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Loss of native brown trout diversity in streams of the continental Croatia

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Introduction: The genetic diversity of brown trout in the Western Balkans has been disrupted by the introduction of non-native Atlantic phylogenetic lineages and non-native haplotypes of the Danubian phylogenetic lineage. The Western Balkans is characterized by the greatest phenotypic and genotypic diversity of trout populations, and a large part of the internal territory belongs to the Black Sea basin, where the Danubian Da1 haplotype is native. Artificial propagation of non-native lineages in the Western Balkans has a long history, and these populations are often the only available material for stocking rivers attractive for fishing.

Material and Methods: Fifteen populations in the Danube basin of the continental Croatia were analysed. The analysis of eight microsatellite loci was performed to determine the structure of brown trout populations, as well as the degree of introgression of non-native genetic material into the native.

Results and Discussion: The results of this study showed significant genetic similarity among brown trout populations, confirming a long history of introduction with non-native genetic material. The main reason was uncontrolled stocking with inadequate material, which is available in fish farms and consists mainly of brown trout of the Atlantic phylogenetic lineage. The results of this study also indicated stocking with brown trout of the non-native haplotypes of the Danubian phylogenetic lineage. The potential breeding origin of brown trout carrying the Danubian Da2 mtDNA haplotype and ways of its introduction into rivers have yet to be investigated. For the survival of the unique gene pool of brown trout in Croatian rivers, it is of fundamental importance to know the structure of wild and farmed populations with the aim of proposing and implementing conservation measures.

KEYWORDS

brown trout, microsatellites, Croatia, stocking, conservation

Introduction

The commercial value of brown trout *Salmo trutta*, along with its significant importance to humans, is reflected in various aspects, particularly its widespread artificial propagation in fish farms globally. Farming is carried out for various purposes, of which stocking notably impacts the survival of native, genetically pure, and geographically unique brown trout populations.

Numerous studies on this species, still considered the *S. trutta* species complex, have shown that the life history, phenotypic plasticity, and high adaptability to diverse environmental conditions have not yet been clarified, more than likely as a consequence of their exceptionally high morphological and genetic variability (Behnke, 1972; Weiss et al., 2001; Sanz, 2018; Škraba Jurlina et al., 2018). The Western Balkans exhibit a high degree of endemism in salmonid genera and species (Behnke, 1986; Oikonomou et al., 2014). Together with the Iberian Peninsula and the Apennines, the Western Balkans is considered the area with the greatest phenotypic and genetic diversity among trout populations (Bernatchez, 2001) with complex evolutionary mechanisms, including the occurrence of secondary contacts between ancestral lineages, as well as local adaptation (Sanz et al., 2002; Snoj et al., 2008; Vera et al., 2010). The Western Balkans is the home range for populations of three phylogenetic lineages of *Salmo* cf. *trutta* which are defined based on the control region (CR) of mitochondrial DNA (mtDNA): Danubian (DA), Adriatic (AD) and *marmoratus* (MA). The DA lineage inhabits the northern, eastern, and central parts of the Balkan Peninsula, i.e., the Black Sea basin, and the AD lineage inhabits the northwestern and southwestern regions, including numerous lotic and some lentic habitats in the continental part of the Adriatic and Ionian Seas' basins (Georgiev, 2003). The MA lineage is found close to the dispersal range of brown trout of the Danubian and Adriatic lineages, in the narrowest area in the Western Balkans, inhabiting the Adriatic drainages in Bosnia and Herzegovina and Montenegro (Snoj et al., 2010; Mrdak et al., 2012; Škraba Jurlina et al., 2020; Righi et al., 2023). The status of the MA lineage, which is still often identified as marble trout, has not yet been resolved in Croatia (Čaleta et al., 2019).

The rivers of the Western Balkans within the Black Sea basin drain a considerable territory and are very important in the context of studying the diversity and life history of different phylogenetic lineages (Škraba Jurlina et al., 2018). The DA lineage, native to the Danube basin, consists of numerous haplotypes, with the Da1 haplotype being the most widespread. Still, not all of the DA haplotypes are considered native to the Western Balkans area/region. The presence of non-native DA haplotypes, as well as other phylogenetic lineages, in most of the studied rivers presents a significant threat to the native population's genetic structure (Snoj, 2004; Marić et al., 2006; Jadan et al., 2007; Tošić et al., 2016; Simonović et al., 2017a; Buj et al., 2020; Ivić et al., 2021).

Breeding brown trout in the Balkans is a profitable enterprise which makes it highly popular across all countries in the region. With the aim of intensive artificial propagation, and thus, a greater commercial profit (Simonović et al., 2017b), the initially domesticated native DA lineage (Gridelli, 1936; Kohout et al., 2012) proved to be insufficient. Consequently, breeders introduced breeding technology from Western Europe and the

Atlantic lineage of brown trout suitable for this purpose (Taler, 1949; Pofuk et al., 2017; Piria et al., 2020).

Human-induced translocations of organisms and anthropogenic alterations of natural habitats are the most common causes of hybridization and introgression of non-native genotypes into native populations (Allendorf et al., 2001). *Salmo trutta* is characterized by genetic diversity at the intrapopulation level, which is the reason for the existence of genetically differentiated local populations. The loss of these unique populations could lead to a significant reduction in genetic variability within the species (Laikre et al., 1999). Stocking became a common practice to increase and support wild stocks of many fish species, including brown trout (Righi et al., 2023). In the beginning of the 20th century, it was pointed out that supportive breeding programs could also pose a threat to gene-level biodiversity, by releasing cultured individuals into the wild, with the aim of increasing the size of natural populations (Laikre et al., 2008). These activities could lead to manipulation of reproductive rates, causing reduced effective population size, increased rate of genetic drift and loss of genetic variability (Ryman et al., 1995; Wang and Ryman, 2001; Laikre et al., 2008), which is observed in many populations of brown trout in the Balkans (Jadan et al., 2007; Buj et al., 2020; Piria et al., 2020; Škraba Jurlina et al., 2020).

Although the research on genetic variability of brown trout has been extensively conducted to assess the introgression of farmed lineages, gene-level monitoring programs are still notably absent, despite repeated calls for their implementation (Laikre et al., 2008; Tošić et al., 2016; Simonović et al., 2017a; Škraba-Jurlina et al., 2020). To ensure the survival of a population or species, it is necessary to preserve its genetic diversity, and thus its evolutionary potential (Ryman et al., 1995). Unfortunately, the genotyping of brown trout individuals in brood stocks has been limited to a small number of fish farms in the Western Balkans.

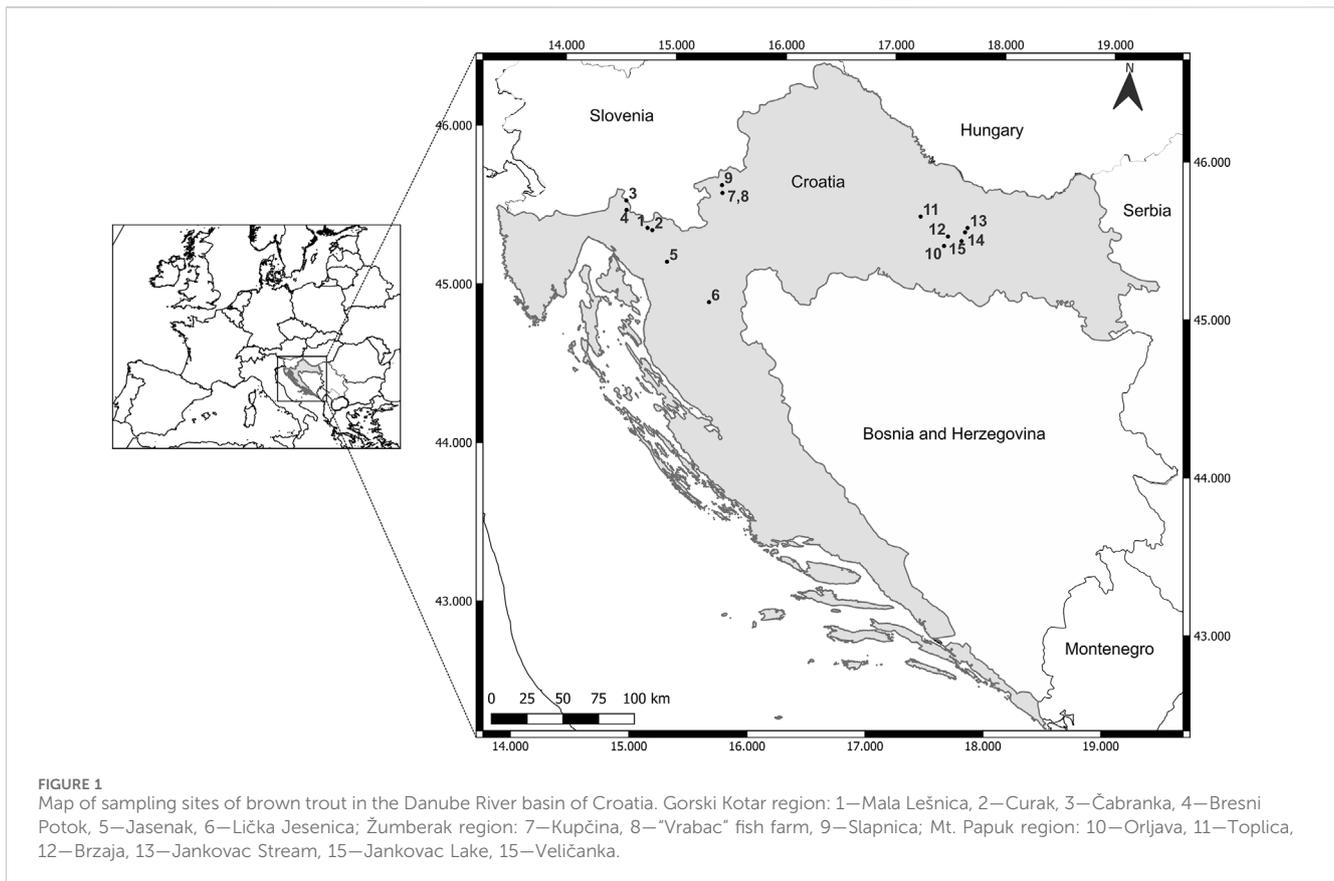
In Croatian fish farms, brown trout of the AT lineage are mainly bred, so consequently the rivers are most likely stocked with these individuals due to their availability (Marić et al., 2022). In this way, non-native genetic material enters natural watercourses (Piria et al., 2020). Brown trout of the AT lineage are already recognized as the main cause of the loss of the original genetic diversity of brown trout (Weiss et al., 2001; Simonović et al., 2017b) and show an invasive character (Simonović et al., 2013; 2015; Piria et al., 2020), but their stocking is not regulated by any law in Croatia.

The goal of this research was to determine the genetic structure of brown trout populations in the Croatian part of the Danube basin, and to assess the degree of introgression of non-native genetic material within studied populations, as well as its possible origin.

Material and methods

A total of 141 samples were collected and analyzed from 15 locations: in the Gorski Kotar area (six locations), in Žumberak (three locations) and on the slopes of Mount Papuk (six locations) (Figure 1). The sampling was carried out in the period of April/May during 2017 and 2018.

For the purposes of genetic analyses, the anal fin clips of each individual were taken and stored in 96% ethanol. The extraction of



DNA was conducted according to the protocol of manufacturer of the Quick-gDNA™ MiniPrep (Zymo Research Corporation, United States) extraction kits.

The analysis of the control region of mitochondrial DNA (CR mtDNA) was used to identify the phylogenetic lineages of brown trout and their haplotypes, while restriction analysis of the L-lactate dehydrogenase nuclear locus (LDH-C*) was used to detect if hybrids between AT and DA lineages exist. For a detailed description of the procedures and software analyses refer to Kanjuh et al. (2020).

Population diversity was accessed by analyses of eight microsatellite loci: Str73INRA (Estoup et al., 1993) and Ssa410Uos (Cairney et al., 2000), SsaD190 and SsaD71 (King et al., 2005), Ssa85 (O'Reilly et al., 1996) and SSp2216 (Paterson et al., 2004), and OMM1064 (Rexroad et al., 2002) and SsoSL438 (Slettan et al., 1995), amplified in four duplex reactions. The total volume of the mixture for all four duplex reactions was 10 μ L, and its composition was as follows: forward and reverse primers for a specific locus (with final concentrations: 0.2 μ M for SsaD190, SsaD71, SSp2216 and OMM1064, 0.4 μ M for Ssa410Uos and SsoSL438, 0.1 μ M for Ssa85 and 0.15 μ M for Str73INRA), dH₂O, 1X PCR buffer, 0.2 mM dNTP, 1.5 mM MgCl₂, 0.5 U Taq polymerase, and 2 μ L DNA sample. The amplification conditions for the first three duplex reactions (Str73INRA and Ssa410Uos, SsaD190 and SsaD71, Ssa85 and SSp2216) were identical: initial denaturation (94°C, 3 min), 30 cycles of denaturation (94°C, 45 s), primer binding (60°C, 1 min), elongation (72°C, 30 s) and final elongation (72°C, 1 h). An exception in the protocol was the fourth duplex reaction

(OMM1064 and SsoSL438), in which the primer binding was set to 57°C, and the number of cycles was 35. Fragment analysis of microsatellite loci was performed in the laboratories of MACROGEN® Europe. For several samples, fragment analysis was performed at the Center for Human Molecular Genetics, Faculty of Biology, University of Belgrade, where the fragment analysis was performed using GeneScan™ 500 LIZ® Size Standard (Applied Biosystems, United States) on an ABI-3130 Genetic Analyzer (Applied Biosystems, United States). Analysis was done using GeneMapper® ID v3.2.1 software (Applied Biosystems, United States).

Statistical analyses of microsatellite loci

Hardy-Weinberg equilibrium (HWE) and p -values for F_{IS} within samples, based on 120,000 randomizations, were observed in Fstat 2.9.3.2 software (Goudet, 2002). Observed and expected heterozygosity values, the average number of alleles per locus, allele frequencies, fixation indices, Nei's distances between populations per 1,000 permutations and the Factorial Correspondent Analysis (FCA) were calculated using GENETIX 4.05 software (Belkhir et al., 1996). The software POPULATIONS 1.2.31 (Langella, 2002) was used to calculate shared allele distances (DAS) and to generate a Neighbor-Joining (NJ) tree. The results of Nei's distances were used for CLUSTER analysis using the method of Unweighted Pair-group Average (UPGMA) in STATISTICA 8.0 software (StatSoft, Inc, 2001). Population structure analysis was done using

TABLE 1 Results of control region of mitochondrial DNA showing distribution of the brown trout haplotypes detected in the studied rives in Croatia. n—number of individuals, DA—Danubian lineage, AT—Atlantic lineage (from Kanjuh et al., 2020).

River	Drainage	n	DA			AT
			Da1	Da2	Da22	At-H3 (At1)
Mala Lešnica	Kupa	12	9			3
Curak	Kupa	12	10			2
Čabranka	Kupa	11	9	2		
Jasenak	Kupa	6	4			2
Bresni potok	Kupa	11		2		9
Slapnica	Kupa	10	8	2		
Kupčina	Kupa	14	13			1
“Vrabac” fish farm	Kupa	10				10
Lička Jesenica	Lička Jesenica	9	1	8		
Veličanka	Sava	6	3		1	2
Orljava	Sava	5			5	
Brzaja	Sava	9		1	7	1
Toplica	Drava	13	10	2		1
Jankovac Stream	Drava	5	5			
Jankovac Lake	Drava	8	1	7		
Total		141	73	24	13	31

STRUCTURE 2.3.4 software (Pritchard et al., 2000). The structure analysis for all 141 individuals was conducted with assumed number of groups $K = 15$ and the test period was set at 100,000 to 200,000, which was repeated seven times for each group K . The STRUCTURE Harvester tool (Earl and vonHoldt, 2012) was used to estimate the most likely value of K according to Evanno et al. (2005). Additional structure analysis was done only for DA individuals, who according to the LDH-C* molecular marker, were homozygous for the LDH-C*100 allele. The assumed number of groups was $K = 15$ and the test period was set at 100,000 to 200,000, which was again repeated seven times for each value K . The bottleneck effect was evaluated using the BOTTLENECK 1.2.02 software (Cornuet and Luikart, 1996; Piry et al., 1999), within Sign test and Wilcoxon test (Sign-rank test) and two mutation models, two-phase mutation model (TPM) and stepwise mutation model (SMM). As parameters for the assessment of significance for these tests, the proportion of a mutation step in the TPM was set at 95% and the mutation variance for the TPM at 12,000 and 100,000 iterations.

Results

Control region of mitochondrial DNA and L-lactate dehydrogenase nuclear locus

The results of CR mtDNA and LDH-C* locus analyses are presented in detail in Kanjuh et al. (2020). In short, DA and AT phylogenetic lineages were identified in analyzed samples. The presence of non-native brown trout haplotypes was detected in

all sampling areas: At1 (At-H3) in eight localities (including the fish farm), Da2 in seven localities and Da22 in three localities (Table 1). The only river where all individuals (five in total) carried the native Da1 haplotype was the Jankovac Stream. Analysis of the LDH-C* locus revealed interbreeding between DA and AT individuals, indicating the presence of hybrids. Even in the Jankovac Stream, the crossbreeding between Da1 and At1 individuals was recorded, with one exception that was homozygous for the LDH-C*100 allele. The only genetically pure population was the one from the fish farm, where all the analysed individuals were non-native AT lineage by maternal (CR mtDNA) and paternal (LDH-C*) inheritance. The results obtained with these two molecular markers indicate a strong introgression of non-native genetic material into native, which is most likely a consequence of crosses in the F1 generation, resulting from uncontrolled, long-term stocking with farmed trout. This is supported by the results showing that 21 individuals of the DA lineage were homozygous for the LDH-C*90 allele, which is characteristic trait of the AT lineage, while six individuals of AT haplogroups were homozygous for the LDH-C*100 allele typically found in the DA haplogroup.

Microsatellite loci

The highest expected (Hexp.) and observed (Hobs.) heterozygosities were recorded in the Jasenak River population (Table 2), in which the highest number of alleles, i.e., the highest allelic richness at the loci OMM1064 (16 alleles) and SsoSL438 (14 alleles) was observed (Supplementary Table S1). The lowest

TABLE 2 Average heterozygosity of microsatellite loci in the studied populations. Hexp.—expected heterozygosity, Hn.b.—objective heterozygosity, Hobs.—observed heterozygosity, SD—standard deviation, p—probability, Ān—average number of alleles per locus, LE—Mala Lešnica, CZ—Curak, ČA—Čabranka, JA—Jasenak, BP—Bresni potok, SL—Slapnica, KČ—Kupčina, KČR—“Vrabac” fish farm, LJ—Lička Jesenica, VE—Veličanka, OR—Orljava, TO—Toplica, JP—Jankovac Stream, JJ—Jankovac Lake, BR—Brzaja

	LE	CZ	ČA	JA	BP	SL	KČ	KČR	LJ	VE	OR	TO	JP	JJ	BR
Hexp	0.765	0.659	0.630	0.780	0.764	0.682	0.683	0.663	0.581	0.670	0.542	0.685	0.537	0.653	0.568
SD	0.144	0.200	0.167	0.071	0.090	0.160	0.270	0.121	0.288	0.147	0.269	0.213	0.326	0.144	0.306
Hn.b	0.798	0.687	0.660	0.850	0.800	0.718	0.708	0.698	0.615	0.731	0.603	0.713	0.597	0.697	0.601
SD	0.150	0.209	0.175	0.077	0.095	0.169	0.280	0.128	0.305	0.161	0.298	0.221	0.362	0.153	0.324
Hobs	0.688	0.458	0.511	0.812	0.784	0.525	0.598	0.670	0.542	0.646	0.475	0.615	0.575	0.609	0.514
SD	0.124	0.244	0.266	0.226	0.175	0.287	0.264	0.225	0.281	0.188	0.301	0.240	0.377	0.245	0.302
p (0.95)	1	1	1	1	1	1	1	1	0.875	1	1	1	0.875	1	1
p (0.99)	1	1	1	1	1	1	1	1	0.875	1	1	1	0.875	1	1
Ān	7.625	6.500	4.750	6.250	7.250	6.375	8.875	5.750	5.250	5.250	3.875	6.750	4.375	4.750	5.875

TABLE 3 Genetic distances for populations pairs, based on allele frequencies across the eight microsatellite loci, are shown by fixation index values (F_{ST}) are shown under diagonal, and estimated gene flow values (Nm) are shown above diagonal. VE—Veličanka, OR—Orljava, LE—Mala Lešnica, TO—Toplica, LJ—Lička Jesenica, SL—Slapnica, JA—Jasenak, JP—Jankovac Stream, JJ—Jankovac Lake, BR—Bresni potok, CZ—Curak, ČA—Čabranka, KČ—Kupčina, KČR—“Vrabac” fish farm, BP—Brzaja

Localities	VE	OR	LE	TO	LJ	SL	JA	JP	JJ	BR	CZ	ČA	KČ	KČR	BP
VE		1.84	4.04	1.93	1.23	1	3.02	0.88	1.26	1.39	1.46	1.33	1.38	0.87	1.5
OR	0.12		2.14	1.02	1.42	0.95	1.66	1.39	1.49	5.26	1.64	1.42	1.85	0.84	1.15
LE	0.06	0.10		2.64	2.37	1.46	19.09	2.13	1.82	2.15	4.69	2.21	3.83	1.46	2.41
TO	0.11	0.20	0.09		1.11	1.23	3.25	0.86	1.4	1	1.48	0.97	1.29	1.35	1.73
LJ	0.17	0.15	0.10	0.18		1.06	1.77	0.86	1.04	1.8	2.94	1.05	2.14	0.84	1.15
SL	0.20	0.21	0.15	0.17	0.19		2.1	0.87	1.02	0.93	1.23	0.88	1.6	1.26	1.77
JA	0.08	0.13	0.01	0.07	0.12	0.11		1.26	1.78	1.37	2.58	1.54	2.48	2.33	5.52
JP	0.22	0.15	0.11	0.22	0.23	0.22	0.17		1.46	1.68	1.56	1.2	2.02	0.69	1.1
JJ	0.17	0.14	0.12	0.15	0.19	0.20	0.12	0.15		1.34	1.26	1.54	1.5	1.14	1.72
BR	0.15	0.05	0.1	0.20	0.12	0.21	0.15	0.13	0.16		2.02	1.47	3.07	0.76	0.95
CZ	0.15	0.13	0.05	0.14	0.08	0.17	0.09	0.14	0.17	0.11		1.49	2.84	1.17	1.32
ČA	0.16	0.15	0.1	0.21	0.19	0.22	0.14	0.17	0.14	0.15	0.14		1.36	0.74	1.28
KČ	0.15	0.12	0.06	0.16	0.10	0.13	0.09	0.11	0.14	0.08	0.08	0.15		1.46	1.41
KČR	0.22	0.23	0.15	0.16	0.23	0.17	0.10	0.27	0.18	0.25	0.18	0.25	0.15		2.3
BP	0.14	0.18	0.09	0.13	0.18	0.12	0.04	0.19	0.13	0.21	0.16	0.16	0.15	0.10	

values of Hexp. were observed for populations from Jankovac Stream (0.537) and Orljava River (0.543). The lowest values of Hobs. were recorded for populations from the Orljava River (0.475) and the Curak River (0.458). The lowest average number of alleles was recorded in the Orljava River population (Table 2).

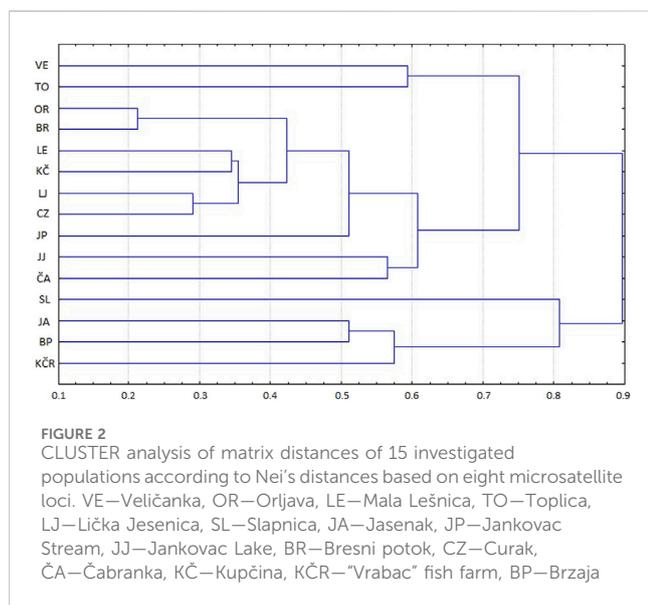
The GENETIX software analysis (Table 3) showed the highest genetic distances (F_{ST}) between “Vrabac” fish farm and the Jankovac Stream (0.27), as well as the smallest gene flow (Nm) between these two populations (0.69). Similarly, Nei’s distances (Table 4) showed the highest values between populations from the “Vrabac” fish farm and those from the Veličanka (1.458) and Čabranka rivers (1.408)

and Jankovac Stream (1.318). The significant genetic distance between the populations from the “Vrabac” fish farm and the Jankovac Stream coincides with the results of the CR mtDNA analysis which showed that all individuals from the fish farm belong to At1 haplotype, while all individuals from the Jankovac Stream belong to Da1 haplotype (Table 1).

The CLUSTER analysis of Nei’s distances and shared allele distances values showed populations’ grouping. Four populations from the Gorski Kotar and Žumberak areas were clustered together (SL, JA, BP and KČR) (Figure 2), and the second cluster consisted of the remaining 11 populations. Two populations stood out compared

TABLE 4 Nei's distances values based on the eight microsatellite loci between pairs of populations. VE—Veličanka, OR—Orljava, LE—Mala Lešnica, TO—Toplica, LJ—Lička Jesenica, SL—Slapnica, JA—Jasenak, JP—Jankovac Stream, JJ—Jankovac Lake, BR—Bresni potok, CZ—Curak, ČA—Čabranka, KČ—Kupčina, KČR—"Vrabac" fish farm, BP—Brzaja

Localities	VE	OR	LE	TO	LJ	SL	JA	JP	JJ	BR	CZ	ČA	KČ	KČR	BP
VE	-														
OR	0.537	-													
LE	0.446	0.513	-												
TO	0.593	0.856	0.482	-											
LJ	0.688	0.470	0.400	0.719	-										
SL	1.395	1.006	0.985	0.890	0.789	-									
JA	0.676	0.750	0.352	0.514	0.608	0.832	-								
JP	1.052	0.492	0.493	1.035	0.741	1.086	0.946	-							
JJ	0.910	0.568	0.706	0.709	0.749	1.139	0.881	0.549	-						
BR	0.583	0.212	0.419	0.781	0.345	0.897	0.766	0.383	0.547	-					
CZ	0.770	0.532	0.306	0.647	0.291	0.852	0.609	0.536	0.783	0.376	-				
ČA	0.740	0.540	0.488	0.987	0.670	1.198	0.853	0.609	0.565	0.458	0.570	-			
KČ	0.810	0.465	0.345	0.769	0.363	0.636	0.619	0.418	0.646	0.264	0.347	0.627	-		
KČR	1.458	1.022	0.863	0.714	0.968	0.822	0.629	1.318	0.858	1.078	0.816	1.408	0.642	-	
BP	1.027	0.953	0.721	0.733	0.866	0.770	0.511	0.972	0.736	1.096	0.979	0.885	0.910	0.520	-

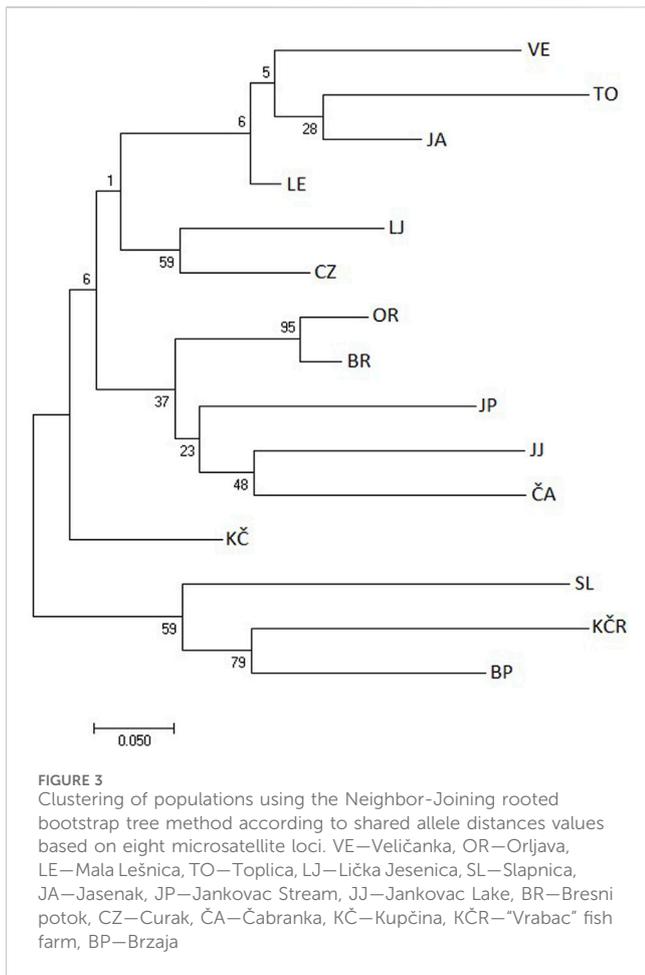


to the others: the Slapnica River and the Jankovac Stream (Figure 2). The shared allele distances values (DAS) showed small deviations from Nei's distances (Figure 3). Analysis of both DAS and Nei's distances values showed sister relationships between the populations of Čabranka and Jankovac Lake, which follows the low F_{ST} value between these two populations (0.14) (Table 3). In contrast to Nei's distances values, DAS values showed clustering of the population from the Jasenak River with populations from the Toplica River, but also the separation of Veličanka, Mala Lešnica and Kupčina rivers.

The corresponding factorial analysis (CFA) (Figures 4A–C) showed a clear separation of the Slapnica River population (Figure 4A), while a large genetic overlap is shown for the other populations.

The evaluation of the bottleneck effect showed that the H_e value was higher than the H_{eq} value in the populations from the Orljava River (at five loci according to TPM), Jankovac Stream (at five loci according to TPM and SMM), Brzaja River (at six loci according to TPM and SMM), and Čabranka River (at six loci according to TPM, and at five loci according to SMM) (Table 5). However, no locus had a probability lower than 0.05. Based on both mutation models (TMP and SMM), a significant heterozygosity deficit was recorded in the rivers Brzaja, Curak and the "Vrabac" fish farm, while for the population from the Kupčina River, a heterozygosity deficit was recorded only according to SMM. These results indicate a recent expansion of the above mentioned populations (Table 5).

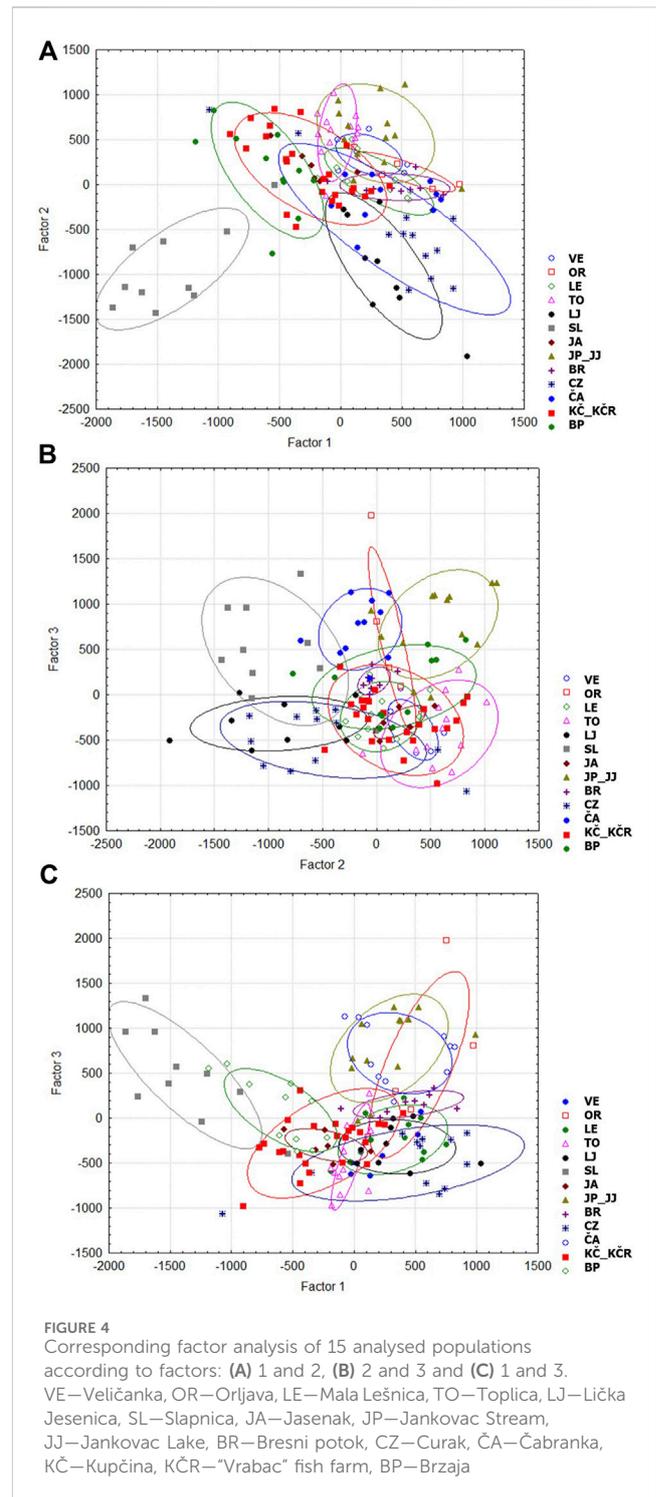
The STRUCTURE analysis revealed two distinct brown trout genetic clusters (when $K = 2$, Figure 5A). The structuring of the populations was very much congruent with the mtDNA haplogroup to which the individuals belong (Kanjuh et al., 2020). One group of individuals colored green were those of the DA lineage, while the others colored red were of the AT lineage (Figure 5B). This genetic population structure can best be seen in the brown trout from the Curak River. Twelve individuals of this population were separated during grouping. Ten of them of the DA lineage were grouped with other individuals of the DA lineage (i.e., colored green). Remaining two individuals from the Curak River belonging to AT lineage (one in homozygous state for LDH-C*90, and the other in heterozygous state) were grouped with individuals of the AT lineage (i.e., colored red). The green



colored group of individuals comprised both DA-AT hybrids and genetically pure DA individuals. Populations from the Čabranka and Lička Jasenica rivers were grouped together (i.e., colored green). All individuals belonged to DA lineage, but among them there were also DA-AT hybrids: two individuals in homozygous state for LDH-C*90 (LJ41 and LJ42), and five individuals in heterozygous state (LJ37, LJ43, ČA5, ČA8 and ČA10) (Figure 5). The grouping with only DA individuals suggested three genetic clusters (K = 3) (Figure 6A). Individuals from the Curak River (CZ) separated into one genetic cluster, while the other two genetic clusters showed overlapping genetic structure (Figure 6B).

Discussion

The results of this study indicated a decrease in genetic variability in the Orljava River and Jankovac Stream. Analysis of CR mtDNA revealed that only one haplotype is present in both populations, non-native Da22 in Orljava River and Da1 in Jankovac Stream (Table 1). Neither Jankovac Stream nor Orljava River population proved to be genetically “pure”, considering that DA-AT hybrids were identified there (Kanjuh et al., 2020). In this regard, a bottleneck effect was observed for these two populations. Therefore, to continuously follow the status of these two populations, ongoing monitoring and further



investigation are necessary. Evaluating the bottleneck effect in other populations revealed a heterozygosity deficit in the Brzaja, Curak and Kupčina rivers, indicating population expansions, most likely due to the introduction of the non-native farmed At1 haplotype. The same was observed in the “Vrabac” fish farm, but that was to be expected knowing that within the fish farm non-native At1 individuals are farmed.

In comparison with all the analysed populations, it was shown that the Slapnica River population is genetically different. The

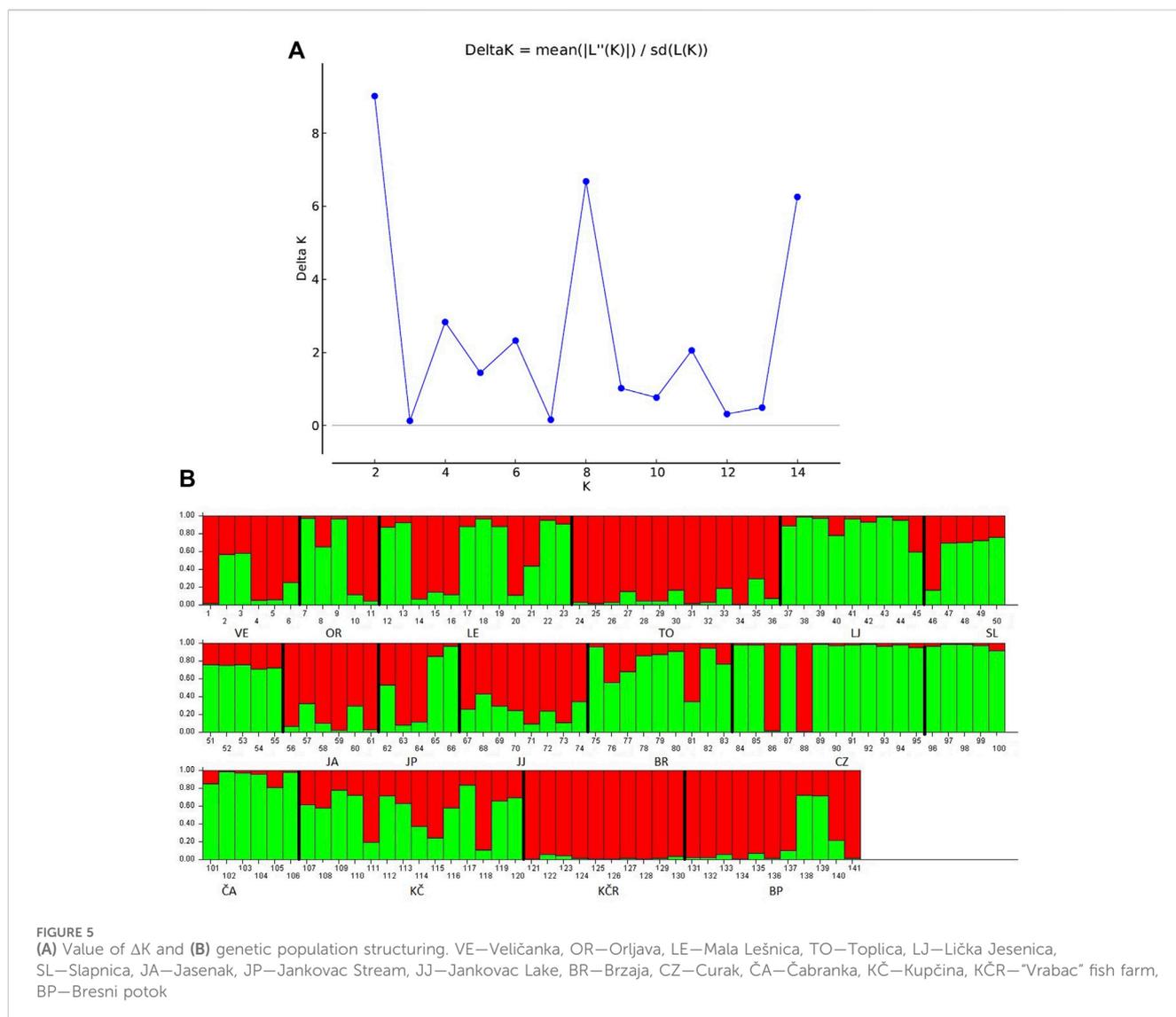
TABLE 5 Wilcoxon test results based on two-phase mutation model (TPM) and stepwise mutation model (SMM). Bold values are significantly lower than 0.05. D—heterozygosity deficiency, E—heterozygosity excess, VE—Veličanka, OR—Orljava, LE—Mala Lešnica, TO—Toplica, LJ—Lička Jesenica, SL—Slapnica, JA—Jasenak, JP—Jankovac Stream, JJ—Jankovac_Lake, BR—Brzaja, CZ—Curak, ČA—Čabranka, KČ—Kupčina, KČR—“Vrabac” fish farm, BP—Bresni potok

	TPM		SMM	
VE	0.125D	0.902E	0.098D	0.963E
OR	0.273D	0.770E	0.273D	0.770D
LE	0.578D	0.473E	0.230D	0.808E
TO	0.156D	0.875E	0.125D	0.902E
LJ	0.289D	0.766E	0.289D	0.766E
SL	0.156D	0.875E	0.125D	0.902E
JA	0.808D	0.230E	0.808D	0.230E
JP	0.234D	0.812E	0.234D	0.812E
JJ	0.726D	0.320E	0.726D	0.320E
BR	0.014D	0.990E	0.014D	0.990E
CZ	0.020D	0.986E	0.010D	0.994E
ČA	0.680D	0.371E	0.629D	0.422E
KČ	0.125D	0.902E	0.027D	0.980E
KČR	0.010D	0.994E	0.006D	0.996E
BP	0.473D	0.578E	0.422D	0.629E

presence of AT lineage individuals was not recorded there, so this population consists only of individuals of the DA lineage. Nevertheless, the presence of non-native Da2 and DA-AT hybrids in the Slapnica River indicates that its native character was disturbed (Kanjuh et al., 2020). Population from the Slapnica River most likely separates due to the presence of private alleles in hybrid individuals, at microsatellite loci Str73INRA, Ssa410Uos, SsaD190 and OMM1064 (Supplementary Table S1). In all other populations, a large overlap was observed (Figure 2; Figure 3), both within geographically close and among geographically distant ones, as clearly evidenced in populations from the Čabranka River (Gorski Kotar) and Jankovac Lake (Mt. Papuk). This level of overlap indicates a high genetic similarity of the populations, which is a consequence of stocking from hatcheries that import and breed, most likely, the same (AT) phylogenetic lineage of brown trout (Piria et al., 2020). In fish farms throughout Croatia, brown trout carrying the At1 mtDNA haplotype is probably the most farmed and used for stocking of streams attractive for fishing. Minor genetic differences between populations observed in this study, most likely caused by stocking with the same genetic material, were additionally confirmed by the results of the population structure analysis. Separation of only two genetic clusters indicates strong introgression of the non-native genotypes, which is consistent with molecular variance results for CR mtDNA haplotypes (Kanjuh et al., 2020), which show little variation among populations. Structuring of only DA individuals also showed genetic similarity between populations. This result indicated the stocking with not only Atlantic (AT) lineage, but also non-native Danubian (DA) haplotypes of brown trout. Two non-native DA

haplotypes were recorded in the analysed rivers of the Danube basin in Croatia (Kanjuh et al., 2020). The Da22 haplotype was limited to three localities (Veličanka River, Orljava River and Brzaja River) at the northern slopes of the Mt. Papuk (Kanjuh et al., 2020). This haplotype is native to the Lohnbach and Daglesbach rivers in Austria (Duftner et al., 2003) and in the Una River drainage in Bosnia and Herzegovina (Škraba et al., 2017). The non-native Da2 haplotype was much more frequent than Da22 haplotype (recorded in seven out of fifteen analysed populations) (Kanjuh et al., 2020). The Da2 haplotype is native to the streams and rivers in southern Germany (Bernatchez, 2001) and streams belonging to the Austrian part of the Danube drainage (Weiss et al., 2001). The stocking of various freshwaters of Europe during the 19th century was carried out with fish material from southern Germany (Kohout et al., 2012). It is most likely that these stockings included the territory of today's Croatia. Precise data on fish farm breeding of non-native Danubian haplotypes of brown trout do not exist, but first data on the presence of the Da2 haplotype in freshwater of Croatia was recorded in the Gacka River, together with the non-native At1 haplotype (Jadan et al., 2007). The separation of DA individuals from the Curak River into a single genetic cluster (Figure 6B) is interesting information from the conservation point-of-view. All DA individuals from the Curak River belong to the native Da haplotype and are homozygous to the native LDH-C*100 allele (Kanjuh et al., 2020). It is possible that the influence of the non-native At1 haplotype was suppressed in this population, considering that out of a total of twelve individuals, only one was genetically pure AT and only one was AT-DA hybrid (Kanjuh et al., 2020).

Genetic research of brown trout and their susceptibility to the presence of the introduced AT lineage in Croatia was until now mainly analysed using *cytb* and CR mtDNA as molecular markers. Buj et al. (2020) and Ivić et al. (2021) emphasized the negative impact of non-native AT lineage on the native populations of the DA lineage (nominal *Salmo labrax*) and the need for conservation of its genetic character in the Plitvice Lakes National Park and Žumberak-Samobor Hills Nature Park, respectively. The presence of the At1 haplotype has been reported for many streams attractive for fishing throughout the Western Balkans (Snoj, 2004; Marić et al., 2006; 2012; Jadan et al., 2007; Simonović et al., 2015; Tošić et al., 2016; Škraba Jurlina et al., 2018; 2020). In this regard, it is significant that apart from the At1 haplotype, the presence of the rare At17 haplotype was recorded in the upper reaches of the Mrtvica River (Adriatic Basin), undoubtedly indicating a unique event of stocking of this river (Škraba Jurlina et al., 2018). As its presence was not detected in the lower part of the river, the authors assume that the trout spawned only in the upper part of the river, producing hybrids with brown trout of the resident Adriatic lineage. They also pointed out that gene flow between the populations in the upper and lower parts of the river was low (despite the absence of physical and/or reproductive barriers), indicating the impossibility of individuals from the upper reaches of the Mrtvica River to migrate downstream and hybridize with the brown trout of the native Adriatic lineage consisting of both resident (predominantly males) and lake-dwelling, migratory (predominantly females) fish (Škraba Jurlina et al., 2018). The negative influence of the At1 haplotype was also recorded in non-fishing rivers, which are characterized by the presence of specific and narrowly distributed haplotypes. One

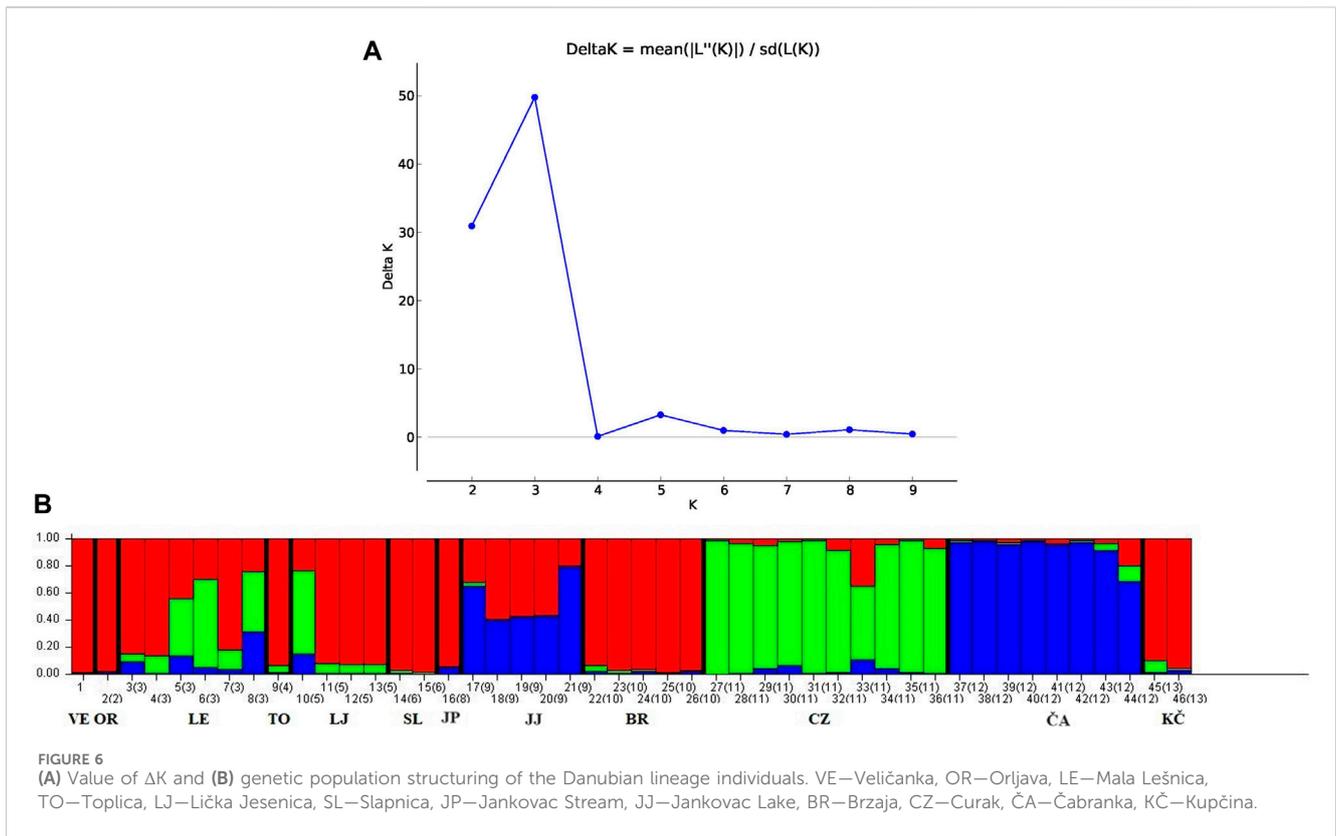


such example is the population in the Vratna River, a tributary of the Danube River at the downstream border of the Iron Gate Gorge, where the rare haplotype Da23c occurs, which is currently threatened by the appearance of the stocked brown trout of the At1 haplotype (Tošić et al., 2016; Škraba Jurlina et al., 2020).

In addition to the genetic analyses, the invasive character of the AT lineage has been highlighted in recent research, which showed the rapid adaptation of this lineage to the different conditions in natural watercourses, the consumption of a wide range of available food and competition for food and space, as well as diet overlap with native DA lineage individuals (Piria et al., 2020). Also, in assessment of invasiveness risk in the Danube and Adriatic basins across four Balkan countries, out of 13 existing and four horizon non-native salmonid species, *S. trutta* of the AT lineage showed a very high risk of invasiveness in the current climate conditions (Marić et al., 2022). In addition to At1, the non-native Da2 haplotype has also a significant impact on disrupting the genetic structure of the native Croatian populations of brown trout. Therefore, it would be important for future research to focus on obtaining data on the

potential breeding origin and ways of introduction of the Da2 haplotype, but also on assessing the risk of invasiveness to the native genetic pool of the brown trout complex. Apart from the presence of the Da2 haplotype in the rivers of Croatia (Kanjuh et al., 2020), the presence of Da2 haplotype was recorded in the rivers of Serbia (Marić et al., 2006; Simonović et al., 2015; Tošić et al., 2016), Bosnia and Herzegovina (Mrdak et al., 2012; Škraba et al., 2017), and Montenegro (Mrdak et al., 2012).

Everything said above points out to a long history and the current management practice of stocking with the hatchery-reared brown trout. There are clear warnings of the negative impact of the AT haplotypes on native and wild brown trout populations. Although the detrimental effects of strong introgression in wild brown trout populations are visible, management procedures in many countries are apparently not strict enough and properly applied. Despite proposals for potential management plans, the native gene pool of brown trout remains under serious threat. Studying the wild, native, and locally specific populations of brown trout on the territory of



Croatia, determining of their status and genetic structure should be a priority to maintain and preserve their original genetic stocks. Genotyping the remaining brown trout stock from inland waters is essential. It is to be implemented into the national legislation, in order to set a mandatory genotyping of hatchery brood stocks and to identify and register native brood fish. This practice is currently weakly enforced in Croatia, and in many countries of the Balkan Peninsula. It is important to examine the genetic structure of brown trout grown in fish farms, as well as the stocks used for stocking, which, so far, has not been legally regulated or implemented as a necessary practice in majority of countries in the Western Balkans. Although some authors point out that stocking as a measure of fisheries management should be re-examined (Škraba Jurlina et al., 2018), for the recovery of the degraded status of the brown trout populations, stocking would be desirable. Stocking should imply a controlled, strictly defined, and planned approach, and the stocked material should be bred as a “foundation brood stock”. In addition, all of the populations that would be studied and marked as genetically “pure” and/or unique to a certain locality should be protected from the negative influence of fishing pressure. Implementing a mandatory unconditional “catch-and-release” angling regime (Simonović et al., 2018) as a compulsory legal protection measure is proposed as a possible solution that would ensure the self-sustainability of the trout stock without the need for restocking, but also the sustainability of the fishery. Many authors additionally pointed out that long-term stabilization of the density of trout populations is often not possible without environmental improvement strategies, such as

habitat restoration, removal of migration barriers, improvement of hydrological regimes, which are generally more effective than stocking (Cowx, 1994; Fjellheim et al., 2003; Oosterhout et al., 2005; Ferguson, 2007; Škraba Jurlina et al., 2018), and the brown trout habitats are certainly special in their characteristics, being very sensitive to all kinds of anthropogenic pressures.

Data availability statement

The datasets presented in this study can be found in online repositories. The names of the repository/repositories and accession number(s) can be found in the article/Supplementary Material.

Ethics statement

The animal study was approved by Ministry of Agriculture and Ministry of Environmental Protection and Energy. The study was conducted in accordance with the local legislation and institutional requirements.

Author contributions

TK: Conceptualization, Formal Analysis, Methodology, Resources, Writing—original draft. AM: Formal Analysis, Methodology, Resources, Writing—review and editing. DŠ: Writing—review and editing. PS: Resources, Writing—review and

editing. IS: Formal Analysis, Resources, Writing–review and editing. MP: Resources, Writing–review and editing. IM: Writing–review and editing.

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The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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Supplementary material

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fenvs.2024.1379878/full#supplementary-material>

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