



Ammonia Oxidation in the Ocean Can Be Inhibited by Nanomolar Concentrations of Hydrogen Peroxide

Bradley B. Tolar^{1,2*†}, Leanne C. Powers^{2†}, William L. Miller², Natalie J. Wallsgrove³, Brian N. Popp³ and James T. Hollibaugh^{1,2*}

¹ Department of Microbiology, University of Georgia, Athens, GA, USA, ² Department of Marine Sciences, University of Georgia, Athens, GA, USA, ³ Department of Geology and Geophysics, University of Hawai'i, Honolulu, HI, USA

Marine Thaumarchaeota were discovered over 20 years ago and although a few isolates from this group are now available for study, we do not yet understand the environmental controls on their growth and distribution. Thaumarchaeotes oxidize ammonia to nitrite, mediating a key step in the global nitrogen cycle, and it is estimated that about 20% of all prokaryotic cells in the ocean belong to this phylum. Despite their almost ubiquitous distribution, marine Thaumarchaeota are rarely abundant in open-ocean surface (<100 m) waters. We tested the hypothesis that this vertical distribution is driven by reactive oxygen species (ROS), specifically H_2O_2 , generated by photochemical and biological processes -- "indirect photoinhibition" rather than light inhibition as previously postulated for ammonia-oxidizing Archaea. Here we show that H₂O₂ can be surprisingly toxic to Thaumarchaeota from the Southern Ocean, with ammonia oxidation inhibited by additions of as little as 10 nM H₂O₂, while temperate Thaumarchaeota ecotypes were more tolerant. This sensitivity could explain the seasonal disappearance of Thaumarchaeota from polar surface waters and the increase in ammonia oxidation rates with depth commonly observed in marine environments. Our results highlight the need for further physiological studies of Thaumarchaeota, and indicate that ROS sensitivity could be used as a characteristic for dividing the group into meaningful ecotypes.

Keywords: Thaumarchaeota, ammonia oxidation, nitrification, nitrogen, ammonia, reactive oxygen species, hydrogen peroxide

INTRODUCTION

Photoinhibition has been hypothesized to reduce ammonia oxidation (AO), the first step of nitrification, in both ammonia-oxidizing Bacteria (AOB; Hooper and Terry, 1973; Olson, 1981; Ward, 1985; Horrigan and Springer, 1990) and Archaea (AOA; Murray et al., 1998, 1999a; Church et al., 2003; Qin et al., 2014). Nitrification links key processes within the nitrogen cycle (e.g., nutrient regeneration, assimilation, respiration) by converting the most reduced form of inorganic N (ammonium) into the most oxidized (nitrate), which can be denitrified and removed from the pool of fixed N (Ward, 2011). Early measurements of nitrification revealed reduced or undetectable rates in surface waters (Alleman et al., 1987), which were attributed to light sensitivity of AOB. AOA are now generally regarded as the dominant ammonia-oxidizing organisms (AOO) in the

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*Correspondence:

Bradley B. Tolar bradley.tolar@gmail.com James T. Hollibaugh aquadoc@uga.edu

[†]Present Address:

Bradley B. Tolar, Department of Earth System Science, Stanford University, Stanford, CA, USA; Leanne C. Powers, Chesapeake Biological Laboratory, University of Maryland Center for Environmental Science, Solomons, MD, USA

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ocean (e.g., Wuchter et al., 2006; Francis et al., 2007; Mincer et al., 2007; Prosser and Nicol, 2008; Santoro et al., 2010; Ward, 2011; Tolar et al., 2013) and belong to the Marine Group I Archaea, now classified as Thaumarchaeota (Spang et al., 2010). However, Thaumarchaeota do not typically achieve high abundances in the surface ocean (Karner et al., 2001; Mincer et al., 2007), with notable exceptions during winter at higher latitudes and in polar regions (Murray et al., 1998, 1999a; Church et al., 2003; Wuchter et al., 2006). Attempts to explain these patterns have focused primarily on competition with Bacteria or phytoplankton and on light inhibition (Murray et al., 1999a; Church et al., 2003; Wells and Deming, 2003; Merbt et al., 2012; Smith et al., 2014b). However, most of these theories are derived from negative correlations between environmental variables and abundance. In the Southern Ocean, Thaumarchaeota populations decrease dramatically-especially in surface waters (Church et al., 2003)—as irradiance increases during the transition from winter to summer (Murray et al., 1998). This could indicate that photoinhibition or a product of increased irradiance (such as reactive oxygen species, or ROS) affects Thaumarchaeota and, perhaps, nitrification in general, as AOB have also been shown to be sensitive to light (Hooper and Terry, 1973; Olson, 1981; Ward, 1985; Horrigan and Springer, 1990).

The photochemical production of ROS [e.g., H₂O₂, superoxide, singlet oxygen, and hydroxyl radical; (Imlay, 2008)] proceeds at significant rates in sunlit surface waters (Figures 1A-D), creating a daily, ubiquitous increase in ROS concentration of 1-2 orders of magnitude. Of these ROS species, H₂O₂ has the longest half-life (Kieber et al., 2003, and references therein; Powers and Miller, 2014). H₂O₂ is primarily introduced by photochemical reactions with colored dissolved organic matter (CDOM; Figure 1B; Cooper et al., 1994; Blough and Zepp, 1995); rates of this production vary with solar irradiance, temperature, and [CDOM], but generally range from 9 to 134 nM h⁻¹ in marine systems (reviewed in Kieber et al., 2003). Superoxide (O_2^-) also reacts with trace metals (Figure 1B) to produce H2O2; reactions with Cu(I) can proceed rapidly with rate constants (k) on the order of 2×10^9 M⁻¹ s⁻¹ (Zafiriou et al., 1998), although [Cu(I)] is relatively low (~0.1 nM; Moffett and Zika, 1988). Other sources include atmospheric input through fluxes at the air-sea interface (0.09-4.0 nM h⁻¹; Thompson and Zafiriou, 1983) and rain (Figure 1C), which can rapidly increase surface water $[H_2O_2]$ due to its high concentration (8.4-82 µM; Zika et al., 1982; Cooper et al., 1987) relative to seawater ($<10 \text{ nM}-4.5 \mu \text{M};$ Kieber et al., 2003). There are also biological sources of ROS; phytoplankton produce H2O2 as part of their metabolism during photosynthesis and respiration (Figure 1A; Palenik et al., 1987; Zepp et al., 1987; Palenik and Morel, 1988) up to 0.46 nmol h^{-1} cell⁻¹ (Twiner and Trick, 2000), though some pathways are not directly linked to photosynthesis. Extracellular biological ROS production $(O_2^- \text{ and } H_2O_2,$ specifically; Figure 1D) in the absence of light has also been observed, with rates 0.01-11 nM h⁻¹ (Palenik and Morel, 1988; Moffett and Zafiriou, 1990; Roe et al., 2016) that have been attributed to microorganisms such as heterotrophic bacteria and phytoplankton (Palenik et al., 1987; Rose et al., 2008; Diaz et al., 2013).

Since ROS can also be formed internally as by-products of biochemical reactions including photosynthesis and respiration (e.g., Marshall et al., 2005; Diaz et al., 2013), they must be removed to prevent cell damage. Prokaryotes possess a number of mechanisms to protect against ROS damage, including detoxifying enzymes, and DNA repair systems (Imlay, 2008; Figure 1E). A survey of the Integrated Microbial Genomes (IMG) database (http://img.jgi.doe.gov/) indicated a potential deficiency of genes from ROS-protection pathways in Thaumarchaeota relative to Bacteria or Euryarchaeota (Supplementary Table 1). All Thaumarchaeota genomes sequenced to date possess at least one copy of superoxide dismutase, which converts superoxide into H₂O₂; however, only Nitrososphaera gargensis (Hatzenpichler et al., 2008) and N. evergladensis (Zhalnina et al., 2014) have an annotated catalase (Supplementary Table 1)-the most effective enzyme known for detoxifying H₂O₂ (Imlay, 2008). These particular catalases have manganese cofactors instead of the more typical heme cofactor, and are less efficient (Mishra and Imlay, 2012). A putative catalase gene was also identified from two Thaumarchaeota single-cell amplified genomes (Luo et al., 2014); a homolog of deoxyribodipyrimidine photolyase, a key enzyme involved in the repair of UV-damaged DNA, was also found in one genome in the same study. Thaumarchaeota have genes annotated as peroxiredoxin; however, peroxiredoxin is less efficient than catalase (Parsonage et al., 2008) and has been shown to remove H_2O_2 effectively only at low intracellular $[H_2O_2]$ (<0.1 μ M; Imlay, 2008). The other peroxidases encoded in Thaumarchaeota genomes may or may not directly detoxify H2O2 (Mishra and Imlay, 2012). It is therefore possible that the distribution of Thaumarchaeota could be explained by the lack of ROS detoxifying enzymes rather than direct photoinhibition or "competition" with Bacteria or phytoplankton for ammonia.

Here we test the hypothesis that this "missing" ROS protection makes Thaumarchaeota highly sensitive to H₂O₂ generated external to the cell. We examined the sensitivity of AO to the ROS species H2O2 in AOO assemblages dominated by Thaumarchaeota (Supplementary Tables 2, 3) from a wide range of marine environments, including the Southern Ocean near Palmer Station (Antarctica), the Gulf of Mexico, the Gulf of Alaska, and coastal waters at Marsh Landing, Georgia, USA (See Materials and Methods for details; Figure 2). Previous work supports the notion that Thaumarchaeota are the abundant ammonia oxidizer at these sites (Hollibaugh et al., 2014; Tolar et al., 2016a,b). Wholeseawater incubations were performed at *in situ* temperatures with additions of ¹⁵N-labeled ammonium to determine rates of AO in the presence of H₂O₂ added at environmentally-relevant (nM) concentrations. We also quantified the effect of H₂O₂ additions on the abundance of transcripts of amoA, a highlytranscribed Thaumarchaeota gene that encodes the α-subunit of the ammonia monooxygenase enzyme as an index of another, more general (transcription broadly) Thaumarchaeota-specific process. Finally, we measured overall microbial activity using



introduced in surface waters (Kieber et al., 2003) via (A) phytoplankton as by-products of photosynthesis, (B) photochemical reactions of colored dissolved organic matter (CDOM) and metals, (C) atmosphere exchange and rain, and (D) production by heterotrophic bacteria (beige). Inset (E) gives a cellular-level view of ROS toxicity: when prokaryotic cells come into contact with ROS, molecules such as DNA, protein, and lipids can become damaged unless enzymes like superoxide dismutase (SOD) or catalase (Cat) are produced to detoxify ROS (Imlay, 2008; detoxification indicated by a red X). As shown in the text, ammonia oxidation by marine Thaumarchaeota (AOA) can be sensitive to ROS.

leucine incorporation to determine the effects of $\rm H_2O_2$ on general cellular function (protein synthesis) for comparison.

MATERIALS AND METHODS

Sample Collection

Seawater samples were collected (**Figure 2**, Supplementary Table 2) from Niskin bottles attached to a CTD rosette along the Palmer LTER 600 line in the Southern Ocean (*ARSV Laurence M. Gould*, September 2010; Station 600.160—64° 12.657′ S, 66° 15.600′ W; and Station 600.080—64° 41.682′ S, 65° 01.849′ W), in the Gulf of Mexico (GoM—*R/V Pelican*, May 2012; Station C4—27° 55.169′ N, 90° 22.072′ W; Station B4—28° 26.551′ N, 89° 45.577′ W; Station A6—28° 39.744′ N, 88° 00.443′ W), in the Gulf of Alaska (GoA—*R/V Melville*, August 2013; Station 004—49° 34.194′ N, 138° 40.188′ W; Station 015—58° 59.968′ N, 140° 59.956′ W; Station 033—49° 57.343′ N, 132° 40.094′ W), and from coastal surface waters from a floating dock at Marsh Landing at the mouth of the Duplin River (a tidal creek draining 1200 ha of salt marsh) adjacent to Sapelo Island, Georgia,

USA (Marsh Landing-August 2011 and 2012, September 2013; 31° 25.068' N, 81° 17.721' W). Two water masses (Antarctic Surface Water at 10-75 m and Circumpolar Deep Water at 300 m) were sampled in the Southern Ocean for comparison as these have shown different Thaumarchaeota populations in a previous study (Kalanetra et al., 2009). For GoM and GoA samples, water was collected at 200 m as this depth was likely to have high abundance of Thaumarchaeota relative to other organisms, except for GoA Stn 015 (50 m), which was from a more shallow, coastal station (max 180 m depth). Additional Marsh Landing samples were collected over a 24 h period at high tide and low tide (Supplementary Table 3) from Marsh Landing (August 2012, September 2013) and from a dock at Hunt Camp (August 2012; 31° 28.012' N, 81° 16.006' W), on the Duplin River upstream of Marsh Landing. We collected 0.22 µm filtered sample water for nutrient analysis, including the measurement of ammonia concentration (Supplementary Tables 2, 3) using previously described methods (Holmes et al., 1999). Photosynthetically-active radiation (PAR) data was gathered to pair to samples collected over 24 h at Marsh Landing (2012,



2013) from the NOAA National Estuarine Research Reserve System (NERRS) meteorological station on Sapelo Island near the floating dock we sampled at Marsh Landing (data accessed from the NOAA NERRS Centralized Data Management Office website: http://www.nerrsdata.org/; accessed 12 Oct 2014).

Experimental Setup

Seawater from each site (**Figure 2**) was collected from Niskin bottles attached to CTD rosettes (Southern Ocean, GoM, GoA) or directly from surface water (Marsh Landing) into 4 L (Southern Ocean) or 10 L (GoM, GoA, Marsh Landing) carboys and kept at *in situ* temperature (on-board Percival incubator—Southern Ocean; seawater flow-through tank—GoM, GoA, Marsh Landing) in the dark. ¹⁵N-labeled ammonium (¹⁵NH₄Cl; Cambridge Isotope Laboratories) was added at a concentration of 50 nM (Santoro et al., 2010; Beman et al., 2012) to determine ammonia oxidation (AO) rates, and hydrogen peroxide (H₂O₂; Fisher Scientific—Southern Ocean; J. T. Baker—GoM, GoA, Marsh Landing) was added to experimental treatments at

target concentrations designated as either "low" (10-100 nM) or "high" (30-300 nM) amendments; actual H2O2 additions varied by sample site (Table 1, Supplementary Table 4). For these addition experiments, H₂O₂ and ¹⁵N-labeled ammonium were added directly to carboys for incubation; however, in the Southern Ocean ¹⁵NH₄Cl was added to duplicate, 250 mL subsamples after an initial incubation (6 h) with H₂O₂. H₂O₂ concentrations in stock solutions were determined by measuring absorbance at 240 nm using the measured molar absorptivity of 38.1 M⁻¹ cm⁻¹ (Miller and Kester, 1988). Initial $[H_2O_2]$ were also estimated by modeling H_2O_2 decay (Figure 3, Supplementary Table 4) when possible. We did not add H_2O_2 to samples collected for diurnal studies at Marsh Landing and Hunt Camp (2012, 2013), as we were more interested in the impact on the in situ [H₂O₂] on AO rates and the variability over a day and tidal cycle. We also took advantage of a rainstorm during sample collection (Marsh Landing 2012, 2013) to gather rainwater (collected over 15 min in a square plastic container) for use in an incubation experiment to compare the effects of H₂O₂ from chemical vs. natural sources.

TABLE 1 | Summary of ROS incubation experiments.

Sample	Depth (m)	in situ [H ₂ O ₂]	Nominal [H ₂ O ₂] Addn.	Modeled [#] [H ₂ O ₂] Addn.	AO Rate (% control)	p-value (t-test
Southern Ocean 600.160	10	ND	30 nM	ND	0.61	<0.001
			100 nM	ND	-0.13	<0.001
	55	ND	30 nM	ND	-0.91	<0.001
			100 nM	ND	-0.28	<0.001
	300	ND	30 nM	ND	3.4	<0.001
			100 nM	ND	0.85	<0.001
Southern Ocean 600.080	10	14.3	30 nM	ND	-8.5	0.002
			100 nM	ND	-1.7	0.007
	75	2.9	30 nM	ND	32	0.13
			100 nM	ND	-12	0.007
	300	7.0	30 nM	ND	8.4	0.001
			100 nM	ND	-50	<0.001
GoM C4	200	16.3	100 nM	163	76	0.38
			300 nM	464	76	0.12
GoM B4	200	28.1	100 nM	234	75	0.17
			300 nM	731	34	0.037
GoM A6	200	14.6	100 nM	197	74	0.26
			300 nM	556	34	0.032
GoA Stn 004	200	6.9	100 nM	112	99	0.93
			300 nM	312	68	0.37
GoA Stn 010	200	8.8	30 nM	73.4	170	0.089
			100 nM	168	103	0.87
			300 nM	419	40	0.036
GoA Stn 015	50	15.6	30 nM	52.9	93	0.26
			100 nM	160	94	0.37
			300 nM	433	94	0.41
GoA Stn 033	200	14.3	30 nM	69.2	63	0.034
Marsh Landing 2011	<1	ND	100 nM	ND	150	0.091
			300 nM	ND	33	0.023
Marsh Landing 2012	<1	201	100 nM	176	91	0.69
			300 nM	248	63	0.24
			100 nM Rain	221	80	0.44
Marsh Landing 2013	<1	108	10 nM	61.1	99	0.90
			30 nM	80.1	96	0.54
			100 nM	138	96	0.36
			300 nM	297	91	0.32
			100 nM Rain	167	86	0.16

All $[H_2O_2]$ in nM; t-test against control rates, p-values in bold are significant (p < 0.05). ND, not determined. Rates are reflected as a percentage of the control; negative rates (italicized) reflect δ^{15} N values that were higher at T = 0 than at the end of incubation (these rates are essentially zero or the limit of detection, defined as 0.001 nM d^{-1}). t-tests were re-calculated using the detection limit instead of the negative rate, which had no effect on the significance of these results. #Modeled $[H_2O_2]$ (nM) at time = 0 h for each experiment determined from y-int of the regression as discussed in Materials and Methods.

Rainwater was filtered prior to $[H_2O_2]$ measurement and addition to prevent decay, and diluted to target nM additions. For comparison to general microbial community activity, leucine

incorporation rates were measured in samples from all sites using 13 C- or 3 H-labeled leucine (See Supplementary Methods for details).

Nucleic Acid Extraction and Quantification

After incubation for 6 h with H₂O₂, samples were filtered directly onto 47 mm (Southern Ocean) or 142 mm (GoM, GoA, Marsh Landing) 0.22 µm pore size GVWP filters (Millipore), which were frozen in RNAlater (Ambion) at -20°C (GoM, GoA) or -80°C (Southern Ocean, Marsh Landing) before transport back to the laboratory, where they were kept frozen at -80° C until extraction. RNA was extracted following previously described methods (Poretsky et al., 2008; Gifford et al., 2011) with the replacement of 200 µm zirconium beads (OPS Diagnostics) for the initial bead-beating step. DNA was removed after extraction with the TURBO DNase-Free Kit (Ambion) following manufacturer's instructions with an additional enzyme treatment at 2X concentration. RT-qPCR targeting amoA transcripts was performed using the One-Step RT-qPCR SYBR Mix (BioRad), Archaeal amoA primers (Wuchter et al., 2006), and 1-2 µL of RNA template. AOB amoA transcripts were not detected in any of our samples (data not shown). qPCR reaction conditions follow previous methods (Kalanetra et al., 2009) with the addition of a 15 min incubation step at 50°C for reverse transcription of RNA to cDNA prior to the initial denaturation at 95°C. Raw transcript abundance (copies μL^{-1} of RNA extract) was converted to transcripts L^{-1} using the volume filtered and an elution volume of 100 µL, assuming 100% extraction efficiency. DNA extraction with phenol:chloroform, quantification of Archaeal, and Bacterial genes using qPCR, and calculation of gene copies L^{-1} were performed as previously described (Tolar et al., 2013). Further characterization of the diversity of Thaumarchaeota communities from this study was done using next-generation sequencing (See Supplementary Methods for details).

Ammonia Oxidation Rate Measurements

Ammonia oxidation (AO) rates were measured during dark incubations at in situ temperature for 24 h (GoM, Marsh Landing), 48 h (GoA), or 96 h (Southern Ocean) with 50 nM 15 NH₄Cl added (mean = 12%, range = 1.7-45% of total [NH₄]; Supplementary Table 2). We did not find significant effects on longer incubation time for AO rate measurements made in the Southern Ocean at the same time (Tolar et al., 2016a). For each rate measurement, an initial sample was filtered and frozen immediately after addition of tracer; for samples from the Southern Ocean, parallel control incubations were performed using the addition of tracer to filtered sample water. Incubations were terminated by freezing samples at -20 or $-80^{\circ}C$ and samples remained frozen at -80° C until analysis. Samples were analyzed using the "denitrifier" method (Sigman et al., 2001) as described previously (Popp et al., 1995; Dore et al., 1998). Briefly, NO2 and NO3 present in each sample were converted to N2O gas through denitrification by Pseudomonas aureofaciens, transferred from the reaction vial, cryofocused, and separated from other gases using a CP-PoraBOND Q capillary column (0.32 mm inner diameter $\times 25 \text{ m} \times 5 \text{ }\mu\text{m}$) at 20–25°C. δ^{15} N values of N₂O were measured using a MAT-252 mass spectrometer (Finnigan) and used to determine how much ¹⁵NH₃ tracer was introduced to the NO_x pool through AO (Beman et al., 2008; Christman et al., 2011). Catalase (Sigma Aldrich) was added to thawed samples (10 μ L of 150 units L⁻¹ catalase per 50 mL sample) that were incubated at room temperature for 60 min to prevent any inhibition of *P. aureofaciens* growth or denitrification ability by residual H₂O₂ (no difference observed in N₂O production with catalase addition; *t*-test, p = 0.43; Supplementary Table 5). AO rates were calculated from δ^{15} N values as previously described (Christman et al., 2011; Beman et al., 2012); the detection limit for this measurement (0.001 nM d⁻¹) was determined by using the detection limit for [NH₄] and [NO_x], as well as the uncertainty in stable isotope measurements with this setup (Beman et al., 2011).

Hydrogen Peroxide Concentrations

H₂O₂ was measured in the Southern Ocean using the POHPAA method (Miller and Kester, 1988) on samples that were first derivatized, then frozen at -20 to -80°C for subsequent measurement on an Aqualog fluorometer (Horiba). Since H₂O₂ concentrations were below the limit of detection (5 nM) in some of these samples (Supplementary Table 6), we switched to a more sensitive method (King et al., 2007) using a FeLume chemiluminescence system (Waterville Analytical; Supplementary Figure 1) for the remaining experiments (GoM, Marsh Landing, GoA). Full details of the setup for the FeLume system are available in the Supplementary Methods. Briefly, H₂O₂ reacts with 10-methyl-0-(p-formylphenyl)-acridinium carboxylate trifluoromethanesulfonate (acridinium ester; provided by James Kiddle, University of Western Michigan) at alkaline pH, forming a chemiluminescent product that can be quantified using flow injection analysis (Powers et al., 2015); See Supplementary Methods for more details. A 2 mM H₂O₂ stock solution was prepared from 30% H₂O₂ (J. T. Baker) in ultrapure water and checked spectrophotometrically as above. Standards were prepared in aged 0.2 µm filtered seawater and checked for low H₂O₂ with H₂O₂ blanks of aged seawater with added catalase (20 μ L of 100 units L⁻¹ catalase to 20 mL seawater). Under these conditions, the H2O2 detection limit, defined as three times the standard deviation of the blank, was 2-5 nM.

While the overall decay of $[H_2O_2]$ has been reported as second order (Zepp et al., 1987), our experiments generally followed first-order kinetics with respect to hydrogen peroxide concentration, using the calculations from Zepp et al. (1987)

$$\frac{\mathrm{d}\left[\mathrm{H}_{2}\mathrm{O}_{2}\right]}{\mathrm{d}t} = -k_{obs}[\mathrm{H}_{2}\mathrm{O}_{2}] \tag{1}$$

where k_{obs} is the observed rate constant (h⁻¹). By integrating Equation (1), H₂O₂ decay can be modeled with the following equation:

$$[H_2O_2] = [H_2O_2]_0 e^{-k_{obs}t}$$
(2)

where $[H_2O_2]_0$ is the initial peroxide concentration. For each experiment, H_2O_2 data were fit to Equation (2) with a nonlinear regression of H_2O_2 vs. time using the curve fitting toolbox in MATLAB[®] (MathWorks) and both $[H_2O_2]_0$ and k_{obs} were used as fitting parameters. Decay rates for GoM samples were determined using H_2O_2 concentration data from the first 24 h of incubation; GoA decay rates from Stations 004 and 015 were determined over 76 h, Station 010 over 147 h, and Station 033 over 24 h (note that all AO rate incubation samples were



(green), 100 nM H_2O_2 (orange), 300 nM H_2O_2 (blue), and rain (gray) incubations. Symbols represent measured H_2O_2 concentrations, while curves represent modeled decay rates; all panels contain error bars representing standard deviation from the mean (note that in some cases error bars are too small to be seen outside of the filled circle). Variability was greater in Marsh Landing 2012 **(G)** experiments due to particulate matter in sample water interfering with H_2O_2 measurement.

terminated at 48 h by subsampling the incubation bottle; the remaining sample was incubated for longer periods as necessary to detect complete decay of H_2O_2); and Marsh Landing 2012 decay rates were determined over 3.5 h, while Marsh Landing 2013 decay rates were determined over 2 h. Half-lives varied for each environment, ranging from \sim 3 to 8 h (Marsh Landing) to over 120 h (GoA; Supplementary Table 4).

Statistical Analysis

Statistical tests were performed with R (http://www.r-project.org/) using the commands "aov" for one-way analysis of variance (ANOVA), "TukeyHSD" for the Tukey's honestly significant difference (HSD) test, "*t*-test" for Student's *t*-test, and "cor.test" for calculations of Pearson's

correlation coefficient (all from the stats package included with R). Stripcharts were used to assess normal distribution of data and a Bartlett test was run to evaluate equality of variance (both within R). Rate data was log-transformed (after scaling to remove negative values) to prevent bias from higher rates obtained in some samples. Percent reduction of rates (AO and leucine incorporation) was calculated by dividing a given rate by the average rate from a non- H_2O_2 addition control.

RESULTS

A summary of results from all experiments is available in **Table 1**. AO rates were below detection in the Southern Ocean

following the addition of 6-80 nM H₂O₂ to samples above in situ concentrations (Figures 4A,B; Table 1, Supplementary Table 6), indicating that AOO in this region are extremely sensitive to H₂O₂. Concentrations of H₂O₂ in Southern Ocean samples were low (2.9-14 nM; Supplementary Table 6) and match previous measurements (Resing et al., 1993; Yocis, 1995; Sarthou et al., 1997). Despite this relatively small increase in [H₂O₂], AO rates decreased 67 to >100% of the control rate (Table 1; negative rates indicate samples where the δ^{15} N value was higher at T = 0, and are considered zero or the limit of detection). The change in abundance of AOA amoA transcripts did not reflect the drastic decrease in AO rates observed in treatments with H₂O₂ additions (this was observed at all sites sampled; Figure 5). Southern Ocean leucine incorporation rates at Station 600.160 were also reduced with increasing H_2O_2 concentration (Figure 6A), but this reduction was only significant at 10 m (100 nM addition only) and 55 m (both 30 and 100 nM additions; Supplementary Table 2). The reduction in leucine incorporation (mean = 17%) was significantly less than that of AO rates (mean = 67%; *t*-test, p < 0.01).

The AO response to H₂O₂ additions was variable in both the Gulf of Mexico (GoM; Figure 4C, Table 1; average 38% AO reduction) and the Gulf of Alaska (GoA; Figure 4D, Table 1; average 8.5% AO reduction), with the greatest inhibition observed at Station B4 (GoM) and Station 033 (GoA). At the furthest offshore GoM site (Station C4), there was no difference in AO between samples with 0 or 100 nM added H_2O_2 , and no change was observed with amendments of up to 300 nM H₂O₂ in the coastal GoA sample (Station 015, 50 m). An analysis of Thaumarchaeota 16S rRNA and amoA genes (Supplementary Figure 2) showed that the clades found in the GoA sample most resilient against H_2O_2 additions (Station 015) were the least similar to other GoA communities. Inhibition of leucine incorporation (protein synthesis) also varied with sample location; for example, we found no significant difference with increased [H₂O₂] at Stations C4 or A6 (GoM; Figure 6B), whereas AO rates were reduced at both sites (24-66%; Figure 4C, Table 1).

Samples from Marsh Landing were the least inhibited by H_2O_2 with only one out of ten measurements showing a significant reduction in AO across all 3 years sampled (Figures 4E,F, Supplementary Figure 3): the addition of 300 nM H₂O₂ in 2011 (67% reduction of AO rate; Table 1; ttest, p < 0.05). These coastal microbial communities were capable of reducing H_2O_2 concentrations ~6X more quickly (Figures 3G,H) than communities from open ocean sites (Supplementary Table 4, Figures 3A-F), and were markedly distinct in both Thaumarchaeota 16S rRNA and amoA diversity (Supplementary Figure 2). Time courses (2011; Supplementary Figure 3) revealed that $300 \text{ nM} \text{ H}_2\text{O}_2$ additions significantly reduced AO within 6 h (*t*-test, p < 0.03; Supplementary Table 2), with no additional reduction or recovery of AO rates after 24 h (p < 0.05; Supplementary Table 2). As with other experiments, Marsh Landing amoA transcript abundance was not reduced by H_2O_2 (Figures 5E,F). Leucine incorporation in these samples was also significantly reduced by H2O2 additions after 6 h (ttest, p < 0.03; Figures 6D-F, Supplementary Table 2), but the reduction was not significantly different from the reduction in AO rates (mean = 18 vs. 5% reduction; *t*-test, $p \ge 0.2$). We also found that additions of rainwater inhibited AO rates (**Figures 4E,F**), and a rain event during the 24 h sampling study in 2012 further suppressed rates at one site, despite the absence of PAR and thus photochemical H₂O₂ production (Supplementary Figure 4B). Rainwater H₂O₂ concentrations were 30–50 μ M (diluted to target nM concentrations for experiments; Supplementary Table 3), consistent with previous measurements (e.g., Zika et al., 1982; Cooper et al., 1987; Willey et al., 1996).

Additional samples collected from Marsh Landing and Hunt Camp over 24h (Figure 7, Supplementary Figure 4) revealed a daily increase in [H2O2] from 10 to 500 nM [Supplementary Table 3; corresponding to increase in sunlight or "photosynthetically active radiation," (PAR)]-1 to 2 orders of magnitude greater than what was detected in the Southern Ocean. Samples collected at and immediately after the time of greatest [H₂O₂] showed decreased AO rates (Figure 7, Supplementary Figure 4), although there was higher variability due to impact of mixing and other processes inherent to an estuarine tidal creek. There was also no clear pattern based on tide sampled (Supplementary Table 3), with no significant difference in rates measured in August 2012 at Marsh Landing except between the two low tides sampled (t-test, p = 0.043; Supplementary Figure 4A), equal rates at high and low tide at Hunt Camp (p = 0.21; Supplementary Figure 4B), and the highest rate in September 2013 at the high tide sampled during the night (p < 0.005 for all comparisons; Figure 7). There was no discernable trend based on time of day or [H₂O₂]. Thaumarchaeota amoA genes did not vary for experiments in 2012 (Supplementary Figure 4), but amoA transcripts peaked at dawn at Marsh Landing (Supplementary Figure 4A) and showed no trend at Hunt Camp (Supplementary Figure 4B). For the Marsh Landing 2013 diurnal study, there was greater variation across the 24-h period than in 2012; amoA gene abundance peaked after noon and midnight, while *amoA* transcripts were most abundant right after sunset and before sunrise (Figure 7).

Our data show that nanomolar concentrations of H_2O_2 can be harmful to open-ocean AOO, with the most sensitive communities found in polar waters (Figure 4, Table 1). In contrast, AOO in coastal samples from Marsh Landing appear to be more resistant to H_2O_2 toxicity, as do AOO in samples from the coastal GoA. Although the results were variable apart from the Southern Ocean experiments, a significant reduction of AO rates following H₂O₂ amendments was observed when the entire dataset was evaluated together (ANOVA, n = 140, F = 11.2, p < 0.001). The same analysis revealed no significant difference in leucine incorporation rates (all sites: n = 169, F = 2.24, p = 0.09; excluding Marsh Landing: n = 128, F = 2.6, p = 0.06), suggesting that AO is more sensitive to H2O2 toxicity than general microbial metabolism. Comparing the percent reduction in AO rates based on site (ANOVA, n = 140, F = 18.7, p < 0.001) revealed a significant difference between samples from the Southern Ocean $(100 \pm 19\%)$ and all other sites [Marsh Landing $(12 \pm 30\%)$, GoM (38 \pm 21%), GoA (8.5 \pm 36%); Tukey's test, p < 0.001for all three pairs). This indicates that our hypothesis that ROS



600.080; symbols represent sample depth), Gulf of Mexico and Gulf of Alaska (**C**,**D**, respectively; symbols represent stations sampled; sample depth was 200 m except at Station 015, which was sampled at 50 m), and coastal waters from Marsh Landing in (**E**) 2012 and (**F**) 2013 (symbols represent H₂O₂ origin, either chemical H₂O₂ or from rainwater; surface water samples collected). Error bars represent standard deviation from the mean of replicate AO measurements [n = 4(**A**,**B**), 2 (**C**-**F**)]. An * to the right of a point indicates a significant difference (*t*-test, p < 0.05) from the control rate (for **A**,**B**, overlapping points were all significantly different, and are represented by only one *). H₂O₂ concentrations measured for each individual sample; non-control sample [H₂O₂], color-coded to match legend for each panel (note that regressions were not done for Station 600.160 (**A**) as [H₂O₂] was not measured).

can inhibit AO is valid only in certain ocean regions, which warrants further investigation, as site-specific environmental factors or overall microbial community composition may play a role.

DISCUSSION

Culture-based experiments testing light inhibition have indicated that AO by both AOB (Hooper and Terry, 1973; Horrigan



and Springer, 1990; Hyman and Arp, 1992; Guerrero and Jones, 1996a) and AOA (Merbt et al., 2012; Qin et al., 2014) is decreased in the presence of increased light; yet none of these studies measured ROS generation during their incubations, which can occur at rates of 10–600 nM $\rm H_2O_2~h^{-1}$ under simulated light conditions (Kieber et al., 2003, 2014; Powers and Miller, 2015). However, Hooper and Terry (1974) postulated

that the mechanism of this damage was likely due to ROS ("a peroxide or free radical"), based on their observation that reduced photoinhibition was observed when oxygen levels were low or absent in their experiments. Additionally, a very recently published study showed that H_2O_2 concentrations \geq 200 nM completely inhibited AO by a newly isolated marine thaumarchaeote, strain DDS1 (Kim et al., 2016).



We have shown here that nanomolar $[H_2O_2]$ can inhibit AO in bacterioplankton where Thaumarchaeota are the dominant AOO, and can also inhibit protein synthesis by the overall microbial community, which includes other nitrogen-cycling organisms. H_2O_2 is produced by abiotic and biotic processes in all sunlit marine waters where it can enter and damage cells that do not have detoxifying enzymes (**Figure 1**). Surface water $[\rm H_2O_2]$ ranges from 8 nM to 4.5 μM in marine environments (reviewed in Kieber et al., 2003), and varies from <10 to 250 nM in open ocean waters (e.g., Yuan and Shiller, 2005). Generally, lower H_2O_2 concentrations (10–20 nM) occur in areas with both low solar irradiance and CDOM like the Southern Ocean (Resing et al., 1993; Yocis, 1995; Sarthou et al., 1997), while higher $[\rm H_2O_2]$ occurs in coastal regions with both increased irradiance and



CDOM, such as in the Amazon River plume (up to 300 nM; Yuan and Shiller, 2001), Orinoco River plume (140-470 nM; Moore et al., 1993), or coastal Gulf of Mexico (90-240 nM; Zika et al., 1985; Cooper et al., 1987). Our study examined Thaumarchaeota populations from both coastal and offshore marine waters that covered this range of natural H₂O₂ concentrations. Our observation of increased sensitivity in Southern Ocean samples could reflect microorganisms that can survive without ROS detoxification systems in the typically low $[H_2O_2]$ present there (2.1-33 nM; Resing et al., 1993; Yocis, 1995; Sarthou et al., 1997). Although H₂O₂ is only one type of ROS, it can both be produced from superoxide and form the hydroxyl radical via reactions with metals, and also is the longest-lived of the ROS species (Kieber et al., 2003, and references therein). Therefore, focusing our work on H₂O₂ can inform the potential impacts of ROS in general on these communities.

Recent work has shown that Thaumarchaeota are active and oxidize ammonia in sunlit surface waters (Smith et al., 2014a), and culture-based Thaumarchaeota photoinhibition experiments indicated that the specific intensity of light that was inhibitory depended on the strain used (Merbt et al., 2012; Qin et al., 2014). This was also true for AOB, with different strains showing different light intensities needed to inhibit growth (Horrigan and Springer, 1990; Guerrero and Jones, 1996a). Additionally, the wavelength of light used also showed a variable inhibitory effect, with UV wavelengths (<400 nm) the most damaging to AO by AOB cultures (Hooper and Terry, 1974; Guerrero and Jones, 1996a). Recovery after photoinhibition was also found to be wavelength-dependent, as AOB exposed to longer wavelengths (>400 nm) regained activity faster than those

exposed to shorter wavelengths or sunlight (Guerrero and Jones, 1996b). Strong absorption of UV radiation (280–400 nm) leads to the production of H_2O_2 (Miller, 2000; Yocis et al., 2000), with only 10–20% of H_2O_2 produced by wavelengths >400 nm (Miller, 2000). Therefore, observed wavelength-dependent biological photosensitivity may simply reflect higher [H_2O_2] production.

In one AOA photoinhibition study, Thaumarchaeota strains did not recover growth in the dark following light exposure (Merbt et al., 2012); vet in another, Qin et al. (2014) observed recovery in two strains, one of which was included in the Merbt et al. (2012) experiments ("Candidatus Nitrosopumilus maritimus" SCM1). Additionally, while Merbt et al. (2012) saw 82% of growth inhibition at 15 μ E m⁻² s⁻¹ for strain SCM1, a 12-fold increase in light intensity (180 μ E m⁻² s⁻¹) was required to match this inhibition in experiments performed by Qin et al. (2014). Although the authors had no explanation for this discrepancy, a recent study by Kim et al. (2016) has provided a potential answer. Thaumarchaeota cultures have shown enhanced growth with additions of α-ketoglutarate (Tourna et al., 2011; Qin et al., 2014), which was determined to scavenge endogenously-produced H₂O₂ that otherwise would inhibit growth (Kim et al., 2016), much like the inhibition of AO shown here. The study by Qin et al. (2014) is the first to report the supplement of α -ketoglutarate to media for growth of this particular strain, and its addition could explain their photoinhibition experimental results by protecting SCM1 against H₂O₂. Interestingly, the two other thaumarchaeote strains examined Qin et al. (2014) were unable to grow without addition of α -ketoglutarate, indicating enhanced sensitivity to

ROS and highlighting differences even amongst closely related isolates.

The sensitivity of ammonia oxidation by AOA to H_2O_2 could explain their typical absence from oceanic surface waters, consistent with early theories for photoinhibition of AO by AOB (Alleman et al., 1987). A mid-summer bloom of Thaumarchaeota in Georgia coastal waters (Hollibaugh et al., 2014) despite high in situ [H₂O₂] (up to 500 nM) probably reflects the lower sensitivity of this population of AOA to H₂O₂, and is likely related to the distinct "cotype" of Thaumarchaeota that resides in coastal Marsh Landing waters compared to open-ocean sites (Supplementary Figure 2). AO has been shown to exhibit less photoinhibition in an estuarine AOB isolate than in marine isolates (Horrigan and Springer, 1990), which is consistent with our findings. However, as shown here, the dynamics of Thaumarchaeota populations in a coastal estuary are more complex (Figure 7, Supplementary Figure 4), with compounding influences of sunlight, tides, physical mixing, and storm events all impacting results. Consequently, it is important to collect samples at shorter time scales to reveal how such complicating factors may individually affect the data. For example, we included incubations with rainwater-derived H₂O₂ that showed an inhibitory effect on AO matching chemical additions (Figures 4E,F) indicating that storm events could impact bacterioplankton and nitrification that occurs in coastal surface waters (Smith et al., 2014a). Although our diurnal experiments were done in the dark after initial sample collection, we were able to observe natural variation in AO rates and archaeal amoA genes and transcripts under in situ conditions that included H₂O₂ concentration. We did not find clear patterns between time of day or tide and H₂O₂ inhibition in these samples, but can now compare our results to future experiments at in situ light levels and sample shorter time scales during the transition between tides.

On a broader scale, our results were location-dependent, with the most sensitive communities found in the Southern Ocean and the most resilient taken from coastal surface waters at Marsh Landing, suggesting that sample location—and thus inherently microbial community composition—is important for predicting ROS sensitivity of AOO. We also observed inconsistencies between sample sites in both the Gulf of Mexico and the Gulf of Alaska, conceivably due to the efficiency with which the surrounding microbial community removes H_2O_2 (Morris et al., 2012), although differences in net decay rate (Supplementary Table 4, **Figures 3A–F**) do not fully explain this variability. Alternatively, this may indicate a difference in the composition of thaumarchaeote populations where some "ecotypes" could be more or less resistant to ROS stress.

The question remains as to why some Thaumarchaeota would be sensitive to H_2O_2 (especially compared to Bacteria), given that they likely evolved in an oxygenated world (Spang et al., 2010; Kelly et al., 2011). Most Thaumarchaeota reside in deep waters with very low *in situ* H_2O_2 concentrations, but some clades have been found in surface waters, primarily at higher latitudes (Murray et al., 1998, 1999a; Wuchter et al., 2006). However, we have documented an abundant and active community of Thaumarchaeota in coastal Georgia surface waters (Hollibaugh et al., 2011, 2014), where $[H_2O_2]$ is high

(10–500 nM), especially during summer (Supplementary Tables 2, 3, **Figure** 7, Supplementary Figure 4). These Thaumarchaeota are comparably less sensitive to H_2O_2 (**Figures 2E,F, Table 1**), and the overall community at Marsh Landing is distinct from open-ocean "ecotypes" (Supplementary Figure 2) and more similar to "*Ca. N. maritimus*" SCM1 (Hollibaugh et al., 2011, 2014). Thus, a more direct test of H_2O_2 sensitivity using isolates or enrichments could assist in division of Thaumarchaeota into meaningful ecotypes, and may also support the previously described phylogenetic differences between "Group A" (near-surface) and "Group B" (deep) clades (Francis et al., 2005; Beman et al., 2008, 2010; Luo et al., 2014) and their activity (Smith et al., 2014a).

We also observed a discrepancy between H₂O₂ sensitivity of AO rates vs. amoA transcripts, where rates were reduced despite unchanged transcript levels (Figure 5). mRNA has a much shorter half-life (minutes) compared to protein (hours) such as enzymes, which coordinate biochemical reactions measured as rates (reviewed in Moran et al., 2013). Despite this disconnect, some correlations have been found between rates and gene abundances (Smith et al., 2014a), but most have been weak (Beman et al., 2008; Tolar et al., 2016a,b) or non-existent (Santoro et al., 2010). We do not understand cell-level regulation of transcription or translation in these organisms well enough to predict why this disconnect occurs, although it may simply be due to biases introduced by sample collection or processing time (Feike et al., 2012; Moran et al., 2013). Here, we have minimized the effect of such biases by including non-addition controls, and so any inhibition observed is relative to the control. As amoA transcript abundance was less impacted by H2O2 additions, this may indicate that ammonia monooxygenase is not transcriptionally inhibited by ROS, but rather by damage to the enzyme itself. In Nitrosomonas europaea, direct inhibition of the ammonia monooxygenase enzyme by light has been observed and recovery only occurred if de novo protein synthesis was possible, indicating that in AOB transcription was not affected (Hyman and Arp, 1992).

Mechanisms of H2O2-induced stress include the Fenton reaction, which depends on the presence of reduced Fe and can cause oxidative damage to Thaumarchaeota by the formation of the toxic hydroxyl radical (Imlay, 2008) and inactivation of ironcontaining enzymes (Sobota and Imlay, 2011; Anjem and Imlay, 2012). This process occurs within cells; however, membranebound enzymes or transporters could also be damaged by exogenous ROS. Damage to enzymes can be reversed (or prevented) by replacement of Fe(III) with a divalent metal, such as manganese (Anjem and Imlay, 2012), and analysis of Thaumarchaeota genomes indicates widespread replacement of Fe by Cu in the active site of many enzymes, including ammonia monooxygenase and proteins in the electron transport system (Walker et al., 2010; Amin et al., 2013). Additionally, the only catalase found in Thaumarchaeota genomes has a manganese cofactor instead of the typical Fe (heme). It is therefore possible that Thaumarchaeota evolved to use alternative metal cofactors in place of Fe to reduce the burden of ROS stress, rather than as a response to low Fe bioavailability (Amin et al., 2013).

H₂O₂ Inhibits Ammonia Oxidation

We have shown here that ammonia oxidation in the open-ocean is more sensitive to exogenous H₂O₂ than in coastal environments (average 57% reduction in AO, excluding Marsh Landing; Table 1), and that distinct Thaumarchaeota communities inhabit these samples (Supplementary Figure 2). Yet, we do not have enough physiological information on individual Thaumarchaeota "ecotypes" to understand what makes them different, only correlations of diversity with depth (Francis et al., 2005), habitat (Biller et al., 2012), or ammonia concentration (Sintes et al., 2013). It is therefore possible that ROS sensitivity could be a defining trait for division of Thaumarchaeota into meaningful ecotypes, although further work with a diverse group of cultures is necessary. Mesocosmbased manipulation studies on environmental samples should provide a useful approach and crucial first step for our understanding of how AOA will respond to a changing environment.

AUTHOR CONTRIBUTIONS

BT, LP, WM, and JH generated hypotheses and planned experiments. WM and LP provided methods and equipment for H_2O_2 analysis, and BT and LP performed experiments and measured H_2O_2 . BP and NW provided laboratory support and training for ¹⁵N-ammonia oxidation measurements, and NW assisted BT with ¹⁵N-ammonia oxidation sample processing. BT and JH analyzed the data and wrote the paper.

REFERENCES

- Alleman, J. E., Keramida, V., and Panteakiser, L. (1987). Light-Induced *Nitrosomonas* inhibition. *Water Res.* 21, 499–501. doi: 10.1016/0043-1354(87)90199-0
- Amin, S. A., Moffett, J. W., Martens-Habbena, W., Jacquot, J. E., Han, Y., Devol, A., et al. (2013). Copper requirements of the ammoniaoxidizing archaeon *Nitrosopumilus maritimus* SCM1 and implications for nitrification in the marine environment. *Limnol. Oceanogr.* 58, 2037–2045. doi: 10.4319/lo.2013.58.6.2037
- Anjem, A., and Imlay, J. A. (2012). Mononuclear iron enzymes are primary targets of hydrogen peroxide stress. J. Biol. Chem. 287, 15544–15556. doi: 10.1074/jbc.M111.330365
- Beman, J. M., Chow, C. E., King, A. L., Feng, Y. Y., Fuhrman, J. A., Andersson, A., et al. (2011). Global declines in oceanic nitrification rates as a consequence of ocean acidification. *Proc. Natl. Acad. Sci. U.S.A.* 108, 208–213. doi: 10.1073/pnas.1011053108
- Beman, J. M., Popp, B. N., and Alford, S. E. (2012). Quantification of ammonia oxidation rates and ammonia-oxidizing archaea and bacteria at high resolution in the Gulf of California and eastern tropical North Pacific Ocean. *Limnol. Oceanogr.* 57, 711–726. doi: 10.4319/lo.2012.57.3.0711
- Beman, J. M., Popp, B. N., and Francis, C. A. (2008). Molecular and biogeochemical evidence for ammonia oxidation by marine *Crenarchaeota* in the Gulf of California. *ISME J.* 2, 429–441. doi: 10.1038/ismej.2007.118
- Beman, J. M., Sachdeva, R., and Fuhrman, J. A. (2010). Population ecology of nitrifying archaea and bacteria in the Southern California Bight. *Environ. Microbiol.* 12, 1282–1292. doi: 10.1111/j.1462-2920.2010.02172.x
- Biller, S. J., Mosier, A. C., Wells, G. F., and Francis, C. A. (2012). Global biodiversity of aquatic ammonia-oxidizing archaea is partitioned by habitat. *Front. Microbiol.* 3:252. doi: 10.3389/fmicb.2012.00252

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SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: http://journal.frontiersin.org/article/10.3389/fmars. 2016.00237/full#supplementary-material

- Blough, N. V., and Zepp, R. G. (1995). "Reactive oxygen species in natural waters," in Active Oxygen in Chemistry, eds C. S. Foote, J. S. Valentine, A. Greenberg, and J. F. Liebman (New York, NY: Blackie Academic), 280–333.
- Christman, G. D., Cottrell, M. T., Popp, B. N., Gier, E., and Kirchman, D. L. (2011). Abundance, Diversity, and activity of ammonia-oxidizing prokaryotes in the Coastal Arctic Ocean in summer and winter. *Appl. Environ. Microbiol.* 77, 2026–2034. doi: 10.1128/AEM.01907-10
- Church, M. J., Delong, E. F., Ducklow, H. W., Karner, M. B., Preston, C. M., and Karl, D. M. (2003). Abundance and distribution of planktonic Archaea and Bacteria in the waters west of the Antarctic Peninsula. *Limnol. Oceanogr.* 48, 1893–1902. doi: 10.4319/lo.2003.48.5.1893
- Cooper, W. J., Saltzman, E. S., and Zika, R. G. (1987). The contribution of rainwater to variability in surface ocean hydrogen peroxide. J. Geophys. Res. 92, 2970–2980. doi: 10.1029/JC092iC03p02970
- Cooper, W. J., Shao, C. W., Lean, D. R. S., Gordon, A. S., and Scully, F. E. (1994). Factors affecting the distribution of H_2O_2 in surface waters. *Environ. Chem. Lakes Reserv.* 237, 391–422. doi: 10.1021/ba-1994-0237.ch012
- Diaz, J. M., Hansel, C. M., Voelker, B. M., Mendes, C. M., Andeer, P. F., and Zhang, T. (2013). Widespread production of extracellular superoxide by heterotrophic bacteria. *Science* 340, 1223–1226. doi: 10.1126/science.1237331
- Dore, J. E., Popp, B. N., Karl, D. M., and Sansone, F. J. (1998). A large source of atmospheric nitrous oxide from subtropical North Pacific surface waters. *Nature* 396, 63–66. doi: 10.1038/23921
- Feike, J., Jürgens, K., Hollibaugh, J. T., Krüger, S., Jost, G., and Labrenz, M. (2012). Measuring unbiased metatranscriptomics in suboxic waters of the central Baltic Sea using a new *in situ* fixation system. *ISME J.* 6, 461–470. doi: 10.1038/ismej.2011.94
- Francis, C. A., Beman, J. M., and Kuypers, M. M. M. (2007). New processes and players in the nitrogen cycle: the microbial ecology of anaerobic and archaeal ammonia oxidation. *ISME J.* 1, 19–27. doi: 10.1038/ismej.2007.8

- Francis, C. A., Roberts, K. J., Beman, J. M., Santoro, A. E., and Oakley, B. B. (2005). Ubiquity and diversity of ammonia-oxidizing archaea in water columns and sediments of the ocean. *Proc. Natl. Acad. Sci. U.S.A.* 102, 14683–14688. doi: 10.1073/pnas.0506625102
- Gifford, S. M., Sharma, S., Rinta-Kanto, J. M., and Moran, M. A. (2011). Quantitative analysis of a deeply sequenced marine microbial metatranscriptome. *ISME J.* 5, 461–472. doi: 10.1038/ismej.2010.141
- Guerrero, M. A., and Jones, R. D. (1996a). Photoinhibition of marine nitrifying bacteria. I. Wavelength-dependent response. *Mar. Ecol. Prog. Ser.* 141, 183–192. doi: 10.3354/meps141183
- Guerrero, M. A., and Jones, R. D. (1996b). Photoinhibition of marine nitrifying bacteria. II. Dark recovery after monochromatic or polychromatic irradiation. *Mar. Ecol. Prog. Ser.* 141, 193–198. doi: 10.3354/meps141193
- Hatzenpichler, R., Lebedeva, E. V., Spieck, E., Stoecker, K., Richter, A., Daims, H., et al. (2008). A moderately thermophilic ammonia-oxidizing crenarchaeote from a hot spring. *Proc. Natl. Acad. Sci. U.S.A.* 105, 2134–2139. doi: 10.1073/pnas.0708857105
- Hollibaugh, J. T., Gifford, S. M., Moran, M. A., Ross, M. J., Sharma, S., and Tolar, B. B. (2014). Seasonal variation in the metratranscriptomes of a Thaumarchaeota population from SE USA coastal waters. *ISME J.* 8, 685–698. doi: 10.1038/ismej.2013.171
- Hollibaugh, J. T., Gifford, S., Sharma, S., Bano, N., and Moran, M. A. (2011). Metatranscriptomic analysis of ammonia-oxidizing organisms in an estuarine bacterioplankton assemblage. *ISME J.* 5, 866–878. doi: 10.1038/ismej.2010.172
- Holmes, R. M., Aminot, A., Kerouel, R., Hooker, B. A., and Peterson, B. J. (1999). A simple and precise method for measuring ammonium in marine and freshwater ecosystems. *Can. J. Fish. Aquat. Sci.* 56, 1801–1808. doi: 10.1139/f99-128
- Hooper, A. B., and Terry, K. R. (1973). Specific inhibitors of ammonia oxidation in *Nitrosomonas. J. Bacteriol.* 115, 480–485.
- Hooper, A. B., and Terry, K. R. (1974). Photoinactivation of ammonia oxidation in *Nitrosomonas. J. Bacteriol.* 119, 899–906.
- Horrigan, S. G., and Springer, A. L. (1990). Oceanic and estuarine ammonium oxidation: effects of light. *Limnol. Oceanogr.* 35, 479–482. doi: 10.4319/lo.1990.35.2.0479
- Hyman, M. R., and Arp, D. J. (1992). ¹⁴C₂H₂-labeling and ¹⁴CO₂-labeling studies of the *de novo* synthesis of polypeptides by *Nitrosomonas europaea* during recovery from acetylene and light inactivation of ammonia monooxygenase. *J. Biol. Chem.* 267, 1534–1545.
- Imlay, J. A. (2008). Cellular defenses against superoxide and hydrogen peroxide. Annu. Rev. Biochem. 77, 755–776. doi: 10.1146/annurev.biochem. 77.061606.161055
- Kalanetra, K. M., Bano, N., and Hollibaugh, J. T. (2009). Ammonia-oxidizing Archaea in the Arctic Ocean and Antarctic coastal waters. *Environ. Microbiol.* 11, 2434–2445. doi: 10.1111/j.1462-2920.2009.01974.x
- Karner, M. B., Delong, E. F., and Karl, D. M. (2001). Archaeal dominance in the mesopelagic zone of the Pacific Ocean. *Nature* 409, 507–510. doi: 10.1038/35054051
- Kelly, S., Wickstead, B., and Gull, K. (2011). Archaeal phylogenomics provides evidence in support of a methanogenic origin of the Archaea and a thaumarchaeal origin for the eukaryotes. *Proc. R. Soc. B. Bio* 278, 1009–1018. doi: 10.1098/rspb.2010.1427
- Kieber, D. J., Miller, G. W., Neale, P. J., and Mopper, K. (2014). Wavelength and temperature-dependent apparent quantum yields for photochemical formation of hydrogen peroxide in seawater. *Environ. Sci. Process. Impacts* 16, 777–791. doi: 10.1039/c4em00036f
- Kieber, D. J., Peake, B. M., and Scully, N. M. (2003). "Reactive oxygen species in aquatic ecosystems," in UV Effects in Aquatic Organisms and Ecosystems, eds E. W. Helbling and H. Zagarese (Cambridge: The Royal Society of Chemistry), 251–290.
- Kim, J.-G., Park, S.-J., Sinninghe Damsté, J. S., Schouten, S., Rijpstra, W. I. C., Jung, M.-Y., et al. (2016). Hydrogen peroxide detoxification is a key mechanism for growth of ammonia-oxidizing archaea. *Proc. Natl. Acad. Sci. U.S.A.* 113, 7888–7893. doi: 10.1073/pnas.1605501113
- King, D. W., Cooper, W. J., Rusak, S. A., Peake, B. M., Kiddle, J. J., O'sullivan, D. W., et al. (2007). Flow injection analysis of H_2O_2 in natural waters using acridinium ester chemiluminescence: method development and optimization using a kinetic model. *Anal. Chem.* 79, 4169–4176. doi: 10.1021/ac06 2228w

- Luo, H. W., Tolar, B. B., Swan, B. K., Zhang, C. L. L., Stepanauskas, R., Moran, M. A., et al. (2014). Single-cell genomics shedding light on marine Thaumarchaeota diversification. *ISME J.* 8, 732–736. doi: 10.1038/ismej.2013.202
- Marshall, J. A., De Salas, M., Oda, T., and Hallegraeff, G. (2005). Superoxide production by marine microalgae. *Mar. Biol.* 147, 533–540. doi: 10.1007/s00227-005-1596-7
- Merbt, S. N., Stahl, D. A., Casamayor, E. O., Martí, E., Nicol, G. W., and Prosser, J. I. (2012). Differential photoinhibition of bacterial and archaeal ammonia oxidation. *FEMS Microbiol. Lett.* 327, 41–46. doi: 10.1111/j.1574-6968.2011.02457.x
- Miller, G. W. (2000). Wavelength and Temperature Dependent Quantum Yields for Photochemical Formation of Hydrogen Peroxide in Seawater. M.S. thesis, State University of New York.
- Miller, W. L., and Kester, D. R. (1988). Hydrogen-peroxide measurement in seawater by (Para-Hydroxyphenyl) acetic acid dimerization. Anal. Chem. 60, 2711–2715. doi: 10.1021/ac00175a014
- Mincer, T. J., Church, M. J., Taylor, L. T., Preston, C., Karl, D. M., and Delong, E. F. (2007). Quantitative distribution of presumptive archaeal and bacterial nitrifiers in Monterey Bay and the North Pacific Subtropical Gyre. *Environ. Microbiol.* 9, 1162–1175. doi: 10.1111/j.1462-2920.2007.01239.x
- Mishra, S., and Imlay, J. (2012). Why do bacteria use so many enzymes to scavenge hydrogen peroxide? Arch. Biochem. Biophys. 525, 145–160. doi: 10.1016/j.abb.2012.04.014
- Moffett, J. W., and Zafiriou, O. C. (1990). An investigation of hydrogen peroxide chemistry in surface waters of Vineyard Sound with H¹⁸₂O₂ and ¹⁸O₂. *Limnol. Oceanogr.* 35, 1221–1229. doi: 10.4319/lo.1990.35.6.1221
- Moffett, J. W., and Zika, R. G. (1988). Measurement of copper(I) in surface waters of the subtropical Atlantic and Gulf of Mexico. *Geochim. Cosmochim. Acta* 52, 1849–1857. doi: 10.1016/0016-7037(88)90008-7
- Moore, C. A., Farmer, C. T., and Zika, R. G. (1993). Influence of the Orinoco River on hydrogen peroxide distribution and production in the Eastern Caribbean. *J. Geophys. Res.* 98, 2289–2298. doi: 10.1029/92JC02767
- Moran, M. A., Satinsky, B., Gifford, S. M., Luo, H., Rivers, A., Chan, L.-K., et al. (2013). Sizing up metatranscriptomics. *ISME J.* 7, 237–243. doi: 10.1038/ismej.2012.94
- Morris, J. J., Lenski, R. E., and Zinser, E. R. (2012). The Black queen hypothesis: evolution of dependencies through adaptive gene loss. *MBio* 3:e00036–12. doi: 10.1128/mBio.00036-12
- Murray, A. E., Preston, C. M., Massana, R., Taylor, L. T., Blakis, A., Wu, K., et al. (1998). Seasonal and spatial variability of bacterial and archaeal assemblages in the coastal waters near Anvers Island, Antarctica. *Appl. Environ. Microbiol.* 64, 2585–2595.
- Murray, A. E., Wu, K. Y., Moyer, C. L., Karl, D. M., and Delong, E. F. (1999a). Evidence for circumpolar distribution of planktonic Archaea in the Southern Ocean. Aquat. Microb. Ecol. 18, 263–273. doi: 10.3354/ame018263
- Olson, R. J. (1981). Differential photoinhibition of marine nitrifying bacteria a possible mechanism for the formation of the primary nitrite maximum. *J. Mar. Res.* 39, 227–238.
- Palenik, B., and Morel, F. M. M. (1988). Dark production of H₂O₂ in the Sargasso Sea. *Limnol. Oceanogr.* 33, 1606–1611. doi: 10.4319/lo.1988.33.6_part_2.1606
- Palenik, B., Zafiriou, O. C., and Morel, F. M. M. (1987). Hydrogen peroxide production by a marine phytoplankter. *Limnol. Oceanogr.* 32, 1365–1369. doi: 10.4319/lo.1987.32.6.1365
- Parsonage, D., Karplus, P. A., and Poole, L. B. (2008). Substrate specificity and redox potential of AhpC, a bacterial peroxiredoxin. *Proc. Natl. Acad. Sci. U.S.A.* 105, 8209–8214. doi: 10.1073/pnas.0708308105
- Popp, B. N., Sansone, F. J., Rust, T. M., and Merritt, D. A. (1995). Determination of concentration and carbon isotopic composition of dissolved methane in sediments and Nearshore Waters. *Anal. Chem.* 67, 405–411. doi: 10.1021/ac00098a028
- Poretsky, R. R., Bano, N., Buchan, A., Moran, M. A., and Hollibaugh, J. T. (2008). "Environmental transcriptomics: a method to access expressed genes in complex microbial communities," *in Molecular Microbial Ecology Manual*, eds G. A. Kowalchuk, F. J. De Bruijn, I. M. Head, A. D. L. Akkermans and J. D. Van Elsas (Dordrecht: Springer), 1892–1904.
- Powers, L. C., Babcock-Adams, L. C., Enright, J. K., and Miller, W. L. (2015). Probing the photochemical reactivity of deep ocean refractory carbon (DORC):

lessons from hydrogen peroxide and superoxide. *Mar. Chem.* 177, 306–317. doi: 10.1016/j.marchem.2015.06.005

- Powers, L. C., and Miller, W. L. (2014). Blending remote sensing data products to estimate photochemical production of hydrogen peroxide and superoxide in the surface ocean. *Environ. Sci. Process. Impacts* 16, 792–806. doi: 10.1039/c3em00617d
- Powers, L. C., and Miller, W. L. (2015). Hydrogen peroxide and superoxide photoproduction in diverse marine waters: a simple proxy for estimating direct CO₂ photochemical fluxes. *Geophys. Res. Lett.* 42, 7696–7704. doi: 10.1002/2015GL065669
- Prosser, J. I., and Nicol, G. W. (2008). Relative contributions of archaea and bacteria to aerobic ammonia oxidation in the environment. *Environ. Microbiol.* 10, 2931–2941. doi: 10.1111/j.1462-2920.2008.01775.x
- Qin, W., Amin, S. A., Martens-Habbena, W., Walker, C. B., Urakawa, H., Devol, A. H., et al. (2014). Marine ammonia-oxidizing archaeal isolates display obligate mixotrophy and wide ecotypic variation. *Proc. Natl. Acad. Sci. U.S.A.* 111, 12504–12509. doi: 10.1073/pnas.1324115111
- Resing, J., Tien, G., Letelier, R., Karl, D. M., and Jones, D. (1993). Palmer LTER: hydrogen peroxide in the Palmer LTER region: II. Water column distribution. *Antarct. J. U.S.* 28, 227–229.
- Roe, K. L., Schneider, R. J., Hansel, C. M., and Voelker, B. M. (2016). Measurement of dark, particle-generated superoxide and hydrogen peroxide production and decay in the subtropical and temperate North Pacific Ocean. *Deep Sea Res.* 107, 59–69. doi: 10.1016/j.dsr.2015.10.012
- Rose, A. L., Webb, E. A., Waite, T. D., and Moffett, J. W. (2008). Measurement and implications of nonphotochemically generated superoxide in the equatorial Pacific Ocean. *Environ. Sci. Technol.* 42, 2387–2393. doi: 10.1021/es7024609
- Santoro, A. E., Casciotti, K. L., and Francis, C. A. (2010). Activity, abundance and diversity of nitrifying archaea and bacteria in the central California Current. *Environ. Microbiol.* 12, 1989–2006. doi: 10.1111/j.1462-2920.2010.02205.x
- Sarthou, G. C., Jeandel, C., Brisset, L., Amouroux, D., Besson, T., and Donard, O.F.X. (1997). Fe and H₂O₂ distributions in the upper water column in the Indian sector of the Southern Ocean. *Earth Planet. Sci. Lett.* 147, 83–92. doi: 10.1016/S0012-821X(97)00004-6
- Sigman, D. M., Casciotti, K. L., Andreani, M., Barford, C., Galanter, M., and Böhlke, J. K. (2001). A bacterial method for the nitrogen isotopic analysis of nitrate in seawater and freshwater. *Anal. Chem.* 73, 4145–4153. doi: 10.1021/ac010088e
- Sintes, E., Bergauer, K., De Corte, D., Yokokawa, T., and Herndl, G. J. (2013). Archaeal *amoA* gene diversity points to distinct biogeography of ammoniaoxidizing *Crenarchaeota* in the ocean. *Environ. Microbiol.* 15, 1647–1658. doi: 10.1111/j.1462-2920.2012.02801.x
- Smith, J. M., Casciotti, K. L., Chavez, F. P., and Francis, C. A. (2014a). Differential contributions of archaeal ammonia oxidizer ecotypes to nitrification in coastal surface waters. *ISME J.* 8, 1704–1714. doi: 10.1038/ismej.2014.11
- Smith, J. M., Chavez, F. P., and Francis, C. A. (2014b). Ammonium uptake by phytoplankton regulates nitrification in the sunlit ocean. *PLoS ONE* 9:e108173. doi: 10.1371/journal.pone.0108173
- Sobota, J. M., and Imlay, J. A. (2011). Iron enzyme ribulose-5-phosphate 3-epimerase in *Escherichia coli* is rapidly damaged by hydrogen peroxide but can be protected by manganese. *Proc. Natl. Acad. Sci. U.S.A.* 108, 5402–5407. doi: 10.1073/pnas.1100410108
- Spang, A., Hatzenpichler, R., Brochier-Armanet, C., Rattei, T., Tischler, P., Spieck, E., et al. (2010). Distinct gene set in two different lineages of ammonia-oxidizing archaea supports the phylum Thaumarchaeota. *Trends Microbiol.* 18, 331–340. doi: 10.1016/j.tim.2010.06.003
- Thompson, A. M., and Zafiriou, O. C. (1983). Air-sea fluxes of transient atmospheric species. J. Geophys. Res. 88, 6696–6708. doi: 10.1029/JC088iC11p06696
- Tolar, B. B., King, G. M., and Hollibaugh, J. T. (2013). An analysis of thaumarchaeota populations from the northern Gulf of Mexico. Front. Microbiol. 4:72. doi: 10.3389/fmicb.2013.00072
- Tolar, B. B., Ross, M. J., Wallsgrove, N. J., Liu, Q., Aluwihare, L. I., Popp, B. N., et al. (2016a). Contribution of ammonia oxidation to chemoautotrophy in Antarctic coastal waters. *ISME J.* 10, 2605–2619. doi: 10.1038/ismej.2016.61
- Tolar, B. B., Wallsgrove, N. J., Popp, B. N., and Hollibaugh, J. T. (2016b). Oxidation of urea-derived nitrogen by thaumarchaeota-dominated marine nitrifying communities. *Environ. Microbiol.* doi: 10.1111/1462-2920.13457. [Epub ahead of print].

- Tourna, M., Stieglmeier, M., Spang, A., Könneke, M., Schintlmeister, A., Urich, T., et al. (2011). Nitrososphaera viennensis, an ammonia oxidizing archaeon from soil. *Proc. Natl. Acad. Sci. U.S.A.* 108, 8420–8425. doi: 10.1073/pnas.10134 88108
- Twiner, M. J., and Trick, C. G. (2000). Possible physiological mechanisms for production of hydrogen peroxide by the icthyotoxic flagellate *Heterosigma* akashiwo. J. Plankton Res. 22, 1961–1975. doi: 10.1093/plankt/22.10.1961
- Walker, C. B., De La Torre, J. R., Klotz, M. G., Urakawa, H., Pinel, N., Arp, D. J., et al. (2010). *Nitrosopumilus maritimus* genome reveals unique mechanisms for nitrification and autotrophy in globally distributed marine crenarchaea. *Proc. Nat. Acad. Sci. U.S.A.* 107, 8818–8823. doi: 10.1073/pnas. 0913533107
- Ward, B. B. (1985). Light and substrate concentration relationships with marine ammonium assimilation and oxidation rates. *Mar. Chem.* 16, 301–316. doi: 10.1016/0304-4203(85)90052-0
- Ward, B. B. (2011). "Nitrification in the Ocean," in *Nitrification*, eds B. B. Ward, D. J. Arp, and M. G. Klotz (Washington, DC: ASM Press), 325–345.
- Wells, L. E., and Deming, J. W. (2003). Abundance of Bacteria, the Cytophaga-Flavobacterium cluster and Archaea in cold oligotrophic waters and nepheloid layers of the Northwest Passage, Canadian Archipelago. *Aquat. Microb. Ecol.* 31, 19–31. doi: 10.3354/ame031019
- Willey, J. D., Kieber, R. J., and Lancaster, R. D. (1996). Coastal rainwater hydrogen peroxide: concentration and deposition *J. Atmos. Chem.* 25, 149–165. doi: 10.1007/BF00053789
- Wuchter, C., Abbas, B., Coolen, M. J. L., Herfort, L., van Bleijswijk, J., Timmers, P., et al. (2006). Archaeal nitrification in the ocean. *Proc. Natl. Acad. Sci. U.S.A.* 103, 12317–12322. doi: 10.1073/pnas.0600756103
- Yocis, B. H. (1995). Dynamics of the Hydrogen Peroxide Cycle in Antarctic Waters. M.S. thesis, State University of New York.
- Yocis, B. H., Kieber, D. J., and Mopper, K. (2000). Photochemical production of hydrogen peroxide in Antarctic Waters. *Deep Sea Res.* 47, 1077–1099. doi: 10.1016/S0967-0637(99)00095-3
- Yuan, J., and Shiller, A. M. (2001). The distribution of hydrogen peroxide in the southern and central Atlantic Ocean. *Deep Sea Res. II* 48, 2947–2970. doi: 10.1016/S0967-0645(01)00026-1
- Yuan, J., and Shiller, A. M. (2005). Distribution of hydrogen peroxide in the northwest Pacific Ocean. *Geochem. Geophys. Geosyst.* 6, Q09MQ02. doi: 10.1029/2004GC000908
- Zafiriou, O. C., Voelker, B. M., and Sedlak, D. L. (1998). Chemistry of the superoxide radical (O₂⁻) in seawater: reactions with inorganic copper complexes. J. Phys. Chem. A 102, 5693–5700. doi: 10.1021/jp980709g
- Zepp, R. G., Skurlato, Y. I., and Pierce, J. T. (1987). "Algal-induced decay and formation of hydrogen peroxide in water: its possible role in oxidation of anilines by algae," in Photochemistry of Environmental Systems, eds R. G. Zika and W. J. Cooper (Washington, DC: American Chemical Society), 215–224.
- Zhalnina, K. V., Dias, R., Leonard, M. T., Dorr de Quadros, P., Camargo, F. A., Drew, J. C., et al. (2014). Genome sequence of *Candidatus Nitrososphaera evergladensis* from group I.1b enriched from Everglades soil reveals novel genomic features of the ammonia-oxidizing archaea. *PLoS ONE* 9:e101648. doi: 10.1371/journal.pone.0101648
- Zika, R. G., Moffett, J. W., Petasne, R. G., Cooper, W. J., and Saltzman, E. S. (1985). Spatial and temporal variations of hydrogen-peroxide in Gulf of Mexico Waters. *Geochim. Cosmochim. Acta* 49, 1173–1184. doi: 10.1016/0016-7037(85)90008-0
- Zika, R. G., Saltzman, E. S., Chameides, W. L., and Davis, D. D. (1982). H₂O₂ levels in rainwater collected in South Florida and the Bahama Islands. *J. Geophys. Res.* 87, 5015–5017. doi: 10.1029/JC087iC07p05015

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