



The Application of PMA (Propidium Monoazide) to Different Target Sequence Lengths of Zebrafish eDNA: A New Approach Aimed Toward Improving Environmental DNA Ecology and Biological Surveillance

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Environmental DNA (eDNA) analysis has enabled more sensitive and efficient biological monitoring than traditional methods. However, since the target species is not directly observed, interpretation of results cannot preclude process Type I errors. Specifically, there may be a spatial or temporal gap between the target eDNA and the eDNA source in the sampled area. Moreover, eDNA surveillance lacks the ability to distinguish whether eDNA originated from a living or non-living source. This kind of Type I error is difficult to control for, in part, because the relationship between the state of eDNA (i.e., intracellular or extracellular) and the degradation rate is still unclear. Here, we applied PMA (Propidium monoazide) to eDNA analysis which enabled us to differentiate “intact cells” from “disrupted cells.” PMA is a dye that has a high affinity for double-stranded DNA and forms a covalent bond with double-stranded DNA and inhibits amplification of the bonded DNA molecules by PCR. Since PMA is impermeable to the cell membrane, DNA protected by an intact cell membrane can be selectively detected. In this study, we investigated the workability of PMA on vertebrate eDNA using zebrafish, *Danio rerio*. Aquarium water was incubated for 1 week to monitor the eDNA degradation process of both intracellular and extracellular eDNA. We developed ten species-specific quantitative PCR assays for *D. rerio* with different amplification lengths that enabled independent quantification of total eDNA (sum of intracellular and extracellular eDNA, commonly measured in other studies) and intracellular eDNA (DNA in intact cells) and allow for analyses of sequence length-dependent eDNA degradation in combination with PMA. We confirmed that PMA is effective at differentiating “intact” and “disrupted” fish cells. We found that total eDNA and intracellular eDNA have different degradation processes

that are dependent on the length of target sequence. For future conservation efforts using eDNA analyses, it is necessary to increase the reliability of the analysis results. The research presented here provides new analysis tools that expand our understanding of the ecology of eDNA, so that more accurate and reliable conclusions can be determined.

Keywords: environmental DNA, decay rate, propidium monoazide, quantitative PCR, zebrafish

INTRODUCTION

Aquatic ecosystems are disproportionately affected by anthropogenic influences, such as pollution, habitat degradation, introduction of invasive species, and overuse of natural resources (Abell et al., 2008; Strayer and Dudgeon, 2010; Collen et al., 2014). Conservation efforts that are used to help mitigate the damage caused to aquatic ecosystems have to be empirically monitored to determine which strategies are effective. One way to monitor the effectiveness of various conservation strategies is to survey the biodiversity within a system. Surveillance of biodiversity includes monitoring the distribution of species of interest (e.g., alien or endangered species), estimating population abundance or biomass of targeted species, estimating occupancy at a site, or assessing the presence or absence of species in targeted systems (Lodge et al., 2012; Pilliod et al., 2013, 2014; Thompson, 2013). Traditional monitoring of the activity ranges and habitat usages of aquatic organisms typically includes a variety of surveillance approaches (e.g., direct capture or visual surveys), which employ different gears and techniques that can be difficult to learn and standardize, and require a considerable amount of labor and cost (Hayes et al., 1996; Rees et al., 2014; Hajibabaei et al., 2016; Evans et al., 2017).

There has been growing interest over the last decade in eDNA (environmental DNA) surveillance as a biological monitoring method for aquatic species, due to several advantages over traditional methods. Costs, labor, and environmental disturbance associated with eDNA surveillance are often smaller compared to direct capture or visual surveys because eDNA surveys only require small volumes of water to be collected at targeted sites (Jerde et al., 2011; Takahara et al., 2013; Smart et al., 2015, 2016; Evans et al., 2017). Furthermore, eDNA monitoring has also been demonstrated to have higher detection sensitivity compared to traditional methods, especially at low target species densities (Dejean et al., 2012; Pilliod et al., 2013; Takahara et al., 2013). Since the target species is not directly observed, however, the results of eDNA analyses can potentially include errors in ecological interpretation (process Type I errors) (Darling and Mahon, 2011; Rees et al., 2014; Barnes and Turner, 2016; also see Taberlet et al., 2018). For example, in both natural and experimental systems, legacy eDNA of the target species can still be accurately amplified and detected even though there are no living individuals of the target species in the system (Merkes et al., 2014; Lance et al., 2017; Tsuji et al., 2017; Kamoroff and Goldberg, 2018). Simply, positive detection of legacy eDNA reported as “presence” of the target species are not analytical false-positives (method-based Type I error), but ecological misinterpretations (*sensu* Darling and Mahon, 2011). While method-based Type I errors are fairly easy to control for in eDNA designs (e.g.,

inclusion of negative controls during all field, extraction, and PCR processes, multiple biological and technical replicates) (Champlot et al., 2010; Darling and Mahon, 2011; Ficetola et al., 2015), there are no empirically tested protocols that allow eDNA surveys to reduce or eliminate process Type I errors, aside from tandem traditional surveillance methods.

Our understanding of eDNA ecology regarding the existing states and degradation rates that influence eDNA detection is limited, in part, due to the inability to demarcate spatial or temporal relationships between eDNA and the source of eDNA. For example, positive eDNA detection could result from DNA derived from dead individuals or transported from another site via bird droppings (Merkes et al., 2014). Furthermore, it is unknown when the detected eDNA was released from the target species or if the individual is still in the immediate surveyed area (Barnes and Turner, 2016; Jo et al., 2017). Several studies using experimental systems have demonstrated that eDNA signals can be detected many days (17–25) after the target organism is removed from the system (Dejean et al., 2011, 2012; Goldberg et al., 2013). Environmental DNA from common carp (*Cyprinus carpio*) remained detectable in sediment over 132 days (Turner et al., 2015). Moreover, eDNA that has an anthropogenic origin could be a source of process Type I error as shown in a reported case where effluent from fish markets along the Maizuru Bay area (Japan) yielded positive eDNA detection for Japanese jack mackerel (*Trachurus japonicus*) (Yamamoto et al., 2016). Clearly, a contributing factor of process Type I error in eDNA analysis and interpretation is the limited spatio-temporal nature of eDNA caused by various aspects associated with the origin, state, and fate of eDNA (Barnes and Turner, 2016).

In addition to the aforementioned issues related to eDNA decay, various biotic/abiotic factors associated with the state of eDNA also impact the detectability and persistence of eDNA in various aquatic environments (Thomsen et al., 2012a,b; Maruyama et al., 2014; Pilliod et al., 2014; Lance et al., 2017; Mächler et al., 2018; Murakami et al., 2019). Turner et al. (2014) suggested that eDNA is released from organisms as relatively large particles (1–10 μm), indicating that eDNA shed from fish is likely within cells and mitochondria, at least at the time point of release. In other words, eDNA is released as intracellular eDNA, which are relatively large particles, that undergo degradation and change their physical state and structure to become smaller (extracellular) particles (Jo et al., 2019). Moreover, Jo et al. (2017) clarified that long eDNA fragments degraded faster than shorter eDNA fragments using two different qPCR assays for Japanese jack mackerel and suggested the potential of longer eDNA fragments as a better proxy for the presence of the target fish. The concentrations of longer eDNA fragments also gave better correlation with estimations of fish distribution/biomass (based

on quantitative echo intensity) than shorter eDNA fragments (Jo et al., 2017). These conclusions suggest that the length of a target sequence in eDNA analysis is a key factor for reliable detection and determination of true presence for the target organism. Thus, the degradation process as it relates to DNA fragment length requires further experimental testing and clarification.

The eDNA degradation patterns observed so far have been measured as the degradation of total eDNA (DNA of target sequence in samples commonly measured in other studies) containing both extra- and intracellular eDNA. It is suggested that DNA molecules within intact cell membranes (intracellular DNA) would be less vulnerable to the attacks from microbes and extracellular enzymes in the environment than extracellular DNA (Levy-Booth et al., 2007). Therefore, the independent examination of both intra- and extracellular eDNA would also be beneficial to understanding the mechanisms of the eDNA degradation process. Clarification of the state of eDNA and the dynamics of degradation will contribute to improve the reliability of eDNA analysis and interpretations for more accurate decision making in future conservation efforts.

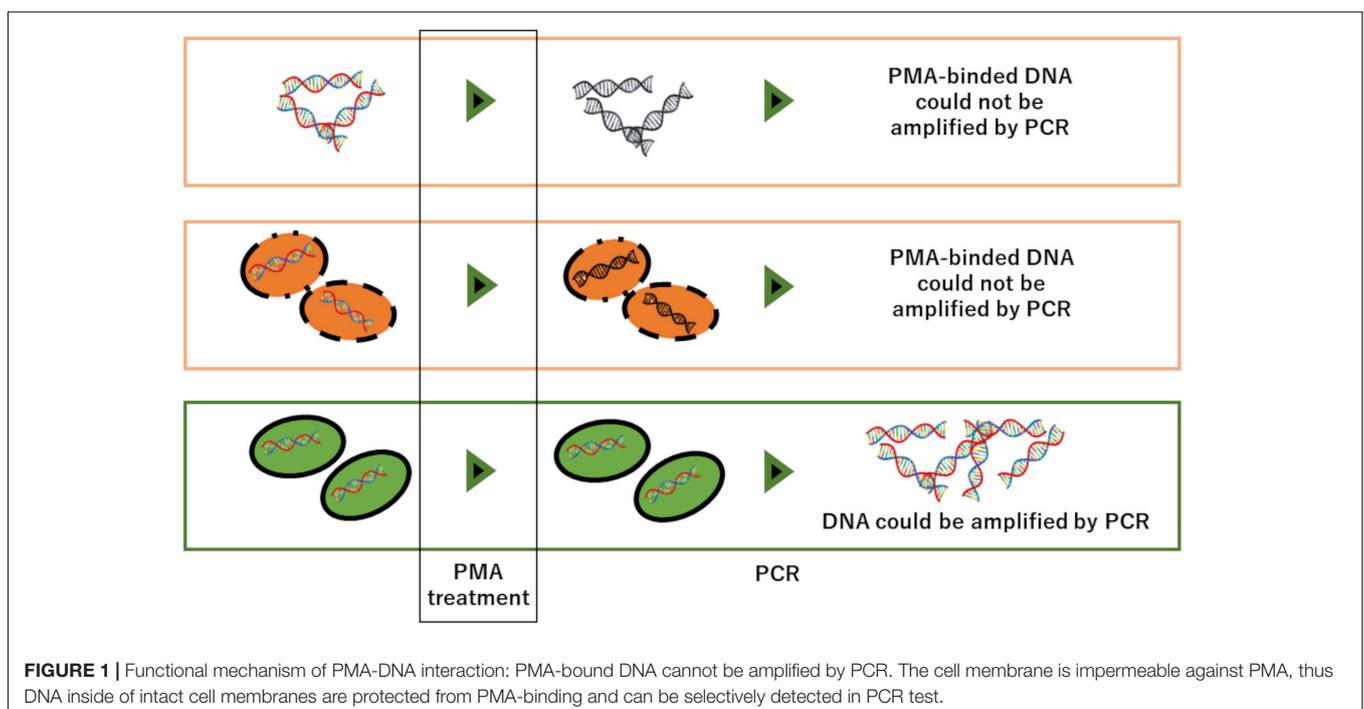
Here, we applied PMA (propidium monoazide) to eDNA analysis to differentiate “intact cells” from “disrupted cells,” which has been mainly used in microbial research. PMA is a photoreactive dye that has a high affinity for double-stranded DNA and forms a covalent bond with double-stranded DNA when exposed to strong visible light, and inhibits amplification of the bonded DNA molecules by PCR (Nocker et al., 2007). In addition, since PMA is impermeable to the cell membrane, DNA protected by an intact cell membrane can be selectively detected (Figure 1; García-Fontana et al., 2016; Emerson et al., 2017). Although there are a few research examples which used PMA on other taxa such as phytoplankton (*Microcystis*

aeruginosa, *Anabaena* sp., *Aphanizomenon* sp., *Synechocystis* sp., *Cryptomonas ovata*, *Scenedesmus obliquus*, and *Nitzschia apiculata*; Joo et al., 2019) and shellfish (*Dreissena polymorpha*; Lance and Carr, 2012), to the best of our knowledge, there are no research examples which applied PMA on vertebrates such as fish.

In this study, we investigated the workability of PMA on vertebrate eDNA using zebrafish, *Danio rerio*. Water samples collected from aquaria were boiled to disrupt the membrane of fish cells in the water to make mock “damaged” samples. PMA was applied to each of the boiled samples and the non-boiled samples, and detectable eDNA copy numbers were quantified using quantitative real-time PCR (qPCR). Next, water samples were collected from the aquaria and incubated for 1 week to monitor the eDNA degradation process between intracellular and extracellular eDNA. For examining the length-dependent degradation of eDNA, we developed 10 species-specific qPCR assays for *D. rerio* with different amplification lengths. The combined usage of PMA and the 10 qPCR assays enabled independent quantification of total eDNA (commonly measured in other studies) and intracellular eDNA (DNA in intact cells). We believe that clarification on the eDNA state (i.e., eDNA from intact vs. disrupted cells) and their respective degradation rates will help provide a basis for reducing process Type I errors and the continual improvement of eDNA techniques.

MATERIALS AND METHODS

In this study, two types of experiments were conducted using *D. rerio* as a model organism: (1) confirming if PMA could be effective for preventing amplification of extracellular DNA, and limiting amplification to DNA from intact cells or organelles, (2)



determining the degradation rate for total eDNA (intracellular and extracellular) and for eDNA in intact cells only, using PMA. In experiment 1, a control experiment was performed to quantify eDNA in PMA-treated and non-treated samples by qPCR. In experiment 2, the time-dependent degradation of eDNA was measured for total eDNA and intracellular eDNA by analyzing PMA-treated and non-treated eDNA samples from *D. rerio* tank water. Water samples obtained from the *D. rerio* tank were stored in a water bath and incubated at the same temperature as the *D. rerio* tank. Thereafter, the incubated water samples were sampled in a time series (0, 1, 2, 3, 5, and 7 days). After filtration, extraction and purification of eDNA from the sample, DNA copy numbers were quantified by qPCR using 10 newly developed qPCR assays with different amplification lengths which share the same probe and reverse primer but have 10 different forward primers. The DNA concentrations from all samples were analyzed to compare the degradation rate and degradation pattern of each treatment. Copy numbers of intracellular eDNA and total eDNA (intra- and extracellular eDNA) were all measured as the copy numbers of target sequence fragments in the eDNA samples by qPCR using the new assays.

Assay Design

Sequences of the mitochondrial genes including *cytb* (cytochrome *b*), tRNA-Glu and ND6 of the target species *D. rerio* as well as the other non-target species were downloaded from NCBI (the National Center for Biotechnology Information)¹. The accession numbers for both target and non-target species used are shown in **Supplementary Table 1**. The non-target species were *Cyprinus carpio*, *Carassius auratus auratus*, *Opsariichthys platypus*, which are kept in our fish keeping facility on a regular basis and belong to the same family (*Cyprinidae*) as *D. rerio*. The sequences were aligned using MAFFT v. 7² with default settings. We searched for base pair differences between *D. rerio* and other non-target species in the 3' end of each primer to generate

species-specific assays. Expected amplicon lengths generated by each assay were 132, 225, 333, 430, 529, 621, 715, 823, 935, and 1,021 bp, respectively (**Table 1**). All primers had a minimum of 2 bp differences between the target species and all non-target species, with at least 1 bp difference in the last 5 bases of the 3' end, except the forward primer for 529 bp assay (**Supplementary Figure 1**). The probe had 2 bp mismatches between the target species and other non-target species. The probe was designed between the primers for the shortest target sequence length (132 bp) and therefore can be used for all assays (**Supplementary Figure 1**). We confirmed the specificity of the primer pairs *in silico* using the Primer-BLAST with default settings³. The workability of the assays was tested on the target *in vitro* by qPCR in triplicate with a 10-fold dilution series generated from an artificially synthesized DNA fragment (gBlocksTM: Integrated DNA Technologies, Inc., Coralville, IA, United States). We defined the limit of detection (LOD) as the lowest number of copies that could be detected in one of the three qPCR replicas and the limit of quantification (LOQ) as the lowest number of copies that could be detected in all of the three qPCR replicas (**Table 1**). The details of qPCR settings are shown in a section "qPCR Conditions."

Experiment 1: PMA Confirmation Test

An acrylic 54 L aquarium was filled with 20 L of dechlorinated tap water and used as an experimental tank and *D. rerio* was held at a density of 5 individuals · L⁻¹. Water temperature was adjusted to 25°C and a photoperiod was set to light: dark = 12:12 h. An air stone and a heater were installed, and water sampling was performed after fish were acclimatized for 3 days. All equipment, including the aquarium used for the sampling, were all decontaminated with 10% bleach solution prior to use.

A disposable plastic cup was used to transfer 6 L of water from the aquarium into a 10 L plastic container. The 6 L water sample was thoroughly mixed by shaking, and was divided into two 3 L samples (contained in 4 L plastic containers). One 3 L sample

¹<https://www.ncbi.nlm.nih.gov>

²<https://mafft.cbrc.jp/alignment/server/index.html>

³<https://www.ncbi.nlm.nih.gov/tools/primer-blast/>

TABLE 1 | Sequences of primers and a probe used in this study.

Target species	Primer/Probe	Name of primer/probe	Oligo name	Sequence (5'→3')	Amplicon length	LOD (Copies)	LOQ (Copies)
<i>Danio rerio</i>	Forward	Dre-cytb-V132F	F1	CTTACGTGGGAGATACCTAGTG	132 bp	30	30
	Forward	Dre-cytb-V225F	F2	TAATAATAACAGCTTTTGTGGGCTACG	225 bp	3	3
	Forward	Dre-cytb-V333F	F3	CTTCCTTCTTTCATCTGCCTG	333 bp	30	30
	Forward	Dre-cytb-V430F	F4	TACACCTCAGACATCTCAACAGC	430 bp	3	3
	Forward	Dre-cytb-V529F	F5	CCAACGCCACTAAATATTTTCAGCG	529 bp	30	30
	Forward	Dre-tRNA-Glu-V621F	F6	TGTTGTAGTTCAACTACAAGAAGCTGC	621 bp	3	3
	Forward	Dre-ND6-V715F	F7	CATACCCCCAACTAGAGCTGC	715 bp	3	30
	Forward	Dre-ND6-V823F	F8	AGACAAAAATGAACCCCATAACTAAC	823 bp	3	3
	Forward	Dre-ND6-V935F	F9	GTCAAAACACCACCGGTCAC	935 bp	3	3
	Forward	Dre-ND6-V1021F	F10	AGCATCAACCGATATTAATAACCAAGTG	1021 bp	1	30
	Reverse	Dre-cytb-VR	R	GCAAGTGTAATAACTATGGCGATG			
	Probe	Dre-cytb-VPr-FAMZEN	Pr	FAM-ACAATGCAA-ZEN-CCCTTACACGATTCTT CGCATTCC-3IABkFQ			

was boiled for 10 min to disrupt the membrane of fish cells in the water (Martin et al., 2013). The other 3 L sample was left at room temperature while the boiling treatment was performed. Both of the 3 L samples were divided into six subsamples with 500 mL each in new disposable plastic cups, and filtered through Sterivex cartridge filters (pore size 0.45 μm ; Merck, Darmstadt, Germany) using a sterile 50 mL luer-lock syringe (SS-50LZ; Terumo Co., Tokyo, Japan). The water was aspirated into the syringe from the disposable cup and then the Sterivex filter is connected to the syringe. The water was then slowly forced through the filter that was inside the Sterivex filter. The water inside the Sterivex filter was removed by pushing air into the Sterivex filter three times using the syringe. The inlet and outlet port on the Sterivex filter were capped (VRMP6 and VRSP6, respectively, ISIS Co., Osaka, Japan). This operation was repeated until all water samples were processed. The Sterivex filters were all stored at room temperature until all filtrations were complete.

PMA treatment (see section “PMA Treatment” for details) was performed on 3 of 6 samples for each of the boiled and non-boiled controls following the procedure shown in **Figure 2A**. The remaining three samples from each of the boiled and non-boiled controls were not treated with PMA. All DNA extractions from all samples were performed just after the PMA processing and stored at -20°C until qPCR testing. The details of PMA treatment and DNA extraction are shown in sections “PMA Treatment” and “DNA Extraction” respectively. **Figure 2A** shows the flowchart of experiment 1.

Experiment 2: Elucidation of Degradation Process of Intracellular/Extracellular eDNA

Aquarium conditions for experiment 2 matched those in experiment 1 (section “Experiment 1: PMA Confirmation Test”) with two exceptions. The volume of water was 30 L, and the fish were kept at a density of 3 individuals $\cdot \text{L}^{-1}$. We reduced the density of fish in experiment 2 based on the results of experiment 1, because a sufficient DNA copy number was expected to be obtained for quantification by qPCR.

Using a disposable plastic cup, 18 L of water was transferred from the aquarium into a 20 L plastic container. The 18 L water sample was thoroughly mixed by shaking and 500 mL samples ($n = 36$) were dispensed into disposable plastic packs (DP16-TN1000; Yanagi Co., Nagoya, Japan) and stored in a water bath maintained at $25.14 \pm 0.12^{\circ}\text{C}$ using a heater for incubation. We randomly selected six samples and one FNC (filtration negative control: 500 mL ultrapure water) to process for each point in our time-series (day 0, 1, 2, 3, 5, and 7). All six samples and one FNC for each time-point were filtered through a Sterivex filter. The filtration was performed following Miya et al. (2016) with minor modifications. Briefly, the lid of the plastic pack was replaced with a rubber cap before filtration. A plastic needle with a silicon tube (4987458150067; NIPRO Co., Osaka, Japan) was put into the rubber cap and the end of the other side of the tube was connected to the inlet port of Sterivex filter via plastic luer connector (VR306; ISIS Co.). The outlet port of the Sterivex filter was connected to a vacuum tube with a 10 μL

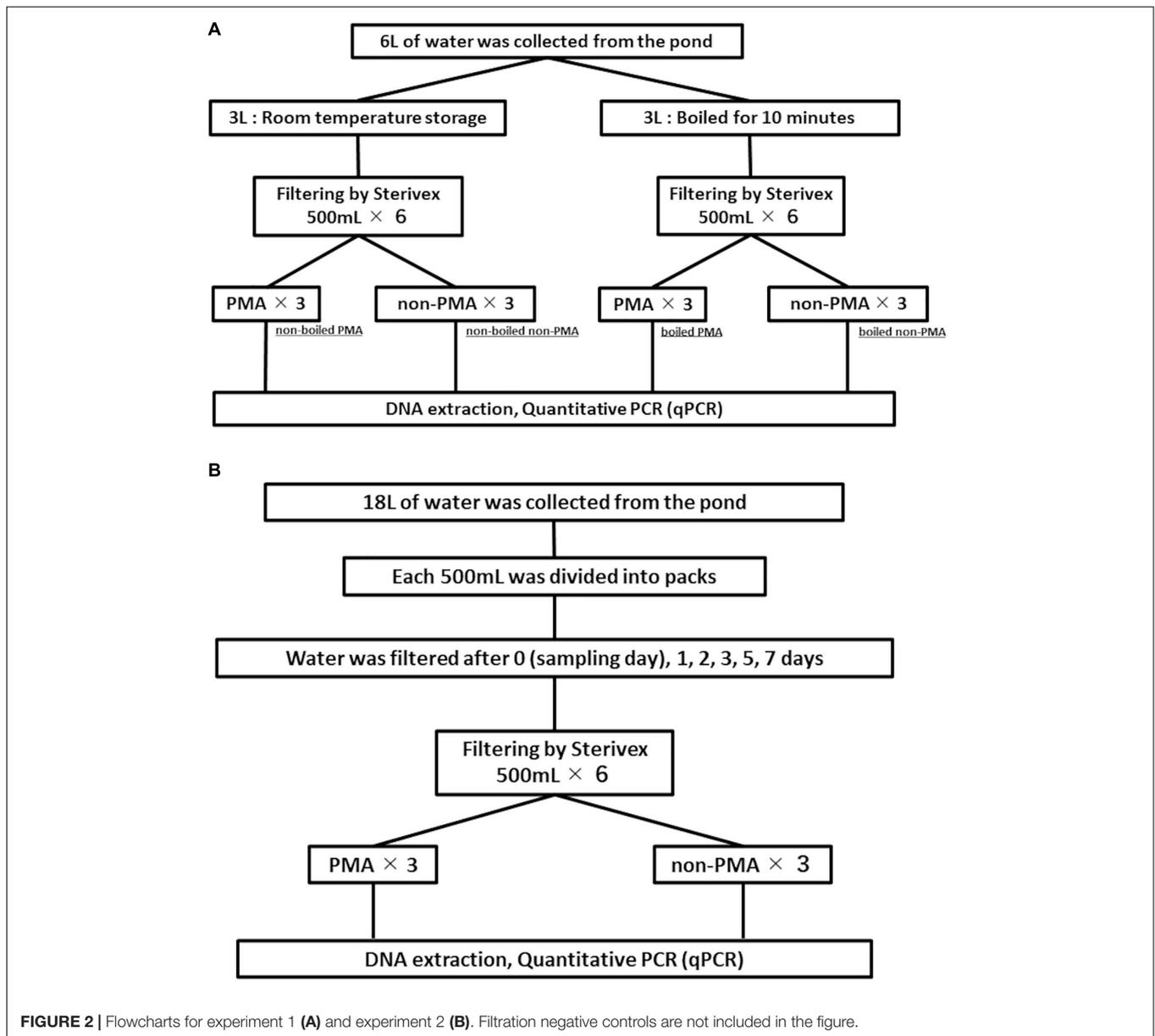
pipette tip. The inlet and outlet ports of the Sterivex filters were capped with VRMP6 and VRSP6, respectively, after filtration and stored at room temperature until all filtrations were completed. PMA treatment was performed on three of the six subsamples from each time-point. The remaining three subsamples from each time-point were not treated with PMA for comparison. DNA was extracted from all samples just after the PMA treatment (see section “PMA Treatment” for details) and stored at -20°C until qPCR testing. **Figure 2B** shows the flowchart of experiment 2.

PMA Treatment

Two milliliters of 50 μM PMA (PMAxxTM; Biotium, Fremont, CA, United States; adjusted to the concentration with ultrapure water) was added into each Sterivex filter, which was assigned to PMA treatment, from the inlet port using a thin pipette tip. Each Sterivex filter was incubated at room temperature for 10 min while covered by aluminum foil to avoid premature exposure to the light. Then, each Sterivex filter assigned to PMA treatment was exposed to strong visible light (465–470 nm wavelength) for 15 min, activating PMA to form a covalent bond with double-stranded DNA, while non-PMA samples (including FNC) were continuously covered by aluminum foil to avoid premature exposure to the light and stored at room temperature during the PMA treatment. We confirmed that there was not a significant difference in DNA copy numbers between the controls kept in dark and light during non-PMA treatment in a preliminary experiment. Adequate wavelength and sufficient luminance were attained using LED bulbs that met the requirements for the PMA reaction based on information provided from the manufacturer (Biotium). The functional mechanism of the PMA-eDNA interaction is shown in **Figure 1**. As all the residues containing eDNA are trapped on the surface of the tubular filter in the Sterivex cartridge, we performed PMA treatment by irradiating the filter surface with visible light from all directions from the outside of the cartridge, referring to Ribeiro et al. (2019) that performed PMA treatment on the eDNA samples trapped on sieve clothes.

DNA Extraction

DNA was extracted from all Sterivex filters using a commercial DNA extraction kit (DNeasy Blood and Tissue Kit; Qiagen, Hilden, Germany) following Miya et al. (2016) with minor modifications. Briefly, we removed the cap from the outlet port of the Sterivex filter and inserted the outlet port into a 3 mL test tube (cat.No.5821-255, WATSON Co., Tokyo, Japan). The Sterivex filter was connected in the test tube using surgical grade tape (1530SP-1, 3M Japan Limited, Tokyo, Japan). The combined Sterivex-test tube unit was centrifuged at $4,000 \times g$ for 2 min to remove PMA solution from PMA-treated samples and the remaining water from non-PMA treated samples. The 3 mL test tube was removed from Sterivex filter and discarded, and the outlet port of Sterivex filter was recapped. Then, the inlet port cap was removed, and 20 μL of proteinase-K and 200 μL of Buffer AL was added to Sterivex filter using a thin pipette tip. The inlet port of Sterivex filter was recapped and the Sterivex filter was incubated for 20 min at 56°C on a shaker at 20 rpm (ROLLER6 digital, IKA, Staufenberg, Germany). After the incubation, the



inlet cap was removed and the inlet port side of the Sterivex filter was connected to a 3 mL test tube. The Sterivex-test tube unit was tightly connected using surgical grade tape and centrifuged at $4,000 \times g$ for 2 min to collect the DNA solution from Sterivex filter. The Sterivex filter was discarded and 200 μL of absolute ethanol was added to the filtrate in the 3 mL test tube. The DNA in the mixture was purified using a DNeasy kit following the manufacturer's instruction. During the final elution step, DNA trapped on the silica membrane of the spin column was eluted with 100 μL of Buffer AE. The buffers (Buffer AL and AE) and the proteinase K were provided from the DNeasy kit.

qPCR Conditions

We used StepOnePlus[®] Real-Time PCR System (Life Technologies, Carlsbad, CA, United States) for qPCR. All

qPCR reactions were performed in a total volume of 12 μL , which included 6- μL of $2 \times \text{TaqPath}^{\text{TM}}$ qPCR Master Mix (Thermo Fisher Scientific, Waltham, United States), 900 nM of each primer and 125 nM of a probe at final concentrations, a 2 μL DNA template and ultrapure water to adjust the total volume. All qPCR was conducted as singleplex with triplicated technical replications. A 1,085 bp fragment of the *D. rerio* mitochondrial DNA sequence was synthesized as gBlocksTM Gene Fragments (Integrated DNA Technologies Inc.) and used to develop a standard curve for DNA quantification. The synthesized fragment crosses small regions of the three genes (cytb, tRNA-Glu, and ND6) and includes the 10 qPCR assays described in section "Experiment 1: PMA Confirmation Test." The standards were adjusted to the copy numbers of $3 \times 10^1 - 3 \times 10^4$ copies per reaction and were included in triplicate in each qPCR run.

Thermal conditions were as follows: 2 min at 50°C and 10 min at 95°C followed by 55 cycles of 95°C for 15 s and 64°C for 90 s. The cycling temperature was determined based on a preliminary experiment to enable all assays to work properly under the same conditions. Negative controls (NTC: non-template control) were conducted in triplicate in all qPCR assays for all qPCR runs, to assess the occurrence of unintended cross contamination using ultrapure water instead of the DNA template. The r^2 values of the standard curve for qPCR exceeded 0.98 in all runs in experiment 1 and 2. In addition, the average slope and y-intercept of the standard curve for qPCR were -3.54 ± 0.16 and 37.28 ± 1.38 (average \pm standard deviation), respectively.

Statistical Analysis

The eDNA copy number was calculated by averaging technical replicates for each sample. If any of the triplicate reactions for any sample had no amplification, then the copy number for that reaction was regarded as zero and included into the calculation of the average (Ellison et al., 2006). Also, data below the LOQ were excluded from the following analysis. We used R version 3.6.1 for all the statistical analyses (R Core Team, 2019). For experiment 1, analysis of variance (ANOVA) was conducted to test the effect of boiling treatment, PMA treatment, and target sequence length on the eDNA copy number. All of the interactions among the factors were also included to the test as factors.

For experiment 2, we performed multivariate analysis of variance (MANOVA) and *post hoc* ANOVAs to test the effect of sampling time and target sequence length on the eDNA copy numbers in the non-PMA and PMA treated samples. MANOVA can simultaneously evaluate the effects of each factor on multiple response variables and reduce the likelihood of Type I errors and increase the statistical power (Warne, 2014). A general linear model (GLM) was used to examine how eDNA copy number changes depending on the target sequence length. Furthermore, in order to analyze the eDNA degradation, the degradation constant and the half-life were calculated as indices instead of the time to be undetected which should depend on the initial copy number of eDNA, allowing comparisons between studies (Maruyama et al., 2014; Sassoubre et al., 2016). We assumed that DNA degrades at a constant rate over time, exhibiting exponential degradation, which can be modeled by the following equation:

$$N(t) = N_0 * e^{-\lambda t}$$

where $N(t)$ is the estimated number of copies of eDNA at time t , N_0 is the number of copies of eDNA at time zero, λ is the decay rate (sometimes denoted as β ; Thomsen et al., 2012b; Maruyama et al., 2014), and t is time (i.e., the number of days). Using the *nls* function in R, the eDNA copy number obtained over the time of each target sequence length in experiment 2 was fitted to an exponential decay curve to calculate the decay rate. The half-life was calculated as follows:

$$t_{1/2} = \frac{\ln(2)}{\lambda}$$

RESULTS

Primer-Probe Design

Both qPCR and agarose gel electrophoresis results indicated that *D. rerio* DNA was successfully amplified by all 10 assays. However, *in silico* tests suggested that a single *D. nigrofasciatus* sequence was potentially amplified with the primer sets F6–F10 (Accession number: KR606519.1). This species had never been kept at our facility nor is it common as an experimental organism, so it was determined that it did not affect the results of this study.

Experiment 1: PMA Confirmation Test

DNA of *D. rerio* was detected by all assays (F1–F10) in non-boiled-non-PMA, non-boiled-PMA and boiled-non-PMA treatments. Environmental DNA was detected and quantified with only two assays (F1 and F2; 132 and 225 bp) in boiled-PMA samples, while the eDNA copy numbers for assays F3 – F10 (333 bp or longer) fell below the LOQ (Tables 2, 3 and Figure 3). Samples that fell below the LOQ were excluded from further analyses. For non-boiled-PMA treated samples, the percentage decrease in mean eDNA copy number ranged from (95.04–99.17%), where greater decreases in copy number corresponded with increasing target sequence length (Table 2 and Figure 3). The mean eDNA copy number for non-boiled-non-PMA treated samples did not significantly change as target sequence length increased (Table 2 and Figure 3). Comparatively, boiled-PMA treated samples had a greater decrease in mean eDNA copy number, 98.72% (132 bp) and 99.55% (225 bp), while the remaining assays (333 bp or longer) fell below the LOQ. Similar to the non-boiled-non-PMA samples, the mean eDNA copy number for boiled-non-PMA treated samples also did not statistically change (Figure 3) across all target sequence lengths (Table 2 and Figure 3), although the overall mean eDNA copy number was lower. The ANOVA test indicated that eDNA copy number was significantly affected by boiling treatment, where both total eDNA (non-PMA) and intracellular eDNA (PMA) copy number decreased significantly after boiling (all $p < 0.05$; Table 3 and Figure 3). All FNCs and NTCs were negative for the amplification of *D. rerio* DNA. All of the raw data obtained by qPCR are shown in Supplementary Table 3.

Experiment 2: Elucidation of the Degradation Process of Intracellular eDNA

eDNA copy number data was used as a response to assess the effects that sampling time and target sequence length had on PMA and non-PMA treated samples. Time and target sequence length had significant effects on eDNA copy number (MANOVA, all $p < 0.05$ Table 4). *Post hoc* test confirmed that eDNA copy number was significantly affected by target sequence length only in PMA treated samples (intracellular eDNA; ANOVA, $p < 0.05$ Table 4). While eDNA copy number gradually decreased in both PMA and non-PMA treatments as target sequence length increased (Figure 4), intracellular eDNA copy numbers (PMA treated samples) were significantly lower in 225, 333, 430, 621,

TABLE 2 | Percentage decrease of average eDNA copy number (shown as $\Delta\%$) for each assay by PMA treatment in non-boiled and boiled treatments.

Amplicon length	Non-boiled			Boiled		
	Avg. eDNA copy number		$\Delta\%$	Avg. eDNA copy number		$\Delta\%$
	Non-PMA	PMA		Non-PMA	PMA	
132 bp	15637.78	859.678	94.50	5994.187	103.558	98.27
225 bp	15520.94	651.186	95.80	5950.311	35.873	99.4
333 bp	16203.333	390.317	97.59	4899.785	0	–
430 bp	16144.768	317.031	98.04	4873.065	0	–
529 bp	19064.72	336.607	98.23	5205.062	0	–
621 bp	17562.675	346.196	98.03	4818.851	0	–
715 bp	17110.21	318.856	98.14	4551.071	0	–
823 bp	19574.029	322.742	98.35	4779.743	0	–
935 bp	18788.451	287.817	98.47	4270.684	0	–
1021 bp	19991.794	237.982	98.81	4568.729	0	–

The average eDNA copy number was calculated by averaging the copy number generated during qPCR amplification from each of three technical replicates from three biological replicates. For 333 bp and above, eDNA was below the LOQ in boiled-PMA control and they were excluded from the analyses. Percentage decrease ($\Delta\%$) was calculated as the reduction from eDNA copies detected in non-PMA treatment to the one detected in PMA treatment.

and 823 bp assay when using eDNA copy number from 132 bp as a baseline (GLM, **Table 5**). As for the total eDNA copy number (non-PMA samples), no significant difference was found among the target sequence lengths (MANOVA, $p > 0.05$ **Table 4**), so GLM was performed only on the result for intracellular eDNA. The decay rates and the half-life calculations are shown in **Supplementary Table 2** and **Figure 5**, respectively. The half-life of the total eDNA was almost unchanged with respect to the target sequence length (To be exact, it decreased with the target sequence length, but the p -value was not significant; MANOVA, $p > 0.05$ **Table 4**), whereas the half-life of intracellular eDNA tended to decrease with increasing target sequence length (**Figure 5**). The calculated half-life was longer for intracellular eDNA at all (testable) target sequence lengths, and nearly double at 132 bp (**Figure 5**). In addition, as target sequence length increased, the difference in half-life tended to decrease. All of FNC and NTC were negative for the amplification of *D. rerio* DNA. All of the raw data obtained by qPCR are shown in **Supplementary Table 4**.

TABLE 3 | ANOVA results for the difference in eDNA copy number as a response to boiled and non-boiled treatment.

ANOVAs		
Response	Factor	P-value
eDNA conc.	Boiled	0.0000
	PMA	0.0000
	Assay length	0.0003
	Boiled: PMA	0.0000
	Boiled: assay length	0.0000
	PMA: assay length	0.0000
	Boiled: PMA: assay length	0.0000

Let $p < 0.05$ be the threshold of significance. All eDNA copy numbers were log-transformed.

DISCUSSION

Experiment 1: PMA Confirmation Test

For the first time, we have shown that PMA is effective for selectively amplifying DNA from intact vertebrate cells. However, while PMA could not completely exclude extracellular and disrupted cell-derived DNA from PCR amplification, we confirmed that PMA could work on damaged fish cells to cause a significant reduction in the detected amount of DNA (**Tables 2, 3** and **Figure 3**). As we expected, there was a drastic difference in the mean eDNA copy number between the boiled and non-boiled, PMA treated samples. Boiling disrupted cell membranes thereby allowing PMA to bind to DNA and either prevent amplification all together or reduce amplification below the LOQ (**Figure 3**) for all the assays, except the two shortest target sequence lengths (132 and 225 bp). The significant decrease in eDNA copy number of the boiled and non-boiled samples after PMA treatment, and the (expected) small decrease of eDNA copy number for non-PMA treated samples (**Figure 3**) suggests that boiling caused DNA degradation. It is probable that longer fragments in our samples that were boiled (100°C) for 10 min would have been degraded, which would cause an increase in smaller fragments (i.e., < 333 bp). Studies looking at the effects of thermal degradation on DNA for plants and bacteria indicate that prolonged exposure to high temperatures (50–200°C) will cause longer fragments to degrade, but smaller amplifiable fragments can persist (Zhang and Wu, 2005; Hrnčířová et al., 2008; Lo et al., 2015; Bitskinashvili et al., 2018). Given the short thermal exposure time in our experiment, any DNA degradation caused by high temperature is likely small and probably associated with other degradation processes (e.g., lipid peroxidation; sensu Zhang and Wu, 2005); however, this mechanism is outside the scope of our study, and our hypotheses are limited. Furthermore, there is evidence to suggest that PMA cannot completely suppress PCR amplification in cases where the target fragment is too short (e.g., 190 bp) (Luo et al., 2010). Given the increase in smaller fragments

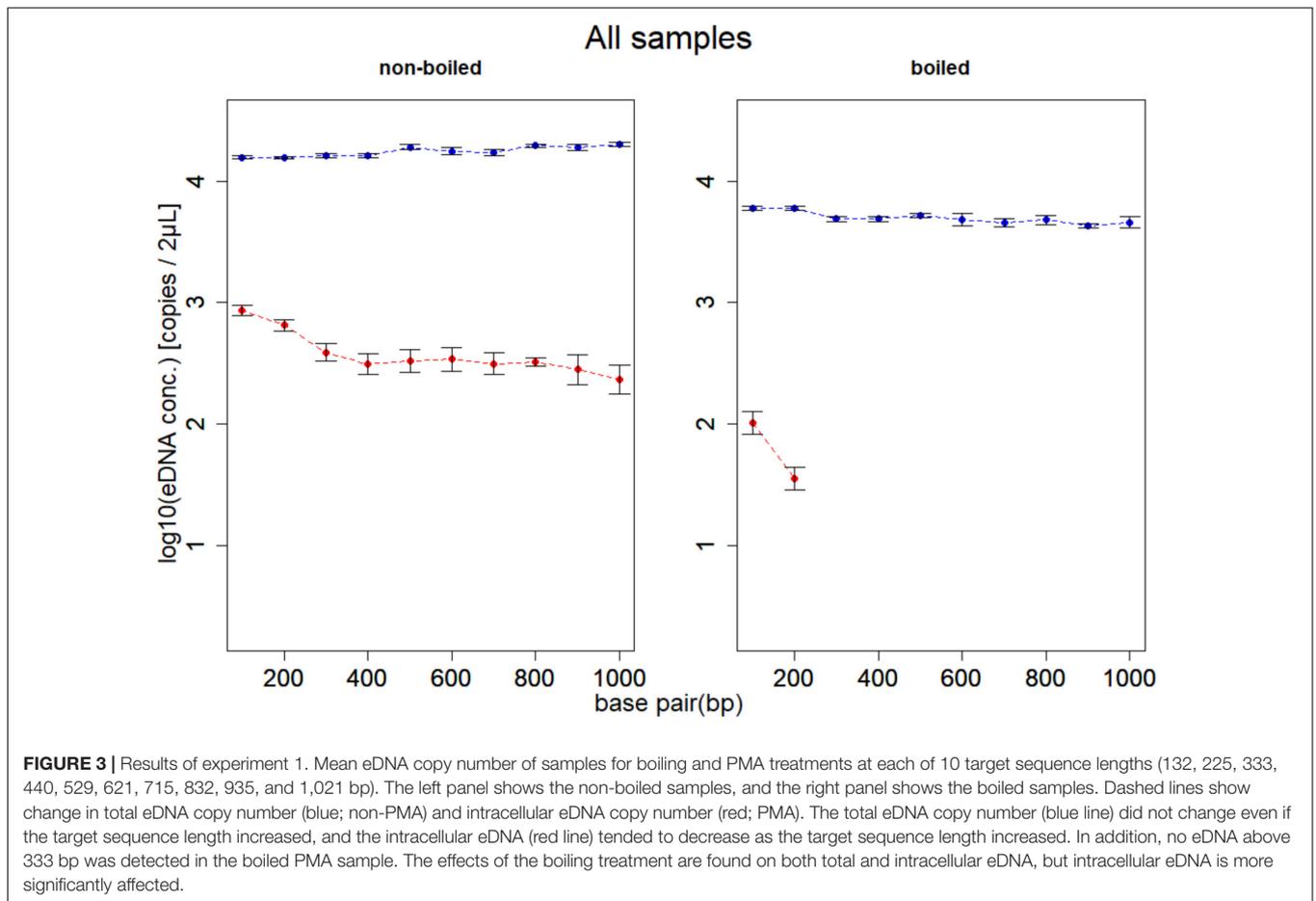


TABLE 4 | Results of MANOVA (upper) and *post hoc* test (lower) for the relationships between eDNA copy numbers at each treatment (PMA/intracellular eDNA vs. non-PMA/total eDNA) and each factor (days and assay length).

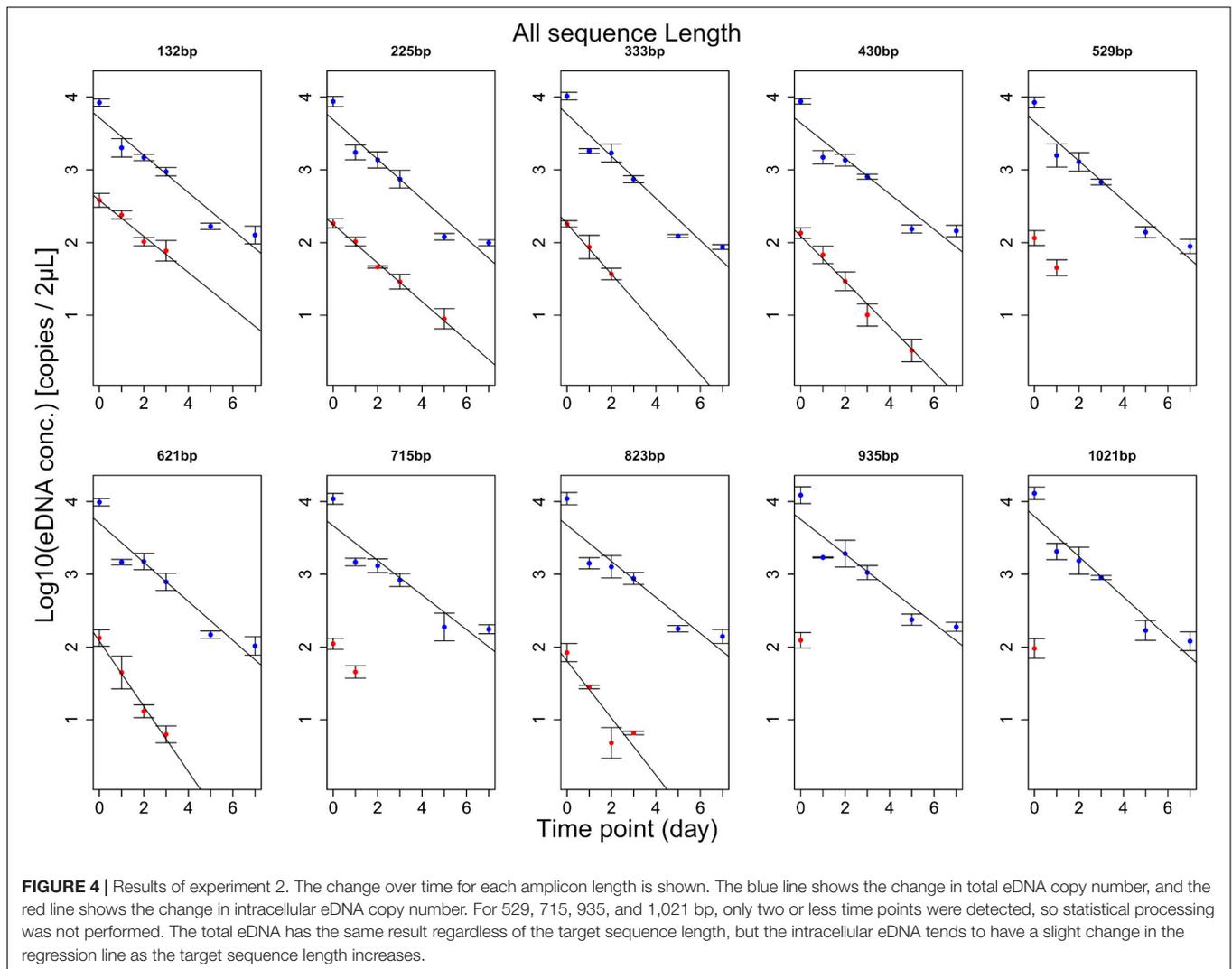
MANOVA				
Response	Factor		F-value	P-value
eDNA conc.(Including total eDNAand intracellular eDNA)	Days		715.33	0.0000
	Assay length		63.21	0.0000
	Days: assay length		9.20	0.0002
Response	Factor	Sum Sq	F-value	P-value
eDNA conc. (Total eDNA)	Days	68.881	1420.364	0.0000
	Assay length	0.160	3.3052	0.0708
	Days: assay length	0.025	0.5151	0.4739
	Residuals	8.535		
eDNA conc. (Intracellular eDNA)	Days	86.146	363.504	0.0000
	Assay length	21.218	89.534	0.0000
	Days: assay length	4.127	17.416	0.0000
	Residuals	41.710		

Let $p < 0.05$ be the threshold of significance. All the eDNA copy numbers were log-transformed.

as extracellular DNA caused by boiling and the inability of PMA to complete inhibit PCR for short fragments, this may explain why amplification was still observed with the 132 and 225 bp fragments after PMA treatment was applied.

Experiment 2: Elucidation of Degradation Process of Intracellular eDNA

Also, for the first time, we were able to selectively amplify and quantify intracellular eDNA, and model the associated



degradation rate. The MANOVA results suggested that the total eDNA present in the sample maintained comparable copy numbers between different target sequence lengths, even in the longest sequence (1,021 bp), and this pattern was maintained during the 7-day incubation period. Conversely, the intracellular eDNA copy number showed significant decreases as assay length increased over the 7-day incubation period. An interesting pattern that was observed was the significant interactions between the incubation time and three of the target sequence lengths, 430, 621, 823 bp for PMA treated samples (MANOVA and GLM; **Tables 4, 5**). There is a distinct difference in the slopes for these three assays that indicates a drastic decrease in mean eDNA copy number with increased time. However, the cause of the drastic slope of these three assays may affect the decay rates of smaller fragments (132, 225, and 333 bp). At any given timepoint, a leptokurtic pattern is expected regarding the number of fragments, such that there will be a significantly larger number of small fragments compared to a smaller number of larger fragments (*sensu* Deagle et al., 2006). Specifically, when larger fragments degrade, they are broken into smaller fragments

and these smaller fragments add to the overall quantity of smaller fragments, which lend themselves to “slowing” the decay rate for smaller fragments.

Our results for total eDNA follow similar patterns seen in other studies that assessed target sequence length and time as a factor of eDNA degradation (Sassoubre et al., 2016; Lance et al., 2017; Bylemans et al., 2018; Jo et al., 2019). GLM was not performed on total eDNA, as no effect of target sequence length was observed (**Table 4**), indicating that the degradation process of total eDNA might be similar regardless of the target sequence length, at least in the aquarium setup of the present study. Alternatively, our results demonstrate that intracellular eDNA may have a different degradation process from that of total eDNA (**Table 4** and **Figures 4, 5**). The half-life calculated for each target sequence length was greater for intracellular eDNA, but was negatively correlated with increasing target sequence length (**Figure 5** and **Supplementary Table 2**). While lack of amplification (or amplification meeting the LOQ) prevented statistical analyses of half-life indices for larger fragment sizes in intracellular eDNA, it might be assumed that total and

TABLE 5 | GLM analysis results for testing response of intracellular (PMA) eDNA copy numbers to days and assay length.

Response	Explanatory	Estimate	SE	P-value
eDNA conc. (Intracellular eDNA)	Intercept	2.5986	0.0703	0.0000
	Days	-0.2591	0.0198	0.0000
	Assay length (225 bp)	-0.2917	0.0986	0.0041
	Assay length (333 bp)	-0.4232	0.1038	0.0001
	Assay length (430 bp)	-0.4871	0.1038	0.0000
	Assay length (621 bp)	-0.6180	0.1070	0.0000
	Assay length (823 bp)	-0.7901	0.1135	0.0000
	Days: assay length (225 bp)	-0.0426	0.0268	0.1153
	Days: assay length (333 bp)	-0.0454	0.0338	0.1821
	Days: assay length (430 bp)	-0.0723	0.0338	0.0353
	Days: assay length (621 bp)	-0.1025	0.0407	0.0138
Days: assay length (823 bp)	-0.1331	0.0551	0.0182	

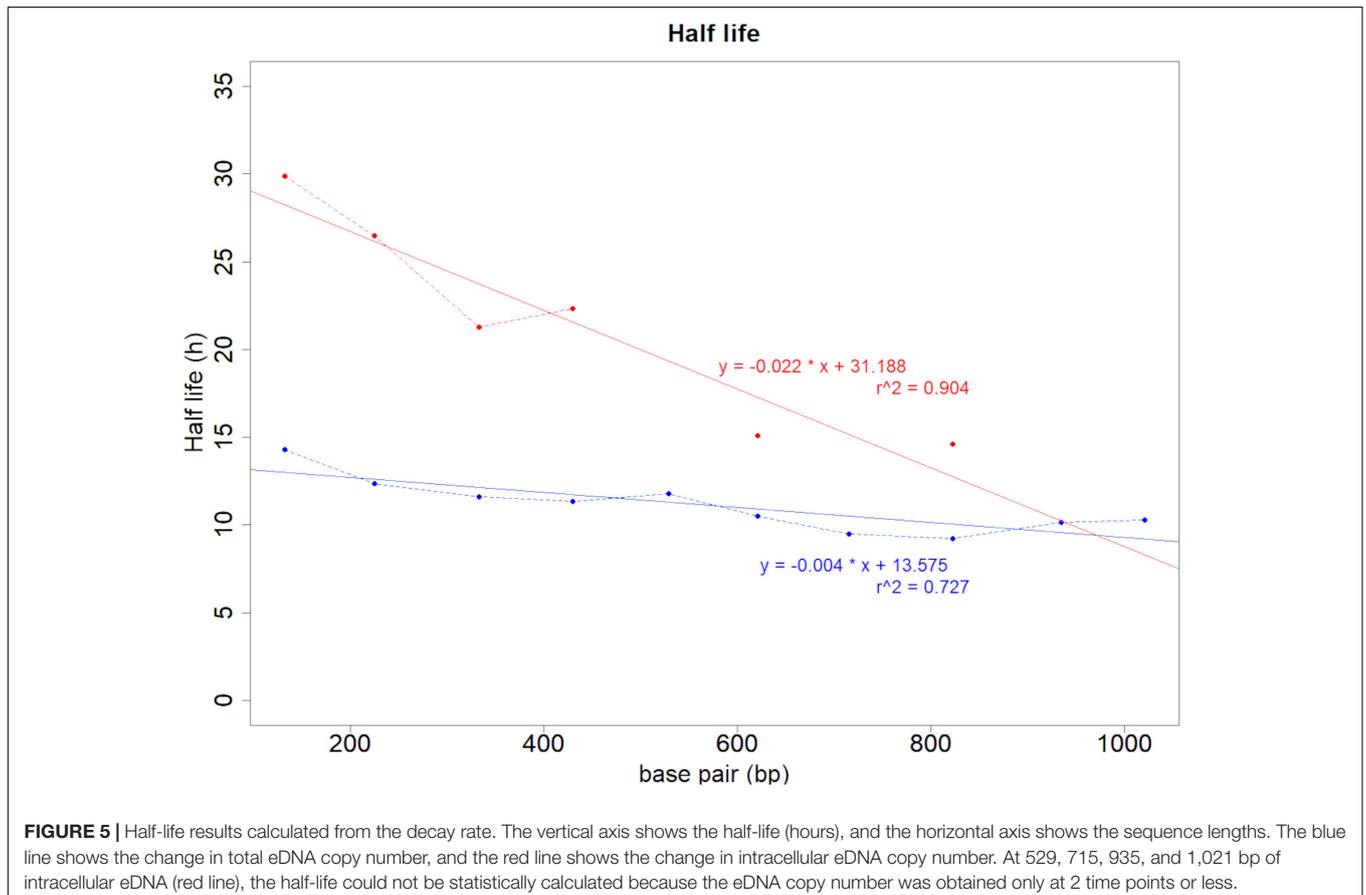
Let $p < 0.05$ be the threshold of significance. All the eDNA copy numbers were log-transformed. Responses of each assay length and interactions was calculated using 132 bp as a baseline.

intracellular eDNA decay rates are similar at larger fragments sizes based on regression (**Figure 5**). This is, in part, supported by our inability to detect larger fragments after the initial sampling (Day 0) which indicates a short half-life (**Figure 4**). Other studies have reported that longer amplicons amplified by PCR have lower quantity and faster degradation when analyzing total eDNA (Deagle et al., 2006; Jo et al., 2017), although the fraction of

intracellular eDNA as part of the total eDNA was much smaller in the case of our aquarium water compared to the larger systems used in Deagle et al. (2006) and Jo et al. (2017). Furthermore, the interpretation of eDNA degradation results could be better supported if accurate estimates of intact vertebrate cells in total eDNA could be calculated.

Synthesis

In this study, PMA was confirmed to be effective to differentiate “intact cells” from “disrupted cells” in vertebrate eDNA analysis, enabling the independent detection of intracellular eDNA among the total eDNA. It has been suggested that intracellular DNA represents a specific fraction of total eDNA (Turner et al., 2014; Jo et al., 2019) but this study is the first to present a method to quantify the fraction more clearly using PMA. Our results suggested that the decay rate is dependent on both the existing state of eDNA, i.e., intracellular vs. extracellular, and the length of the target sequence. The process of eDNA degradation may differ between experimental conditions and in the natural environment, and thus it is important to confirm the degradation pattern of intracellular eDNA in environmental water samples under various environmental conditions (e.g., pH, trophic state, temperature and biomass; Jo et al., 2019). For example, intracellular DNA is suggested to degrade much less efficiently than extracellular DNA, by extracellular enzymes,



due to the presence of cell membranes (Paul et al., 1987; Levy-Booth et al., 2007; Corinaldesi et al., 2008). Our results from experiment 2, where the half-life of intracellular eDNA was longer than the half-life of total eDNA, suggest that intracellular eDNA is protected from external degradation factors by the cell membrane. Many cells begin to degrade by apoptosis before shedding from the individual (Hotchkiss et al., 2009). However, during normal apoptotic shedding of epithelial cells, intact mitochondria may be released from the cell and mtDNA may be protected from endonuclease degradation (Murgia et al., 1992; Tepper and Studzinski, 1993). In aquatic animals, this process releases the entire mitochondria into the water column, where the bilayer resists degradation and mitochondrial nucleoids further protect mtDNA (Rickwood and Chambers, 1981). The mitochondrial bilayer and tissue cell membrane could provide a greater level of protection against DNA degradation that is not seen with just one cell membrane alone (Foran, 2006).

Conversely, while protected from external decomposition factors, internal degradation factors may act on eDNA before it is released from the target species. In Experiment 1, the total eDNA copy number of the non-boiled samples did not change as the target sequence length increased, but the intracellular eDNA copy number changed. Although intracellular eDNA is protected from external degradation factors in the environment and has a longer half-life, there is a possibility that DNA fragmentation is progressing due to internal degradation factors associated with apoptotic processes (Nagata, 2005; Hotchkiss et al., 2009). This indicates that extracellular eDNA (free eDNA), damaged intracellular eDNA and intracellular eDNA may be subject to separate degradation processes. In the future, in order to further clarify mechanisms driving eDNA degradation rates and processes, it is necessary to study the dynamics of each existing state, as well as, total eDNA. This will ultimately provide a basis for developing more robust analyses that help limit or prevent process Type I and II errors.

PMA has been effective as a molecular surveillance tool for human health issues, but has a great potential to assist with future conservation efforts of aquatic systems. Removing the “noise,” such as eDNA resuspended from sediments or transported from upstream sites, can increase the accuracy eDNA surveillance by allowing researchers to detect contemporary signals from of targeted species. As discussed earlier, process Type I errors in current eDNA analyses may results due to the (unintentional) misinterpretation of the eDNA sources (e.g., the transport of eDNA by predator droppings, boats, or wetland birds) (Deiner and Altermatt, 2014; Merkes et al., 2014). In cases where rapid and accurate biosurveillance information is needed, like determining if an invasive species has entered a new body of water, combined PMA-treatment with multi-assay surveillance (especially assays with different amplicon lengths) may help avoid misinterpretation and determine whether the signal obtained by eDNA analysis is new or old. To this end, resource management agencies deploy various eradication measures to remove invasive species from invaded systems (Bonar et al., 2009; Booy et al., 2017). PMA-eDNA surveillance would allow researchers to discern between intact and degraded cells, to identify the effectiveness of eradication strategies, i.e.,

detection of intracellular eDNA after a set time period might indicate that an eradication strategy has failed. The timing of seasonal migratory patterns of fish (long-term stabilization by sediment adsorption) could be more easily and reliably assessed (Maruyama et al., 2014). Critical habitat and spawning habitat (and season) are unknown for many aquatic organisms, and thus PMA-eDNA can help researchers hone in on specific areas of an aquatic environment without the need visual surveys that require many hours and resources. To realize these applied ideas in future research, we hope that the combination of PMA with multiple-sized amplicons for target species will facilitate new research for clarifying other mechanistic underpinnings of eDNA ecology.

DATA AVAILABILITY STATEMENT

The datasets presented in this study can be found in online repositories. The names of the repository/repositories and accession number(s) can be found in the article/**Supplementary Material**.

ETHICS STATEMENT

Ethical review and approval was not required for the animal study in accordance with the local legislation and institutional requirements, but we followed the instructions suggested by Japanese animal welfare regulations and the “Act on Welfare and Management of Animals” (Ministry of Environment of Japan) for husbandry and handling of fish.

AUTHOR CONTRIBUTIONS

TH, RP, and HY: concept and design and data analysis. TH, KT, and KM: lab work. All authors: writing final version.

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SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fevo.2021.632973/full#supplementary-material>

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Conflict of Interest: The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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