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EDITED BY
Jaime A Collazo,
North Carolina State University,
United States

REVIEWED BY Lucian Pârvulescu, West University of Timișoara, Romania

*CORRESPONDENCE
Hanno Sandvik,

⋈ hanno.sandvik@nina.no

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For quantitative criteria in alien species assessment

Hanno Sandvik^{1,2}*

¹Centre for Biodiversity Dynamics (CBD), Norwegian University of Science and Technology (NTNU), Trondheim, Norway, ²Norwegian Institute for Nature Research (NINA), Trondheim, Norway

Impact or risk assessments of alien species can use qualitative criteria (such as verbally described categories) or quantitative criteria (numerically defined threshold values of empirically measurable quantities). According to a common misconception, the use of qualitative criteria in invasion biology is justified by uncertainty in the available data. Yet qualitative criteria have the effect of *increasing* uncertainty. In contrast, assessments using quantitative criteria are testable, transparent, highly repeatable and comparable. Most of these characteristics do not even depend on the availability of numerical data. Although quantitative criteria do not necessarily make assessments *correct*, they do make them *correctable*, which is the benchmark of science.

KEYWORDS

impact assessment, invasive species, qualitative assessment, quantitative assessment, risk assessment, testability

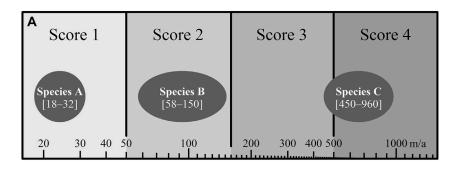
Introduction

Ecological risk assessments can be carried out using quantitative or qualitative criteria. IUCN's (2012) Red List criteria are an example of a quantitative set of criteria. A species may for instance be listed as endangered (EN) or vulnerable (VU), depending on whether its area of occupancy (AOO) is below or above 500 km².

In contrast, many assessment schemes addressing the ecological impact of alien (invasive) species are still based on qualitative sets of criteria (for an overview, see e.g. Roy et al., 2018). However, there is no obvious reason why red-listing of threatened species and impact assessments of alien species should follow different scientific standards, and the need for quantitative sets of criteria for alien species assessments has been stated before (e.g. Lodge et al., 2006; Suedel et al., 2007). I would here like to reinforce this conviction by clarifying the advantages of, and refuting some worries about, quantitative criteria. For example, it has been stated that "the lack of [numerical] data" and "challenges in interpretation and communication" constitute "Major hurdles preventing the use of [...] quantitative risk assessment method[s]" (BNZ, 2006:26; Roy et al., 2018:528). However, such statements confound quantitative *criteria* with the quantitative nature of the *data* used.

When are criteria quantitative?

An assessment criterion is here regarded as *quantitative* if and only if it provides a numerically defined threshold value of an empirically measurable quantity for each assessment category. In the Red List example, according to criterion B2, an AOO of 500 km² is the threshold value distinguishing species in the EN category from species in the VU category (IUCN, 2012). Because such threshold values are defined for all assessment categories (CR, EN, VU, NT, LC) for all assessment criteria (A–E) and sub-criteria, the Red List constitutes a fully quantitative set of criteria.



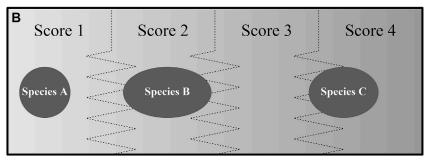


FIGURE 1
Concept figure illustrating three alien species in an (A) quantitative and a (B) qualitative set of criteria. Species are assigned to one of four scores based either (A) on the estimated numerical value of their expansion speed (in square brackets, meters per year) or (B) on expert judgment. The length of the ellipses illustrates epistemic uncertainty (e.g. natural variation and measurement error), whereas semantic uncertainty is illustrated by the (A) sharpness or (B) fuzziness of the thresholds between scores.

A set of criteria that does not meet this definition is here referred to as *qualitative*. This obviously includes cases where the assessment categories are described using a scale of subjective terms such as 'low', 'limited', 'moderate', 'medium', 'high', 'severe', without any translation into numerical thresholds.

Sometimes, 'semi-quantitative' sets of criteria are said to occupy a middle-ground. From the perspective taken here, they do not represent a separate category, however. They are either a mixture of quantitative and qualitative criteria; or they use categories that are *referred* to by numbers (scores) but *defined* verbally.

It is thus important to understand that the use of numerical scores does not make a criterion quantitative: Measurement theory distinguishes between different scale types, for which different operations are admissible (Houle et al., 2011). For instance, an AOO of 300 km² is three times larger than an AOO 100 km², but a score of three does not need to be three times as substantial as a score of 1. Ignoring these differences can lead to erroneous or meaningless inferences (Wolman, 2006).

Conversely, a set of criteria can be quantitative even though it uses verbal descriptors for its categories, as long as the underlying thresholds are numeric. For instance, referring to ecological effects as 'local' or 'large-scale', while defining the threshold between these terms as affecting "less than" or "at least 5% of the population size or AOO", respectively (Sandvik et al., 2019), is fully compatible with a quantitative set of criteria.

Yes/no questions may or may not be quantitative criteria. "Is the species a parasite," may be regarded as quantitative, because the answer can be established empirically. The question whether a species has "prolific seed production" is quantitative only if a numerical definition of 'prolific' is provided (e.g. 1,000 seeds m^{-2} a^{-1} ; Gordon et al., 2010).

Handling uncertainty in quantitative sets of criteria

Risk assessments, as all other scientific activities, are imbued with uncertainty. Uncertainty may be epistemic (e.g. natural variation, observation error, measurement error, estimation error) and semantic (e.g. linguistic vagueness) in its nature (Akçakaya et al., 2000; Regan et al., 2002). Although uncertainty in risk assessments of alien species has received quite some attention (McGeoch et al., 2012; Latombe et al., 2019; Clarke et al., 2021), and even though some of these studies acknowledge that uncertainty is partly due to criterion delimitations being "a matter of degree" (Clarke et al., 2021, p. 13), the nature of the criteria has been conspicuous by its absence from recommendations for reducing uncertainty. However, quantitative criteria take an obvious step towards reducing semantic uncertainty, i.e. towards reducing the scope of potential interpretations that can be given to the questions posed, definitions used, and criteria applied. This is an important gain, even though the epistemic portion of uncertainty is not affected by the nature of the criteria.

This gain is visualized in Figure 1: epistemic uncertainty about the expansion speed (Sandvik, 2020) of three alien species is illustrated using the lengths of ellipses, while semantic uncertainty is symbolized by the borders between categories (scores). In a quantitative set of criteria (Figure 1A), some species will unambiguously fit into a certain category (A and B), while the category may be uncertain for others (C).

The results of the assessment are		The set of criteria is	
		quantitative	qualitative
The data available are	numerical estimates	testable repeatable	not testable not repeatable
	expert judgments	testable not repeatable	not testable not repeatable

FIGURE 2

The nature of the results obtained (repeatable and/or testable), given a certain set of criteria (quantitative or qualitative) and a certain type of data (numerical or expert judgment).

In other words, although uncertainty is present in all three cases, it only affects the classification of species C. If one chooses a qualitative set of criteria instead (Figure 1B), species A may still fit into one score and C will be uncertain; in addition, however, it has become unclear whether species B should be scored as 2 or 3 (or even 1). As can be seen from Figure 1A, this uncertainty is not due to poor data; it is due to fuzzy thresholds, i.e., scores allowing different interpretations. Some experts may even argue, based on Figure 1B, that C fits perfectly into score 4. In the latter case, qualitative criteria, by increasing semantic uncertainty, result in an *underestimate* of the epistemic uncertainty.

Combining quantitative criteria with expert judgment

A quantitative set of criteria is at its best when it can be combined with quantitative data, i.e. when numerical estimates are available for the parameters underlying the criteria. However, this is a not a prerequisite for using quantitative criteria. If good data are absent, estimates have to be replaced by 'expert judgments', based on the personal scientific expertise of specialists. As expert judgments are often associated with qualitative criteria, it is important to point out that expert judgments are fully compatible with a quantitative set of criteria, provided that the judgments are based on the threshold values. This means that the experts do not need to produce a numerical point estimate; it suffices that they substantiate that the true parameter value is likely to lie between two specified threshold values.

For example, an expert judgment is based on the threshold values of Figure 1A when the expert assesses an alien species to have an expansion speed between 50 and 160 m per year, even if no specific expansion speed (e.g. 144 m a⁻¹) has been estimated. Although this judgment is subjective and may turn out to be erroneous, it has an important advantage over a qualitative one: *the result is testable*. The search for more knowledge can thus concentrate on the question: are there any empirical findings suggesting that the species may have an expansion speed of less than 50 or more than 160 m per year?

It is comparatively easy to envisage quantitative criteria for aspects of invasiveness, such as expansion speed or likelihood of establishment (Sandvik et al., 2019). However, even ecological effects, such as interactions with native species, can be quantified if expressed e.g. in terms of the reduction in population growth rate, carrying capacity or AOO (Laska and Wootton, 1998). It is true that such quantification

requires extensive long-term study (Doak et al., 2008; Novak and Wootton, 2008) and is thus difficult in practice. However, quantitative criteria do not presuppose that numerical data are actually available, but only that it is possible *in principle* to obtain such data and that the effect *criteria*, therefore, can be defined numerically. For example, defining a 'local impact' as "affecting less than 5% of the population size" provides the expert with a clear guideline for her judgment. For many species, it will be unknown whether more or less than 5% of the population are actually affected. However, the precisely defined threshold permits an unambiguous understanding of the criterion in question.

Figure 2 shows the nature of the results obtained, given a certain set of criteria and a certain type of data. A quantitative set of criteria does *not* guarantee that the results are *repeatable* (for this to be the case, the additional requirement of numerical data needs to be fulfilled). However, it does guarantee that the results are *testable*. A qualitative set of criteria can neither guarantee repeatability nor testability—not even when numerical data are available. In combination with numerical data, the quantitative set of criteria even allows a quantification of the uncertainty (see Figure 1A), an option that does not exist in the qualitative alternative.

Preferring a qualitative set of criteria because numerical estimates are not always available, essentially amounts to maintaining (in terms of Figure 2): "When results can have low repeatability, they do not need to be testable either." Which is equivalent to stating (in terms of Figure 1): "When the epistemic uncertainty is large, we can just as well increase the semantic uncertainty." However, there is no conceivable reason to accept such reasoning. All means of reducing uncertainty ought to be welcome.

Advantages of quantitative criteria

The advantages of quantitative over qualitative sets of criteria may be summarized as increasing testability, repeatability, comparability, and transparency.

- 1) Testability.—An assessment is testable if there are objective criteria for deciding whether it is compatible with or contradicted by empirical findings. If, for example, a species has been classified as endangered based on its AOO, it is obvious what kind of information would be required to change this conclusion: if novel findings indicate that the AOO is actually larger than 500 km², the previous assessment has been refuted. In contrast, if the AOO of a species is qualitatively assessed to be 'small', it is far from obvious what kind of information would be needed to change or refute this supposition.
- 2) Repeatability.—An assessment is repeatable if different experts would have ended up with the same result, given that the same information was available to them. For example, different experts will agree that a species should be classified as endangered, given that the best estimate of its AOO is 100 km². In contrast, different experts will not necessarily agree whether an AOO of 100 km² qualifies as 'limited' or 'moderate', simply because different experts can have different understandings of such terms (Burgman, 2001). To be sure, repeatability is a matter of degree, since no quantitative method will ever reach a repeatability of strictly 100%, whereas methods are available for improving the repeatability of qualitative assessments (Burgman, 2001; McBride et al., 2012).

- 3) Comparability.—The interpretation of qualitative criteria is always affected by the reference framework of the experts carrying out the assessments (Tversky and Kahneman, 1974; Martin et al., 2012). What an expert means by saying that the expansion speed of an alien species is 'moderate', may for instance depend on whether she is an expert on birds or on snails. Methods to increase repeatability do not necessarily increase comparability, because reaching a consensus among e.g. bird experts is one thing (intersubjectivity), whereas reaching consensus across bird and snail experts would be a rather different matter (objectivity).
- 4) Transparency.—The result of a quantitative assessment is transparent for all scientists and end users. In other words, it is obvious how the scores of each criterion were derived from the underlying data, and how the final category is derived from the scores. Likewise, it is possible to test whether and how specific assumptions (e.g. regarding detection rate; Christy et al., 2010; Sandvik, 2020) affect the assessment.

Of course, qualitative criteria are not necessarily characterized by the total *absence* of testability, repeatability, comparability, and transparency. The crucial point is that, other things being equal, quantitative criteria are *more* testable, repeatable, comparable, and transparent than qualitative ones. This plea for quantitative risk assessments may be summarized in two propositions.

- The advantage of a quantitative over a qualitative set of criteria is not that the risk assessment becomes *certain*, but that, by reducing one specific source of uncertainty, it becomes *testable*.
- The advantage of a quantitative over a qualitative set of criteria is not necessarily that the risk assessment becomes *more correct*, but that it becomes *more correctable*.

Discussion

Testability is the benchmark of science (Popper, 1934). In conservation biology, Mace and Lande (1991) introduced testability to Red List assessments. It is about time to apply the same scientific standards to impact assessments of alien species. This does not necessarily require large changes. For example, EICAT's (Hawkins et al., 2015) and SEICAT's (Bacher et al., 2018) criteria might be turned into quantitative ones by providing numerical definitions of key terms such as 'decline in population size' (e.g. "a reduction of at least 15% in the population size of at least 1 native species over a 10-year period"; cf. Sandvik et al., 2019, p. 2806) or 'irreversibility' (e.g. a reversion requiring more than x US\$ or y years). Several fully (Sandvik et al., 2019) and partially (e.g. Baker et al., 2005; Gordon et al., 2010;

References

Akçakaya, H. R., Ferson, S., Burgman, M. A., Keith, D. A., Mace, G. M., and Todd, C. R. (2000). Making consistent IUCN classifications under uncertainty. *Conserv. Biol.* 14, 1001–1013. doi:10.1046/j.1523-1739.2000.99125.x

Bacher, S., Blackburn, T. M., Essl, F., Genovesi, P., Heikkilä, J., Jeschke, J. M., et al. (2018). Socio-economic impact classification of alien taxa (SEICAT). *Methods Ecol. Evol.* 9, 159–168. doi:10.1111/2041-210X.12844

Baker, R., Hulme, P., Copp, G. H., Thomas, M., Black, R., and Haysom, K. (2005). *UK non-native organism risk assessment scheme user manual: Version 3.3.* York: Great Britain Non-native Species Secretariat.

D'hont et al., 2015; USDA, 2019) quantitative sets of criteria are already available, and it is to be hoped for that other assessment schemes will follow by defining numerical threshold values for their criteria

Data availability statement

The original contributions presented in the study are included in the article/supplementary material, further inquiries can be directed to the corresponding author.

Author contributions

HS conceived of and wrote the article.

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Conflict of interest

The author declares that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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BNZ (2006). Risk analysis procedures. Wellington: Ministry of Agriculture and Forestry.

Burgman, M. A. (2001). Flaws in subjective assessments of ecological risks and means for correcting them. *Aust. J. Environ. Manag.* 8, 219–226. doi:10.1080/14486563.2001.

Christy, M. T., Yackel Adams, A. A., Rodda, G. H., Savidge, J. A., and Tyrrell, C. L. (2010). Modelling detection probabilities to evaluate management and control tools for an invasive species. *J. Appl. Ecol.* 47, 106–113. doi:10.1111/j.1365-2664.2009. 01753.x

Clarke, D. A., Palmer, D. J., McGrannachan, C., Burgess, T. I., Chown, S. L., Clarke, R. H., et al. (2021). Options for reducing uncertainty in impact classification for alien species. *Ecosphere* 12, e03461. doi:10.1002/ecs2.3461

D'hont, B., Vanderhoeven, S., Roelandt, S., Mayer, F., Versteirt, V., Adriaens, T., et al. (2015). Harmonia⁺ and Pandora⁺: Risk screening tools for potentially invasive plants, animals and their pathogens. *Biol. Invasions* 17, 1869–1883. doi:10.1007/s10530-015-0843-1

Doak, D. F., Estes, J. A., Halpern, B. S., Jacob, U., Lindberg, D. R., Lovvorn, J., et al. (2008). Understanding and predicting ecological dynamics: Are major surprises inevitable? *Ecology* 89, 952–961. doi:10.1890/07-0965.1

Gordon, D. R., Mitterdorfer, B., Pheloung, P. C., Ansari, S., Buddenhagen, C., Chimera, C., et al. (2010). Guidance for addressing the Australian weed risk assessment questions. *Plant Prot. Q.* 25, 56–74.

Hawkins, C. L., Bacher, S., Essl, F., Hulme, P. E., Jeschke, J. M., Kühn, I., et al. (2015). Framework and guidelines for implementing the proposed IUCN environmental impact classification for alien taxa (EICAT). *Divers. Distrib.* 21, 1360–1363. doi:10.1111/ddi. 12379

Houle, D., Pélabon, C., Wagner, G. P., and Hansen, T. F. (2011). Measurement and meaning in biology. *Q. Rev. Biol.* 86, 3–34. doi:10.1086/658408

IUCN (2012). IUCN red list categories and criteria: Version 3.1. 2nd ed. Gland and Cambridge: IUCN.

Laska, M. S., and Wootton, J. T. (1998). Theoretical concepts and empirical approaches to measuring interaction strength. *Ecology* 79, 461–476. doi:10.1890/0012-9658(1998)079 [0461:TCAEAT]2.0.CO;2

Latombe, G., Canavan, S., Hirsch, H., Hui, C., Kumschick, S., Nsikani, M. M., et al. (2019). A four-component classification of uncertainties in biological invasions: Implications for management. *Ecosphere* 10, e02669. doi:10.1002/ecs2.2669

Lodge, D. M., Williams, S., MacIsaac, H. J., Hayes, K. R., Leung, B., Reichard, S., et al. (2006). Biological invasions: Recommendations for U.S. policy and management. *Ecol. Appl.* 16, 2035–2054. doi:10.1890/1051-0761(2006)016[2035: BIRFUP]2.0.CO;2

Mace, G. M., and Lande, R. (1991). Assessing extinction threats: Toward a reevaluation of IUCN threatened species categories. *Conserv. Biol.* 5, 148–157. doi:10.1111/j.1523-1739. 1991.tb00119.x

Martin, T. G., Burgman, M. A., Fidler, F., Kuhnert, P. M., Low-Choy, S., McBride, M., et al. (2012). Eliciting expert knowledge in conservation science. *Conserv. Biol.* 26, 29–38. doi:10.1111/j.1523-1739.2011.01806.x

McBride, M. F., Garnett, S. T., Szabo, J. K., Burbidge, A. H., Butchart, S. H. M., Christidis, L., et al. (2012). Structured elicitation of expert judgments for threatened species assessment: A case study on a continental scale using email. *Methods Ecol. Evol.* 3, 906–920. doi:10.1111/j.2041-210X.2012.00221.x

McGeoch, M. A., Spear, D., Kleynhans, E. J., and Marais, E. (2012). Uncertainty in invasive alien species listing. *Ecol. Appl.* 22, 959–971. doi:10.1890/11-1252.1

Novak, M., and Wootton, J. T. (2008). Estimating nonlinear interaction strengths: An observation-based method for species-rich food webs. Ecology 89, 2083–2089. doi:10.1890/08-0033.1

Popper, K. (1934). Logik der Forschung. Wien: Springer.

Regan, H. M., Colyvan, M., and Burgman, M. A. (2002). A taxonomy and treatment of uncertainty for ecology and conservation biology. *Ecol. Appl.* 12, 618–628. doi:10.1890/1051-0761(2002)012[0618:ATATOU]2.0.CO;2

Roy, H. E., Rabitsch, W., Scalera, R., Stewart, A., Gallardo, B., Genovesi, P., et al. (2018). Developing a framework of minimum standards for the risk assessment of alien species. *J. Appl. Ecol.* 55, 526–538. doi:10.1111/1365-2664.13025

Sandvik, H. (2020). Expansion speed as a generic measure of spread for alien species. *Acta Biotheor.* 68, 227–252. doi:10.1007/s10441-019-09366-8

Sandvik, H., Hilmo, O., Finstad, A. G., Hegre, H., Moen, T. L., Rafoss, T., et al. (2019). Generic ecological impact assessment of alien species (GEIAA): The third generation of assessments in Norway. *Biol. Invasions* 21, 2803–2810. doi:10.1007/s10530-019-02033-6

Suedel, B. C., Bridges, T. S., Kim, J., Payne, B. S., and Miller, A. C. (2007). Application of risk assessment and decision analysis to aquatic nuisance species. *Integr. Environ. Assess. Manag.* 3, 79–89. doi:10.1002/ieam.5630030107

Tversky, A., and Kahneman, D. (1974). Judgment under uncertainty: Heuristics and biases. *Science* 185, 1124–1131. doi:10.1126/science.185.4157.1124

USDA (2019). Guidelines for the USDA-APHIS-PPQ weed risk assessment process. Raleigh: USDA.

Wolman, A. G. (2006). Measurement and meaningfulness in conservation science. *Conserv. Biol.* 20, 1626–1634. doi:10.1111/j.1523-1739.2006.00531.x