree.2014.10.006
- Bloomfield, K. J., Farquhar, G. D., and Lloyd, J. (2008). Photosynthesis nitrogen relationships in tropical forest tree species as affected by soil phosphorus availability: a controlled environment study. *Funct. Plant Biol.* 41, 820–832. doi: 10.1071/FP13278
- Bucher, M. (2007). Functional biology of plant phosphate uptake at root and mycorrhiza interfaces. *New Phytol.* 173, 11–26. doi: 10.1111/j.1469-8137.2006. 01935.x
- Bulgarelli, D., Schlaeppi, K., Spaepen, S., Ver Loren, van Themaat, E., and Schulze-Lefert, P. (2013). Structure and functions of the bacterial microbiota of plants. *Annu. Rev. Plant Biol.* 64, 807–838. doi: 10.1146/annurev-arplant-050312-120106
- Bunemann, K., Oberson, A., and Frossard, E. (2004). Phosphorus in Action: Biological Processes in Soil Phosphorus Cycling, Vol. 26, Berlin: Springer-Verlag.
- Burns, R. G., DeForest, J. L., Marxsen, J., Sinsabaugh, R. L., Stromberger, M. E., Wallenstein, M. D., et al. (2013). Soil enzymes in a changing environment: Current knowledge and future directions. *Soil Biol. Biochem.* 58, 216–234. doi: 10.1016/j.soilbio.2012.11.009
- Calvert, F., Christiansen, J., Ofsoski, N., and Smillie, J. (1993). *The Resin P Soil Test*. Christchurch: Hill Laboratories.
- Caporaso, J. G., Kuczynski, J., Stombaugh, J., Bittinger, K., Bushman, F. D., Costello, E. K., et al. (2010). QIIME allows analysis of high-throughput community sequencing data. *Nat. Methods* 7, 335–336. doi: 10.1038/nmeth. f.303
- Carney, K. M., and Matson, P. A. (2006). The influence of tropical plant diversity and composition on soil microbial communities. *Microb. Ecol.* 52, 226–238. doi: 10.1007/s00248-006-9115-z
- Cernusak, L., Winter, K., Dalling, J., Holtum, J., Jaramillo, C., Korner, C., et al. (2013). Tropical forest responses to increasing CO₂: current knowledge and opportunities for future research. *Funct. Plant Biol.* 40, 531–551. doi: 10.1071/ FP12309
- Chacon, N., Silver, W. L., Dubinsky, E. A., and Cusack, D. F. (2006). Iron reduction and soil phosphorus solubilization in humid tropical forests soils: the roles of labile carbon pools and an electron shuttle compound. *Biogeochemistry* 78, 67–84. doi: 10.1007/s10533-005-2343-3
- Cleveland, C. C., Townsend, A. R., Taylor, P., Alvarez-Clare, S., Bustamante, M. M. C., Chuyong, G., et al. (2011). Relationships among net primary productivity, nutrients and climate in tropical rain forest: a pan-tropical analysis. *Ecol. Lett.* 14, 939–947. doi: 10.1111/j.1461-0248.2011.01658.x
- Collins, C. G., Wright, S. J., and Wurzburger, N. (2016). Root and leaf traits reflect distinct resource acquisition strategies in tropical lianas and trees. *Oecologia* 180, 1037–1047. doi: 10.1007/s00442-015-3410-7
- Comas, L. H., Callahan, H. S., and Midford, P. E. (2014). Patterns in root traits of woody species hosting arbuscular and ectomycorrhizas: Implications for the

the DOE Public Access Plan (http://energy.gov/downloads/doe-public-access-plan).

SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: https://www.frontiersin.org/articles/10.3389/fpls.2017.01834/ full#supplementary-material

evolution of belowground strategies. *Ecol. Evol.* 4, 2979–2990. doi: 10.1002/ece3. 1147

- Condit, R., Engelbrecht, B. M. J., Pino, D., Pérez, R., and Turner, B. L. (2013). Species distributions in response to individual soil nutrients and seasonal drought across a community of tropical trees. *Proc. Natl. Acad. Sci. U.S.A.* 110, 5064–5068. doi: 10.1073/pnas.1218042110
- Cornwell, W. K., Bedford, B. L., and Chapin, C. T. (2001). Occurrence of arbuscular mycorrhizal fungi in a phosphorus-poor wetland and mycorrhizal response to phosphorus fertilization. Ann. J. Bot. 88, 1824–1829. doi: 10.2307/3558359
- Costa, R., Götz, M., Mrotzek, N., Lottmann, J., Berg, G., and Smalla, K. (2006). Effects of site and plant species on rhizosphere community structure as revealed by molecular analysis of microbial guilds. *FEMS Microbiol. Ecol.* 56, 236–249. doi: 10.1111/j.1574-6941.2005.00026.x
- Daims, H., Lebedeva, E. V., Pjevac, P., Han, P., Herbold, C., Albertsen, M., et al. (2015). Complete nitrification by *Nitrospira bacteria*. *Nature* 528, 504–509. doi: 10.1038/nature16461
- Dakora, F. D., and Phillips, D. A. (2002). Root exudates as mediators of mineral acquisition in low-nutrient environments. *Plant Soil* 245, 35–47. doi: 10.1023/A: 1020809400075
- Dalling, J. W., Heineman, K., Lopez, O. R., Wright, S. J., and Turner, B. L. (2016). "Nutrient availability in tropical rain forests: the paradigm of phosphorus limitation," in *Tropical Tree Physiology*, eds G. Goldstein and L. S. Santiago (Cham: Springer), 261–273.
- Edgar, R. C. (2010). Search and clustering orders of magnitude faster than BLAST. *Bioinformatics* 26, 2460–2461. doi: 10.1093/bioinformatics/btq461
- Fierer, N., and Jackson, R. B. (2006). The diversity and biogeography of soil bacterial communities. *Proc. Natl. Acad. Sci. U.S.A.* 103, 626–631. doi: 10.1073/ pnas.0507535103
- Goll, D. S., Brovkin, V., Parida, B. R., Reick, C. H., Kattge, J., Reich, P. B., et al. (2012). Nutrient limitation reduces land carbon uptake in simulations with a model of combined carbon, nitrogen and phosphorus cycling. *Biogeosciences* 9, 3547–3569. doi: 10.5194/bg-9-3547-2012
- Haichar, F. Z., Marol, C., Berge, O., Rangel-Castro, J. I., Prosser, J. I., Balesdent, J., et al. (2008). Plant host habitat and root exudates shape soil bacterial community structure. *ISME J.* 2, 1221–1230. doi: 10.1038/ismej.2008.80
- Han, W., Fang, J., Guo, D., and Zhang, Y. (2004). Leaf nitrogen and phosphorus stoichiometry across 753 terrestrial plant species in China. *New Phytol.* 168, 377–385. doi: 10.1111/j.1469-8137.2005.01530.x
- Hidaka, A., and Kitayama, K. (2011). Allocation of foliar phosphorus fractions and leaf traits of tropical tree species in response to decreased soil phosphorus availability on Mount Kinabalu, Borneo. J. Ecol. 99, 849–857. doi: 10.1111/j. 1365-2745.2011.01805.x
- Hinsinger, P., Herrmann, L., Lesueur, D., Robin, A., Trap, J., Waithaison, K., et al. (2015). Impact of roots, microorganisms, and microfauna on the fate of soil phosphorus in the rhizosphere. *Annu. Plant Rev.* 48, 377 doi: 10.1002/ 9781118958841.ch13
- Hodge, A. (2004). The plastic plant: root responses to heterogeneous supplies of nutrients. New Phytol. 162, 9–24. doi: 10.1111/j.1469-8137.2004.01015.x
- Hofhansl, F., Andersen, K. M., Fleischer, K., Fuchslueger, L., Rammig, A., Schaap, K. J., et al. (2016). Amazon forest ecosystem responses to elevated atmospheric CO₂ and alterations in nutrient availability: filling the gaps with modelexperiment integration. *Front. Earth Sci.* 4:19. doi: 10.3389/feart.2016.00019
- Iversen, C. M. (2014). Using root form to improve our understanding of root function. New Phytol. 203, 707–709. doi: 10.1111/nph.12902

- Keller, A. B., Reed, S. C., Townsend, A. R., and Cleveland, C. C. (2013). Effects of canopy tree species on belowground biogeochemistry in a lowland wet tropical forest. Soil Biol. Biochem. 58, 61–69. doi: 10.1016/j.soilbio.2012.10.041
- Kong, D., Ma, C., Zhang, Q., Li, L., Chen, X., Zeng, H., et al. (2014). Leading dimensions in absorptive root trait variation across 96 subtropical forest species. *New Phytol.* 203, 863–872. doi: 10.1111/nph.12842
- Lambers, H., Mougel, C., Jaillard, B., and Hinsinger, P. (2009). Plant-microbesoil interactions in the rhizosphere: an evolutionary perspective. *Plant Soil* 321, 83–115. doi: 10.1007/s11104-009-0042-x
- Lambers, H., Shane, M. W., Cramer, M. D., Pearse, S. J., and Veneklaas, E. J. (2006). Root structure and functioning for efficient acquisition of phosphorus: matching morphological and physiological traits. *Ann. Bot.* 98, 693–713. doi: 10.1093/aob/mcl114
- Liu, B., Li, H., Zhu, B., Koide, R. T., Eissenstat, D. M., and Guo, D. (2015). Complementarity in nutrient foraging strategies of absorptive fine roots and arbuscular mycorrhizal fungi across 14 coexisting subtropical tree species. *New Phytol.* 208, 125–136. doi: 10.1111/nph.13434
- Liu, L., Gundersen, P., Zhang, T., and Mo, J. (2012). Effects of phosphorus addition on soil microbial biomass and community composition in three forest types in tropical China. Soil Biol. Biochem. 44, 31–38. doi: 10.1016/j.soilbio.2011.08.017
- Lundberg, D. S., Yourstone, S., Mieczkowski, P., Jones, C. D., and Dangl, J. L. (2013). Practical innovations for high-throughput amplicon sequencing. *Nat. Methods* 10, 999–1002. doi: 10.1038/nmeth.2634
- Lynch, J. P., and Brown, K. M. (2001). Topsoil foraging an architectural adaptation of plants to low phosphorus availability. *Plant Soil* 237, 225–237. doi: 10.1023/A:1013324727040
- Mage, S. M., and Porder, S. (2013). Parent material and topography determine soil phosphorus status in the Luquillo Mountains of Puerto Rico. *Ecosystems* 16, 284–294. doi: 10.1007/s10021-012-9612-5
- Marschner, P., Crowley, D., and Rengel, Z. (2011). Rhizosphere interactions between microorganisms and plants govern iron and phosphorus acquisition along the root axis - model and research methods. *Soil Biol. Biochem.* 43, 883–894. doi: 10.1016/j.soilbio.2011.01.005
- McCormack, M. L., Dickie, I. A., Eissenstat, D. M., Fahey, T. J., Fernandez, C. W., Guo, D., et al. (2015). Redefining fine roots improves understanding of below-ground contributions to terrestrial biosphere processes. *New Phytol.* 207, 505–518. doi: 10.1111/nph.13363
- McMurdie, P. J., and Holmes, S. (2013). Phyloseq: an R package for reproducible interactive analysis and graphics of microbiome census data. *PLOS ONE* 8:e61217. doi: 10.1371/journal.pone.0061217
- Mommer, L., and Weemstra, M. (2012). The role of roots in the resource economics spectrum. New Phytol. 195, 725–727. doi: 10.1111/j.1469-8137.2012.04247.x
- Nannipieri, P., Giagnoni, L., Landi, L., and Renella, G. (2011). "Role of phosphatase enzymes in soil," in *Phosphorus in Action. Soil Biology*, Vol. 26, eds E. Bünemann, A. Oberson, and E. Frossard (Berlin: Springer), 215–243.
- Nautiyal, C. S. (1999). An efficient microbiological growth medium for screening phosphate solubilizing microorganisms. *FEMS Microbiol. Lett.* 170, 265–270. doi: 10.1016/S0378-1097(98)00555-2
- Niu, Y. F., Chai, R. S., Jin, G. L., Wang, H., Tang, C. X., and Zhang, Y. S. (2013). Responses of root architecture development to low phosphorus availability: a review. Ann. Bot. 112, 391–408. doi: 10.1093/aob/mcs285
- Oksanen, A. J., Blanchet, F. G., Kindt, R., Legendre, P., Minchin, P. R., Hara, R. B. O., et al. (2015). Package 'vegan '. Available at: https://cran.r-project.org, https://github.com/vegandevs/vegan
- Pan, Y., Birdsey, R. A., Phillips, O. L., and Jackson, R. B. (2013). The structure, distribution, and biomass of the world's forests. *Annu. Rev. Ecol. Evol. Syst.* 44, 593–622. doi: 10.1146/annurev-ecolsys-110512-135914
- Panhwar, Q. A. (2012). Isolation and characterization of phosphate-solubilizing bacteria from aerobic rice. *Afr. J. Biotechnol.* 11, 2711–2719. doi: 10.5897/AJB10. 2218
- Patton, C., and Kryskalla, J. (2003). U.S. Geological Survey National Water Quality Laboratory: Evaluation of Alkaline Persulfate Digestion as an Alternative to Kjeldahl Digestion for Determination of Total and Dissolved Nitrogen and Phosphorous in Water. USGS Water-Resources Investigations Report No. 03–4174. Denver, CO: National Water Quality Laboratory.
- Philippot, L., Raaijmakers, J. M., Lemanceau, P., and van der Putten, W. H. (2013). Going back to the roots: the microbial ecology of the rhizosphere. *Nat. Rev. Microbiol.* 11, 789–799. doi: 10.1038/nrmicro3109

- Pii, Y., Mimmo, T., Tomasi, N., Terzano, R., Cesco, S., and Crecchio, C. (2015). Microbial interactions in the rhizosphere: beneficial influences of plant growthpromoting rhizobacteria on nutrient acquisition process. a review. *Biol. Fertil. Soils* 51, 403–415. doi: 10.1007/s00374-015-0996-1
- Plassard, C., and Dell, B. (2010). Phosphorus nutrition of mycorrhizal trees. Tree Physiol. 30, 1129–1139. doi: 10.1093/treephys/tpq063
- Raven, J. A. (2015). Interactions between nitrogen and phosphorus metabolism. Annu. Plant Rev. 48, 187–214. doi: 10.1002/9781118958841.ch7
- Reasoner, D. J., and Geldreich, E. E. (1985). A new medium for the enumeration and subculture of bacteria from potable water. *Appl. Environ. Microbiol.* 49, 1–7.
- Reed, S. C., Yang, X., and Thornton, P. E. (2015). Incorporating phosphorus cycling into global modeling efforts: a worthwhile, tractable endeavor. *New Phytol.* 208, 324–329. doi: 10.1111/nph.13521
- Rejmánková, E., Sirová, D., and Carlson, E. (2011). Patterns of activities of root phosphomonoesterase and phosphodiesterase in wetland plants as a function of macrophyte species and ambient phosphorus regime. *New Phytol.* 190, 968–976. doi: 10.1111/j.1469-8137.2011.03652.x
- Richardson, A. E., Hocking, P. J., Simpson, R. J., and George, T. S. (2009). Plant mechanisms to optimise access to soil phosphorus. *Crop Pasture Sci.* 60, 124–143. doi: 10.1071/CP07125
- Richardson, A. E., and Simpson, R. J. (2011). Soil microorganisms mediating phosphorus availability update on microbial phosphorus. *Plant Physiol.* 156, 989–996. doi: 10.1104/pp.111.175448
- Saghir, K., Zaidi, A., and Ahmad, E. (2014). "Phosphate solubilizing microorganisms," in *Bioaugmentation Biostimulation and Biocontrol (LIBRO)*, Vol. 108, M. Saghir Khan, A. Zaidi, and J. Musarrat (Berlin: Springer).
- Satta, G., Grazi, G., Varaldo, P. E., and Fontana, R. (1979). Detection of bacterial phosphatase activity by means of an original and simple test. J. Clin. Pathol. 32, 391–395. doi: 10.1136/jcp.32.4.391
- Smith, S., and De Smet, I. (2012). Root system architecture: insights from Arabidopsis and cereal crops. Philos. Trans. R. Soc. Lond. B Biol. Sci. 367, 1441–1452. doi: 10.1098/rstb.2011.0234
- Smith, S. E., Jakobsen, I., Grønlund, M., and Smith, F. A. (2011). Roles of arbuscular mycorrhizas in plant phosphorus nutrition: interactions between pathways of phosphorus uptake in arbuscular mycorrhizal roots have important implications for understanding and manipulating plant phosphorus acquisition. *Plant Physiol.* 156, 1050–1057. doi: 10.1104/pp.111.174581
- Smith, S. E., and Smith, F. A. (2011). Roles of arbuscular mycorrhizas in plant nutrition and growth: new paradigms from cellular to ecosystem scales. *Annu. Rev. Plant Biol.* 62, 227–250. doi: 10.1146/annurev-arplant-042110-103846
- Spohn, M., Simone, N., Cormann, M., Schloter, M., and Fischer, D. (2015). Distribution of phosphatase activity and various bacterial phyla in the rhizosphere of *Hordeum vulgare* L. depending on P availability. *Soil Biol. Biochem.* 89, 44–51. doi: 10.1016/j.soilbio.2015.06.018
- Stone, M. M., and Plante, A. F. (2014). Changes in phosphatase kinetics with soil depth across a variable tropical landscape. *Soil Biol. Biochem.* 71, 61–67. doi: 10.1016/j.soilbio.2014.01.006
- Tabatabai, M. A., and Bremner, J. M. (1969). Use of *p*-nitrophenyl phosphate for assay of soil phosphatase activity. *Soil Biol. Biochem.* 1, 301–307. doi: 10.1016/ 0038-0717(69)90012-1
- Townsend, A. R., Cleveland, C. C., Houlton, B. Z., Alden, C. B., and White, J. W. C. (2011). Multi-element regulation of the tropical forest carbon cycle. *Front. Ecol. Environ.* 9:9–17. doi: 10.1890/100047
- Treseder, K. K., and Vitousek, P. M. (2001). Effects of soil nutrient availability on investment in acquisition of N and P in Hawaiin rain forests. *Ecology* 82, 946–954. doi: 10.1890/0012-9658(2001)082[0946:EOSNAO]2.0.CO;2
- Turner, B. L., and Engelbrecht, B. M. J. (2011). Soil organic phosphorus in lowland tropical rain forests. *Biogeochemistry* 103, 297–315. doi: 10.1007/s10533-010-9466-x
- Ushio, M., Kitayama, K., and Balser, T. C. (2010). Tree species-mediated spatial patchiness of the composition of microbial community and physicochemical properties in the topsoils of a tropical montane forest. *Soil Biol. Biochem.* 42, 1588–1595. doi: 10.1016/j.soilbio.2010.05.035
- Ushio, M., Wagai, R., Balser, T. C., and Kitayama, K. (2008). Variations in the soil microbial community composition of a tropical montane forest ecosystem: does tree species matter? *Soil Biol. Biochem.* 40, 2699–2702. doi: 10.1016/j.soilbio. 2008.06.023

- Vacheron, J., Desbrosses, G., Bouffaud, M.-L., Touraine, B., Moënne-Loccoz, Y., Muller, D., et al. (2013). Plant growth-promoting rhizobacteria and root system functioning. *Front. Plant Sci.* 4:356. doi: 10.3389/fpls.2013. 00356
- Valverde-Barrantes, O. J., Smemo, K. A., Feinstein, L. M., Kershner, M. W., and Blackwood, C. B. (2013). The distribution of below-ground traits is explained by intrinsic species differences and intraspecific plasticity in response to root neighbours. J. Ecol. 101, 933–942. doi: 10.1111/1365-2745. 12087
- Van Raij, B. B., Cantarella, H., and Quaggio, J. A. (2009). Ion exchange resin for assessing phosphorus availability in soils. *Better Crops* 93, 23–25.
- Vitousek, P. M., and Sanford, R. L. (1984). Nutrient cycling in moist tropical forest. Annu. Rev. Ecol. Syst. 17, 137–167. doi: 10.1146/annurev.es.17.110186.001033
- Vos, M., Wolf, A. B., Jennings, S. J., and Kowalchuk, G. A. (2013). Micro-scale determinants of bacterial diversity in soil. *FEMS Microbiol. Rev.* 37, 936–954. doi: 10.1111/1574-6976.12023
- Warren, J. M., Hanson, P. J., Iversen, C. M., Kumar, J., Walker, A. P., and Wullschleger, S. D. (2015). Root structural and functional dynamics in terrestrial biosphere models-evaluation and recommendations. *New Phytol.* 205, 59–78. doi: 10.1111/nph.13034
- Weintraub, M. N. (2011). "Biological phosphorus cycling in arctic and alpine soils," in *Phosphorus in Action. Soil Biology*, Vol. 26, eds E. Bünemann, A. Oberson, and E. Frossard (Berlin: Springer).

- White, P. J., and Hammond, J. P. (2008). "Phosphorus nutrition of terrestrial plants," in *The Ecophysiology of Plant-Phosphorus Interactions*, Vol. 7, eds P. White and J. P. Hammond (Dordrecht: Springer), 51–81. doi: 10.1007/978-1-4020-8435-5_4
- Yang, X., Post, W. M., Thornton, P. E., and Jain, A. (2013). The distribution of soil phosphorus for global biogeochemical modeling. *Biogeosciences* 10, 2525–2537. doi: 10.5194/bg-10-2525-2013
- Yang, X., Thornton, P. E., Ricciuto, D. M., and Hoffman, F. M. (2016). Phosphorus feedbacks constraining tropical ecosystem responses to changes in atmospheric CO₂ and climate. *Geophys. Res. Lett.* 43, 7205–7214. doi: 10.1002/ 2016GL069241.Received

Conflict of Interest Statement: The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

Copyright © 2017 Cabugao, Timm, Carrell, Childs, Lu, Pelletier, Weston and Norby. This is an open-access article distributed under the terms of the Creative Commons Attribution License (CC BY). The use, distribution or reproduction in other forums is permitted, provided the original author(s) or licensor are credited and that the original publication in this journal is cited, in accordance with accepted academic practice. No use, distribution or reproduction is permitted which does not comply with these terms.