



Pilot-Scale Testing of UV-A Light Treatment for Mitigation of NH₃, H₂S, GHGs, VOCs, Odor, and O₃ Inside the Poultry Barn

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Poultry farmers are producing eggs, meat, and feathers with increased efficiency and lower carbon footprint. Technologies to address concerns about the indoor air quality inside barns and the gaseous emissions from farms to the atmosphere continue to be among industry priorities. We have been developing and scaling up a UV air treatment that has the potential to reduce odor and other gases on the farm scale. In our recent laboratory-scale study, the use of UV-A (a less toxic ultraviolet light, a.k.a. “black light”) and a special TiO₂-based photocatalyst reduced concentrations of several important air pollutants (NH₃, CO₂, N₂O, O₃) without impact on H₂S and CH₄. Therefore, the objectives of this research were to (1) scale up the UV treatment to pilot scale, (2) evaluate the mitigation of odor and odorous volatile organic compounds (VOCs), and (3) complete preliminary economic analyses. A pilot-scale experiment was conducted under commercial poultry barn conditions to evaluate photocatalyst coatings on surfaces subjected to UV light under field conditions. In this study, the reactor was constructed to support interchangeable wall panels and installed on a poultry farm. The effects of a photocatalyst’s presence (photocatalysis and photolysis), UV intensity (LED and fluorescent), and treatment time were studied in the pilot-scale experiments inside a poultry barn. The results of the pilot-scale experiments were consistent with the laboratory-scale one: the percent reduction under photocatalysis was generally higher than photolysis. In addition, the percent reduction of target gases at a high light intensity and long treatment time was higher. The percent reduction of NH₃ was 5–9%. There was no impact on H₂S, CH₄, and CO₂ under any experimental conditions. N₂O and O₃ concentrations were reduced at 6–12% and 87–100% by both photolysis and photocatalysis. In addition, concentrations of several VOCs responsible for livestock odor were reduced from 26 to 62% and increased with treatment time and light intensity. The odor was reduced by 18%. Photolysis treatment reduced concentrations of N₂O, VOCs, and O₃, only. The initial economic analysis has shown that LEDs are more efficient than fluorescent lights. Further scale-up and research at farm scale are warranted.

Keywords: air pollution, odor, indoor air quality, ammonia, emissions, poultry production, photocatalysis, concentrated animal feeding operations

INTRODUCTION

Poultry farmers are producing eggs, meat, and feathers with increasing efficiency and a lower carbon footprint. Technologies to address concerns about the air quality inside barns and the gaseous emissions from farms to the atmosphere continue to be among industry priorities. Only ~25% of research on technologies to mitigate emissions from animal production systems has been tested on farms, and scaling up technologies from the laboratory to farm scales has proven to be a challenge. However, farmers prefer technologies that are simple to adopt and low in cost. Therefore, pilot- and farm-scale experiments should be preceded for the technical application on the livestock farm. According to previous studies (Maurer et al., 2016), the treatment of odorous emissions with UV light has the potential to be a relatively simple adaptation to existing and new facilities. Yet many practical questions remain to be addressed to scale up this technology from the laboratory to the farm.

Proposed UV-based methods consist of direct irradiation and photocatalytic treatment. In the former, light is directly absorbed by the target gases (or potentially other photoreactive gases). By contrast, photocatalysis consists of the light being principally absorbed by a photoactive coating with secondary reactions between target gases and reactive intermediates taking place mainly at or near the coated surfaces. UV irradiation treatment has a variety of mitigation effects on the target gas, depending on several parameters. Direct irradiation generally requires shorter wavelength light (e.g., 254 nm), in that few target molecules absorb in the near UV (e.g., 360 nm), whereas photocatalytic treatment can proceed with the longer wavelengths because of choice of coating. Light flux is another obvious parameter. Other critical variables in the real application are the presence of other agents in the gas mixture, such as the degree of humidity, and presence or absence of ozone.

The wavelength of irradiation is an important consideration for the application of UV technology because of its potential effects on humans and animals. The shortest, easily accessible wavelengths (e.g., 185 nm) will cause a build-up of ozone and N₂O because the light is absorbed by atmospheric O₂. Traditional bactericidal light sources (often 254 nm) do not cause that problem but would cause rapid sunburn for exposed skin. Near-UV light (often called UV-A) that is sufficient for photocatalysis is also not appropriate for long-term skin exposure but is otherwise the most benign. The most common photocatalyst is nano-particulate TiO₂, which is chosen for its relatively broad application, comparatively high efficiency, durability, lack of toxicity, and low cost (Hashimoto et al., 2005; Zaleska, 2008; Rockafellow et al., 2012; Schneider et al., 2014).

Photocatalysis is initiated when photons of sufficient energy (i.e., greater than the semiconductor band gap) are absorbed by the TiO₂ particles, resulting in electron/hole (e⁻/h⁺) generation (Vautier et al., 2001; Schneider et al., 2014; Lee et al., 2018; Maurer and Koziel, 2019). For commonly available TiO₂, the threshold is roughly 380 nm. Under atmospheric moisture conditions, HO• (hydroxyl) radicals are produced by the interaction of hole (h⁺) with H₂O molecular (Vautier et al., 2001; Nakata and

Fujishima, 2012; Lee et al., 2018; Maurer and Koziel, 2019). The most common electron sink is molecular oxygen. Either through these reactive intermediates or by direct interaction with the e⁻ or h⁺, the target materials are degraded. Full mineralization (conversion to CO₂, H₂O, and inorganic ions) can usually be achieved through exhaustive treatment. Although the detailed mechanism of photocatalysis varies with different target pollutants and treatment conditions, it is commonly agreed that the primary reactions responsible are these interfacial oxidation and reduction reactions (Abe, 2010; Maeda and Domen, 2010; Nakata and Fujishima, 2012; Schneider et al., 2014).

The applicability of UV photocatalytic technology to the farm has been investigated in previous studies. Photocatalysis based on TiO₂ has been evaluated to reduce odorous gases and fine particulate concentrations as well as for increased feed conversion rates (Guarino et al., 2008; Costa et al., 2012; Zhu et al., 2017; Maurer and Koziel, 2019; Yang et al., 2020). In addition, optimal conditions and parameters affecting target gas mitigation have been investigated (Lee et al., 2020). Moreover, economic analysis has shown that it is reasonable compared with other technologies (Koziel et al., 2008; Liu et al., 2015).

Here, we report a study of the mitigation of odorous target gases, volatile organic compounds (VOCs), greenhouse gases (GHGs), and odor using UV-A treatment in actual poultry farms. This study was brought up to a pilot scale based on the lessons learned about UV-A performance of the photocatalysis demonstrated in a recent laboratory-scale study (Lee et al., 2020). The results provided evidence that photocatalysis with TiO₂ coating and UV-A light can reduce gas concentrations of NH₃ (3–19%), CO₂ (4%), N₂O (7–12%), and O₃ (12–48%) without significant effect on H₂S and CH₄. In addition, the mitigation of target gases was generally improved with parameters that sensibly dictate higher effective dosages: the presence of photocatalyst, relative humidity (RH, 12%), higher light intensity, longer treatment time, and low dust accumulation on the photocatalyst surface. However, it was found that the optimum mitigation conditions (RH and dust accumulation) and the effect of parameters (light intensity and treatment time) depend on the type of targeted gas.

Poultry barns manage lighting very precisely owing to bird physiological and production needs. Thus, the poultry industry is generally prepared to consider the adoption of light-based technologies. However, carefully scaled-up studies are still needed to move forward UV-based treatment for air quality improvements without jeopardizing current production practices. To date, no studies report on applying actual photocatalysis technology in real poultry farms. Therefore, the objectives of this research are (1) to scale up the UV treatment to pilot scale in the poultry farm, (2) to evaluate the mitigation of NH₃, H₂S, GHGs, odor, and odorous VOCs, and (3) to complete a preliminary economic analysis. In addition, the effects of photocatalyst presence (in comparison with direct photolysis), UV intensity (based on LED vs. fluorescent light sources), and treatment time were studied in the pilot-scale experiments inside a poultry barn.

Experimental Fan Calibration

The fan calibration was performed by measuring the velocity of each 0.012-m distance triplicated from the top of the fan and averaging a total of 8 points air velocity (Table 1). The dimmer switch was adjusted to measure the maximum, medium, and minimum air velocity of the fan. These three velocities were named as setting 1, 2, and 3 from the fastest speed order. In the experiments, fan velocity used settings 2.0, 2.5, and 3.0 to set typical air exchange rates inside mechanically ventilated barns. The treatment times in the inside of the reactor corresponding to the three air velocities are 40, 100, and 170 s, respectively. The air velocity of the fan was measured using a wind speed sensor (Modern Device, Providence, RI, USA) with Arduino Uno (Arduino LLC, Boston, MA, USA). The wind sensor was calibrated using the value of volts from the sensor at the WTM-1000 mini wind tunnel (Omega Engineering, Norwalk, CT, USA) from 0 to 10 m·s⁻¹ (Figure 2).

UV-A Light Sources

Two fluorescent light bulbs were combined with one lamp, and a total of two lamps were used in this study. The LED used an aluminum board equipped with 108 LED chips. Light intensity was measured with an ILT-1700 radiometer (International Light Technologies, Peabody, MA, USA) equipped with an NS365 filter and SED033 detector (International Light Technologies, Peabody, MA, USA). For economic analysis, power consumption was measured using a wattage meter (P3, Lexington, NY, USA). The LED had ~10 times greater intensity and lower power consumption compared with the fluorescent lamp (Table 2 and Figure A1). The effective exposure of light intensity was mapped on the interior surfaces of the reactor (Figure 3).

Teaching Poultry Farm

Pilot-scale testing was conducted at ISU Poultry Teaching Farm (Ames, IA, USA). The study was not using animals, and they were not exposed to UV-A. The performance of UV-A reactor was tested in realistic barn conditions where the environmental parameters (temperature, relative humidity, ventilation), dust, and gases were representative of the conditions inside poultry barns in general. The teaching farm is a caged facility with about 200 laying hens (Figure 4A). Animal density is 0.045–0.056 m³ head⁻¹. Once a day, the manure was cleaned manually using a scraper. The teaching farm was set up with a side ventilation system in which the flow rate of the fan was automatically

changed according to the temperature. The UV-A reactor was located in a nearby manure collector site (Figure 4B).

Ammonia and Hydrogen Sulfide

A gas monitoring system (OMS-300; Smart Control & Sensing, Daejeon, Republic of Korea) equipped with electrochemical gas sensors of Membrapor (Wallisellen, Switzerland) was used to measure NH₃ (NH₃/CR-200) and H₂S (H₂S/C-50) concentrations. Both gas sensors were calibrated with standard gases before the experiment. The calibration result of both sensors was $R^2 > 0.98$.

Volatile Organic Compounds

Samples were collected and analyzed in the same method as in the previous study (Maurer and Koziel, 2019). Air samples for VOC measurements were collected using 1-L glass gas sampling bulbs (Supelco, Bellefonte, PA, USA). Air samples were taken using a portable vacuum sampling pump (Leland Legacy; SKC, Eighty-Four, PA, USA) with a set flow rate of 5 L min⁻¹ for 1 min and analyzed in the same day. Chemical analyses of poultry odorants were completed using a solid-phase microextraction (SPME) (50/30 μm DVB/CAR/PDMS, 2-cm-long fibers; Supelco) using static extraction for 1 h at room temperature and gas chromatography–mass spectrometry (GC-MS) system for analyses (Agilent 6890 GC; Microanalytics, Round Rock, TX, USA).

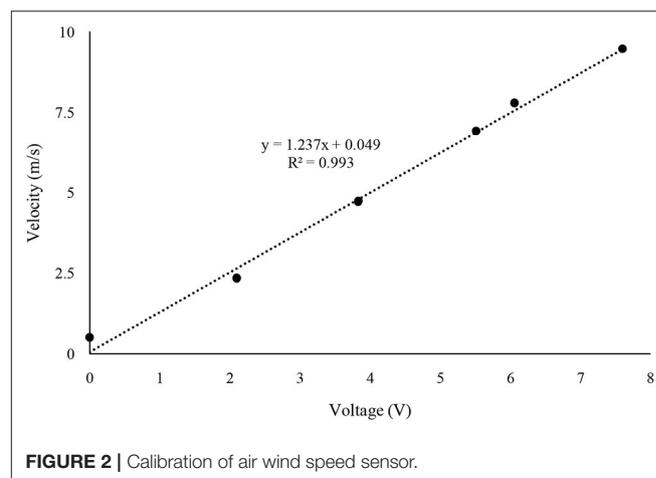
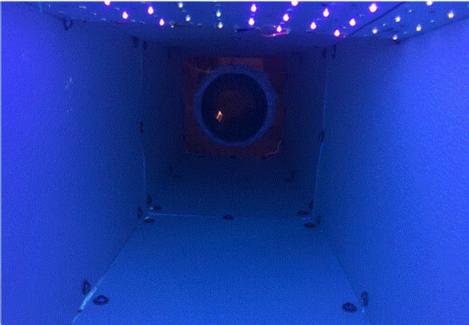


FIGURE 2 | Calibration of air wind speed sensor.

TABLE 1 | Calibration of air velocity.

Velocity control setting	Distance from top of fan (m)								Average (m s ⁻¹)
	0.013	0.025	0.037	0.05	0.062	0.074	0.087	0.099	
1.0	1.54	1.50	1.22	0.81	0.81	1.03	1.00	1.09	1.14 ± 0.31
2.0	1.27	1.21	0.92	0.45	0.49	0.55	0.64	0.60	0.71 ± 0.32
2.5	0.54	0.35	0.26	0.17	0.19	0.25	0.29	0.29	0.29 ± 0.12
3.0	0.30	0.23	0.21	0.11	0.10	0.13	0.16	0.17	0.18 ± 0.07

TABLE 2 | Comparison of experimental UV-A lamps.

	Fluorescent	LED
Total light intensity ($\text{mW}\cdot\text{cm}^{-2}$)	0.44	4.85
Power (W)	48.2	43.3
Lamp's position inside the reactor		

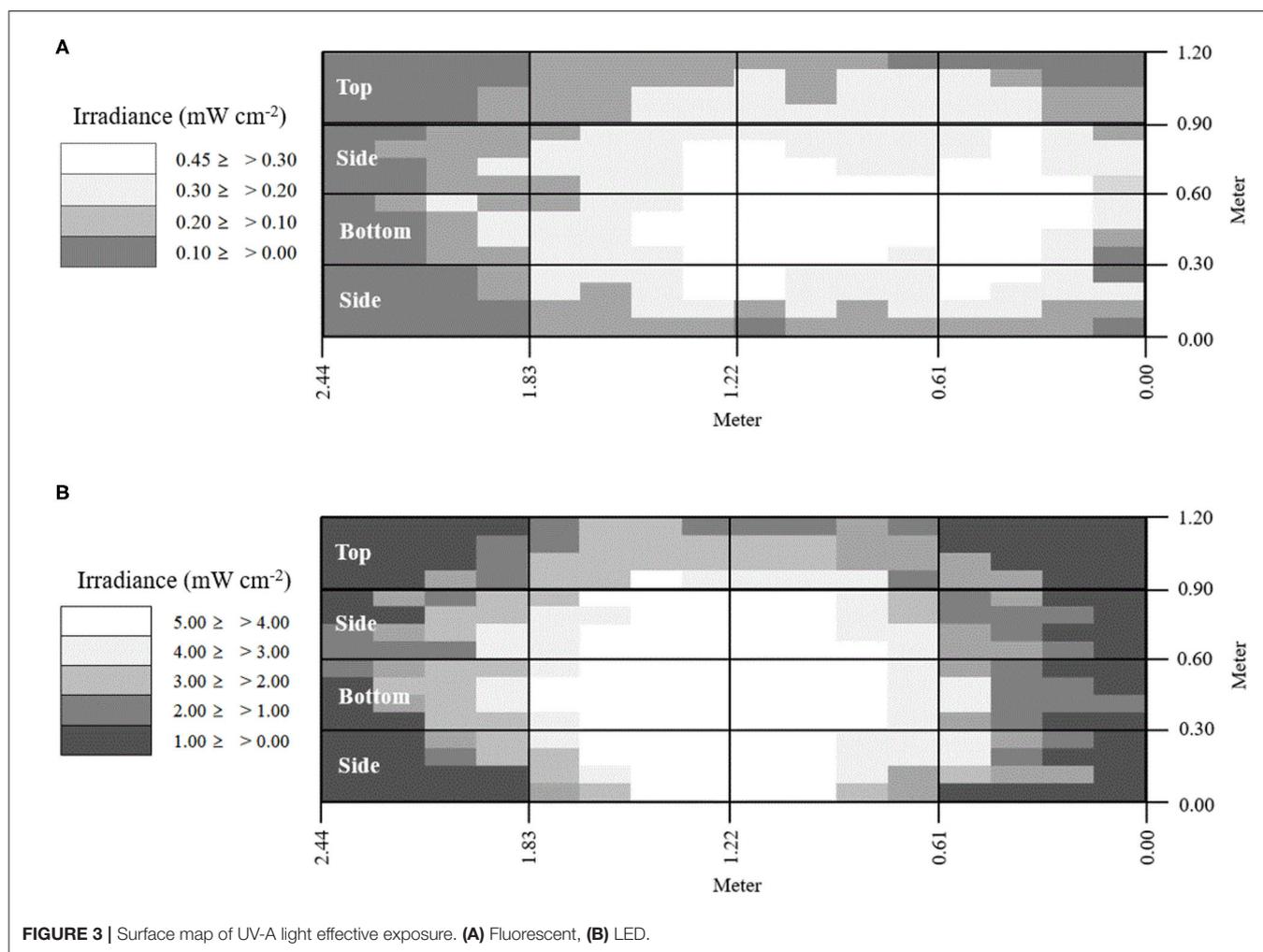




FIGURE 4 | Pilot-scale mitigation was tested on a teaching poultry farm. **(A)** Interior of the teaching poultry barn, **(B)** UV-A reactor.

Greenhouse Gases

Greenhouse gases (GHGs) methane (CH_4), carbon dioxide (CO_2), and nitrous oxide (N_2O) were measured. GHG samples were collected using syringes and 5.9-ml Exetainer vials (Labco Limited, UK) and were analyzed for GHG concentrations on a GC equipped with flame ionization and electron capture detectors (SRI Instruments, Torrance, CA, USA). Samples were analyzed on the day of collection. Standard calibrations were constructed daily using 10.3 and 20.5 ppm CH_4 , 1,005 and 4,010 ppm CO_2 , and 0.101 ppm and 1.01 ppm N_2O , and pure helium used to 0 ppm (Air Liquide America, Plumsteadville, PA, USA).

Ozone

The O_3 detector was connected to the monitoring system (Series 500 monitor; Aeroqual, New Zealand) and installed at the sampling site inside the reactor. The ozone sensor (OZS, Aeroqual) was sent to the professional company (Gas Sensing, IA, USA) for certified calibration before use. The detection range was from 0 to 50 ppb.

Odor

Odor samples were collected from the incoming and outgoing air sampling ports of the UV reactor in 10-L Tedlar sample bags using a Vac-U-Chamber (SKC) and sampling pump (SKC). Tedlar sampling bags were pre-cleaned by flushing with clean air three times before use. Odor samples were analyzed using a dynamic triangular forced-choice olfactometer (St. Croix Sensory, Stillwater, MN, USA). Four trained panelists at two repetitions each were used in the analysis of each sample. Each sample was presented to the panelists from low concentration to higher concentrations; each dilution level doubled the concentration of the odor.

Data Measurement and Analysis

Gas samples were collected after 30 min of equilibration time under each treatment condition. The overall mean

percent reduction (mitigation) for each measured gas was estimated using:

$$\% \text{ Reduction} = \frac{E_{\text{con}} - E_{\text{Treat}}}{E_{\text{con}}} \times 100 \quad (1)$$

where E_{Con} and E_{Treat} are the mean measured concentrations in control and treated air, respectively.

Emission rates were calculated as a product of measured gas concentrations and the total airflow rate through the wind tunnel, adjusted for standard conditions and dry air using collected environmental data. The overall mean mitigation of each measured gas emission was estimated using

$$\text{Mitigation of } E \left(\text{g} \cdot \text{min}^{-1} \right) = \left(C_{\text{con}} \times V \times \frac{273.15 \times MW}{(K_{\text{con}}) \times 2.24 \times 10^4} - C_{\text{Treat}} \times V \times \frac{273.15 \times MW}{(K_{\text{Treat}}) \times 2.24 \times 10^4} \right) \quad (2)$$

where mitigation of $E \text{ (g min}^{-1}\text{)}$ is the mitigation of gas emission. C_{Con} and C_{Treat} are the mean measured concentrations in control and treated air (mL m^{-3}), respectively. V is the ventilation rate ($\text{m}^3 \text{min}^{-1}$). MW is the molecular weight of target gas (g mol^{-1}). K_{Con} and K_{Treat} are the temperature in control and treated air (K), respectively. The 2.24×10^4 is an ideal gas conversion factor for liters to moles at 273.15 K.

The electric power consumption was calculated as the measured power at each UV-A light source. Electric power consumption (kWh) was calculated using Equation 3.

$$\text{Electric power consumption (kWh)} = P \times T_s \times 3,600^{-1} \times 10^{-3} \quad (3)$$

where P is the power (W) and T_s is the treatment time (s).

Energy efficiency was calculated using the power consumption per mitigation of gas emission at treatment time (Equation 4), and the power consumption was changed as cost. The cost (\$) was calculated by converting the power consumption (kWh) into dollars (12 cents is 1 kWh).

$$\text{Energy efficiency (kWh}\cdot\text{g}^{-1}\cdot\text{min)} = \frac{\text{The electric power consumption (kWh)}}{\text{Reduction of E (g}\cdot\text{min}^{-1})} \quad (4)$$

Statistical Analysis

The program of R (version 3.6.2) was used to analyze the mitigation effect of UV-A irradiative treatment for target gases on the poultry farm. The parameters of catalyst, lamp type, and treatment time between control concentration and treatment concentration were statistically analyzed using one-way ANOVA. The statistical difference was confirmed by obtaining the *p*-value through the paired Tukey test. A significant difference was defined for a *p* < 0.05 in this study.

RESULTS

Environmental Parameters

The average temperature inside the poultry barn was $25 \pm 3^\circ\text{C}$, and the average RH was $59 \pm 4\%$. The average temperature of the gas after the fluorescent UV-A light irradiation and LED was 27 ± 2 and $28 \pm 3^\circ\text{C}$, respectively. In addition, RH decreased by about 3% after UV irradiation. Therefore, after the UV-A light irradiation, the temperature was increased, but RH was decreased. It was confirmed that the concentration of target gases was significantly reduced because of the increased ventilation in the farm from 12:00 (e.g., NH_3 concentration is 10 ± 1 ppm at 12:00). Therefore, all samples were collected before 10:00.

Ammonia and Hydrogen Sulfide

The average NH_3 concentration in the poultry barn without treatment was 23 ± 3 ppm (Table A1). In the case of photolysis (Table 3 and Figure A2), there was no statistically significant NH_3 mitigation in all experimental conditions (*p* > 0.05). For photocatalysis (Figure A4 and Table A4), the percent NH_3 reduction was 5% when using fluorescent UV-A and 9% when using LED UV-A (*p* < 0.05). However, there was no significant mitigation difference between treatment time of 100 s and 170 s (*p* > 0.05) when using the LED source. The H_2S concentrations were too low for measurement with the instrumentation used (0–5 ppb). Thus, reductions could not be estimated.

Volatile Organic Compounds

In the case of photolysis (Table 4), three VOCs showed statistically significant percent reductions and were limited to the longest treatment time (170 s), i.e., dimethyl disulfide (25.8% under fluorescent UV-A), *p*-cresol (35.6%), and indole (31.4%) under LED UV-A. In the case of photocatalysis, five VOCs showed significant percent reductions, i.e., *p*-cresol (32.2% under fluorescent UV-A), and dimethyl disulfide, butanoic acid, *p*-cresol, indole, and skatole showed 31.2–61.9% reductions under LED UV-A. In general, treatment with the LED at the

TABLE 3 | Summary of NH_3 mitigation with UV-A treatment.

Treatment time (s)	Photolysis		Photocatalysis	
	Fluorescent	LED	Fluorescent	LED
40	1.7 (0.35)	0.1 (0.95)	−0.2 (0.89)	2.5 (0.07)
100	1.1 (0.29)	0.9 (0.22)	2.6 (0.07)	8.1 (0.04)
170	0.7 (0.23)	1.2 (0.25)	5.2 (0.04)	8.7 (0.01)

Values in table report % reduction (*p*-value). Bold font means a statistical difference (*p* < 0.05).

longest treatment time (170 s) was most effective in reducing the concentrations of most VOCs measured.

Greenhouse Gases

The average GHGs concentration in the poultry barn was 2.5 ± 0.2 (CH_4), 465 ± 48 (CO_2), and 0.28 ± 0.03 (Table A2, N_2O) ppm. For CH_4 and CO_2 , there was no statistically different change in their concentration (Table 5). However, N_2O showed under photocatalysis an average 6% reduction with fluorescent UV-A and 9% with LED UV-A. In addition, The photolysis using an only LED UV light source still showed a 7% reduction. The N_2O mitigation showed a higher percent reduction with the high light intensity and with longer treatment time (*p* < 0.05).

Odor

Except for one of 12 treatments, the odor mitigation was represented (Table 6). Odor percent reduction was statistically different only in long treatment time and high light intensity conditions (LED, 170 s). The odor percent reduction was presented at 18% (*p* < 0.05). The odor unit ($\text{OU}_e\cdot\text{m}^{-3}$) decreased from 582 ± 25 to 475 ± 38 in condition with statistical differences (Table A3).

Ozone

The average O_3 concentration in control samples was 9.0 ± 4.7 ppb (Table 7). The average O_3 concentration after treatments was 0.3 ± 1.2 (ppb). Both photolysis and photocatalysis showed a similar tendency to decrease O_3 concentration (Figures A3, A5 and Table A5). The LED-based irradiations showed a greater reduction than fluorescent (100 vs. 87%). However, there was no statistically significant difference between them (*p* > 0.05).

DISCUSSION

Comparison With Laboratory-Scale and Pilot-Scale Experiments

This pilot-scale study showed the percent reduction in the concentrations of NH_3 , N_2O , O_3 , and some types of VOCs in poultry barn exhaust as a result of direct and indirect photolytic treatment. The results show similar trends as the laboratory experiment that used the same UV-A light source and TiO_2 density. A comparison of the laboratory and pilot scales are summarized in Table 8. On the laboratory scale (Lee et al., 2020), an RH of 12% was reported as the optimal humidity condition for reducing target gases. As the treatment time decreased, the

TABLE 4 | Mitigation of VOCs.

Target compound	Fluorescent						LED					
	Treatment time (s)											
	40		100		170		40		100		170	
	Photo lysis	Photo catalysis	Photo lysis	Photo catalysis	Photo lysis	Photo catalysis	Photo lysis	Photo catalysis	Photo lysis	Photo catalysis	Photo lysis	Photo catalysis
DMDS	16 (0.62)	16 (0.69)	-24 (0.66)	22(0.30)	34 (0.20)	26 (0.54)	21 (0.13)	38 (0.37)	18 (0.14)	33 (0.26)	40 (0.06)	35 (0.30)
DEDS	14 (0.26)	23 (0.26)	-2 (0.90)	13 (0.67)	26 (0.03)	19 (0.06)	-9 (0.38)	18 (0.66)	14 (0.28)	26 (0.08)	19 (0.51)	47 (0.02)
DMTS	-7 (0.81)	18 (0.27)	11 (0.82)	52 (0.05)	30 (0.21)	36 (0.34)	-64 (0.42)	1 (0.99)	10 (0.36)	8 (0.51)	-16 (0.41)	43 (0.08)
AA	26 (0.69)	-46 (0.68)	-10 (0.68)	39 (0.68)	26 (0.62)	59 (0.23)	27 (0.52)	-44 (0.16)	0 (0.99)	44 (0.37)	-17 (0.49)	75 (0.42)
PA	-47 (0.56)	29 (0.59)	1.0 (0.99)	33 (0.54)	-8 (0.88)	51 (0.09)	-53 (0.40)	-5 (0.92)	-2 (0.97)	4 (0.92)	-14 (0.23)	36 (0.47)
IA	1 (0.95)	20 (0.76)	-33 (0.39)	-21 (0.82)	10 (0.74)	-16 (0.44)	6 (0.56)	6 (0.76)	-6 (0.87)	-44 (0.53)	-21 (0.23)	29 (0.53)
BA	-51 (0.50)	28 (0.06)	-38 (0.61)	6 (0.87)	27 (0.27)	22 (0.71)	17 (0.75)	22 (0.11)	-7 (0.88)	16 (0.79)	-4 (0.57)	62 (<0.01)
IVA	-33 (0.70)	-21(0.20)	-14 (0.84)	34 (0.09)	-19 (0.57)	38 (0.30)	-36 (0.66)	-40 (0.64)	-44 (0.14)	23 (0.70)	-49 (0.07)	40 (0.41)
Phenol	-42 (0.69)	-25 (0.39)	-39 (0.50)	-37 (0.61)	33 (0.14)	-3 (0.92)	-22 (0.57)	-5 (0.84)	16 (0.11)	-3 (0.94)	28 (0.31)	-27 (0.63)
<i>p</i> -Cresol	3 (0.96)	31 (0.56)	23 (0.79)	32 (0.04)	26 (0.26)	11 (0.90)	5 (0.64)	22 (0.64)	8 (0.73)	52 (0.04)	36 (0.01)	49 (0.03)
Indole	4 (0.85)	-6 (0.90)	5 (0.95)	45(0.17)	19 (0.59)	54 (0.11)	0 (0.98)	15 (0.23)	12 (0.27)	31 (0.03)	31 (0.02)	21 (0.69)
Skatole	-3 (0.91)	47 (0.06)	-34 (0.83)	-36 (0.42)	3 (0.92)	19 (0.80)	10 (0.78)	54 (0.18)	8 (0.85)	12 (0.90)	-7 (0.93)	35 (0.04)

Dimethyl disulfide (DMDS), diethyl disulfide (DEDS), dimethyl trisulfide (DMTS), acetic acid (AA), propanoic acid (PA), isopentanoic acid (IA), butanoic acid (BA), isovaleric acid (IVA). Values in table report % reduction (*p*-values), bold font signifies statistical significance.

TABLE 5 | Comparison of greenhouse gas mitigation.

T time ^a (s)	CH ₄				CO ₂				N ₂ O			
	Photolysis		Photocatalysis		Photolysis		Photocatalysis		Photolysis		Photocatalysis	
	Flu ^b	LED	Flu	LED	Flu	LED	Flu	LED	Flu	LED	Flu	LED
40	-3.9 (0.61)	-2.2 (0.56)	2.4 (0.15)	-4.3 (0.07)	7.4 (0.15)	-1.4 (0.65)	2.3 (0.12)	5.3 (0.28)	2.2 (0.21)	3.8 (0.31)	1.2 (0.45)	5.3 (0.06)
100	-0.6 (0.31)	-6.8 (0.14)	-2.5 (0.28)	2.6 (0.31)	1.9 (0.09)	-2.7 (0.22)	-0.2 (0.55)	1.8 (0.42)	2.3 (0.35)	2.6 (0.42)	5.6 (0.04)	6.9 (0.02)
170	-1.7 (0.09)	0.3 (0.93)	3.8 (0.25)	1.7 (0.48)	0.4 (0.88)	11.8 (0.32)	-1.2 (0.58)	7.3 (0.05)	1.2 (0.32)	7.3 (0.02)	7.5 (0.01)	12.5 (0.01)

^aTreatment time; ^bFluorescent. Values in table report % reduction (*p*-value). Bold font means a statistical difference (*p* < 0.05).

light intensity decreased, the relative humidity increased, and dust accumulation increased, the percent reduction gradually decreased. As a representative example, when the RH increased from 12 to 60%, the percent reduction of NH₃ declined from 19 to 6%. Therefore, the mitigation for the target gas was expected to be low, considering the actual RH and dust inside the poultry farm.

However, unlike these concerns, there was a statistically significant decrease in NH₃, N₂O, and O₃. In the case of NH₃ concentration, the percent reduction increased with the increase of the light intensity. However, it did not show a statistical

difference with the treatment time (100 vs. 170 s). For N₂O concentration, the percent reduction improved as the light intensity, and the treatment time increased. O₃ was detected in very low concentrations in the poultry farm; then, it was almost mitigated after UV irradiation. Interestingly, CO₂ did not show a percent reduction effect, which is a different result compared with that from the laboratory scale. It is considered as a result reflecting that there is no chemical reason for the reduction of CO₂ under general photocatalysis conditions with normal TiO₂. Pilot studies also showed that photocatalysis has the effect of

improving indoor air quality by reducing some types of VOC and odor. Therefore, the application of photocatalysis in the poultry farm through laboratory-scale and pilot-scale results is considered to be a potential and positive technology in terms

of reducing odorous gases, decreasing GHGs, and improving indoor air quality.

Economic Analysis

The advent of inexpensive LED lighting has reduced the cost per photon significantly below what it was with previous fluorescent lamp technology. Clearly, there remains a direct relationship between treatment time and energy cost, however. Here, we consider the cost of reducing 1 g of the target gas, considering only statistically significant reduction results. If NH₃ is the target gas, the cost was found to be \$0.7 to \$3.6 to remove 1 g of material per minute (Table 9) depending on the light intensity. A cost of \$32.1–\$79.0 was estimated to remove 1 g of N₂O per minute (Table 10), depending on the light intensity. For NH₃ and N₂O, however, removal is not predicted with statistical significance, for periods of high ventilation rates, such as summer, i.e., when treatment times are short (i.e., 40 s in the current apparatus). It is likely that increased light intensity or methods that would increase contact (treatment) time in the barn (with correspondingly higher energy costs) could be made effective. Importantly, the estimated cost is considered reasonably expensive compared with other odor treatment techniques and equipment presented in previous studies (Koziel et al., 2008; Maurer et al., 2017). In terms of cost, shorter treatment time is better (100 s) (Tables 9, 10). However, it is longer treatment time (170 s), which results in a percent reduction for more target gases (VOCs and odor, Table 8), that would be considered more suitable for farm application.

In 170 s of treatment time with LED lighting, we estimate that ~548 mg of NH₃, ~21 mg of N₂O, and ~3.3 mg of O₃ could be removed from the atmosphere for the cost of \$1 per minute (Table 11). For reference, the NH₃ emissions in a poultry barn can be ~1 g NH₃ bird⁻¹ day⁻¹ (Wheeler et al., 2003; Xin

TABLE 6 | Mitigation of odor.

Treatment time (s)	Odor			
	Photolysis		Photocatalysis	
	Fluorescent	LED	Fluorescent	LED
40	0.1 (0.98)	2.2 (0.52)	6.8 (0.43)	15.7 (0.22)
100	15.0 (0.31)	-21.9 (0.14)	9.3 (0.15)	7.3 (0.25)
170	4.9 (0.59)	9.6(0.64)	7.6 (0.38)	18.4 (0.01)

Values in table report % reduction (*p*-value). Bold font means a statistical difference (*p* < 0.05).

TABLE 7 | Mitigation of ozone concentration (ppb).

Treatment time (s)	UV type	Photolysis		Photocatalysis	
		Control	Treatment	Control	Treatment
		40	Fluorescent	10.0 ± 2.7	0.0 ± 0.0
	LED	11.3 ± 1.5	0.0 ± 0.0	7.7 ± 1.2	0.0 ± 0.0
100	Fluorescent	17.3 ± 2.5	2.3 ± 4.0	6.7 ± 1.5	0.0 ± 0.0
	LED	8.7 ± 1.2	0.0 ± 0.0	2.7 ± 1.5	0.0 ± 0.0
170	Fluorescent	6.3 ± 1.5	0.7 ± 1.2	13.3 ± 2.1	0.0 ± 0.0
	LED	12.3 ± 1.2	0.0 ± 0.0	5.3 ± 1.2	0.0 ± 0.0

Values in table report average concentration ± standard deviation. All figures are statistically significant.

TABLE 8 | Percent reduction of target gases with UV-A photocatalysis in the poultry farm condition.

References	Experiment conditions	Average % reduction of target gases					
		NH ₃	CO ₂	N ₂ O	O ₃	VOCs	Odor
Lee et al. (2020)	Laboratory scale Temp ^a : 25 ± 3°C RH ^b : 12% T time ^c (s): 200	18.7	3.8	9.5	48.4	Not investigated	Not investigated
	Laboratory scale Temp ^a : 25 ± 3°C RH ^b : 60% T time ^c (s): 40	5.1	1.1	5.2	27.6	Not investigated	Not investigated
	Laboratory scale Temp ^a : 25 ± 3°C RH ^b : 60% T time ^c (s): 200	6.2	9.3	4.9	37.5	Not investigated	Not investigated
This study	Pilot scale Temp ^a : 28 ± 3°C RH ^b : 56% T time ^c (s): 40	2.5	5.3	5.3	100	No % reduction	15.7
	Pilot scale Temp ^a : 28 ± 3°C RH ^b : 56% T time ^c (s): 100	8.1	1.8	6.9	100	31.4 (Indole)	7.3
	Pilot scale Temp ^a : 28 ± 3°C RH ^b : 56% T time ^c (s): 170	8.7	7.3	12.5	100	47.2 (DEDS) 61.9 (BA) 49.3 (p-Cresol) 35.4 (Skatole)	18.4

^aTemperature; ^bRelative humidity; ^cTreatment time; diethyl disulfide (DEDS), butanoic acid (BA); bold font signifies a statistical difference (*p* < 0.05).

et al., 2003) or ~ 0.7 mg NH_3 bird $^{-1}$ min $^{-1}$. UV-A technology with 170 s treatment time can remove 66 mg of NH_3 per minute while consuming 1 kWh of electricity (Table 11), an equivalent of 100% mitigation of NH_3 emissions from ~ 100 birds. Therefore, NH_3 emitted from the farm with ~ 100 birds can be removed by changing the light bulb installed on the farm to UV-A light. However, research on reducing the lighting cost while reducing emissions by replacing the ordinary lights with LED UV-A lights is warranted.

Evaluation of Photocatalysis With UV-A Light Based on TiO_2 in the Livestock Environment

Table 8 summarizes previous research on the mitigation of selected target gases via photocatalysis with UV-A in livestock-relevant environmental conditions. The combined results show that photolysis with UV-A with assistance by TiO_2 coatings yields significant reductions of some target gases. However, in all instances of Table 12, mitigation is more effective with the use of TiO_2 coatings and photocatalytic assistance.

The combination of UV-A with TiO_2 showed a reduction in NH_3 , H_2S , GHGs, O_3 , VOCs, and odor. NH_3 reduced about 9–31% depending on the light intensity, the treatment time, and the coating density, whereas H_2S showed a 4–14% reduction depending on light intensity. In the case of NH_3 , the mitigation was reported to be 31% at the farm scale, whereas at the laboratory scale and pilot scale it showed 10 to 20%. The effect

of NH_3 reduction is difficult to directly compare in both of the previous researches because of the difference of TiO_2 coating thickness and light intensity. However, in the farm-scale study, the NH_3 concentration in the farm (1.9 ppb) is lower than the general NH_3 concentration in the swine farm. Therefore, further research is needed. H_2S showed a reduction effect in the experiment of the laboratory scale using a standard gas of 1,000 ppm as a control. However, the exact reduction could not be determined in the pilot-scale and the other laboratory-scale study because of a low concentration (below 5 ppm).

For the GHGs, the previous laboratory-scale study did not find a significant reduction in CH_4 using photocatalysis,

TABLE 11 | Economic analysis when using photocatalysis with 170 s treatment time under LED.

Target gases	NH_3	N_2O	O_3
Reduction rate (%)	8.7 ± 1.0	12.4 ± 2.1	100 ± 0.0
Reduction of target gases emission ($\mu\text{g}\cdot\text{min}^{-1}$)	135 ± 22.6	5.1 ± 0.9	0.8 ± 0.2
Energy efficiency of mitigating target gases ($\text{kWh}\cdot\text{g}^{-1}\cdot\text{min}$)	15.5 ± 2.4	407 ± 70.2	2,610 ± 645
Reduction of gas emission-power consumption ($\text{mg}\cdot\text{min}^{-1}\cdot\text{kWh}^{-1}$)	65.8 ± 11.0	2.5 ± 0.5	0.4 ± 0.1
Reduction of gas emission per one dollar ($\text{mg}\cdot\text{min}^{-1}\cdot\text{dollar}^{-1}$)	548 ± 91.9	20.9 ± 3.8	3.3 ± 0.7

TABLE 9 | Economic analysis of NH_3 .

Treatment time (s)	Fluorescent			LED		
	40	100	170	40	100	170
Reduction rate (%)	-0.2 ± 2.4	2.6 ± 1.2	5.2 ± 1.7	2.5 ± 1.2	8.1 ± 3.2	8.7 ± 1.0
Reduction of NH_3 emission ($\mu\text{g}\cdot\text{min}^{-1}$)	15.7 ± 109.9	59.8 ± 22.9	81.3 ± 26.1	171.1 ± 62.8	229.1 ± 77.0	134.5 ± 22.6
Energy efficiency ($\text{kWh}\cdot\text{g}^{-1}\cdot\text{min}$)	3.2 ± 7.2	25.2 ± 11.4	30.4 ± 11.3	3.0 ± 1.0	5.8 ± 2.4	15.5 ± 2.4
Cost per reduction of NH_3 emission ($\text{dollar}\cdot\text{g}^{-1}\cdot\text{min}$)	0.4 ± 0.9	3.0 ± 1.4	3.6 ± 1.4	0.4 ± 0.1	0.7 ± 0.3	1.9 ± 0.3

Bold font signifies a statistical difference ($p < 0.05$).

TABLE 10 | Economic analysis of N_2O .

Treatment time (s)	Fluorescent			LED		
	40	100	170	40	100	170
Reduction rate (%)	1.2 ± 2.3	5.5 ± 1.9	7.5 ± 1.5	5.3 ± 2.2	6.9 ± 1.5	12.4 ± 2.1
Reduction of N_2O emission ($\mu\text{g}\cdot\text{min}^{-1}$)	2.1 ± 3.8	4.1 ± 1.4	3.6 ± 0.7	8.6 ± 3.8	4.6 ± 1.0	5.1 ± 0.9
Energy efficiency ($\text{kWh}\cdot\text{g}^{-1}\cdot\text{min}$)	17 ± 287	362 ± 150	658 ± 137	66 ± 34	268 ± 62	407 ± 70
Cost per reduction of N_2O emission ($\text{dollar}\cdot\text{g}^{-1}\cdot\text{min}$)	2.0 ± 34.5	43.5 ± 18.0	79.0 ± 16.5	7.9 ± 4.0	32.1 ± 7.4	48.8 ± 8.4

Bold font signifies a statistical difference ($p < 0.05$).

TABLE 12 | Summary of % reduction for selected target gas with TiO₂ and UV-A light in the livestock farm condition.

Reference	Experiment conditions	Coating material	UV type (wavelength)	Light intensity	Target gas (% ave ^a reduction)	
					Photolysis	Photocatalysis
Yao and Feilberg (2015)	Simulated livestock facility (laboratory) Tem ^b : Not reported RH ^c : 50% T time ^d (s): 0.24	TiO ₂ (1.5 m ² · g ⁻¹)	UV-A (315–400 nm) Main: 368 nm	2.3–5.6 mW · cm ⁻²	Not reported	H₂S (4.2–14.0) MT (80.1–87.4) DMS (92.2–95.8) DMDS (82.5–90.8) 1-Butanol (92.6–95.1) AA (81.0–88.5) PA (96.5–97.8) BA (98.4–99.2) VA (98.8–99.0)
Costa et al. (2012)	Swine farm (farm) Tem ^b : 26°C RH ^c : 56% T time ^d (s): 72	TiO ₂ (7 mg · cm ²)	UV-A (315–400 nm)	0.05–0.45 lux	Not reported	NH ₃ (2.0) CH₄ (27.4) CO ₂ (–4.5) N ₂ O (–0.8) PM₁₀ (17.0) PM _{2.5} (–8.1)
Guarino et al. (2008)	Swine farm (farm) Tem ^b : 24°C RH ^c : 54% T time ^d (s): 363	TiO ₂ (7 mg · cm ²)	UV-A (315–400 nm)	Not reported	Not reported	NH₃ (30.5) CH₄ (10.8) CO₂ (15.3) N ₂ O (4.2)
Zhu et al. (2017)	Simulated swine farm (laboratory) Tem ^b : 40°C RH ^c : 40% T time ^d (s): 40, 200	TiO ₂ (10 μg · cm ²)	UV-A (365 nm)	0.06 mW · cm ⁻²	Not reported	DMDS (35.0, 40.4) DEDS (27.7, 81.0) DMTS (37.1, 76.3) BA (62.2, 86.9) Guaiacol (37.4, 100.0) p-Cresol (27.4, 93.8)
(Maurer and Koziel, 2019)	Swine farm (pilot) Tem ^a : 22–26°C RH ^c : 36–80% T time ^d (s): 24, 47	TiO ₂ (10 μg · cm ²)	UV-A (365 nm)	<0.04 mW · cm ⁻²	N₂O (4.2, 7.6)	CH ₄ (–1.4, –2.2) CO₂ (0.0, –3.1) N₂O (7.3, 8.7) AA (–52.9, –19.7) Odor (not reported, 16.3) p-Cresol (–21.4, 22.0)
(Lee et al., 2020)	Simulated poultry farm (laboratory) Tem ^b : 25 ± 3°C RH ^c : 12% T time ^d (s): 40, 200	TiO ₂ (10 μg · cm ²)	UV-A (365 nm)	4.85 mW · cm ⁻²	NH ₃ (0.0, 0.0) CO ₂ (4.0, 2.3) N₂O (3.3, 6.5) O₃ (26.0, 12.9)	NH₃ (10.4, 18.7) CO₂ (4.2, 3.8) N₂O (8.9, 9.5) O₃ (37.0, 48.4)
This study	Poultry farm (pilot) Tem ^b : 28 ± 3°C RH ^c : 56% T time ^d (s): 40, 170	TiO ₂ (10 μg · cm ²)	UV-A (365 nm)	4.85 mW · cm ⁻²	NH ₃ (0.1, 1.2) CO ₂ (–1.4, 11.8) N₂O (3.8, 7.3) O₃ (100, 100) p-Cresol (5.2, 35.6) Indole (0.3, 31.4)	NH₃ (2.5, 8.7) CO₂ (5.3, 7.3) N₂O (5.3, 12.5) O₃ (100, 100) Odor (15.7, 18.4) DEDS (18.1, 47.2) BA (22.1, 61.9) p-Cresol (21.8, 49.3) Skatole (53.6, 35.4)

^aMean; ^bTemperature; ^cRelative humidity; ^dTreatment time; methanethiol (MT), dimethyl sulfide (DMS), dimethyl disulfide (DMDS), diethyl disulfide (DEDS), dimethyl trisulfide (DMTS), acetic Acid (AA), propionic acid (PA), butanoic acid (BA), valeric acid (VA). Bold font signifies a statistical difference ($p < 0.05$).

despite the well-known oxidation of hydrocarbons under other photocatalytic conditions. However, here, moderate reductions (11–27%) were observed with the larger scale and high coating thickness. A surprising slight decrease in CO₂ was observed in the laboratory-scale and farm-scale experiments, but none is reported here. In general, CO₂ is the oxidative endpoint for photocatalytic oxidation of virtually all carbon-containing compounds under conditions like those used here, and thus its mitigation would not derive from its chemical removal. N₂O was found to decrease by 7–13% in the laboratory and pilot scales, but no statistically significant reduction was reported in the farm-scale experiments.

In the case of O₃, the concentration decreases during the UV-A irradiation, which was shown in this study and previous studies on a laboratory scale. The concentration of O₃ showed more reduction at the pilot scale (100%) than at the laboratory scale (48%). It is considered that the O₃ concentration is further reduced to decompose the target gases because more target gases and malodorous organic compounds are present in the farm.

VOCs have been reported to reduce some types of organosulfur compounds, volatile fatty acids, phenols, and alcohol with photocatalysis. In the simulated livestock environment, it was reported that percent reduction was showed in methanethiol, dimethyl sulfide, dimethyl disulfide, and dimethyl trisulfide of organosulfur compounds, in acetic acid, propionic acid, butanoic acid, and valeric acid of volatile fatty acids, in *p*-cresol and guaiacol of phenols, and in 1-butanol. In the case of the poultry environment, the percent reduction was reported in diethyl disulfide of organosulfur compounds, in butanoic acid of volatile fatty acids, in *p*-cresol of phenol, and in indole and skatole, respectively. Odor mitigation was reported to decrease by 16–18%, reflecting the mitigation of VOCs and other target gas.

However, some people have a skeptical view of photocatalysis because of the by-products after photocatalysis. Through the results of previous studies (Table 13), some hazardous by-products such as N₂O, SO₄²⁻ (sulfate), CH₃OH (methanol), methane by-products (C₂H₆, C₂H₄, CH₂O₂), and HCHO

TABLE 13 | List of major by-products after photocatalysis.

References	Target gases	Major by-products after photocatalysis
Jayanty et al. (1976), Levine and Calvert (1977), Mozzanega et al. (1979), Cant and Cole (1992)	NH ₃	N ₂ O and N ₂
Canela et al. (1998), Portela et al. (2007, 2008, 2010), Alonso-Tellez et al. (2012)	H ₂ S	SO ₂ and SO ₄ ²⁻
Taylor (2003), Yuliaty and Yoshida (2008), Chen et al. (2016)	CH ₄	CH ₃ OH, H, C ₂ H ₆ , C ₂ H ₄ , CH ₂ O ₂ , and CO ₂
Tan et al. (2006), Paramasivam et al. (2012)	CO ₂	CH ₄ , CH ₃ OH, HCHO, and CO
Blyholder and Tanaka (1971), Obalová et al. (2013)	N ₂ O	N ₂ and O ₂
Lin et al. (2002)	O ₃	O ₂ and O
Koziel et al. (2008)	VOCs	Partially oxidized species, CO ₂ , and H ₂ O

(formaldehyde) were reported. However, H₂S, CH₄, and CO₂, which produce harmful by-products, did not show statistical mitigation in this study. In addition, no research has been conducted to investigate the by-product concentrations observed in the poultry farm after photocatalysis. It is assumed that the target gases of concern here would be at low concentrations, and the generated by-products will not likely be generated at harmful levels. Further research and scaling up to poultry barn applications are needed. Therefore, although further studies on the by-product and economic suitability are needed, the potential for photocatalytic technology with UV-A in the livestock farm is still favorably considered because of the mitigation of NH₃, GHGs, O₃, VOCs, and odor.

CONCLUSION

The results of the study provide evidence that a photocatalyst using TiO₂ coating and UV-A light can reduce the target gas concentrations in poultry farm conditions. The photocatalysis reduced NH₃, N₂O, O₃, VOCs, and odor. However, it did not affect H₂S, CH₄, and CO₂. In the case of NH₃ concentration, the percent reduction ranged from 5 to 9% and was affected by light intensity. However, there was no statistical increase with increasing treatment time. The percent reduction in N₂O and O₃ concentrations increased with increasing light intensity and treatment time (6–12% for N₂O and 87–100% for O₃). For VOCs, greater light intensity (LED) and longer treatment time (170 s) improved mitigation. The percent reduction was observed for DEDS (26–47%), BA (62%), *p*-cresol (32–49%), indole (31%), and skatole (35%) concentrations. The odor showed a statistical reduction of 18% only under LED with 170 s treatment time. Preliminary economic analysis showed that the cost of removing 1 g of target gas per minute was \$1.9 ± 0.3 for NH₃, \$48.8 ± 8.4 for N₂O, and \$312.8 ± 77.4 for O₃ in the most effective conditions (LED with 170 s treatment time). The application of photocatalysis based on

TiO₂ with UV-A in the poultry farm is therefore considered to be potentially beneficial in terms of reducing odorous gases, decreasing GHGs, removing O₃, and improving indoor air quality. Further research needs to be extended to farm-scale trials for investigating more detailed economic analysis and mitigation of target gases. In addition, it is also necessary to investigate the by-products after photocatalysis treatment for safe technology applications.

DATA AVAILABILITY STATEMENT

All datasets presented in this study are included in the article/supplementary material.

AUTHOR CONTRIBUTIONS

JK and WJ: conceptualization. JK: methodology. ML, JW, HA, and JK: validation. ML: formal analysis and writing—original draft preparation. ML, PL, BC, ZM, and CB: investigation. PL, BC, ZM, CB, and JK: resources. ML and JK: data curation. ML, HA, JK, and WJ: writing—review and editing. ML: visualization. HA and JK: supervision. JK and WJ: project administration and funding acquisition. All authors contributed to the article and approved the submitted version.

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SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fchem.2020.00613/full#supplementary-material>

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Conflict of Interest: The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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