



## OPEN ACCESS

## EDITED BY

Shuisen Chen,  
Guangzhou Institute of Geography,  
China

## REVIEWED BY

Marta Chiesi,  
National Research Council (CNR), Italy  
Shanlei Sun,  
Nanjing University of Information  
Science and Technology, China

## \*CORRESPONDENCE

Li He  
heli2020@cdut.edu.cn

## SPECIALTY SECTION

This article was submitted to  
Environmental Informatics  
and Remote Sensing,  
a section of the journal  
Frontiers in Ecology and Evolution

RECEIVED 19 September 2022

ACCEPTED 14 November 2022

PUBLISHED 06 December 2022

## CITATION

Huang X, He L, He Z, Nan X, Lyu P and  
Ye H (2022) An improved  
Carnegie-Ames-Stanford Approach  
model for estimating ecological  
carbon sequestration in mountain  
vegetation.  
*Front. Ecol. Evol.* 10:1048607.  
doi: 10.3389/fevo.2022.1048607

## COPYRIGHT

© 2022 Huang, He, He, Nan, Lyu and  
Ye. This is an open-access article  
distributed under the terms of the  
[Creative Commons Attribution License  
\(CC BY\)](https://creativecommons.org/licenses/by/4.0/). The use, distribution or  
reproduction in other forums is  
permitted, provided the original  
author(s) and the copyright owner(s)  
are credited and that the original  
publication in this journal is cited, in  
accordance with accepted academic  
practice. No use, distribution or  
reproduction is permitted which does  
not comply with these terms.

# An improved Carnegie-Ames-Stanford Approach model for estimating ecological carbon sequestration in mountain vegetation

Xu Huang<sup>1,2</sup>, Li He<sup>1,3\*</sup>, Zhengwei He<sup>1</sup>, Xi Nan<sup>4</sup>, Pengyi Lyu<sup>1,2,5</sup>  
and Haiyan Ye<sup>6</sup>

<sup>1</sup>State Key Laboratory of Geohazard Prevention and Geoenvironment Protection, Chengdu University of Technology, Chengdu, China, <sup>2</sup>School of Earth Sciences, Chengdu University of Technology, Chengdu, China, <sup>3</sup>College of Tourism and Urban-Rural Planning, Chengdu University of Technology, Chengdu, China, <sup>4</sup>Institute of Mountain Hazards and Environment, CAS, Chengdu, China, <sup>5</sup>Exploration and Development Research Institute of Southwest Oil and Gas Field Company, PetroChina, Chengdu, Sichuan, China, <sup>6</sup>Yongnian Middle School, Zigong, China

The vegetation in mountainous areas is abundant, and its ecological carbon sequestration ability is of great significance to maintain the sustainable and healthy development of the ecological environment. However, when estimating the carbon sequestration of mountain vegetation, the Carnegie-Ames-Stanford Approach (CASA) model assigns a uniform value to the maximum light energy utilization ( $\epsilon_{\max} = 0.389 \text{ gC/MJ}$ ), ignoring the influence of vegetation types and topographic factors on  $\epsilon_{\max}$ , resulting in the low accuracy of the CASA model in estimating the carbon sequestration of mountain vegetation. In this paper, the improved CASA model was combined with Landsat 8 Operational Land Imager (OLI) remote sensing image data to improve the estimation accuracy of carbon sequestration of mountain vegetation. The first was the establishment of a linear link between the terrain characteristics (slope and aspect), vegetation types, and  $\epsilon_{\max}$  in mountainous locations. The second was the improvement of the CASA model's calculation method for key parameters. The different distributions of the estimation results from the two techniques in 2015 and 2016 are then compared using Landsat 8 data as the data source, and the impact of the terrain factors in the improved CASA model on the estimation results is confirmed. Finally, the improved CASA model and the CASA model are used to estimate the Net Primary Productivity (NPP) of the study area from 2000 to 2020, and the estimated results of the two models are compared with the computation results of the MODIS data NPP product. The findings indicate that the improved CASA model's estimation results have a higher

degree of fit and a better correlation. The improved CASA model aids in precisely understanding the ecological carbon sequestration potential of mountain areas and increases the estimation accuracy of vegetation carbon sequestration in mountainous areas.

#### KEYWORDS

improve CASA model, ecological carbon sequestration, topographic influence, maximum utilization of light energy, mountainous areas

## Introduction

The CASA model primarily calculates NPP from light utilization efficiency ( $\epsilon$ ) and the Absorbed Photosynthetic Active Radiation (APAR) absorbed by vegetation. Although it is frequently used to estimate NPP (Field et al., 1998; Hicke et al., 2002; Singh et al., 2002; Xing et al., 2010), there are still some drawbacks. The calculation techniques for several parameters need to be updated because the original CASA model was built using NPP estimation of different types of vegetation in North America, which differs slightly from vegetation types in other countries (Yong et al., 2022). Second, when the same value of the maximum light use efficiency ( $\epsilon_{\max}$ ) = 0.389 gC/MJ is assigned, the CASA model does not consider the impact of vegetation types on  $\epsilon_{\max}$  (Chen et al., 2017). Actually, different vegetation types have very varying  $\epsilon_{\max}$  (Chen et al., 2021). Additionally, the initial CASA model only established a functional relation with the Simple Ratio index (SR) for the Fraction of Absorbed Photosynthetically Active Radiation (FPAR), which was insufficient to identify the internal link between FPAR and vegetation. Last but not least, it is difficult to obtain the soil parameters required for CASA model estimation, with low accuracy and complex calculation, which can easily affect the estimation results (Wang H. et al., 2019; Zhu et al., 2019).

Mountainous vegetation is rich and diverse, and improving the estimation accuracy of the ecological carbon sequestration in mountainous areas can provide a clearer understanding of the ecological carbon sequestration capacity of mountain vegetation (Xu et al., 2020; Jiang et al., 2021; Li H. et al., 2021; Li Y. et al., 2021; Zhang et al., 2021). Wang et al. (2020) showed that from 2010 to 2016 in mainland China, the carbon sequestration of terrestrial ecosystems was 45% of the anthropogenic carbon emissions during the same period, approximately 1.11 billion tons, and these carbon sequestration contributions are mainly due to the strong ecological carbon sequestration capacity in the southwestern mountainous areas of China. The high slope and significant fluctuation of mountain areas, however, make it easy for topographic elements to alter the accuracy of the assessment of vegetation carbon sequestration (He et al., 2018; Yu et al., 2020; Li M. et al., 2021; Li Y. et al., 2021; Li et al., 2022;

Wang et al., 2022). The Surface Water Sub-Index was used to enhance the parameters in the CASA model, which decreased the influence of terrain elements on estimation accuracy (Pan and Xu, 2020; Xia et al., 2020). The precision of the data they used, however, had a significant impact on the estimation outcomes. To lessen the influence of topographic factors on the estimation of vegetation carbon sequestration, Chen and Sun et al. introduced the Shadow Elimination Vegetation Index (SEVI) into the model (Chen and Zeng, 2016; Sun and Xie, 2021). However, this method was difficult to use to eliminate the impact of shadows on slopes with large slopes. To some extent, Geng et al. revised the CASA model's parameters based on the Digital Elevation Model (DEM) data to increase the estimation accuracy of the CASA (Wang H. et al., 2019; Wang Y. et al., 2019), but this method's correcting effect on terrain falling shadow is subpar. Early studies have shown (Sun et al., 2016; Li H. et al., 2021) that changing model parameters is challenging when assessing vegetation carbon sequestration at a wide range of regional scales. They only established the value range of  $\epsilon_{\max}$  based on experimental data but did not establish a functional relation between terrain parameters and the  $\epsilon_{\max}$  (Han, 2020; Peng et al., 2021). Therefore, the estimation results of the improved model are significantly different from those of previous methods (Liu Y. et al., 2013; Fan, 2018; Jiang et al., 2019; He et al., 2019; Li et al., 2019a,b).

The primary goals of this paper, which are based on the aforementioned issues, are to (1) address the issue that the CASA model's  $\epsilon_{\max}$  is uniformly set at 0.389 g C/MJ and (2) establish a functional relationship between  $\epsilon_{\max}$  and vegetation types and terrain factors so that  $\epsilon_{\max}$  changes as the terrain changes; then, (3) lessen the effect of topographic factors on the accuracy of vegetation carbon sequestration estimates in mountainous areas and establish a functional relationship between topographic factors and vegetation NPP in conjunction with the fact that vegetation resources are primarily distributed in mountainous areas with high relief (Shen et al., 2022; Zhao et al., 2022). As a result, this paper estimates the NPP value of the study area using the improved CASA model based on Landsat 8 OLI data sources in Sichuan mountainous areas, analyses the relationship between the estimated results in different periods in the same area and the topographic factors, and confirms the viability of

the improved CASA model. Then, the correlation and the fitting degree of the estimation results of the three methods (CASA model, improved CASA model, and MODIS data NPP product) are compared to verify whether the estimation accuracy of the improved CASA model has been improved.

## Materials and methods

### Study area

The study area should have a mountainous distribution to meet the research objectives. Second, the research region should also have a reasonably flat area distribution to more clearly discern the differences in vegetation estimating findings between the two models for various terrain features. As a result, this article uses some of the neighboring Sichuan mountain areas and Chengdu Plain as its research area and chooses high-quality Landsat 8 OLI data from the geospatial data cloud as its data source. According to previous research results on mountain areas in China (Christian and Jens, 2011; Zhang et al., 2013, 2022; Price et al., 2019; Körner et al., 2021; Zhang and Min, 2022), the selected definition criteria for Sichuan mountain areas are as follows: areas with elevations less than 500m and degrees of relief greater than 50m; elevations between 500 and 2500 m, degrees of relief higher than 100 m or slopes greater than 25°; and elevations greater than 2500 m. According to GIS calculations, the Sichuan mountainous areas is 328,500 square kilometers, and the plain and hilly area is 157,500 square kilometers. The distribution range of mountainous areas and the spatial location of Landsat 8 OLI image data required in this paper are shown in **Figure 1**.

The study area is high in the northwest and low in the southeast, with relatively low elevation in the east and large topographic relief. The west and north have relatively high elevations and relatively flat terrain. Affected by topographic factors, the eastern and southern parts of the study area have more precipitation and higher temperatures, while the western and northern parts of the study area have less precipitation and lower temperature. Therefore, forests, shrubs, and other vegetation are mainly distributed in the east and south of the study area, with rich and diverse vegetation types, while grasslands, shrubs, and other vegetation are mainly distributed in the west and north, with a single vegetation type.

### Data and processing

#### Landsat 8 operational land imager data

Remote sensing data are mainly used to calculate vegetation parameters involved in the model. The remote sensing data required in this paper are Landsat 8 OLI data with good quality in the second quarter of each year from 2000 to 2020, including

data information from 11 bands. All Landsat 8 OLI data are downloaded from the geospatial data cloud.<sup>1</sup> Before data processing, radiometric correction, atmospheric correction, and other preliminary processing work are needed. Then, band 4 and band 5 in the data were used for function calculation to calculate the vegetation index required by the paper, such as the Normalized Difference Vegetation Index (NDVI), the Simple Ratio Index (SR), the Leaf Area Index (LAI), and so on.

#### Digital elevation model

The DEM data required by the paper are from the geospatial data cloud, (see text footnote 1) with a spatial resolution of 90 m. DEM data are mainly used to calculate the spatial distribution, slope, aspect, and other information of mountain areas. The DEM data of Sichuan mountainous areas are obtained by clipping the mountain areas based on the DEM data of Sichuan Province.

#### MOD17A3HGF data product

The NPP product data required in this paper are used to verify the accuracy of the improved CASA model and the estimated results of the CASA model for carbon sequestration. Based on the MOD17A3HGF dataset, the MODIS MOD17A3HGF NPP data of the study area from 2000 to 2020 were generated through the processing of splicing, projection transformation and cropping of the segmentation images covering the study area. This dataset has a spatial resolution of 500 m and a temporal resolution of 8 days.

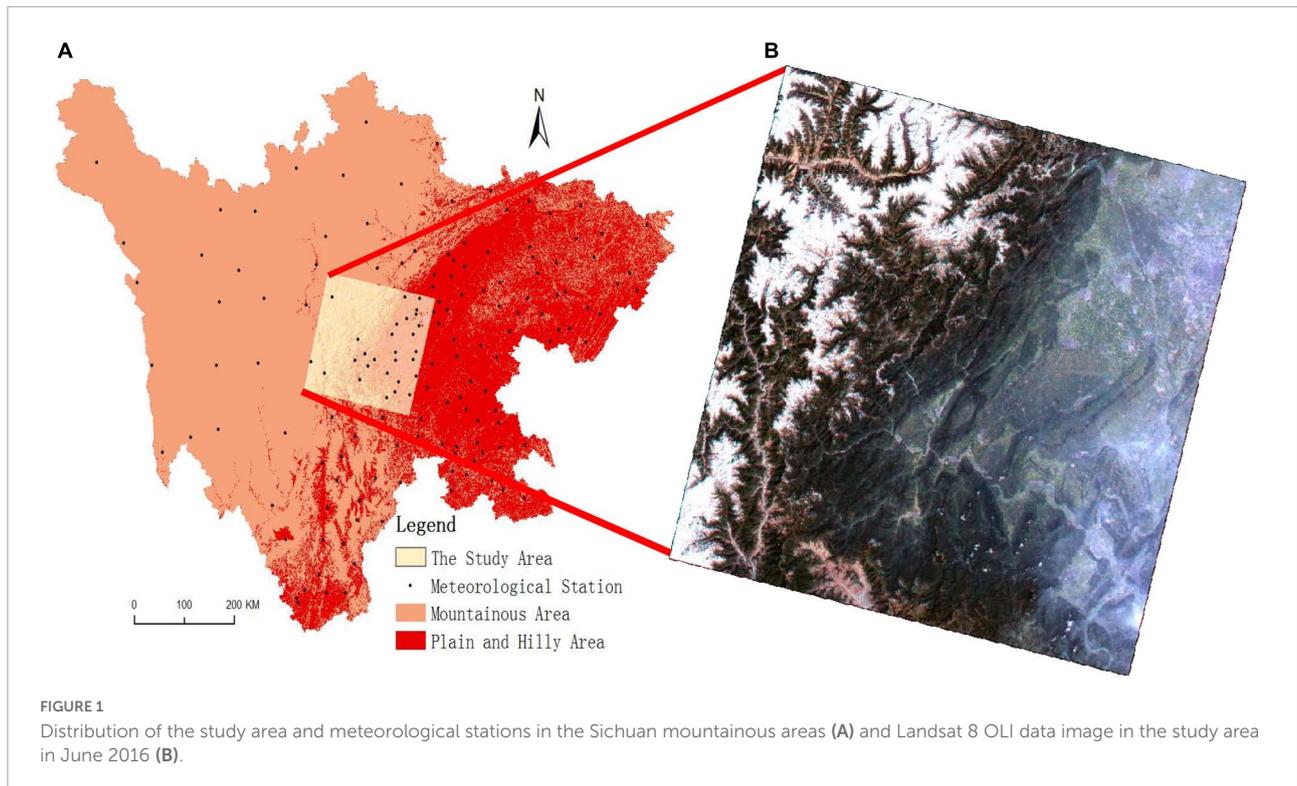
#### Meteorological data

Meteorological data are mainly used to calculate the climate parameters involved in the model. The average daily temperature, hours of sunshine, and daily precipitation used herein were provided by the National Meteorological Information Center<sup>2</sup> (National Meteorological Information Center [NMC], 2018). The paper selects 156 meteorological station data in Sichuan Province (as shown in **Figure 1**), interpolates the meteorological data in Sichuan Province using the Inverse Distance Weight method (IDW) of ARCGIS software, and then uses the mask data in the mountain area to cut them, thus obtaining the meteorological data in Sichuan mountain areas. As seen from **Figure 1**, meteorological stations are distributed less in the west and more in the east, and the overall distribution is relatively uniform, which is conducive to the accurate acquisition of meteorological data. The processing consisted of the following sections:

- (1) Meteorological data from meteorological stations around the research area were selected to obtain monthly average temperature, total sunshine hours and total precipitation.

<sup>1</sup> <http://www.gscloud.cn/>

<sup>2</sup> <http://data.cma.cn/>



- (2) Assign the spatial coordinate information based on the longitude and latitude of each station to the meteorological data.
- (3) Meteorological data with a spatial resolution of 90 m are obtained by the IDW interpolation method of meteorological station data.

### Vegetation types data

Vegetation types data are used to determine  $\epsilon_{\max}$  of different vegetation, and  $\epsilon$  varies greatly with different vegetation types. Papers required for vegetation types data from the resources and environmental science data center of the Chinese Academy of Sciences.<sup>3</sup> According to the data on vegetation distribution in Sichuan mountainous areas,  $\epsilon_{\max}$  among different vegetation types can be calculated.

### Improvement of Carnegie-Ames-Stanford Approach model

The CASA model is one of the most representative sunshine energy utilization models (Du et al., 2021) and it is not solely easy in structure and high in responsiveness; however conjointly

<sup>3</sup> <https://www.resdc.cn/Default.aspx>

widely utilized in vegetation NPP estimation on giant spatial and international scales (Cai et al., 2018; Yang, 2019). To date, it is a comparatively mature calculation method that incorporates a certain theoretical basis, the parameters are straightforward to obtain (Chen et al., 2021), and the calculable NPP accuracy is high (Li et al., 2019a; Liu et al., 2020; Zhang et al., 2020). The CASA model calculates NPP through two variables: APAR and  $\epsilon$ . Therefore, this paper mainly improves their calculation methods.

### Improvement of $\epsilon$

The calculation method of  $\epsilon$  according to the CASA model is shown in Formula 1 (Du et al., 2021), most parameters square measure the typical air temperature ( $T_{\epsilon 1}$  and  $T_{\epsilon 2}$ ) and the average evaporation ( $W_{\epsilon}$ ). However, because of the massive fluctuation of the piece of ground within the study space, the temperature distinction between the highest and bottom of the slope is giant, and the evaporation is completely different.

$$\epsilon = \epsilon_{\max} * T_{\epsilon 1} * T_{\epsilon 2} * W_{\epsilon} \quad (1)$$

In the formula,  $\epsilon_{\max}$  represents the highest utilization rate of sunshine energy (gC/MJ), and  $T_{\epsilon 1}$ ,  $T_{\epsilon 2}$ , and  $W_{\epsilon}$  represent the parameters of  $\epsilon_{\max}$  at vasoconstrictor, heat, and water stress, respectively.

This research improves the calculation method of  $\epsilon$  by merging it with other carbon sequestration estimating models (Vegetation Photosynthesis Model, VPM; Vertically Generalized Productivity Model, VGPM, etc.) to increase the

estimation accuracy of the CASA model in mountainous areas. To lessen the influence of terrain characteristics on the estimation accuracy, the parameters involved in  $\epsilon$  are obtained using band reflectance as much as feasible. The improved equation is:

$$\epsilon = \epsilon_{\max} * \text{Tem} * \text{Wat} \quad (2)$$

In the formula,  $\epsilon_{\max}$  represents the maximum light energy utilization (gC/MJ), Tem is the temperature parameter and Wat is the vegetation leaf age and water content parameter.

### The improvement of $\epsilon_{\max}$

According to previous research results, the size of  $\epsilon_{\max}$  depends on vegetation types, NDVI and LAI, and their geometric relationship (Equation 3) is established (Anikó et al., 2017). However, the effect of topographic factors on  $\epsilon_{\max}$  is ignored. Therefore, based on previous studies, this paper added topographic factors (slope and aspect) into the geometric relationship and constructed the functional relationship between  $\epsilon_{\max}$  and slope, aspect, vegetation types, NDVI, and LAI (Equations 3–5). To build a functional relationship between terrain factors and  $\epsilon_{\max}$ , first, the point information of at least 3 points location (PT) should be determined in the same coordinate system (slope as abscissa,  $\epsilon_{\max}$  as vertical coordinate).  $\epsilon_{\max}$  on sunny and shady slopes fluctuates depending on the angle of irradiation and the area of light received under the same sunshine due to the influence of slope and aspect (Li Y. et al., 2021). For plants on the sunny slope, when the slope is small,  $\epsilon_{\max}$  does not change much (PT 1) (Zhang and Ren, 2016). When the slope is larger than 15°,  $\epsilon_{\max}$  will decline exponentially as the slope rises (PT 2) (Xu et al., 2019); at 30°,  $\epsilon_{\max}$  is only 0.5 times the slope's bottom (PT 3) (Li H. et al., 2021; Li M. et al., 2021); and until the slope is approximately 70°,  $\epsilon_{\max}$  is essentially zero (PT 4). The reason is that when the slope increases, the soil water content and thickness decrease due to the effects of gravity, which has a direct impact on how well flora use solar energy (Sun et al., 2014; Wang B. et al., 2021; Wang J. et al., 2021). For vegetation on shady slopes,  $\epsilon_{\max}$  decreases with the increasing slope (Zhang et al., 2021). When the slope increases to 15°,  $\epsilon_{\max}$  is only half of the slope bottom (PT 5) (Zhu and Yuan, 2019), and when the slope increases to 30°,  $\epsilon_{\max}$  is only 0.2 times the slope bottom (PT 6) (Du et al., 2018). When the slope increases to 60° (PT 7),  $\epsilon_{\max}$  is 0 (Zhang et al., 2020). The reason is that with the increase in slope, the soil thickness and water content decrease, while the shielding areas between vegetation increase, which reduces the illumination areas of shady slope vegetation, resulting in  $\epsilon_{\max}$  decreasing gradually (Zhu et al., 2006; Li et al., 2022). According to the rules summarized by predecessors above, for the functional relationship between the two in a sunny slope area, the formula  $y = (1.2-x/75)^3$  can well couple the relationship between the two. For the shady slope area, the formula  $y = (1-x/90)^3$  is found to be able to couple the relationships. Therefore, these two formulas were put into

Equation 3 to establish the functional relationship between the slope aspect and  $\epsilon_{\max}$  (Equations 4 and 5). However, this is only an ideal function.  $\epsilon_{\max}$  is also affected by vegetation types, the spatial distribution of vegetation types is affected by altitude, and the vegetation types change regularly with altitude. According to the above analysis, the relationship between  $\epsilon_{\max}$  and slope and facet is shown in Figure 2, and the value of  $\epsilon_{\max}$  is not fixed but changes with the change in slope and aspect.

To estimate the carbon sequestration of mountain vegetation more accurately, it is necessary to establish the geometric relationship between  $\epsilon_{\max}$  and slope and aspect. According to previous studies,  $\epsilon_{\max}$  varies greatly with different vegetation types, and its size depends on vegetation types, NDVI, and LAI (Yang et al., 2015; Yan et al., 2022). In this paper, the slope and aspect are taken as the influence parameters to improve the previous calculation method of  $\epsilon_{\max}$ . The improved calculation method of  $\epsilon_{\max}$  is:

The calculation equation of  $\epsilon_{\max}$  of sunny slope vegetation is:

$$\begin{aligned} \epsilon_{\max} &= k_1 * a_{\max}^n + k_2 * \text{NDVI} * \text{LAI}, 0 \leq S \leq 15, \\ n &= 1, 2, 3... \end{aligned} \quad (3)$$

$$\begin{aligned} \epsilon_{\max} &= k_1 * a_{\max}^n * \left(1.2 - \frac{x}{75}\right)^3 + \\ &k_2 * \text{NDVI} * \text{LAI}, S > 15, n = 1, 2, 3... \end{aligned} \quad (4)$$

The calculation formula of  $\epsilon_{\max}$  for shady slope vegetation is:

$$\begin{aligned} \epsilon_{\max} &= k_1 * a_{\max}^n * \left(1 - \frac{x}{90}\right)^3 + \\ &k_2 * \text{NDVI} * \text{LAI}, S > 0, n = 1, 2, 3... \end{aligned} \quad (5)$$

In the formula,  $a_{\max}^n$  is the maximum value of light energy utilization efficiency of the same vegetation; S is the slope, taking a positive integer; and n is the species and quantity of vegetation. k is the adjustment coefficient. When  $\text{NDVI} \leq 0.1$ ,  $k_1 = 0$ ,  $k_2 = 0$ ; when  $\text{NDVI} > 0.1$ ,  $k_1 = 1$ ,  $k_2 = 0.1$ .

According to the vegetation distribution data in Sichuan mountain areas, the spatial distribution of different types of mountain vegetation can be obtained, and  $\epsilon_{\max}$  of different types of mountain vegetation in Sichuan can be obtained by referring to previous studies (Table 1) (Zhu et al., 2006; Wang B. et al., 2013; Cai et al., 2014; Wang et al., 2015; Du et al., 2017; Tang et al., 2017; Lan et al., 2018).

The  $\epsilon_{\max}$  value of different vegetation will change according to the external conditions (Zhu et al., 2006; Li et al., 2012). The functional relationship between  $\epsilon_{\max}$  and slope and aspect was established according to Equations 3–5 (Figure 2). On this basis, according to Landsat 8 data (Figure 3B), the spatial distribution of hillside vegetation types and the relationship between altitude (Figure 3A) and  $\epsilon_{\max}$  values of different

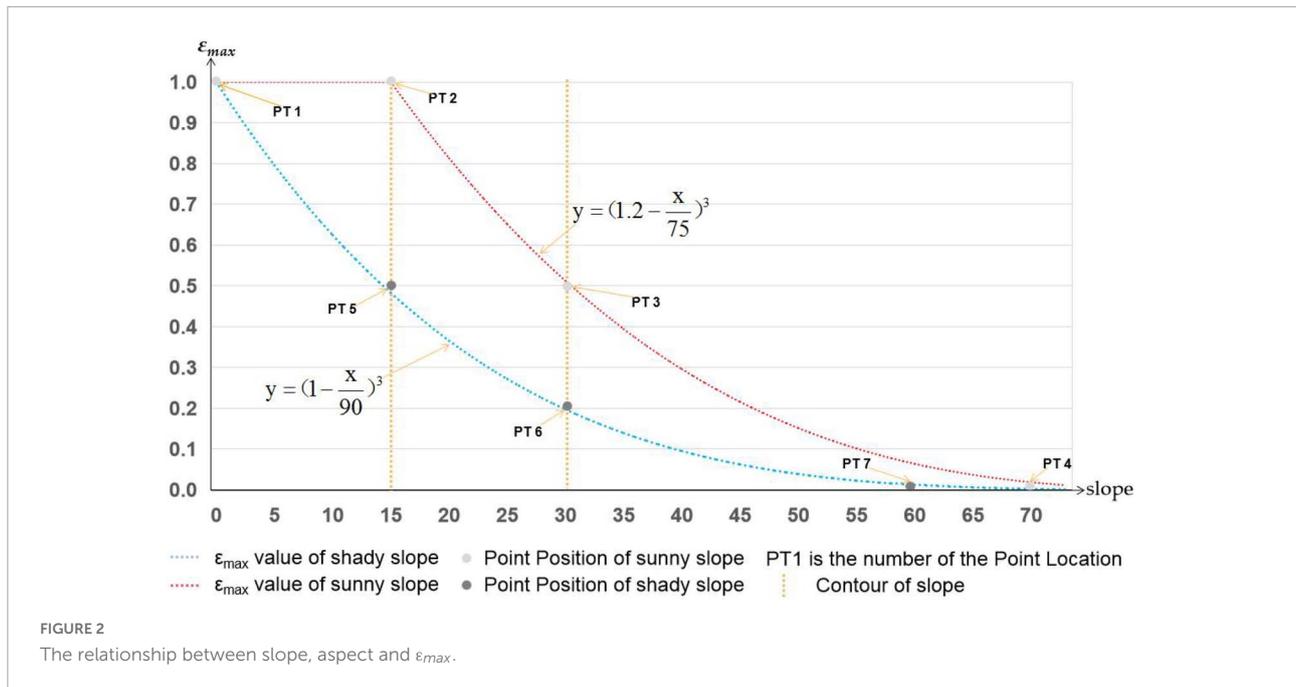


TABLE 1  $\epsilon_{max}$  of different vegetation types in the study area under ideal conditions.

Broad-leaved forest	Coniferous forest	Mixed forest	Deciduous shrub	Cultivated vegetation	Other
1.14	1.05	1.12	0.768	0.89	0.6

vegetation types (Table 1), the functional relation between  $\epsilon_{max}$  values of different vegetation types and slope (Figure 3C) was calculated through Equations 3–5. In Figure 3A, the vegetation types are divided into seven categories from low to high altitude, and their  $\epsilon_{max}$  values under ideal conditions are shown in Table 1. According to the slope size and slope orientation, only suitable equations are selected from Equations 3–5 to calculate  $\epsilon_{max}$ . The calculation results are shown in c in Figure 3. The uniform value of  $\epsilon_{max}$  of the CASA model is 0.389 (gC/MJ), while  $\epsilon_{max}$  of the improved CASA model changes with the change in vegetation types, slope, and aspect. This characteristic of the improved CASA model is more consistent with the actual situation of vegetation  $\epsilon_{max}$ . When the sunny slope is less than 35° or the shaded slope is less than 30°, the  $\epsilon_{max}$  value of the improved CASA model is larger than that of the CASA model. In this paper, Equations 3–5, and  $\epsilon_{max}$  in the ideal state of vegetation in Table 1 are used to establish the functional relation between the  $\epsilon_{max}$  value of mountain vegetation and vegetation types, slope, and aspect.

The improvement of tem

Tem indicates the stress parameter of the external temperature on light energy utilization. Its calculation formula

is (Wang Y. et al., 2019):

$$Tem = \frac{(T-T_{min})(T-T_{max})}{(T-T_{min})(T-T_{max}) - (T-T_{opt})^2} \tag{6}$$

$T_{min}$ ,  $T_{max}$ , and  $T_{opt}$  are also known as the three fundamental temperatures, which are the lowest, highest and best temperatures of vegetation during photosynthesis. The three fundamental temperatures of different vegetation types are usually obtained by a literature collection (Chen et al., 2014). After consulting relevant documents (Peng et al., 2000; Zhang et al., 2012, 2014; Li, 2013; Wang L. et al., 2013), the three base temperatures of the main vegetation in the study area are shown in Table 2.

The improvement of wat

Wat is the product of vegetation leaf age and leaf water content parameters, indicating the impact of vegetation leaf age and water content on light energy utilization. The calculation formula of Wat is (Zhang and Zhang, 2017; Yan et al., 2018):

$$Wat = \frac{(1+LSWI)^2}{2(1+LSWI_{max})}, LSWI = \frac{(NIR-SWIR)}{(NIR+SWIR)} \tag{7}$$

In the formula, NIR represents the reflectance of vegetation in the near-infrared band, and SWIR represents the reflectance of vegetation in the infrared band.

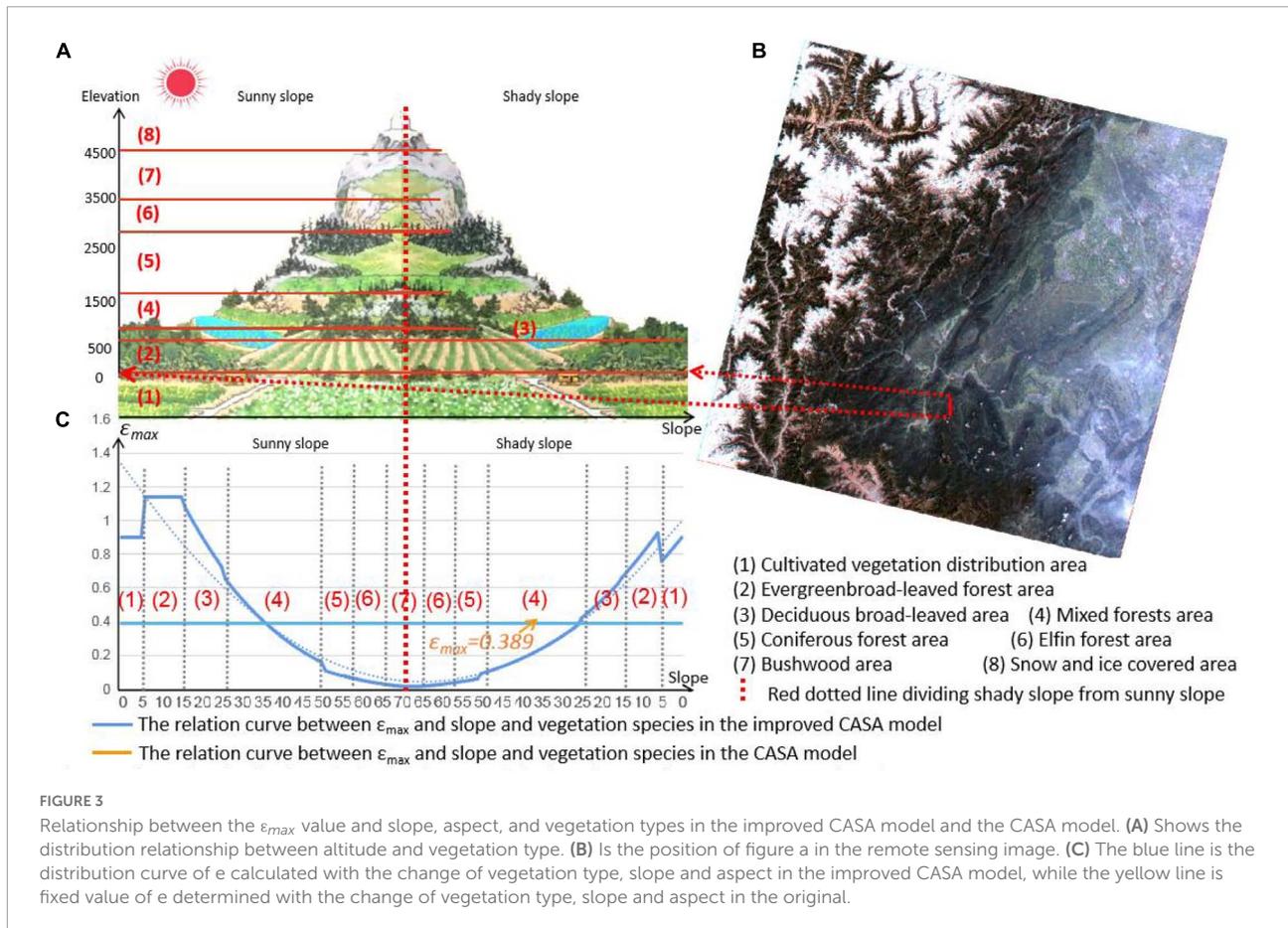


TABLE 2 Three benchmark temperatures for the main vegetation in the study area.

Vegetation types	Minimum temperature (°C)	Optimum temperature (°C)	Maximum temperature (°C)
Coniferous forest	10	19	27
Broad-leaved forest	13	21.5	29
Mixed forest	11.5	20.7	31
Deciduous shrub	11	22.6	32
Cultivated vegetation	12	25	35
Other	11	22	30

### Improvement of absorbed photosynthetic active radiation

In this paper, the calculation method of APAR is improved, and the improvement of the calculation method of APAR is mainly the improvement of FPAR (Zhu et al., 2019). In the CASA model, the functional relationship between the FPAR and SR index (Equation 9) is used to calculate the FPAR. However, according to the research findings, the calculated results of FPAR based on SR are usually lower than the measured values because the NDVI index is inconsistent with the measured values due to the shadow caused by the ground potential factor (Chen et al., 2018; Wang H. et al., 2019; Wang et al., 2020; Xia et al.,

2020). Therefore, in the calculation of FPAR, instead of using the CASA model to calculate the formula of FPAR, the paper selected the average value of the calculated results (Equation 12) of the SR index, EVI index, and NDVI (Equations 9–11) that had a functional relationship with FPAR as the final FPAR value (Li H. et al., 2021; Li M. et al., 2021; Li Y. et al., 2021).

$$APAR = SOL * FPAR * 0.5 \tag{8}$$

$$FPAR_{SR} = \min \left( \frac{SR - SR_{min}}{SR_{max} - SR_{min}}, 0.95 \right), SR = \frac{NIR}{R} \tag{9}$$

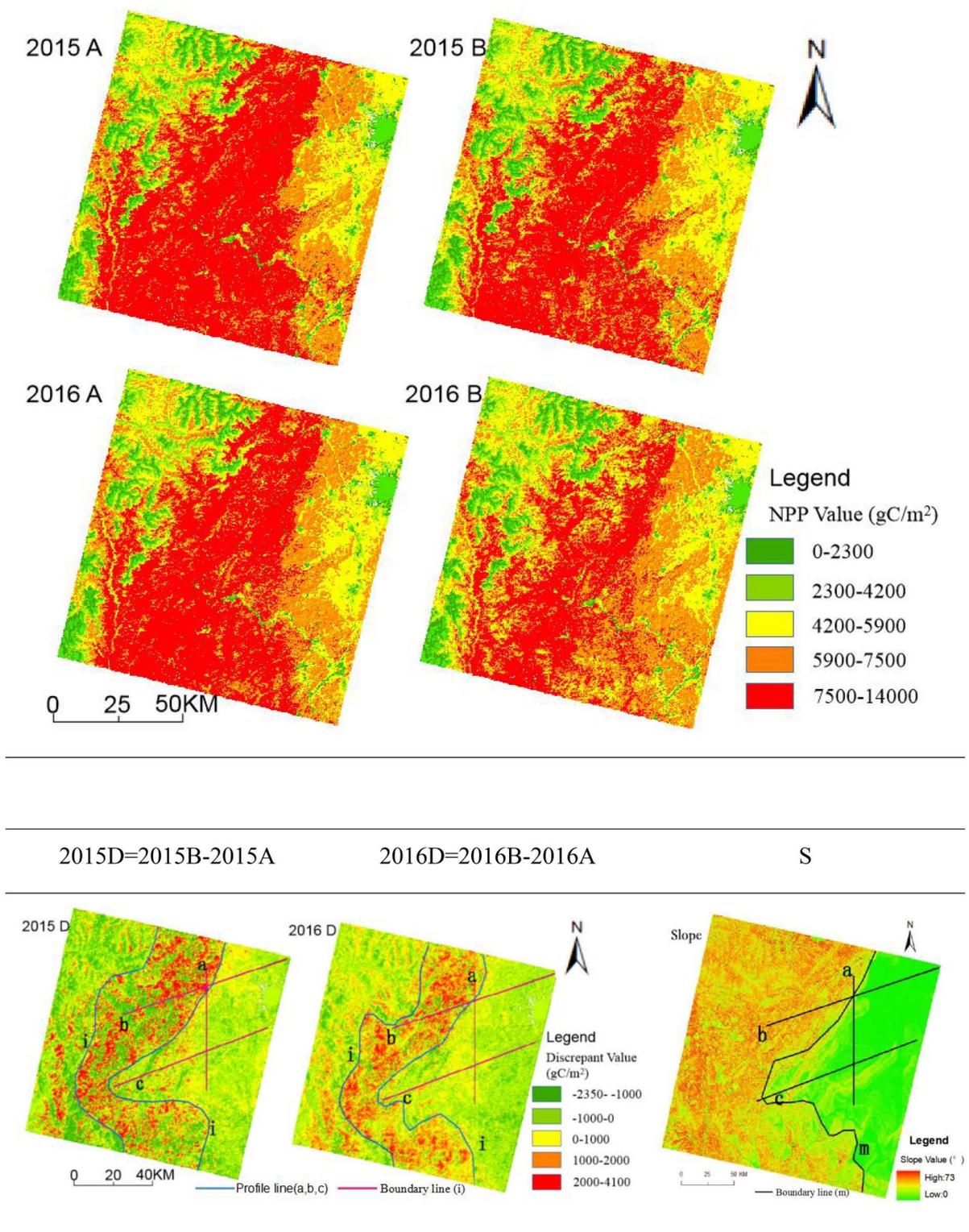


FIGURE 4  
Discrepant data and profile line distribution of the improved CASA model and CASA model estimation results.

$$FPAR_{EVI} = a * EVI \tag{10}$$

$$FPAR_{NDVI} = \frac{(NDVI - NDVI_{min}) (FPAR_{max} - FPAR_{min})}{(NDVI_{max} - NDVI_{min})} + FPAR_{min} \tag{11}$$

$$FPAR = \frac{(FPAR_{SR} + FPAR_{NDVI} + FPAR_{EVI})}{3} \tag{12}$$

In the formula, *a* is the empirical coefficient,  $FPAR_{max}$  and  $FPAR_{min}$  are the maximum and minimum values of FPAR, and the values taken as are 0.95 and 0.001.

### Validation of the model

To verify the feasibility of estimating vegetation carbon sequestration by the improved CASA model, it is necessary to verify the spatial relationship between slope, slope aspect, and the difference in estimating carbon sequestration by the two models (the improved CASA model and the CASA model).

### Verification of slope

In this paper, the improved CASA model and CASA model are used to estimate the vegetation NPP of Landsat 8 OLI image data in the same area at different periods (2015 and

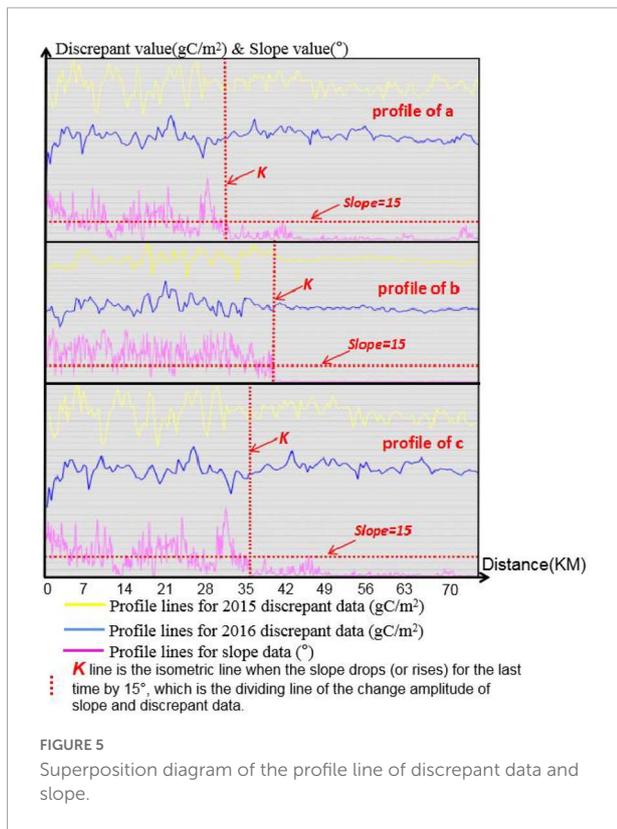


TABLE 3 Statistical table of the distribution of discrepant data on shady slopes and sunny slopes.

Year	Area	Total Of D value	Sunny slope		Shady slope		DRS1/DRS2		
			Area	D value of sunny slope ratio	Area ratio	D value of Shady slope ratio	D Value ratio (DRS1)	D Value ratio (DRS2)	
2015	37206	174608	18505	49.74	115127	65.93	59481	34.07	1.935
2016	37206	246636	18505	49.74	162756	65.99	83880	34.01	1.940

2016). The CASA model estimation result (2015B and 2016B in Figure 4) is subtracted from the improved CASA model estimation result (2015A and 2016A in Figure 4) to obtain the discrepant data (2015D and 2016D in Figure 4) between the two model estimation results. Then, the profile line (a, b, c) of the discrepant data and slope (S in Figure 4) and the boundary line (i,m) of high and low values are established by ArcGIS software to analyse the change rule of slope and discrepant data.

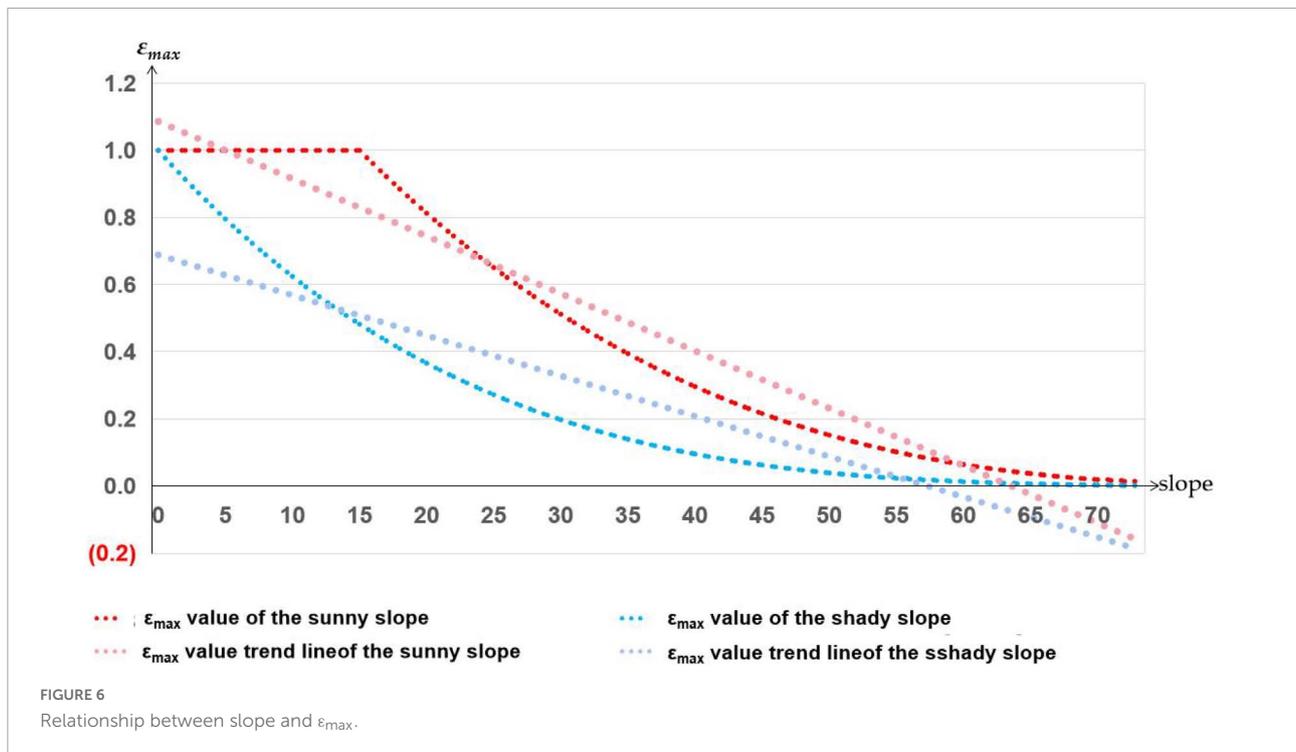
The discrepant data between 2015 and 2016 show that while the negative value of the discrepant data is primarily dispersed in the east and northwest of the research region with a modest slope, the positive value is primarily concentrated in the area with a high slope. The reason is that, as seen in Figure 3, the  $\epsilon_{max}$  value of the improved CASA model is greater than that of the CASA model when the sunny slope is less than  $35^\circ$  or the shady slope is less than  $26^\circ$  and smaller than that of the CASA model when the slope is greater than  $35$  or  $26^\circ$ . As a result, the area with a large slope is where the positive value of the resulting discrepant data is primarily spread, whereas the area with a small slope is where the negative value is disseminated. Second, the high value areas of the discrepant data are concentrated in the areas with a large slope, as can be observed from D (2015), D (2016), and slope data (S), but the eastern parts with relatively smooth terrain are rarely distributed. Despite having a considerable slope, the northwest study area has a low value of different data due to the greater altitude and colder temperatures, which cause less vegetation and lower discrepant data. The borderline (i) west of the discrepant data is the same, showing that other factors, such as climate, have taken the place

of the topographic component as the progressive rise in altitude has affected the estimation results. The borderline (i) east of the discrepant data is essentially consistent with the boundary line (m) of the slope, indicating that slope is one of the main factors affecting the estimation results and that the change in the discrepant data is positively correlated with the slope in the easternmost areas of the study area.

It can be seen from the profiles of the three profile lines (Figures 5a–c) that when the slope rises for the first time or drops to  $15^\circ$  for the last time, the boundary line (K line) of the discrepant data fluctuation will be generated. On the left side of the K line, the discrepant data between 2015 and 2016 and the undulation of the slope are large; on the right side of the K line, the discrepant data and gradient between 2015 and 2016 fluctuate slightly. It shows that the variation value of discrepant data is larger in areas with larger slopes and smaller in areas with smaller slopes. At the same time, the improved CASA model also uses  $15^\circ$  as the dividing point to divide the impact of slope on the  $\epsilon_{max}$ . Therefore, Figures 4, 5 show the distribution characteristics of the estimation results and slopes, which shows the feasibility of the improved CASA model in carbon sequestration estimation.

### Verification of slope aspect

The second difference between the improved CASA model and the CASA model is that the improved model divides the slope aspect into the shady slope and the sunny slope and establishes the functional relationship between the shady slope, sunny slope, and  $\epsilon_{max}$ . To verify the relationship between the



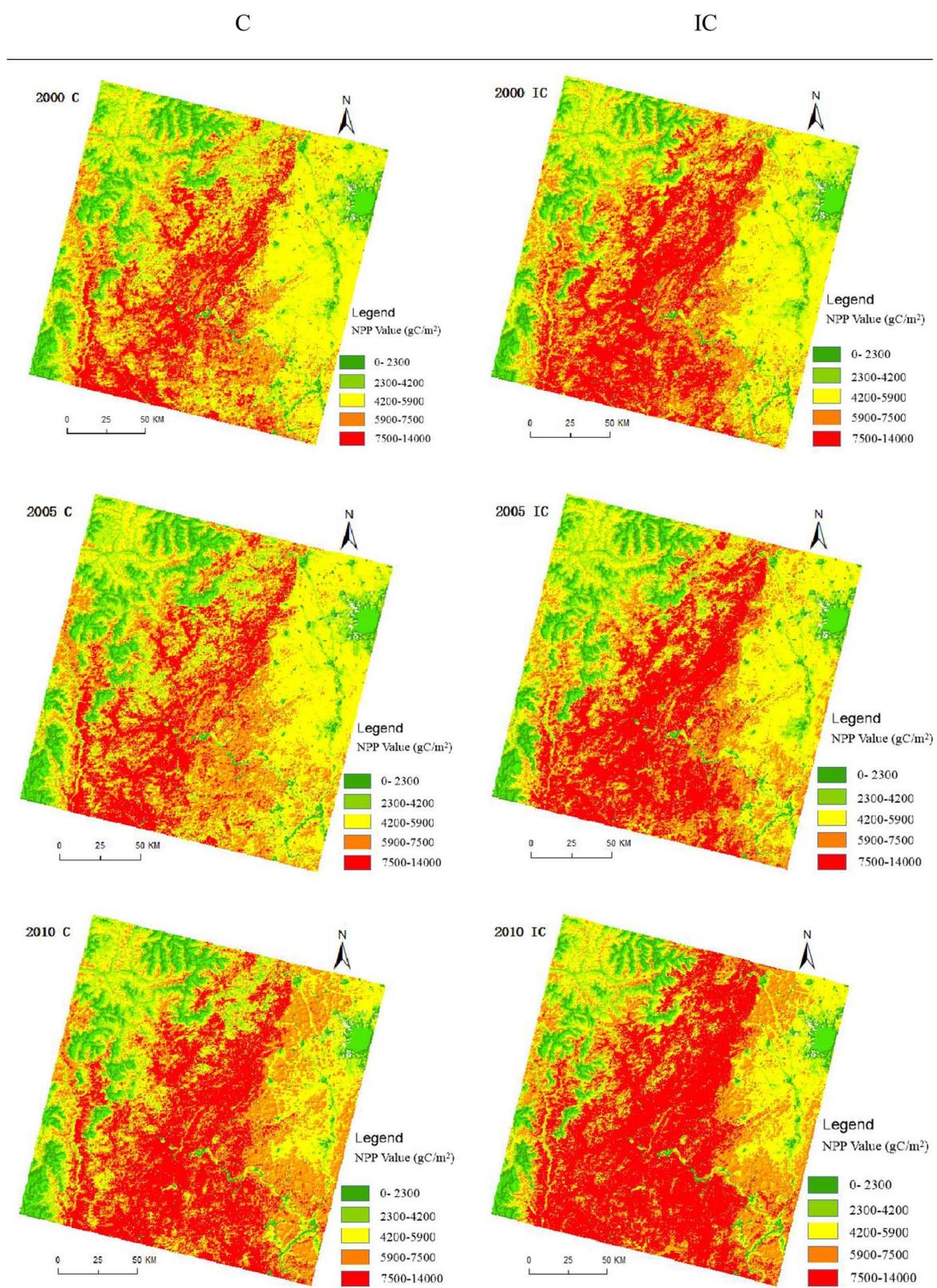
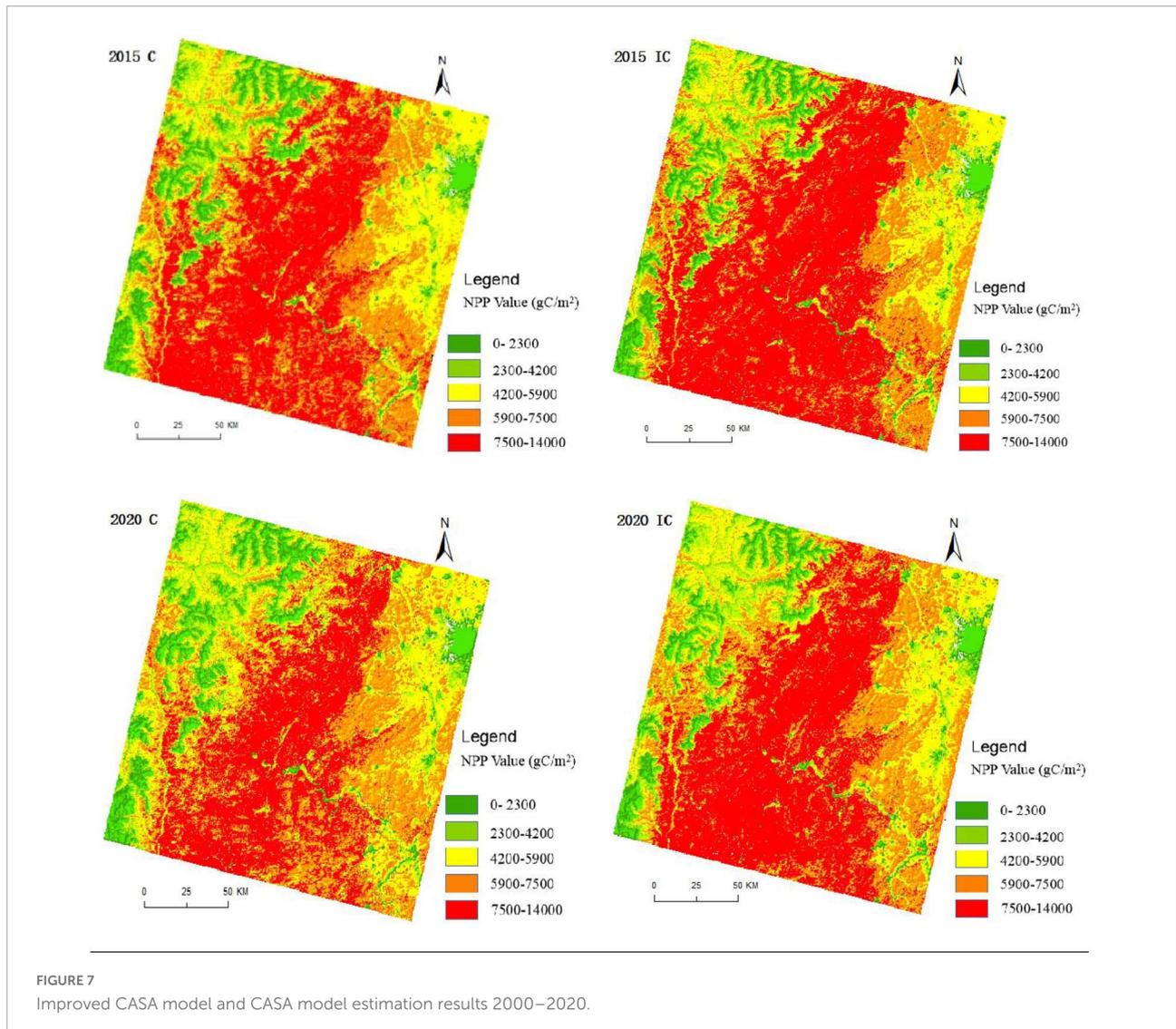


FIGURE 7  
(Continued)



discrepant data and the slope aspect, the study counts the area of sunny slopes and shady slopes of the study area, superimposes the discrepant data with the slope aspect map of the study area, and counted the distribution of the discrepant data on the sunny slopes and shady slopes. According to statistical [Table 3](#), the areas of shady slopes (50.26%) and sunny slopes (49.74%) in the study area are the same. If the influence of the slope aspect on  $\epsilon_{\max}$  is not considered, the distribution of discrepant data on the shady slope and sunny slope should also be the same. However, the geometric calculation formula of slope aspect and  $\epsilon_{\max}$  was established by the improved CASA model, and the sum of  $\epsilon_{\max}$  values of the sunny slopes is 1.57 times the sum of  $\epsilon_{\max}$  values of the shady slopes ([Figure 6](#)). Therefore, without considering the influence of other factors, the ratio of the data value of the discrepant data on the sunny slopes to the data value of the discrepant data on the shady slope is greater than 1.57 because vegetation coverage on the sunny slope is higher than that on

the shady slope. According to the data statistics, the distribution discrepant data value of the sunny slope in the study area is 1.94 times that of the shady slope, which proves the influence of the improved slope aspect of the CASA on the estimation result of carbon sequestration.

## Results

To verify the accuracy of the estimation results of the improved CASA model, the results of vegetation carbon sequestration (IC and C in [Figure 7](#)) estimated by the improved CASA model and CASA model in the study area from 2000 to 2020 were compared with the calculation results of NPP products from MODIS data. It should be noted that the estimation result here is not the annual NPP value of the study area, but the author selects remote sensing images with good

quality each year according to the quality of remote sensing images in the study area from 2000 to 2020 to estimate the carbon sequestration in the study area. In contrast, MODIS NPP data are synthetic data of 8 days that overlap with the acquisition time of remote sensing images to reduce the estimation error. The accuracy of the improved CASA model was verified by analysing the correlation and the degree of fit among the three. **Figure 7** shows only the estimates for 2000, 2005, 2010, 2015, and 2020.

According to the basic principle that chloroplasts need to absorb 1.63 g of carbon from the air for every 1 g of organic material produced during photosynthesis (Liu X. et al., 2013; Liu Y. et al., 2013), the total forest carbon sequestration in Sichuan mountain areas from 2000 to 2020 can be estimated by the CASA model, improved CASA model and MODIS data products (**Table 4**).

**Table 4** shows that the calculated value of MODIS data (M) > estimated value of improved CASA model (IC) > estimated value of CASA model (C), the estimated average value of improved CASA model is 554,48 tons, the estimated average value of CASA model is 538,800 tons, and the mean value of MODIS data is 583,000 tons. The improved model increased the estimated mean value by 2.92%. This assumes that the estimated value of MODIS data is the measured value and compares the improved CASA model and the relationship between the estimated value of the CASA model and MODIS

data to determine the accuracy of the estimation of the improved CASA model.

Because the impact of slope and aspect on vegetation carbon sequestration estimation is relatively stable, its impact will not be affected by the external temperature and precipitation. Therefore, if the difference in the estimation results of the three methods is relatively stable and the variation trend is the same, it can show the accuracy of the improved model in the estimation of forest carbon sequestration in mountainous areas. As shown in **Figure 8**, the trend lines of vegetation carbon sequestration estimated by the three methods are consistent, and the changing trend is consistent. The trend line fitting degree index of the improved CASA model ( $R^2 = 0.5531$ ) is the largest and the fitting degree is the best.

Assuming that the MODIS data are measured values, the correlation between the improved CASA model, the estimated results of the CASA model, and the calculated results of the MODIS data is calculated to determine whether the estimated results of the improved CASA model are accurate. In **Figure 9**, the correlation between the estimated results of the improved CASA model and MODIS data ( $R^2 = 0.9596$ ) is higher than that between the estimated results of the CASA model and the calculated results of MODIS data ( $R^2 = 0.9465$ ). Thus, the accuracy of the improved CASA model in estimating forest carbon sequestration in mountainous areas was illustrated.

TABLE 4 Estimation of vegetation carbon sequestration in the study area from 2010 to 2020 (unit: 10,000 tons).

Year/Carbon sequestration	Estimated value of MODIS (M)	Estimated value of improve the model (IC)	Estimated value of CASA model (C)	(IC-C)/C*100
2000	52.854	50.078	48.887	2.44
2001	54.239	51.507	50.328	2.34
2002	57.179	54.153	52.91	2.35
2003	57.372	54.32	53.116	2.27
2004	53.961	50.921	49.782	2.29
2005	55.447	52.497	51.271	2.39
2006	59.64	56.828	55.515	2.37
2007	58.501	55.518	54.226	2.38
2008	58.049	54.937	53.622	2.45
2009	60.246	57.227	55.816	2.53
2010	55.321	53.204	51.086	4.15
2011	59.465	56.039	55.158	1.60
2012	54.488	52.907	50.138	5.52
2013	63.253	59.975	58.517	2.49
2014	59.464	56.647	54.765	3.44
2015	62.157	59.572	57.27	4.02
2016	62.095	58.434	57.347	1.90
2017	59.543	57.554	54.886	4.86
2018	57.148	55.777	52.536	6.17
2019	62.734	59.1	57.977	1.94
2020	61.151	57.205	56.337	1.54
Average Value	58.3	55.448	53.88	2.92

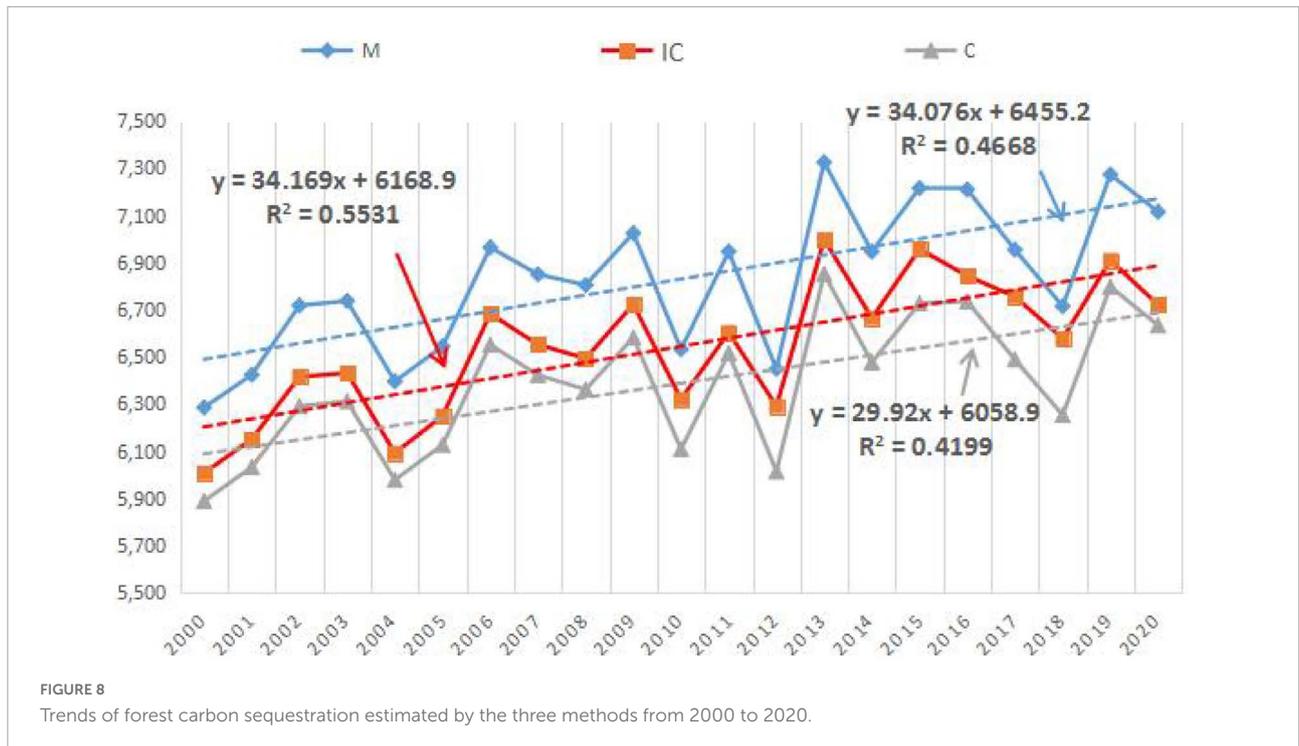


FIGURE 8 Trends of forest carbon sequestration estimated by the three methods from 2000 to 2020.

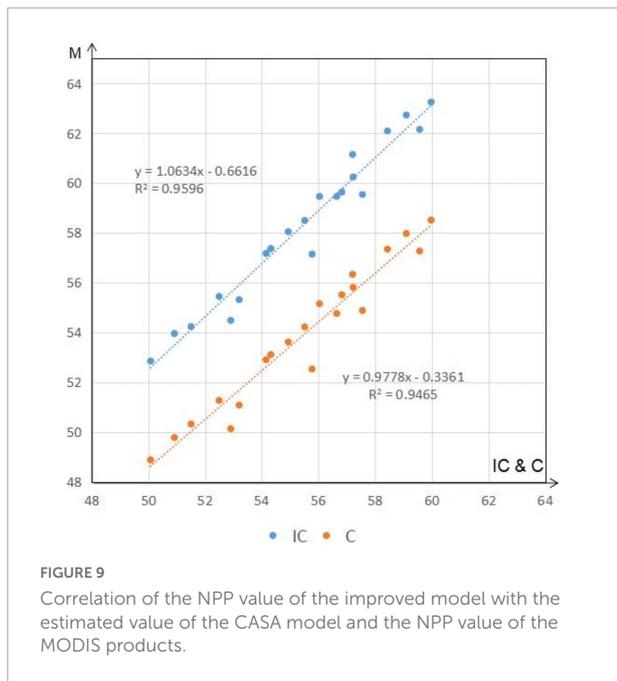


FIGURE 9 Correlation of the NPP value of the improved model with the estimated value of the CASA model and the NPP value of the MODIS products.

## Discussion

Compared with other improved CASA models, the improved CASA model's biggest characteristic is to build the  $\epsilon_{max}$  and vegetation types, slope, and slope to the function relation, overcome the CASA model for the  $\epsilon_{max}$  unified assigned to estimate the precision of the impact of makesup

for the CASA model to estimate does not consider terrain factors. After verifying the feasibility of the improved CASA model through statistical analysis of the data and comparing the estimation results of the improved CASA model, CASA model, and MODIS NPP product, it is found that the calculation results of the improved CASA model are between the CASA model and MODIS data product. Because MODIS data apply to the global-large scale spatial range, the estimated result of carbon sequestration is larger than the actual (Steinberg et al., 2006; Chen et al., 2008; Li et al., 2010; Zhu et al., 2018). However, the estimation results of the CASA model are mainly combined with climate conditions such as temperature and precipitation, and due to the influence of mountainous terrain factors, the estimated results are slightly smaller than the actual carbon sequestration value (As-syakur et al., 2010; Li et al., 2019a,b; Meraj et al., 2022). Therefore, the estimation results of the improved CASA model have the best fitting degree and good correlation between the CASA model and MODIS data products, indicating that the improved CASA model has improved the estimation accuracy of the model to a certain extent.

However, the improved CASA model also has shortcomings: the improved CASA model requires high-quality remote sensing image data, especially a low proportion of cloud pixels in the image (Kuang et al., 2020; Yong et al., 2022). Because cloud pixels have a great impact on the acquisition of the vegetation index, the number of cloud pixels directly affects the accuracy of estimation, and remote sensing images more or less have cloud pixels. Therefore, to improve the accuracy of carbon sequestration estimation, it is necessary to select

remote sensing images with fewer cloud pixels and good quality to estimate vegetation sequestration, which increases the workload to a certain extent; Second, in mountainous areas with a large degree of relief, there are large differences in temperature, humidity, and precipitation between the top and bottom of the slope (Wang et al., 2018). Therefore, when the improved CASA model is used to analyse meteorological data, it is easy to cause inconsistencies and uncertainties in the accuracy of interpolation results. The greater the degree of relief is, the greater the uncertainty. The improved CASA model has improved the accuracy of carbon sequestration estimation of vegetation in mountainous areas with a large degree of relief but has not significantly improved the accuracy of carbon sequestration estimation of vegetation in plains and hills with a relatively high degree of relief. The solution to this problem also depends on appropriately increasing the number of meteorological stations and the reasonable spatial layout of meteorological stations. Moreover, the change in land use types will lead to the change in vegetation types in space (Li, 2021; Wang et al., 2022), thus leading to the change in the maximum light energy utilization rate of vegetation. On a short time scale, due to the small change in vegetation types, the estimated result image of the improved CASA model is small. If it is on a longer time scale, it will have a greater impact on the estimation results. The impact of slope on the carbon sequestration of vegetation is the impact of soil properties on the carbon sequestration of vegetation. Soil properties include soil organic matter content, soil thickness, soil moisture, and other factors. In general, the greater the slope is, the smaller the soil thickness and the lower the soil organic matter and soil water content. Due to the flourishing vegetation in mountain areas, it is difficult to obtain soil property information by using remote sensing technology. However, the spatial resolution of soil information today is low, and the precision is not high, which cannot meet the research needs. Therefore, the paper uses slope to reflect the impact of soil properties on the carbon sequestration of vegetation from the side, which will also affect the estimation accuracy to a certain extent. The next step will be to strengthen the research of remote sensing technology in information acquisition of high vegetation areas to overcome the shortcomings of the improved CASA model.

## Conclusion

The improved CASA model overcomes the shortcomings of the CASA model in that the  $\epsilon_{\max}$  is uniformly assigned and the influence of terrain factors is not considered. The calculation method of key parameters is improved, and the functional relation between vegetation, aspect, slope, and  $\epsilon_{\max}$  of vegetation is established. After verifying the feasibility of the improved model, and comparing the estimation results of the three methods, it is found that the improved CASA model has

the best fitting degree and high correlation. The improved CASA model is helpful to more accurately grasp the ecological carbon sequestration capacity of mountain vegetation.

## Data availability statement

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

## Author contributions

XH, LH, and ZH: conceptualization and methodology. XH and LH: writing—original draft preparation and writing—review and editing. XN and PL: visualization. HY and PL: data curation. All authors have read and agreed to the published version of the manuscript.

## Funding

This research was supported by Natural Science Foundation of Sichuan Province (2022NSFSC1040) and supported by the Independent Research Project of the State Key Laboratory of Geohazard Prevention and Geoenvironment Protection Independent Research Project (SKLGP2021Z003).

## Acknowledgments

Thank you for the free data provided by the platform and the scholars who provided references, as well as the State Key Laboratory of Geohazard Prevention and Geoenvironment Protection and Chengdu University of Technology for their support.

## Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

## Publisher's note

All claims expressed in this article are solely those of the authors and do not necessarily represent those of their affiliated organizations, or those of the publisher, the editors and the reviewers. Any product that may be evaluated in this article, or claim that may be made by its manufacturer, is not guaranteed or endorsed by the publisher.

## References

- Anikó, K., Hrvoje, M., Laura, D., Mislav, A., Tomáš, H., and Zoltán, B. (2017). Identification of years with extreme vegetation state in central Europe based on remote sensing and meteorological data. *South East Eur. For.* 8, 1–20.
- As-syakur, A. R., Osawa, T., and Adnyana, I. W. S. (2010). Medium spatial resolution satellite imagery to estimate gross primary production in an Urban Area. *Remote Sens.* 2:1496.
- Cai, W., Wenping, Y., Shunlin, L., Liu, S., Dong, W., Chen, Y., et al. (2014). Large differences in terrestrial vegetation production derived from satellite-based light use efficiency models. *Remote Sens.* 6, 8945–8965. doi: 10.3390/rs6098945
- Cai, Y., Zheng, J., Du, M., Mu, C., and Peng, J. (2018). “Grassland npp monitoring based on multi-source remote sensing data fusion,” in *Proceedings of the ISPRS technical commission III midterm symposium on developments, technologies and applications in remote sensing*, 371–376.
- Chen, F., Li, H., and Liu, Y. (2018). Temporal and spatial differentiation of vegetation NPP in mountainous areas of Southern Yunnan and its influencing factors based on GIS and CASA: Taking Yuanyang County, Yunnan province as an example. *J. Ecol.* 37, 2148–2158. doi: 10.13292/j.1000-4890.201807.003
- Chen, J., Yan, H., Wang, S., Gao, Y., Huang, M., Wang, J., et al. (2014). Estimation of total primary productivity of China's terrestrial ecosystem by VPM remote sensing model. *Quat. Res.* 34, 732–742.
- Chen, L., Gao, Y., Li, L., Liu, Q., and Gu, X. (2008). Forest NPP estimation based on MODIS data under cloudless condition. *Sci. China Ser. D Earth Sci.* 51, 331–338.
- Chen, S., Guo, B., Zhang, R., Zang, W., Wei, C., Wu, H., et al. (2021). Quantitatively determine the dominant driving factors of the spatial-temporal changes of vegetation NPP in the Hengduan Mountain area during 2000–2015. *J. Mt. Sci.* 18, 427–445.
- Chen, T., Peng, L., Liu, S., and Wang, Q. (2017). Spatio-temporal pattern of net primary productivity in Hengduan Mountains area, China: Impacts of climate change and human activities. *Chin. Geogr. Sci.* 27, 948–962.
- Chen, X., and Zeng, Y. (2016). Temporal and spatial variation of vegetation NPP and its relationship with climatic factors in subtropical mountainous and hilly region: A case study of Hunan province. *Acta Geogr. Sin.* 1, 35–48.
- Christian, K., and Jens, P. (2011). A definition of mountains and their bioclimatic belts for global comparisons of biodiversity data. *Alp. Bot.* 121:73. doi: 10.1007/s00035-011-0094-4
- Du, M., Zheng, J., Ren, X., Cai, Y., Mu, C., and Yan, K. (2018). Effects of topography on the distribution pattern of grassland net primary productivity in Changji prefecture, Xinjiang. *Acta Zool. Sin.* 38, 4789–4799.
- Du, B., Alattentua, Bao, G., Wang, N., and Lin, K. (2021). Simulation of net primary productivity of Xilin Gol grassland based on CASA model. *Res. Soil Water Conserv.* 28, 293–300. doi: 10.13869/j.cnki.rswc.2021.05.034
- Du, S., Liu, L., Liu, X., and Hu, J. (2017). Response of canopy solar-induced chlorophyll fluorescence to the absorbed photosynthetically active radiation absorbed by chlorophyll. *Remote Sens.* 9:911. doi: 10.3390/rs9090911
- Fan, Y. (2018). *Temporal and spatial variation of vegetation carbon source/sink and its response to climate change in Hexi Corridor*. Lanzhou: Northwest Normal University.
- Field, C. B., Behrenfeld, M. J., Randerson, J. T., and Falkowski, P. (1998). Primary production of the biosphere: Integrating terrestrial and oceanic components. *Science* 281, 237–240. doi: 10.1126/science.281.5374.237
- Han, H. (2020). *Relationship between vegetation phenology and net primary productivity (NPP) and its response to climate change in the Loess Plateau*. Shaanxi: Shaanxi Normal University.
- He, L., Li, A., Yin, G., and Bian, J. (2019). Retrieval of grassland aboveground biomass through inversion of the prosoil model with modis imagery. *Remote Sens.* 11:1597. doi: 10.3390/rs11131597
- He, N., Zhang, J., Liu, C., Xu, L., Chen, Z., Liu, Y., et al. (2018). Research progress on spatial pattern and influencing factors of forest ecosystem traits-integrated analysis based on transect in eastern China. *J. Ecol.* 38, 6359–6382.
- Hicke, J. A., Asner, G. P., Randerson, J. T., and Tucker, C. J. (2002). Trends in North American net primary productivity derived from satellite observations, 1982–1998. *Glob. Biogeochem. Cycles* 16, 1019–1040.
- Jiang, H., Yu, J., Jiang, S., Huang, B., and Li, Y. (2021). Remote sensing estimation and analysis of NPP of vegetation in Fujian province based on CASA model and SEVI index. *J. Hainan Univ.* 39, 372–382.
- Jiang, H., Yuan, Y., and Wang, S. (2019). The performance evaluation and application of shadow elimination vegetation index (SEVI) in removing topographic umbra and falling shadow interference. *J. Geoinform. Sci.* 21, 1975–1984.
- Körner, C., Urbach, D., and Paulsen, J. (2021). Mountain definitions and their consequences. *Alp. Bot.* 131, 1–5. doi: 10.1007/S00035-021-00265-8
- Kuang, Q., Yuan, Q., Han, J., Leng, R., Wang, Y., Zhu, K., et al. (2020). A remote sensing monitoring method for alpine grasslands desertification in the eastern Qinghai-Tibetan Plateau. *J. Mt. Sci.* 17, 1423–1437.
- Lan, F., Wenmin, Q., Lunche, W., Aiwen, L., and Ming, Z. (2018). Comparison of artificial intelligence and physical models for forecasting photosynthetically active radiation. *Remote Sens.* 10, 1855–1855. doi: 10.3390/rs10111855
- Li, A., Bian, J., Lei, G., and Huang, C. (2012). Estimating the maximal light use efficiency for different vegetation through the CASA model combined with time-series remote sensing data and ground measurements. *Remote Sens.* 4, 3857–3876. doi: 10.3390/rs4123857
- Li, C. (2021). *Accuracy improvement method and influencing factors of aboveground biomass estimation of mid-subtropical forests by remote sensing*. Ph.D. thesis. Jiangsu: Nanjing Forestry University. doi: 10.27242/-d.cnki.gnjlu.2021.000024
- Li, C., Cao, H., Fan, Y., Han, H., Sun, H., and Wang, Y. (2019a). Remote sensing estimation and analysis of NPP based on corrected CASA model: A case study of Hexi corridor. *Acta Ecol. Sin.* 39, 1616–1626.
- Li, C., Han, H., Fan, Y., Cao, H., Wang, Y., and Sun, H. (2019b). Variation and scenario simulation of NPP in wudaoliang region of qinghai-tibet plateau based on Biome-BGCmodel. *Geogr. Sci.* 39, 1330–1339. doi: 10.13249/-j.cnki.sgs.-2019.08.015
- Li, C., Yang, S., Wu, F., Xu, Y., and Cui, X. (2022). Temporal and spatial differentiation of carbon emissions from energy consumption and carbon sequestration by vegetation in China under the “double carbon” background. *China Environ. Sci.* 42, 1945–1953.
- Li, G., Zhang, H., Fan, W., Zhang, H., and Xin, X. (2010). Accuracy verification and analysis of MODIS / fAPAR in hulunbuir grassland area. *J. Agric. Eng.* 26, 217–224+8.
- Li, H., Qu, Y., Zeng, X., Zhang, H., Cui, L., and Luo, C. (2021). Dynamic response of the vegetation carbon storage in the sanjiang plain to changes in land use/cover and climate. *Herit. Sci.* 9:134. doi: 10.1186/S40494-021-00605-1
- Li, M., Shang, G., Zhou, P., and Deng, L. (2021). Spatial distribution and influencing factors of ecosystem carbon storage in the Loess Plateau. *Acta Ecol. Sin.* 41, 6786–6799.
- Li, X. (2013). *Remote sensing estimation of NPP in Heihe river basin and analysis of its spatiotemporal change characteristics*. Shaanxi: Shaanxi Normal University.
- Li, Y., Tao, S., Li, R., et al. (2021). Spatial and temporal evolution of net primary productivity in Liaoning province and its response to topographic factors. *J. Meteorol. Environ.* 37, 107–112.
- Liu, M., Jiao, J., Pan, J., Song, J., Che, Y., and Li, L. (2020). Changes in the spatiotemporal pattern of net primary productivity (npp) and its driving factors in qinghai province. *Acta Ecol.* 40, 5306–5317.
- Liu, X., Ren, Z., and Lin, Z. (2013). Dynamic evaluation of carbon sequestration and oxygen release value of ecosystem in qinghai tibet plateau. *Geogr. Res.* 32, 663–670.
- Liu, Y., Huang, B., Yi, C., Chen, T., Yu, J., and Qu, L. (2013). Remote sensing simulation and analysis of vegetation net primary productivity based on terrain correction. *J. Agric. Eng.* 29, 130–141.
- Meraj, G., Kanga, S., Ambadkar, A., Kumar, P., Singh, S. K., Farooq, M., et al. (2022). Assessing the yield of wheat using satellite remote sensing-based machine learning algorithms and simulation modeling. *Remote Sens.* 14, 3005–3005.
- National Meteorological Information Center [NMC] (2018). *China meteorological data service centre*. Beijing: National Meteorological Information Center [NMC].
- Pan, J., and Xu, B. (2020). Simulation of NPP spatial distribution of potential vegetation in China. *Chin. J. Ecol.* 39, 1001–1012. doi: 10.13292/j.1000-4890.202003.030
- Peng, C., Zhou, Q., Meng, H., Li, M., Zhou, L., and Liu, X. (2021). Sensitivity analysis of landscape ecology and identification of key restoration areas in Chongqing section of three gorges reservoir based on LECl. *Soil Water Conserv. Res.* 28, 325–330+339+4.
- Peng, S., Guo, Z., and Wang, B. (2000). Estimation of light utilization rate of vegetation in Guangdong by GIS and RS. *Acta Zool. Sin.* 6, 903–909.

- Price, M. F., Arnesen, T., Erik, G., and Metzger, M. J. (2019). Mapping mountain areas: Learning from global, European and Norwegian perspectives. *J. Mt. Sci.* 16, 1–15.
- Shen, X., Wang, P., Wang, W., Wang, S., Xie, R., Wang, Y., et al. (2022). Summary of research methods on net carbon sink capacity of coastal salt marsh. *J. Ecol.* 41, 792–803. doi: 10.13292/j.1000-4890.202203.009
- Singh, R., Semwal, D. P., Rai, A., and Chhikara, R. S. (2002). Small area estimation of crop yield using remote sensing satellite data. *Int. J. Remote Sens.* 23, 49–56. doi: 10.1080/01431160010014756
- Steinberg, D. C., Goetz, S. J., and Hyer, E. J. (2006). Validation of MODIS F/Sub PAR/ products in boreal forests of Alaska. *IEEE Trans. Geosci. Remote Sens.* 44, 1818–1828.
- Sun, Q., Feng, X., and Xiao, X. (2014). Spatial and temporal pattern of net primary productivity of vegetation in Wuling mountain area and its relationship with topographic factors. *J. Geo Inf. Sci.* 16, 915–924.
- Sun, Q., Li, B., Li, F., Zhang, Z., Ding, L., Zhang, T., et al. (2016). Advances in estimation of net primary productivity of vegetation in the three-river headwaters region. *Acta Geogr. Sin.* 71, 1596–1612.
- Sun, Z., and Xie, S. (2021). Spatial-temporal variation and factor detection of net primary productivity in yunnan province based on geographic detector. *Chin. J. Ecol.* 40, 3836–3848. doi: 10.13292/J.1000-4890.202111.033
- Tang, W., Qin, J., Yang, K., Niu, X., Min, M., and Liang, S. (2017). An efficient algorithm for calculating photosynthetically active radiation with MODIS products. *Remote Sens. Environ.* 194, 146–154.
- Wang, B., Wang, J., Yang, Y., Chang, S., Chang, X., and Liu, A. (2013). Improved algorithm of photosynthetic active radiation absorption component and maximum light utilization of vegetation. *Acta Pratacult. Sin.* 22, 220–228.
- Wang, B., Zhou, S., Jing, Y., Song, J., and Wang, R. (2021). Temporal and spatial changes of vegetation phenology in shanxi province and the impact of topography on phenology. *J. Ecol.* 40, 1839–1848. doi: 10.13292/j.1000-4890.202106.030
- Wang, H., Luo, G., Wang, W., Pachikin, K., Li, K., Zheng, H., et al. (2019). Inversion of farmland soil moisture in the middle and lower reaches of the Syr Darya based on multi-source remote sensing data. *J. Nat. Resour.* 34, 2717–2731.
- Wang, J., Feng, L., Palmer, P. I., Liu, Y., Fang, S., Bösch, H., et al. (2020). Large Chinese land carbon sink estimated from atmospheric carbon dioxide data. *Nature* 586, 720–723.
- Wang, J., Zhao, A., and Hu, X. (2021). Spatial and temporal distribution of net primary productivity of vegetation in Beijing. Tianjin and Hebei and analysis of natural driving factors. *J. Ecol. Environ.* 30, 1158–1167. doi: 10.16258/j.cnki.1674-5906.2021.06.006
- Wang, L., Gong, W., Feng, L., Lin, A., Hu, B., and Zhou, M. (2015). Estimation of hourly and daily photosynthetically active radiation in Inner Mongolia, China, from 1990 to 2012. *Int. J. Climatol.* 35, 3120–3131. doi: 10.1002/joc.4197
- Wang, L., Gong, W., Zhang, M., and Ma, Y. (2013). Study on dynamic monitoring of vegetation NPP in Wuhan area. *J. Wuhan Univ.* 38, 548–552. doi: 10.13203/j.whigis2013-05.004
- Wang, W., Zou, W., and Wang, R. (2022). Land space planning path and governance mechanism of urban agglomeration under the goal of “double carbon”. *Environ. Protect.* 50, 64–69. 0253-9705.2022 Z1.011 doi: 10.14026/-j.cnki
- Wang, Y., Cao, G., Wang, Y., Xu, L., Zhang, W., and Wang, X. (2018). Canopy transpiration and its influencing factors of larch huabei artificial forest in the southern side of liupan mountain. *Chin. J. Appl. Ecol.* 29, 1503–1514. doi: 10.13287/j.1001-9332.201805.006
- Wang, Y., Xu, X., Huang, L., Yang, G., Fan, L., Wei, P., et al. (2019). An improved CASA model for estimating winter wheat yield from remote sensing images. *Remote Sens.* 11:1088. doi: 10.3390/rs11091088
- Xia, A., Wang, Y., and Hao, Y. (2020). Advances in estimating carbon storage of grassland vegetation in complex terrain using remote sensing. *Acta Ecol. Sin.* 40, 6338–6350.
- Xing, X., Xu, X., Zhang, X., Zhou, C., Song, M., Shao, B., et al. (2010). Simulating net primary production of grasslands in northeastern Asia using MODIS data from 2000 to 2005. *J. Geogr. Sci.* 20, 193–204.
- Xu, J., Chen, D., Li, W., and Wei, W. (2019). Research on net primary productivity of vegetation in gannan prefecture based on light energy utilization rate model. *Grass Sci.* 36, 2455–2465.
- Xu, Y., Xiao, F., and Yu, Y. (2020). A review on the temporal and spatial distribution of net primary productivity of forest ecosystems in China and its response to climate change. *J. Ecol.* 40, 4710–4723.
- Yan, D., Li, J., Xue, Y., Qi, L., and Luan, Z. (2022). Inversion of vegetation biophysical characteristics in jiangsu province based on sentinel-2. *J. Environ. Sci.* 42, 70–79. doi: 10.13671/j.hjkxxb.2021.0488
- Yan, P., Jin, H., Li, W., Chi, H., and Zhao, Y. (2018). An improved phenology-based CASA model for estimating net primary production of forest in central China based on Landsat images. *Int. J. Remote Sens.* 39, 7664–7692.
- Yang, L. (2019). Evaluating the responses of net primary productivity and carbon use efficiency of global grassland to climate variability along an aridity gradient. *Sci. Total Environ.* 652, 671–682. doi: 10.1016/j.scitotenv.2018.10.295
- Yang, Y., Li, L., Wang, B., Liu, A., and Wang, M. (2015). Simulation of net primary productivity of vegetation in Xilin Gol grassland based on improved CASA model. *J. Ecol.* 34, 2344–2352. doi: 10.13292/j.1000-4890.2015.0203
- Yong, L., Xiu, H., Fu, W., Xu, Y., Zhang, Y., Tang, S., et al. (2022). Estimation of terrestrial net primary productivity in China from fengyun-3D satellite data. *J. Meteorol. Res.* 36, 401–416.
- Yu, H., Ni, W., Zhang, Z., Sun, G., and Zhang, Z. (2020). Regional forest mapping over mountainous areas in northeast china using newly identified critical temporal features of sentinel-1backscattering. *Remote Sens.* 12:1485. doi: 10.3390/rs12091485
- Zhang, J., and Ren, Z. (2016). Spatial and temporal pattern and causes of net primary productivity of vegetation in the Han river basin. *J. Ecol.* 36, 7667–7677.
- Zhang, M., Jiang, W., Chen, Q., and Liu, X. (2012). Simulation of the maximum solar energy utilization of various types of grassland comprehensive sequential classification system based on the improved CASA model. *Grassl. Lawn* 32, 60–66. doi: 10.13817/j.cnki.cyyccp.2012.04.017
- Zhang, Q., Zhang, L., He, H., Han, S., Li, Y., Ou, Y., et al. (2014). Inversion study on maximum solar energy utilization of vegetation based on vorticity related flux data. *Quat. Res.* 34, 743–751.
- Zhang, R., Guo, J., and Zhang, Y. (2020). Spatial distribution pattern of net primary productivity (NPP) of grassland in Xinjiang and its response to climate change. *J. Ecol.* 40, 5318–5326.
- Zhang, S., Deng, W., Hu, M., Zhang, H., Wang, Z., and Peng, L. (2022). Identification and differentiation analysis of human nature interaction in transitional geographical space in mountainous areas. *J. Geogr.* 77, 1225–1243. doi: 10.11821/dlxb202205013
- Zhang, W., Li, A., and Jiang, X. (2013). Quantitative bounds of China's mountain spatial range based on DEM. *Geogr. Geogr. Inf. Sci.* 29, 58–63. doi: 10.13227/j.hjkk.202109031
- Zhang, Y., and Zhang, X. (2017). Estimation of net primary productivity of different forest types based on improved CASA model in Jing-Jin-Ji region, China. *J. Sustain. For.* 36, 568–582. doi: 10.1080/105498-11.2017.1314971
- Zhang, Y., Yuan, Q., Fang, L., Han, Y., Zhu, Q., Qi, L., et al. (2021). Spatial and temporal pattern of net primary productivity of high-altitude spruce fir forests in changbai mountain nature reserve and its driving factors. *J. Ecol.* 40, 3483–3492. doi: 10.13292/j.1000-4890.20211.013
- Zhang, Z., and Min, M. (2022). Study on the spatial-temporal differentiation of vegetation productivity and its relationship with topographic factors in Yili Area. *J. Cent. China Norm. Univ.* 54, 711–720. doi: 10.19603/j.cnki.1000-1190.2020.04.022
- Zhao, R., Huang, X., Yun, W., Wu, K., Chen, Y., Wang, S., et al. (2022). Key issues in the field of natural resource management under the goal of carbon peaking and carbon neutralization. *J. Nat. Resour.* 37, 1123–1136.
- Zhu, W., Pan, Y., He, H., Deyong, Y., and Haibo, H. (2006). Simulation of maximum light use efficiency for some typical vegetation types in China. *Chin. Sci. Bull.* 51, 457–463. doi: 10.1007/s11434-006-0457-1
- Zhu, L., and Yuan, J. (2019). Spatial and temporal distribution of net primary productivity of vegetation in beijing tianjin hebei region and its relationship with topographic factors. *Sci. Technol. Bull.* 35, 197–203. doi: 10.13774/j.cnki.kjtb.2019
- Zhu, Y., Du, L., Xie, Y., Liu, K., Gong, F., Dan, Y., et al. (2018). Accuracy evaluation of different meteorological interpolation methods and their impact on grassland NPP estimation. *Soil Water Conserv. Res.* 25, 160–167. doi: 10.13869/j.cnki.rswc.2018.06.024
- Zhu, Y., Du, L., Xie, Y., Liu, K., Gong, F., Dan, Y., et al. (2019). Temporal and spatial characteristics of grassland net primary productivity and its climate response in Ningxia from 2000 to 2015. *Acta Ecol.* 39, 518–529.