



Sewage Sludge Application in *Eucalyptus urograndis* Plantation: Availability of Phosphorus in Soil and Wood Production

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Abreu-Junior CH, Oliveira MG, Cardoso PHS, Mandu TS, Florentino AL, Oliveira FC, Reis JV, Alvares CA, Stape JL, Nogueira TAR, Capra GF and He Z (2020) Sewage Sludge Application in Eucalyptus urograndis Plantation: Availability of Phosphorus in Soil and Wood Production. Front. Environ. Sci. 8:116. doi: 10.3389/fenvs.2020.00116 Phosphate fertilizers used in planted forests mainly come from scarce and nonrenewable sources, and there is a need for new sources of phosphorus (P). An alternative is the use of sewage sludge (SS), rich in organic matter, nitrogen, and P. The objectives were to evaluate the soil P availability and its effect on wood production in 22month-old Eucalyptus urograndis plantation treated with SS. A field experiment was set up in a randomized complete block design, four replicates, with the following treatments: control without fertilization: mineral fertilization (MF): 14.5 Mg ha⁻¹ of SS + 22 kg ha⁻¹ of P (S1P1); 29 Mg ha⁻¹ of SS (S2); 29 Mg ha⁻¹ of SS + 17.5 kg ha⁻¹ of P (S2P2); and 43.5 Mg ha⁻¹ of SS (S3), dry base. Of the total P in SS, 65% was in organic form and 42% was in labile + moderately labile forms. S2P2 and S3 positively altered the pattern of soil P distribution, with more P in the labile and moderately labile fractions than in the non-labile fraction, along 0-20 cm depths, than MF and control. There were higher microbial and available P as a function of SS dose. Wood volume and biomass were highly related to soil P availability as a function of SS dose. Within the SS treatments, S2P2 resulted in higher gains of volume and biomass of wood. The SS application at the recommend rate, supplemented with 66% of recommended P fertilizer dose in Eucalyptus plantations, can reduce the use of P fertilizer by 33% and N fertilizer by 100% and maintain comparable soil P availability and wood production.

Keywords: alternative fertilizers, biosolids, organic fertilizer, P fractionation, soil P lability

INTRODUCTION

Municipal wastewater treatment plants (WWTPs) have sharply increased worldwide, especially in the industrialized countries as regulated by specific laws (Nascimento et al., 2020). In WWTPs, an important organic residue is generated in huge amounts: the sewage sludge (SS). The management and disposal of SS are complex and expensive, and SS reuse in several fields of application has

tremendously increased worldwide (Cieślik et al., 2015). SS application as soil fertilizer/amendment in agriculture is the most promising practice (Abreu-Junior et al., 2005, 2017; Franco et al., 2010; Florentino et al., 2019; Nascimento et al., 2020). Depending on its origin and quality, SS may contain a huge amount of organic matter and essential elements for plants (Bettiol and Ghini, 2011), thus replacing or integrating mineral fertilizers of non-renewable origin widely used for crop cultivation (Alleoni et al., 2012; Ferraz et al., 2016; Abreu-Junior et al., 2017). This means that the use of non-renewable limited resources, such as phosphate fertilizers, can be significantly reduced by the SS agricultural use (Franco et al., 2010; Abreu-Junior et al., 2017), with a circular-economy perspective.

Phosphorous (P) is one of the nutrients present in high concentration in SS (Abreu-Junior et al., 2005; Costa et al., 2014), representing an important resource for agriculture. Agriculture in tropical areas is mainly conducted in highly weathered, leached, and acidic soils where a rapid loss of P can occur because of fixation by soil components such as 1:1 clays and Fe- and Al-oxyhydroxides (Gustafsson et al., 2012; Rodrigues et al., 2016). Such processes increase the content of legacy-P, P in non-labile forms and not available to plants (Withers et al., 2018). For these soils, higher doses of P fertilizers are required to maintain desired agricultural and forest productivity. Since P fertilizers are of non-renewable origin (mainly extracted from phosphate rock), the constant and increasing exploitation may lead to depletion and scarcity of this resource (Van Vuuren et al., 2010). New and non-conventional P sources are strongly required.

Phosphorus in SS, treated and stabilized for agricultural use, is generally in inorganic forms (Carvalho et al., 2015; Kahiluoto et al., 2015; He et al., 2016). In order to be applied to substitute mineral fertilizers in forestry and agricultural areas, assessment of P availability in SS-amended soil is necessary. The sequential extraction method proposed by Hedley et al. (1982) is an analytical tool for understanding the availability of P in tropical soils, since it segregates organic (P_0) and inorganic (P_i) P fractions, relating their affinity with different soil components. Additionally, it also allows grouping P fractions in labile, moderately labile, and non-labile pool, based on their availability to plants.

By determining P availability in soil, together with the microbial biomass, responsible for SS degradation and P cycling, we can improve our knowledge on P chemistry and recycling in a *Eucalyptus* plantation. Studies on P recycling in forests have been already conducted but for different managements and/or after long periods of residue application. To our knowledge, no studies have been conducted to examine the availability and legacy of P in tropical low fertility soil due to SS application for *Eucalyptus* cultivation. Such information is fundamental to characterize and quantify soil P availability for improved management of SS application to increase fertilizer use efficiency and wood production in the tropical regions.

Forest systems in Brazil, despite their high productivity, are usually characterized by soil with high nutrient deficiency, with particular reference to P in *Eucalyptus* commercial plantations (Alvares et al., 2011; Gonçalves et al., 2013; Binkley et al., 2017). This is true despite the low P requirement of this crop; indeed, *Eucalyptus* has been historically planted in the areas with sandy infertile soils (Bazani et al., 2014). As this commercial plantation has currently covered 7.4 million hectares in the whole Brazil (Sistema Ibge De Recuperação Automática - SIDRA, 2017), high amounts of P are required to sustain the Brazilian roundwood industry, one of the most important and productive in the world. In the last few years, roundwood industries and scholars closely collaborated to find new and sustainable P sources and their application and management in tropical soils.

We hypothesized that the application of SS increases the labile P fraction, P microbial biomass, and P availability in soil, improving the wood production of *Eucalyptus urograndis* plantation. The aims of this work were to evaluate soil P availability and wood production in a 22-month-old commercial *Eucalyptus urograndis* plantation treated with sewage sludge, applied 7 months after planting, combined with or without mineral fertilizer.

MATERIALS AND METHODS

Experimental Area and Soil Characterization

The research was conducted in a commercial *Eucalyptus* field at the Suzano S.A. Company, municipality of Boa Esperança do Sul, State of São Paulo, southern Brazil (21°57′05 S e 48°31′ W). The field experiment was conducted in a stand where *Eucalyptus* trees were commercially cultivated in the last 40 years and SS had been never applied before.

The experimental area has humid-temperate climate, with dry winter and hot summer, and with the rare possibility of frost in early winter (Cwa climate type - Köppen classification) (Alvares et al., 2013, 2018). Average annual rainfall is 1,400 mm, and average minimum and maximum temperature ranges from 15 to 31°C. A detailed pedological survey was carried out before the experiment and the soil was classified as Typic Hapludox (Soil Survey Staff, 2014).

Before the experiment, the soil Ap surface horizon (0-40 cm depth) was sampled at 0-5, 5-10, 10-20, and 20-40 cm depths. Each sub-depth of soil was collected by taking 20 subsamples randomly along the lines of the plantation. Samples were then homogenized, air dried, and sieved to <2.0 mm prior to analysis. All soil samples were characterized according to official physical (Camargo et al., 2009) and chemical (Raij et al., 2001) procedures (Table 1). The densimeter method was used for particle-size analysis. Soil pH was measured potentiometrically with a glass electrode at the soil/solution mixture of 1:2.5 1 mol L^{-1} CaCl₂. Soil organic matter (SOM) content was estimated by the Walkley-Black method. Soil available P, Ca2+, K+, and Mg²⁺ were extracted by an ion-exchange resin procedure. Concentrations of P in the extracts were determined by the colorimetric method, and those of Ca and Mg were determined by atomic absorption spectrophotometer (AAS) using a VARIAN SpectrAA 140 and K using a CORNING 400 flame-photometer. Total acidity (H + AI) was determined by SMP buffer solution at pH 7.0. Cation Exchange Capacity (CEC) was equal to $Ca^{2+} + Mg^{2+} + K^+ + (H + Al)$. Available Cu, Fe, Zn, and Mn

TABLE 1 Soil surface Ap physical ¹ -chemical ² characterization according
to fixed depths.

Attributes	Unit	Depth (cm)			
		0–5	5–10	10–20	20–40
pH _(CaCl2)	_	5.3	5.4	5.7	4.9
Organic matter	g dm ⁻³	14	10	7	3
Resin-P	mg dm ⁻³	7	6	3	2
К	$\rm mmol_{c}~dm^{-3}$	0.3	0.3	0.5	0.4
Ca ²⁺	mmol _c dm ⁻³	23	18	16	9
Mg ²⁺	mmol _c dm ⁻³	2	2	2	1
H + Al	$\rm mmol_{c}~dm^{-3}$	20	19	15	18
Cation exchange capacity	$\rm mmol_{c}~dm^{-3}$	46	40	34	29
Base saturation	%	56	51	54	38
В	mg dm ⁻³	0.3	0.3	0.3	0.2
Cu	mg dm ⁻³	0.3	0.3	0.3	0.3
Fe	mg dm ⁻³	32	28	35	28
Mn	mg dm ⁻³	2.0	1.0	2.0	0.7
Zn	mg dm ⁻³	0.6	0.2	0.6	0.8
Sand (>0.05 mm)	g kg ⁻¹	900	900	900	900
Silt (>0.002 and <0.05 mm)	g kg ⁻¹	10	10	20	20
Clay (<0.002 mm)	g kg ⁻¹	90	90	80	80

¹Camargo et al. (2009). ²Raij et al. (2001).

were extracted by DTPA solution, at pH 7.3, while available B was determined by the hot water method.

Sewage Sludge Characterization

Class B sewage sludge was obtained from the Jundiaí wastewater treatment plant (Jundiaí municipality, São Paulo State). The sludge was generated in a biological system of aerated ponds with a complete mixture followed by sedimentation ponds. The biological sludge was stabilized in the sedimentation ponds for 12 months, which resulted in an OM content of dry solids under 70%. The sludge was then treated with synthetic cationic polymer, centrifuged, and air-dried. Before application, the SS was characterized for both physical properties and potentially toxic elements (PTE) in accordance with method SW-846-3051A (USEPA, 2007) (Table 2). The SS contained high concentrations of total OC, N, and P, while PTE contents were below the limits established by Brasil (2006) and those usually observed in SS (Abreu-Junior et al., 2005; Nascimento et al., 2020), thus allowing its reuse for agricultural purposes (Brasil, 2006). Class B sludges contain detectable levels of pathogens and their agricultural use is regulated by CONAMA Resolution No. 375 (Brasil, 2006).

Field Preparation and Experimental Procedure

Minimum tillage started in January 2015, with subsoiling 0.4 m deep between the lines of the previous rotation over the entire area, where the *Eucalyptus* seedlings were planted. Before planting, 1.8 Mg ha⁻¹ of limestone, to supply calcium and magnesium, was applied at the soil surface along a continuous band close to the seedling. Seedlings of *Eucalyptus urophylla* S.T. Blake × *Eucalyptus grandis* Hill ex Maiden (clone SP5727

TABLE 2	Sewage sludge ch	emical features and	PTE characterization.
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Attribute	Sewage sludge-CSJ	Attribute	Sewage sludge-CSJ
pH-H ₂ O	6.8		mg kg ⁻¹
	%	В	1.0
Moisture	62	As	3.26 (41) ¹
Volatile solids	49	Ba	629 (1300)
Ashes	51	Cd	2.40 (39)
	g kg ⁻¹	Pb	45 (300)
Organic carbon	166	Cu	308 (1500)
Total N	22	Cr	61 (1000)
Total P	30	Hg	<0.05 ² (17)
К	1.7	Мо	8.03 (50)
Na	1.8	Ni	31 (420)
S	24	Se	<0.05 ² (100)
Ca	21	Zn	667 (2800)
Mg	5.0		

 1Limits to SS agricultural use established by Resolution 375 (Brasil, 2006). 2ND , not detected (concentrations $< 0.1 \ mg \ kg^{-1}$).

owned by Suzano, *Eucalyptus urograndis*, hereafter *E. urograndis*) were then planted in February 2015. Seedlings were spaced as $3 \text{ m} \times 2.25 \text{ m}$, with an amount of 1,481 plants per hectare.

The experimental design consisted in totally randomized blocks with six treatments (**Table 3**) and four replicates for each treatment, totaling 24 experimental plots. Each plot had an area of 675 m² planted with 100 trees (10×10). For data collection, we used the 36 trees (6×6) located in the center of each plot, occupying a useful area of 243 m². Mineral fertilizers in the form of triple superphosphate were applied to the planting pit next to the seedling. Seven months after planting, SS was applied in a continuous band of 0.6 m wide, without incorporation into the soil, 0.2 m aside the planting rows.

The studied treatments (Table 3) were control, without SS and mineral fertilizers (C); mineral fertilization for high wood production, as recommended by Suzano S.A. Company (MF); 14.5 Mg ha⁻¹ of SS + 22 kg ha⁻¹ of P (S1P1); 29 Mg ha⁻¹ of SS (S2); 29 Mg ha⁻¹ of SS + 17.5 kg ha⁻¹ of P (S2P2); and 43.5 Mg ha⁻¹ of SS (S3). Sludge doses were calculated according to nitrogen criteria, following CONAMA Resolution No. 375 (Brasil, 2006), in order to provide 192 kg ha⁻¹ of N, as recommended by the Suzano S.A. Company. Thus, the doses of 14.5, 29.0, and 43.5 Mg ha⁻¹ of SS, at dry base, were equivalent to 50, 100, and 150% of the recommended N. Doses of 22 and 17.5 kg ha⁻¹ of P were applied to provide 83 and 66% of mineral P recommendation (26 kg ha^{-1}), combined with SS at 50 and 100% of recommended N doses, S1P1 and 100% S2P2 treatment, respectively, in order to obtain wood production equivalent to the mineral NPK fertilizer application, as recommend by Abreu-Junior et al. (2017). All SS treatments were complemented with mineral K (KCl) and B (ulexite).

Soil Sampling and Preparation

Soil samples were collected in September 2016, 15 months after SS application (22 months after planting and first application of mineral fertilizer), at the depths of 0–5, 5–10, and 10–20 cm,

TABLE 3 Amount of nutrients applied by treatments with mineral fertilizers and
sewage sludge.

Treatment	Ν	Ρ	к	в	Zn	Cu
		kg ha ^{−1}				
Control (C)	0	0	0	0	0	0
Mineral fertilization (MF)	192	26	137	6.5	2.8	2.8
Planting	30	26	25	1.5	1.5	1.5
Fertilization 1 ³	45	0	112	1.8	1.3	1.3
Fertilization 2 ⁴	54	0	0	1.5	0	0
Fertilization 3 ⁵	63	0	0	1.7	0	0
Sludge (50%) ² + P (83%) + B + K (S1P1)	96 ⁶ (319)	22 (435)	137 (290)	6.5 (0.015)	0 (9.7)	0 (4.5)
Planting ¹	0	22	25	3.2	0	0
Fertilization 1 ³	0	0	112	3.3	0	0
Sludge (100%) ² -	192 ⁶	0	137	6.5	0 (19.3)	0
P + B + K (S2)	(638)	(870)	(855)	(0.03)		(9)
Planting ¹	0	0	25	3.2	19.3	9
Fertilization1 ³	0	0	112	3.3	0	0
Sludge (100%) ² + P (66%) + B + K (S2P2)	192 ⁶ (638)	17.5 (870)	137 (855)	6.5 (0.03)	0 (19.3)	0 (9)
Planting ¹	0	17.5	25	3.2	0	0
Fertilization 1 ³	0	0	112	3.3	0	0
Sludge (150%) ² -	288 ⁶	0	137	6.5	0	0
P + B + K (S3)	(957)	(1,305)	(870)	(0.045)	(29)	(13.5)
Planting ¹	0	0	25	3.2	0	0
Fertilization 1 ³	0	0	112	3.3	0	0

Values in parentheses represent the total amount of nutrients supplied by sewage sludge. ¹Planting fertilization (triple superphosphate, KCl, and ulexite) in 02/12/2015. ²Sewage sludge application in 09/12/2015, 207 days after transplantation, in doses of 14.5, 29, and 43.5 Mg ha⁻¹ (dry base) corresponding to 50, 100, and 150% of N recommendation (192 kg ha⁻¹). ³First fertilization, performed in 11/05/2015, 261 days after transplantation. ⁴Second fertilization, performed in 12/21/2015, 307 days after transplantation. ⁶ Bold values represent the total amount of nutrients applied by mineral fertilizers, except for N values that show the dose of available nitrogen applied by sewage sludge, considering the N mineralization rate of 30% for aerobic SS (Brasil, 2006).

in order to assess differences on P pools, considering superficial SS application, and that labile inorganic P is rapidly adsorbed and organic P can move along 0–20 soil depth. Each soil sample consisted of 18 subsamples collected per plot, nine in the planting rows, where mineral fertilizers were applied, and nine in the interrow, where SS was applied, after carefully removing superficial fallen organic material. At the laboratory, samples were air-dried and sieved at 1.0 mm prior to P analysis. To avoid biomass degradation, soon as the 0–5 cm depth samples were taken, they were homogenized and a fresh 1/4 subsample of each sample was collected, transferred to a plastic flask, and stored in a polystyrene thermal box with ice until transported to laboratory, where the subsamples were stored in a freezer until P microbial analysis.

Wood Production Assessment

To estimate the volume of wood and biomass of wood, barks, branches, and leaves, the height of 12 trees and circumference at breast height (CBH) of all trees in the inner plot were measured 22 months after planting. The height of all trees was estimated by the regression model proposed by Curtis (1967) and modified by Ribeiro et al. (2010), using the equation $\ln H = 2.87875$ -4.28769***/CBH; $R^2_{aj.} = 0.55$; n = 431. After that, Schumacher and Hall's logarithmic model (Schumacher and Hall, 1933) was chosen to adjust the data and estimate tree biomass at the age of 22 months (**Supplementary Data**).

Phosphorus Analysis

Phosphorus Fractionation in Sewage Sludge

Phosphorus fractionation in the SS was performed according to the method proposed by Hedley et al. (1982) and modified by Sui et al. (1999). The SS sample was air-dried and sieved to <2 mm, and then portions of 0.5 g in triplicates were placed in 50-ml centrifuge tubes and sequentially extracted with 30 ml of deionized water (water-soluble inorganic P; fraction 1, F1); 0.5 mol L^{-1} NaHCO₃ (both inorganic and organic extracted P; fraction 2, F2); 0.1 mol L⁻¹ NaOH (both extracted P_i and P_0 ; fraction 3, F3); 1 mol L⁻¹ HCl (extracted P_i ; fraction 4, F4); and 0.5 mol L^{-1} NaOH (both extracted P_i and P_o ; fraction 5, F5). After the final extraction, residual P (residual P_i ; fraction 6, F6) was extracted with H₂SO₄ + H₂O₂ + MgCl₂ and determined colorimetrically according to Murphy and Riley (1962). From the alkaline extracts, an aliquot was taken to determine total P (P_t) by digestion with ammonium persulfate [(NH₄)₂S₂O₈] and sulfuric acid (H₂SO₄) in an autoclave. Inorganic P was determined according to Dick and Tabatabai (1977). Organic P was determined as the difference between P_t and P_i . After quantification of each fraction, they were grouped according to predicted lability: the labile P included F1 + F2; moderately labile P was composed of F3 + F4; and non-labile P was F5 + F6.

Soil P Availability

Soil available P was estimated by a resin-extraction method with a mixture of Amberlite IRA-400 anion-exchange resin and P was determined using a spectrophotometer colorimetry at 882 nm. Weighted soil samples from the 0–20 cm depth of each plot were used for the analysis (Raij et al., 2001) and interpretation of resin-P availability was referred to Raij et al. (1997).

Microbial Biomass Phosphorus

The Brookes et al. (1982) and Hedley and Stewart (1982) methods were applied. Microbial biomass P is calculated from the difference between the amount of inorganic P extracted by 0.5 mol L^{-1} NaHCO₃ (pH 8.5) from fresh soil fumigated with CHCl₃ and the amount extracted from unfumigated soil. Additionally, a known amount of P was added to the sample to prevent adsorption of released P by soil colloids (Morel et al., 1996).

P Fractionation in Soil

Phosphorus fractionation was performed according to the method proposed by Hedley et al. (1982) and modified by Condron et al. (1985). Portions of 0.5 g in triplicates were placed in 50-ml centrifuge tubes and sequentially extracted with 2 cm² of anion-exchange resin saturated with NaHCO₃ and immersed in distilled water (extracted P_i), 0.5 mol L⁻¹ NaHCO₃ (both

extracted P_i and P_o), 0.1 mol L⁻¹ NaOH (both extracted P_i and P_o), 1 mol L⁻¹ HCl (extracted P_i), and 0.5 mol L⁻¹ NaOH (both extracted P_i and P_o). Assessment of inorganic and organic phosphorus, as well as its lability fractions, was conducted as described for SS analysis. P extracted from the soil was evaluated as orthophosphate with a spectrometry (colorimetry) method.

Plant Samples

Plant samples were washed with distilled water and then dried in a forced air oven at 60°C until constant weight is obtained. Samples were weighed for moisture determination and digested by an open system nitric-perchloric digestion. P concentration in the digested solution was determined in a spectrophotometer (colorimeter) with wavelength at 420 nm (Malavolta et al., 1997).

Statistical Analysis

Statistics were carried out using the software program R (R Core Team, 2013). Differences among treatments were compared using a one-way analysis of variance by the *F* test on each measured variable and, when a significant effect was observed, multiple comparisons were assessed by a Duncan's *post hoc* test (p < 0.05). When reported, values in this study indicate the mean \pm standard error. Single correlations were also performed.

RESULTS

Phosphorus Fractionation in the Sewage Sludge

The fractionation of P in SS showed that 0.5% of total P (30 g kg⁻¹, **Table 2**) was extractable to water; 14% to 0.5 mol L⁻¹ NaHCO₃; 13% to 0.1 mol L⁻¹ NaOH; 14.5% to 1 mol L⁻¹ HCl; and 57% to 0.5 mol L⁻¹ NaOH; and that 1% represented the residual P (**Figure 1**). The SS contained 15% of P in labile form (4.5 g kg⁻¹); 27% in moderately labile form (8.1 g kg⁻¹); and 58% as non-labile (17.4 g kg⁻¹). The SS had predominant organic P forms (65% of total P), which, in turn, represents 40, 42, and 82%, respectively, of P contained in labile, moderately labile, and non-labile P.

Soil P Lability

In 15 months after application, SS treatments at the dose of 100 and 150% of N criterion, with or without mineral P (S2, S2P2, and S3) promoted a higher level of P_i (39–68 mg kg⁻¹) in the 0–20 cm soil depth, as compared to S1P1 treatment (P_i , 28–40 mg kg⁻¹) or control (P_i , 16–20 mg kg⁻¹) (**Figures 2A–C**). In general, labile P_o concentration was similar among the treatments (2–18 mg kg⁻¹).

The effect of SS application, in increasing P_i/P_o ratio in P fraction, was higher for moderately labile P, especially for S3 ($P_i = 84-126 \text{ mg kg}^{-1}$; $P_o = 3-17 \text{ mg kg}^{-1}$), along 0–20 cm soil horizon depth. The S2P2 and S2 treatments promoted higher concentration of moderately labile P ($P_i = 71-92 \text{ mg kg}^{-1}$; $P_o = 4-36 \text{ mg kg}^{-1}$), at the upper 0–10 cm soil horizon depth, than mineral fertilizer, S1P1, and control ($P_i = 25-47 \text{ mg kg}^{-1}$; $P_o = 2-22 \text{ mg kg}^{-1}$) (**Figures 2D**,E); however, these treatments were similar at 10–20 cm soil horizon depth ($P_i = 24-47 \text{ mg kg}^{-1}$; $P_o = 6-23 \text{ mg kg}^{-1}$) (**Figure 2F**).

Non-labile fraction, which contained the most of total P in soils of mineral fertilizer and control treatments, and same P_i/P_o ratio in all treatments (**Figures 2G–I**), was just increased by S3 treatment, at very upper 0–5 soil horizon depth ($P_i = 89 \text{ mg kg}^{-1}$; $P_o = 13 \text{ mg kg}^{-1}$), compared with other treatments ($P_i = 56-73 \text{ mg kg}^{-1}$; $P_o = 7-16 \text{ mg kg}^{-1}$) (**Figure 2G**).

Soil P in the Microbial Biomass and Availability

Sewage sludge increased P in the microbial biomass, at 0–5 cm depth, and soil available P, at 0–20 cm depth, estimated by resin extraction (**Figure 3**), as assessed 15 months after application. With the increasing dose of SS, the concentration of MB-P was raised from 0.20 mg kg⁻¹, at MF, to 0.95 mg kg⁻¹, at S3. Mineral fertilizer application had no effects on MB-P concentration (**Figure 3A**), but S1P1, S2 and S2P2 had intermediate values (0.45–0.67 mg dm⁻³). Soil resin-P concentration was higher for S2P2 application (43 mg dm⁻³), than MF, S1P1, S2 and S3 treatments (20–31 mg dm⁻³), and the control (7 mg dm⁻³) (**Figure 3B**).

Production and P Nutrition in *E. urograndis* Plantation Treated With Sewage Sludge

Twenty-two months after *E. urograndis* planting, the highest production of wood volume ($62 \text{ m}^3 \text{ ha}^{-1}$) and biomass of wood (22.6 Mg ha^{-1}), bark (2.4 Mg ha^{-1}), branches (7.0 Mg ha^{-1}), and leaves (4.0 Mg ha^{-1}) were achieved by mineral fertilizer application (**Table 4**). However, SS application at the recommended dose (29 Mg ha^{-1}) supplemented with triple superphosphate resulted in similar results, with wood volume of $60.5 \text{ m}^3 \text{ ha}^{-1}$ and biomass of wood, bark, branches, and leaves of $22, 2.4, 6.8, \text{ and } 3.8 \text{ Mg ha}^{-1}$, respectively. Both treatments had a higher production than the control.

Mineral fertilizer combined with SS application increased P concentration and accumulation in branches and total P accumulation in *Eucalyptus* trees (**Table 5**). However, there were no differences among the treatments for P concentration and accumulation in wood, bark, and leaves. For the branches, there was a 60% higher P concentration in the MF treatment relative to the control, remaining equal to the S2P2 treatment. The compartments with higher P concentration followed the order: leaves > bark > branches > wood. There was greater P accumulation in branches due to mineral fertilizer application, with an average of 4.5 kg ha⁻¹, being 71% higher than the control and statistically equal to the SS treatments. The accumulation of P in the tree components followed the order wood > leaves > branches > bark.

Production of *E. urograndis* Plantation in Relation to Soil P Availability and as a Function of SS Application

Production of wood volume and biomass of wood, bark, branches, and leaves were correlated (p < 0.01) with soil P availability, expressed in terms of soil resin-P, in the control and SS application treatments (**Figure 4**).



Based on wood production of $62 \text{ m}^3 \text{ ha}^{-1}$ for *E. urograndis* after 22 months of planting under mineral fertilizer application, corresponding to 100% of the expected relative production (**Table 6**), an equivalent relative wood production given by the application of SS at recommend dose supplement with 66% of mineral P recommend (S2P2 treatment) can be observed. Consequently, this treatment reduces the use of mineral P and

DISCUSSION

Phosphorus Availability in the Sewage Sludge and Soil

N fertilizer by 33 and 100%, respectively.

Our results showed that the class B sewage sludge had 65% of total P as organic compounds, which is slightly higher than previously reported (Carvalho et al., 2015; Kahiluoto et al., 2015; He et al., 2016). The SS also had 15 and 27% of total P in labile and moderately labile compounds, which agrees with literature (Kahiluoto et al., 2015; He et al., 2016). Such properties justify the forest and agricultural use of SS as a potential source of P fertilizer. For a *Eucalyptus* plantation, the application of the recommend dose by the N criteria (29 Mg ha⁻¹, dry base) of SS, with 30 g kg⁻¹ of total P (**Table 2**), was expected to add 566 kg ha⁻¹ of organic P, including 50 kg ha⁻¹ of labile organic P, plus 75 kg ha⁻¹ of labile inorganic P, including 4.35 kg ha⁻¹ of water-soluble P (equivalent to that supplied by a soluble mineral source of P, e.g., triple superphosphate), while suppling 192 kg ha^{-1} of N, equivalent to the recommend mineral N dose. Labile and moderately labile P are mainly responsible for the shortand medium-term soil P supply, since SS contained both P_i

and $P_{\rm o}$ fractions and the mineralization of $P_{\rm o}$ in sludge after application is governed by a combination of physicochemical and biological processes in soil (Condron and Newman, 2011; Weihrauch and Opp, 2018).

Non-labile P fraction in SS might play an important role in a long-term release of P in the soil, since it has a higher proportion of P_0 (F5 P_0 fraction, Figure 1A) than P_i (F5 P_i fraction, **Figure 1A**). Higher P_o in this fraction should be attributed to organic phospholipids, innermost adsorbed P, and other organic contaminants derived from the organic urban sludge and from runoff infiltration/derivation and few illegal sewage pipes. Inorganic P adsorbed to Fe and Al could be enhanced by the chemical treatments in the sewage sludge using metal salts (e.g., FeCl₃) (García-Albacete et al., 2012), but not by treatment with synthetic polymer (Carvalho et al., 2015). These treatments can modify the proportions of P-Fe and P-Al of the SS and subsequently in the soil (Alleoni et al., 2012). In the present study, the SS was treated with synthetic cationic polymer, and most of Al₂O₃ and Fe₂O₃/FeO were derived from soil particles (kaolinite and oxides) from runoff infiltration/derivation and few illegal sewage pipes, so the F5 P_i fraction was low (Figure 1A). The non-labile P_0 in the SS is likely an important source for the formation of legacy P in soil, due to long-term mineralization (Munhoz et al., 2011; Alleoni et al., 2014; Foltran et al., 2019).

Considering P lability in highly weathered tropical soil untreated with organic fertilizer, there is a lower level of labile P compared with moderately labile and non-labile fractions, caused by the high P adsorption capacity (Alleoni et al., 2014; Lopes and Guilherme, 2016; Withers et al., 2018). The SS application changed this pattern and improved soil P lability (**Figure 2**). Fifteen months after application, SS at recommended dose



Duncan test ($p \le 0.05$). Vertical lines on bars represent standard error (n = 4).

with or without soluble triple superphosphate P (S2 and S2P2 treatments), and SS at 150% of dose (S3 treatment) positively altered the pattern of soil P distribution, with more P in the labile and moderately labile fractions than non-labile fraction, at 0–5, 5–10, and 10–20 cm horizon depths, compared with mineral fertilizer, with soluble triple superphosphate P.

For short to medium term, 15 months after SS application, there were higher concentrations of P in labile and moderately

labile pools in S2, S2P2, and S3 treatments, mostly in inorganic P forms, up to 20 cm depth (**Figure 2**), explained with a very high positive correlation (r = 0.97, p < 0.01) with amount of SS-P applied to soil in organic forms, corresponding to 566 and 848 kg ha⁻¹ of organic P. Application of organic matter to highly weathered soils generally decreases P adsorption and increases P lability (Novais and Smyth, 1999; Laboski and Lamb, 2003), as some absorbing sites are occupied by organic ligands with



FIGURE 3 Phosphorus in the microbial biomass (A), at 0–5 cm horizon depth, and P availability, at 0–20 cm horizon depth, assessed by resin (B) in soil, 15 months after sewage sludge application (SS was applied 7 months after *E. urograndis* planting). C, control; MF, mineral fertilizer; S1P1, sludge (50%) + P (83%) + B + K; S2, sludge (100%) - P + B + K; S2P2, sludge (100%) + P (66%) + B + K; S3, sludge (150%) - P + B + K. Bars followed by the same letter do not differ by Duncan test ($p \le 0.05$). Vertical lines on bars represent standard error (n = 4).

TABLE 4 Production of 22-month-old *E. urograndis* plantation in a soil treated with class B sewage sludge 7 months after planting (mean ± standard error, n = 4).

Treatment	Volume (m ³ ha ⁻¹)	Wood (Mg ha ⁻¹)	Bark (Mg ha ⁻¹)	Branch (Mg ha ⁻¹)	Leaves (Mg ha ⁻¹)	Total (Mg ha ⁻¹)
С	55.7 ± 1.2 c	$20.1 \pm 0.5 \text{ c}$	$2.2\pm0.05~\mathrm{c}$	$6.2 \pm 0.15 \text{ c}$	$3.5\pm0.09~\mathrm{c}$	$32.0 \pm 0.7 \text{ c}$
MF	$62.0 \pm 1.1 a$	$22.6 \pm 0.4 a$	$2.4 \pm 0.05 a$	7.0 ± 0.15 a	$4.0 \pm 0.09 \text{ a}$	$36.0 \pm 0.7 \text{ a}$
S1P1	$57.9 \pm 1.8 \ {\rm bc}$	$21.0\pm0.7~\mathrm{bc}$	$2.3\pm0.08~{ m bc}$	6.5 ± 0.23 bc	$3.6\pm0.14~\mathrm{bc}$	$33.4 \pm 1.1 \ {\rm bc}$
S2	$58.2 \pm 0.7 \ \rm bc$	21.1 ± 0.3 bc	$2.3\pm0.03~{ m bc}$	$6.5\pm0.09~{ m bc}$	$3.7\pm0.05~{ m bc}$	$33.6\pm0.4~\mathrm{bc}$
S2P2	$60.5 \pm 1.0 \text{ ab}$	$22.0 \pm 0.4 \text{ ab}$	$2.4 \pm 0.04 \text{ ab}$	$6.8\pm0.13~\text{ab}$	$3.8\pm0.08~\text{ab}$	$35.0 \pm 0.6 \text{ ab}$
S3	$58.0 \pm 0.9 \ \rm bc$	21.0 ± 0.4 bc	2.3 ± 0.04 bc	$6.5 \pm 0.12 \text{ bc}$	$3.7 \pm 0.07 \text{ bc}$	$33.5 \pm 0.6 \text{bc}$
CV (%)	3.6	3.9	3.9	4.2	4.3	4.3

C, control; MF, mineral fertilizer; S1P1, sludge (50%) + P (83%) + B + K; S2, sludge (100%) - P + B + K; S2P2, sludge (100%) + P (66%) + B + K; S3, sludge (150%) - P + B + K. Mean values followed by different letters, in the column, are different by Duncan test ($p \le 0.05$).

high binding energy. In addition, the organic compounds are rapidly mineralized, releasing more P_i to soil solution (Novais et al., 2007). Thus, when SS is applied, there was a greater contribution to the increment of soil P_i in labile and moderately labile pools (**Figure 2**) when there was a greater labile plus moderately labile P input.

Indeed, 15 months after application, the greater contribution of SS application, except at the lower dose (S1P1), was the buildup of a soil moderately labile P pool (**Figures 2D–F**), which has an important nutritional and ecological advantage (Rodrigues et al., 2016; Withers et al., 2018). However, the greater or lesser contribution to soil available P is dependent on the characteristics of soil and applied organic material, and its dose (Abreu-Junior et al., 2005; García-Albacete et al., 2012; Borges et al., 2019).

The microbial community is responsible for the conversion of organic to its inorganic forms, by solubilization and mineralization processes (Shi et al., 2019), so microbial P concentration would be an indicator of short- to long-term potential of soil in maintaining P availability to plant nutrition. There was also a close relation between organic P and soil microbial P concentration (r = 0.89, p < 0.01), as SS dose increased, with no mineral fertilizer effect, decreasing the following order of treatments: S3 > S2P2 = S2 > S1P1 > MF = C(Figure 3A). Rheinheimer et al. (2000) reported a flow from 7 to 26 kg ha^{-1} per year of P in no-tillage system on three different soils of the tropical region, showing the importance of soil microbiota for P availability to plants and cycling. Also, some microorganisms, already present in the soil and/or supplied by SS, have the ability to solubilize P that is chemically bonded to Fe and Al oxides, playing a fundamental role in soil P maintenance and availability, particularly in the tropical and degraded soils (Wu et al., 2007). Despite the great contribution of SS to soil microbial activity, microbial biomass P concentrations were lower than those reported in the literature, for clays and organic soils of temperate regions (Rheinheimer et al., 2000; Wu et al., 2007; Sugito et al., 2010; Shi et al., 2019). So, the lower MB-P values in our study may be explained by field conditions, with soil sampling at the 0-5 cm depth of a superficial sand

TABLE 5 | Concentration and accumulation of P in 22-month-old E. urograndis plantation in a soil treated with class B sewage sludge 7 months after planting (mean \pm standard error, n = 4).

P concentration (g kg ⁻¹)						
Treatment	Wood	Bark	Branches	Leaves		
С	0.36 ± 0.009 a	0.81 ± 0.03 a	$0.40 \pm 0.02 \mathrm{c}$	1.47 ± 0.05 a		
MF	0.33 ± 0.015 a	$0.83 \pm 0.04 \text{ a}$	$0.64 \pm 0.06 \text{ a}$	1.38 ± 0.14 a		
S1P1	0.39 ± 0.011 a	$0.78 \pm 0.02 \text{ a}$	0.60 ± 0.05 a	$1.41 \pm 0.06 \mathrm{a}$		
S2	0.37 ± 0.007 a	0.77 ± 0.03 a	$0.57 \pm 0.05 \text{ ab}$	$1.46 \pm 0.10 \ {\rm a}$		
S2P2	0.35 ± 0.015 a	0.82 ± 0.03 a	$0.44\pm0.05~{ m bc}$	$1.43 \pm 0.03 \ {\rm a}$		
S3	0.36 ± 0.013 a	0.76 ± 0.03 a	$0.56 \pm 0.03 \text{ ab}$	$1.54 \pm 0.11 \mathrm{~a}$		
CV (%)	7.3	7.7	17.0	5.1		

P accumulation (kg ha⁻¹)

Treatment	Wood	Bark	Branches	Leaves	Total
Control	7.2 ± 0.15 a	1.8 ± 0.03 a	$2.5 \pm 0.17 \text{ c}$	5.1 ± 0.16 a	16.6 ± 0.43 b
MF	7.4 ± 0.30 a	$2.0 \pm 0.08 \text{ a}$	$4.5 \pm 0.46 \text{ a}$	5.4 ± 0.28 a	19.3 ± 0.39 a
S1P1	8.1 ± 0.10 a	1.8 ± 0.09 a	3.8 ± 0.19 ab	5.1 ± 0.21 a	18.9 ± 0.18 a
S2	7.8 ± 0.19 a	1.8 ± 0.07 a	3.7 ± 0.35 ab	5.4 ± 0.23 a	18.7 ± 0.66 a
S2P2	7.7 ± 0.21 a	1.9 ± 0.09 a	3.0 ± 0.35 bc	5.5 ± 0.07 a	18.3 ± 0.31 a
S3	7.5 ± 0.35 a	1.7 ± 0.05 a	3.7 ± 0.25 ab	5.6 ± 0.21 a	18.5 ± 0.64 a
CV (%)	6.9	8.2	18.1	6.4	5.4

C, control; MF, mineral fertilizer; S1P1, sludge (50%) + P (83%) + B + K; S2, sludge (100%) - P + B + K; S2P2, sludge (100%) + P (66%) + B + K; S3, sludge (150%) -P + B + K. Mean values followed by different letters, in the column, are different by Duncan test ($p \le 0.05$).



(resin-P), 15 months after sewage sludge application (SS was applied 7 months after Eucalyptus planting) (n = 5). C, control; MF, mineral fertilizer; S1P1, sludge (50%) + P (83%) + B + K; S2, sludge (100%) - P + B + K; S2P2, sludge (100%) + P (66%) + B + K; S3, sludge (150%) - P + B + K. **p < 0.01.

horizon (90% of sand, Table 1), in a very dry season, in September 2016, following 2-3 months with no rain. The influence of seasonality was observed by Haripal and Sahoo (2014), where the rainy season had higher soil P-MB values than the dry season, due to drought stress. This also might have resulted from the pretreatment of our soil samples before analysis (freezing rather than cold storage).

Concerning soil P availability, after 15 months of application, all treatments with SS had resin-P at a high concentration (>16 mg dm⁻³) (**Figure 3B**) for forest plantation (Raij et al., 1997), mainly because of the mineralization of exogenous organic P added to soil by SS (Condron and Newman, 2011; Weihrauch and Opp, 2018), while the control still kept resin-P at an average level (from 6 to 8 mg dm^{-3}), due to **TABLE 6** | Doses of nitrogen, phosphate, potassium, and boron mineral fertilizer applications for *E. urograndis* planting, as a function of the sewage sludge doses, based on the N criterion.

Dose of SS application ¹ (Mg ha ⁻¹)				Estimated relative wood volume productivity ² (%)	
	Ν	Р	к	в	
0	192	26	137	6.5	100 a
29	0	17.5	137	6.5	98 ab
29	0	0	137	6.5	94 bc
43.5	0	0	137	6.5	94 bc
14.5	0	22	137	6.5	93 bc
0	0	0	0	0	90 c

¹The dose of sewage sludge (dry base) based on the N criterion (Brasil, 2006). ²Relative production of 100% is equivalent to a wood volume estimated at 62 m³ ha⁻¹ for E. urograndis grown for 22 months under minimum tillage and conventional application of lime and NPK fertilizers, at the recommended dose. Relative productivity means followed by different letters are different by Duncan test ($p \le 0.05$).

P-litter recycling (Vargas et al., 2019) and low P adsorption capacity attributed to low clay content in the soil (**Table 1**). However, the application of SS, at the recommend dose by N criterion complemented with 66% of the triple superphosphate P recommended dose (S2P2 treatment) had the best improvement on soil P availability. This was attributed to the SS organic P mineralization and the positive effect of the interaction between the organic and mineral sources of P on soil P availability, higher than when they are applied alone (Munhoz et al., 2011; Borges et al., 2019), besides lowering P adsorption in soil as Al and Fe are complexed by exogenous organic matter from the SS (Kahiluoto et al., 2015).

Sludge applied at the 50% of recommended dose supplemented with 83% of triple superphosphate P recommended dose (S1P1 treatment), at the recommend dose (S2 treatments), and at 150% of the recommend dose (S3 treatment) were as effective as mineral fertilization (after 22 months of application, at *Eucalyptus* planting) in increasing soil P availability (Figure 3B). This effect of mineral fertilizer is also explained by the soil low P adsorption capacity at the superficial sand mineral horizon. There was higher P availability as indicated by a larger amount of P-resin in the soil after the application of phosphate fertilizer in sandy soils as compared to the highly weathered clay soils (Sugihara et al., 2016; Martins et al., 2018). SS application at the recommended dose (29 Mg ha^{-1}) added 75 kg ha^{-1} of labile P_i to the soil, corresponding to 75 mg dm⁻³ of labile P_i , at the 0-10 cm soil depth (Figures 2A,B). Moreover, the SS P_0 could be slowly mineralized to contribute to the labile P pool (Figure 2), thus reducing P fixation by adsorption (Novais and Smyth, 1999; Laboski and Lamb, 2003) and increasing soil P_i availability (Figure 3B) even in 15 months after SS application.

As most of soil inorganic P in the SS treatments was in the moderately labile fraction, with higher P_i/P_o ratio (**Figures 2D-F**), it is expected that there will be a continuous long-term release of available inorganic P into the soil solution and taken up by plants, thus reducing the risk of loss by transforming to non-labile forms. This result was corroborated by Florentino et al. (2019) who reported high soil resin-P concentration even after more than 17 years of sewage sludge application in four fields of *Eucalyptus* plantations, as compared to the control and mineral fertilization treatments. These findings reinforce the importance of soil inorganic moderately labile P fraction for the long-term effect of SS application on legacy P.

Our results (Figures 1-3) and those obtained by Florentino et al. (2019) provide evidences that labile, moderately labile, and organic non-labile P in the SS are important sources for short- to long-term soil P_i supply and building up legacy P in soil. Considering that soil P availability is estimated by resin extractable P and the length of *Eucalyptus* rotation, of 6 years, it is reasonable to suggest the following soil P dynamic after SS application: (1) labile SS P continuously supplied the short-term available P to plants (e.g., in the 12 months after SS application); (2) moderately labile SS P fraction maintained the labile plus moderate labile soil P pools, as the medium-term available P to plants (e.g., from 12 to 36 months of SS application); and (3) nonlabile SS P compounds, specially organics through the biological soil cycles, might supply the non-labile, moderately labile, and labile soil P pools as the long-term plant available P sources (e.g., after 36 months of SS application), in other words, the legacy P in soil. Such P dynamics for tropical soils treated with SS agrees with the reports by Munhoz et al. (2011); Alleoni et al. (2014), and Foltran et al. (2019).

Production and P Concentration and Accumulation in *E. urograndis* Plantation as a Function of Sewage Sludge Application

In the present study, for a 22-month-old *E. urograndis* trees, higher production of wood volume and biomass was achieved by application of (**Table 4**) (i) SS at the recommended dose supplemented with mineral P fertilizer (S2P2; 60.5 m³ ha⁻¹) and (ii) mineral fertilizers for high yield (MF; 62 m³ ha⁻¹). This is in accordance with a 23-month-old *E. grandis* with a wood volume production of 62 m³ ha⁻¹, when SS was applied according to N criterion and supplemented with 50% of P recommendation (Abreu-Junior et al., 2017). However, in our study, 56 m³ ha⁻¹ of wood was achieved by the control treatment (**Table 4**), while, in the 23-month-old *E. grandis* experiment (Abreu-Junior et al., 2017), 30 m³ ha⁻¹ of wood was recorded by the control, corresponding to 90 and 48% of the maximum production observed (**Table 6**).

The low growth difference of *E. urograndis* up to 22 months between SS and mineral fertilizer treatments with the control (**Tables 4, 6**) may have been caused by two reasons, among the various biotic and abiotic factors that influence plant development (Chen et al., 2011). The first is related to the soil resin-P concentration, which was 7 mg dm⁻³ at the control, indicating a medium level of P (6–8 mg dm⁻³) for *Eucalyptus* growth (Raij et al., 1997). Thus, contributing to medium response of *E. urograndis* to the fertilization. Ferraz et al. (2016) observed a 75% lower wood volume

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production at the control (without fertilization) in relation to SS treatment. In this case, soil available P concentration was 2 mg dm⁻³, corresponding to low level of P (0-2 mg dm^{-3}) in the surface mineral horizon (0-5 cm) for Eucalyptus growth, thus contributing to high response of E. grandis to the fertilization. The second point is related to the interaction of E. urograndis genotype with precipitation (soil moisture) and nutrient availability in the soil environment. Precipitation during the first and second year after plantation was 1,442 and 926 mm, respectively; they were mainly concentrated between November and March, providing low soil water availability. This may lead to reduced absorption and nutrient efficiency as well as to issues in light use by plants, thus limiting the full expression of potential production of *Eucalyptus* plantation (Stape et al., 2008; Elli et al., 2019). Water stress may lead to physiological and metabolic disorders in the Eucalyptus plants, thus compromising its development (Tariq et al., 2019). The planted material (clone SP5727 owned by Suzano, E. urograndis) was genetically selected for wood production on dry and infertile soil. Consequently, this clone has low fertilizer response under dry periods, differently from other *Eucalyptus* genotypes that have higher dry matter on roots instead of canopy when growing in dry soil, especially with no P fertilization (Tariq et al., 2019).

According to Abreu-Junior et al. (2017), as time goes by and SS mineralization continuously releases N and P to the soil, the application of SS on 100% of the N supplemented with 66% of the P had an increase of 7% in the relative wood volume production for 44-month-old *E. grandis* trees (178 m³ ha⁻¹), compared to mineral NPK fertilization (166 m³ ha⁻¹). For a 42-month-old *E. urograndis* trees, at the same experiment of this study, application of SS on 100% of the N supplemented with 66% of the P (S2P2; 210 m³ ha⁻¹) had an increase of 5 and 11% in relative wood production compared to MF and control treatments (199 and 188 m³ ha⁻¹), respectively (Mandu, 2020). These indicated a high potential for wood productivity in *Eucalyptus* plantations, growing in infertile tropical soils treated with sewage sludge, with yield comparable with those of mineral fertilization for high production.

The soil fertility, E. urograndis genotype, and precipitation, as explained above, also reflected in the P concentration and accumulation in Eucalyptus tree compartments. It was noticed that SS and MF treatments generally had the same pattern for P concentration and accumulation in the wood, bark, branches, leaves, and total biomass. The P concentration in leaves was higher (mean value of 1.45 g kg^{-1}) than those usually representing the appropriate range (0.9–1.3 g kg⁻¹; Raij et al., 1997). This can be related by both available soil P and by the higher volume of soil explored by deep fine Eucalyptus roots. However, these values refer to mature fully developed plants, with leaves sampled at 22 months of age, when there is still high demand and absorption of nutrients by plants (Laclau et al., 2010; Ferraz et al., 2016; Foltran et al., 2019). As a matter of fact, Crous et al. (2019) observed a reduction in Eucalyptus foliar P concentration with increasing planting age in medium P availability soil.

The most important result, besides the relative low growth rate, is that the sludge application, 7 months after Eucalyptus planting, based on N criterion and complemented with mineral P, brought production gains up of 4.8 m^3 ha⁻¹ of wood volume, 1.9 Mg ha⁻¹ of wood, 0.2 Mg ha⁻¹ of bark, 0.6 Mg ha^{-1} of branches, and 0.3 Mg ha^{-1} of leaves biomass, with a total of 3.0 Mg ha⁻¹ of biomass on a 22-month-old E. urograndis trees (Table 4). Additionally, SS application (i) increased concentration of available P from 7 to 43 mg kg $^{-1}$, in soil surface Ap horizons (0-20 cm; Figure 4), while (ii) reducing by 33 and 100% the use of mineral P and N fertilizer, respectively (Table 6). These results demonstrated that the use of sewage sludge in commercial Eucalyptus plantations can improve soil P availability and wood yield by combining sludge and mineral P fertilizer doses, as compared to conventional NPK fertilization. This outcome strongly reinforced previous results and recommendations proposed by Abreu-Junior et al. (2017).

CONCLUSION

The sewage sludge contained different pools of labile and nonlabile P that can be used to support the short- to long-term soil P availability to plants and to build up a legacy P in soil, when applied at the N criterion dose. Application of SS at the recommended dose with or without triple superphosphate P and at the 150% of recommended dose positively altered the pattern of soil P distribution, with more P in the labile and moderately labile fractions than non-labile fraction, along the 0-20 cm depths (corresponding to the mineral Ap surface horizon), as compared with triple superphosphate application alone. Consequently, there were higher P concentration in the microbial biomass and soil available P pools, indicating that P cycling in the system and microbial community is enhanced by SS application and soil P availability is improved. Production of wood volume and biomass of tree compartments were highly correlated with soil P availability, as a function of SS dose. SS application at the recommend dose and supplemented with 66% of mineral P resulted in the highest gains of volume and biomass of wood, while reducing by 33 and 100% the use of mineral P and N fertilizer, respectively. The application of SS combined with mineral P fertilizer in commercial E. urograndis plantations can increase soil P availability and subsequent wood production, comparable to NPK-micronutrient fertilization for high yield.

DATA AVAILABILITY STATEMENT

The datasets generated for this study are available on request to the corresponding author.

AUTHOR CONTRIBUTIONS

AF, CA-J, FO, and MO: conception and design of the work. AF, MO, and TM: acquisition of data. AF, CA-J, JR, MO, and

PC: analysis and interpretation of data. CA, CA-J, FO, GC, JR, JS, PC, TN, and ZH: drafting and revising the work. All authors contributed to the article and approved the submitted version.

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SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: https://www.frontiersin.org/articles/10.3389/fenvs. 2020.00116/full#supplementary-material

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The remaining authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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