



“If You Know the Enemy and Know Yourself”: Addressing the Problem of Biological Invasions in Ports Through a New NIS Invasion Threat Score, Routine Monitoring, and Preventive Action Plans

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Invasive alien species (IAS) are currently considered one of the greatest threats to global marine ecosystems. Thus, ships and maritime activity have been identified as the main factors responsible for the vast majority of accidental species translocations around the world, implying that prevention should be the core of environmental port policies. Preventive port strategies should include analyzing risks based on traffic origins and volumes, revising port policies for inspections, estimating probabilities of non-indigenous species (NIS) appearance, monitoring routine species within ports, and finally implementing management plans and focused actions. Here, we conducted a comprehensive NIS prediction analysis for the port of Gijón (northern Spain), one of the largest ports in the south Bay of Biscay, as a case study that can be extrapolated to other international seaports. An extensive bibliographic search (1953–2020) was conducted and we identified 380 species that have been transported through hull fouling and ballast water around the world. We evaluated their likelihood of arriving (from 14 years of traffic data) and becoming established (from habitat suitability and demonstrated impacts and invasion ability) within the Gijón port, creating a new NIS Invasion Threat Score (NIS-ITS). This new index could help to identify target species that are likely invaders for early detection and prevention policies within the port. The results showed that 15 NIS had >90% likelihood of becoming a biological invasion problem in Gijón Port. At the same time, we reported morphological and genetic analysis of biota found in two successive annual monitoring surveys of Gijón port and ships ($n = 612$ individuals) revealing 18 NIS, including 6 of the NIS predicted from high NIS-ITS. Actually, 80% (12 NIS) of those potentially most dangerous species (NIS-ITS > 90%) have already been detected in the Bay of Biscay area. We propose the use of this new tool for a

risk-reduction strategy in ports, based on accurate predictions that help in promoting specific early detection tests and specific monitoring for NIS that have a high chance of establishment. All international seaports can adopt this strategy to address the problem of biological invasions and become “blueports” in line with EU policy.

Keywords: invasive species, genetic barcoding, ports strategies, blueports, NIS assessment, invasion likelihood

INTRODUCTION

The distribution of marine taxa is limited by natural barriers to their dispersal (Elton, 1958). However, over the past centuries, human activities have modified those barriers (Vemeij, 1978) and, consequently, the distribution of many species (Kotta et al., 2016). At present, based on data from the European Alien Species Information Network (EASIN), there are approximately 14,000 non-indigenous species (NIS) recorded in Europe (EASIN Catalog)¹ and invasive alien species (IAS) represent a major threat to the ecological, economic and social values of marine environments (Vanderploeg et al., 2002; Molnar et al., 2008). They are considered one of the greatest threats to global marine ecosystems, affecting their structure and function (Costello et al., 2010), with negative socioeconomic consequences (e.g., Gherardi et al., 2011). Ballast water and hull fouling, associated with maritime transport, are responsible for the vast majority of accidental marine translocations around the world (Carlton, 2001; Hewitt et al., 2009; Seebens et al., 2013), however, there are other human-mediated pathways that can be used by species to spread to new habitats, such as the opening of new corridors that facilitate the spread of species (Steger et al., 2019; Castellanos-Galindo et al., 2020) or escapes from aquaculture facilities (Thorvaldsen et al., 2015; Ju et al., 2020) among others.

At a global scale, ships are widely recognized as the primary vector of aquatic biological invasions (Carlton, 1985; Ruiz et al., 2000; Hewitt et al., 2004; Gollasch, 2006). Additionally, ships contain sub-vectors, including ballast water (Carlton, 1985), ballast sediments (Bailey et al., 2005), hull biofouling (Allen, 1953; Ashton et al., 2016), internal tank biofouling (Drake et al., 2005), sea chests (Coutts, 1999), and internal seawater piping systems (Carlton, 1985; Lewis and Dimas, 2007). As a result, all ships are different in terms of vector risk, and there are differences in species entrainment, transfer and release among different ship types (Davidson et al., 2018).

Risk analyses and management actions to reduce ship-mediated species dispersal around the world are being developed and have increased as marine biosecurity concerns grow (Davidson et al., 2018). There are several international legislative frameworks focused on managing the introduction and spread of potential NIS. Within the European Union, the main policy and legislative drivers to target marine NIS are the Marine Strategy Framework Directive (MSFD), the Biodiversity Strategy, and EU Regulation no. 1143/2014 on Aquatic Invasive Species (EC, 2014). Within this legislative framework, the EU blue growth strategy advocates for a blue economy that exploits natural resources in a sustainable way and protects the environment

(European Commission, 2012). This includes the *Blueports objective*, a strategy that aims to create an adequate management of resources along with new strategies to avoid biological invasions and the environmental or economic impacts that these can produce (Dopico Rodríguez and Borrell Pichs, 2020). Recently, the first EU-scale Horizon Scanning (HS) focusing on marine NIS was performed, aiming to deliver a ranked list of species that should be of high priority for risk assessment (Tsiamis et al., 2020). These lists of species with invasive potential can be used to update the EU regulation 1143/2014, which shows a very limited coverage of marine invasive species and, this way, preventive actions against these potentially harmful species can be promoted (Tsiamis et al., 2020). These global initiatives are extremely relevant, although they can lose effectiveness if they are not properly interpolated and adjusted to the characteristics of the different marine areas in a much more concrete or specific way within the European Union (areas, countries, even ports). On the other hand, it is necessary to develop accessible and friendly methodologies for stakeholders (usually not experts in biological invasions) with the responsibility of making decisions that can prevent these events (e.g., port authorities).

Outside of Europe, a set of port monitoring guidelines have been developed to maintain up-to-date records of the distribution of NIS, such as those for Australian ports (Hayes et al., 2019). Similarly, in New Zealand, a series of targeted (species-specific) surveys are now implemented in 11 harbors every 6 months (Arthur et al., 2015). Moreover, in order to avoid species introductions via marine traffic, some Australian ports require arriving ships to be inspected and treated for marine pests before they can enter a marina. These inspections are compulsory for vessels arriving from international ports (DPC, 2013).

Recently, the Ballast Water Management Convention (BWM) entered into force. All international sea-going ships under this convention must comply with standards that require discharging ballast water from the last port and replacing it with new sea water at a minimum of 200 nautical miles from shore before arriving at a new port (IMO, 2004, 2011). Despite this, several studies have indicated that in comparison to ballast water, biofouling likely accounts for more NIS introductions (e.g., Fofonoff et al., 2003; Davidson et al., 2009; Hewitt and Campbell, 2010). In this context, other projects such as the GloFouling partnership project (a collaboration among the International Maritime Organization, the United Nations Development Programme and the Global Environment Facility) have been developed with the objective of building capacity in developing countries for implementing the IMO Biofouling and other relevant guidelines for biofouling management and to catalyze overall reductions in the transboundary introduction of biofouling-mediated IAS.

¹<https://easin.jrc.ec.europa.eu/>

Thus, biofouling should be included in a marine NIS risk assessment regardless of the currently low capacity to predict inoculation rates and difficulties associated with sampling (Ware et al., 2014).

Although a management and legislative framework and a marine vector-based risk assessment methodology are well-established in the literature (Campbell and Hewitt, 2011; Keller et al., 2011; Chan et al., 2012; Floerl et al., 2013; Ware et al., 2014), both have many limitations. First, NIS monitoring surveys are not yet performed in all international ports consequently, there is not a global and updated information source about the current state of biodiversity in ports (Blueports Interreg Project Advisory Committee, pers. comm.) and there are not complete baselines for temporal and spatial intra-/inter-port comparisons worldwide. Secondly, vessel arrival details are a poor proxy since the substantial variety in ships and shipping routes make the process stochastic (Verling et al., 2005; Lawrence and Cordell, 2010; Ruiz et al., 2013; Ware et al., 2014).

In this study, an initial biological invasion preventive strategy has been developed for the international port of Gijon, located in the Bay of Biscay, which is the leading port for solid bulk movement in Spain. The approach consisted on a global NIS occurrence likelihood assessment based on an exhaustive review of worldwide previously reported species introductions through shipping in the last 70 years (biofouling and ballast water) and combined with, the habitat suitability for each species in the given port conditions, species invasiveness and traffic data analyses. Together, these factors were considered to produce a specific NIS invasion threat score (NIS-ITS) for the port of Gijon. Two consecutive years of physical monitoring surveys of Gijon Port's biota were performed, and different hull ships were sampled to analyze the biofouling community with the aim of comparing predicted and actual NIS scenarios.

MATERIALS AND METHODS

“El Musel” the International Industrial Port of Gijon

Located on the Cantabrian coast (5°41.9496'W, 43°33.5120'N), the port of Gijon is one of the main seaports in the Atlantic arc and the leading port for solid bulk movement in Spain². The port of Gijon covers 415 hectares and has more than 7,000 linear meters of docks. The spaces provided by the port of Gijon are structured in areas divided by each type of traffic (solid bulk, liquid and container terminals, and multipurpose facilities for various types of traffic). The port also has a small marina with recreational boats located outside the main docks (**Figure 1**).

To identify and control environmental quality, the Port Authority of Gijon has established and implemented an Environmental Surveillance Plan. This plan was developed taking into account the applicable regulations and the Program of Recommendations for Maritime Works (ROM) and based on the standards of the International Standard ISO 14001. For this reason, the Port Authority of Gijon is

focused on identifying its most significant environmental impacts and acting on them by continuously improving its management strategies³.

Expected NIS Scenario

To evaluate the threats of invasions in the port of Gijon, we created a NIS-Invasion Threat Score (NIS-ITS). A review of the scientific literature from 1953 to 2020 was conducted considering reports of species dispersal via maritime traffic. This search was conducted considering all living organisms, from microscopic species such as dinophytes that have been studied traveling in ballast tanks to macroscopic organisms such as bryozoans or ascidians that have been observed traveling via hull fouling. All translocation reports were collected, and a list of species with evidence of using marine traffic as a vector for expansion was generated. This dataset was analyzed for the case study of the port of Gijon and species were categorized as native, NIS, IAS or cryptogenic in the area. These classifications were carried out by using the Invasive Species Specialist Group (ISSG) global database of the International Union for Conservation of Nature⁴, the DAISIE (Roy et al., 2020) and the AquaNIS (AquaNIS, 2015) databases. Once this categorization was done, the NIS-ITS was calculated for the NIS in the area under study. On the other hand, those species with evidences of using maritime traffic as a vector, but classified as native in the area under study were listed apart in order to facilitate a checklist to other regions where these species could arrive and might become a NIS.

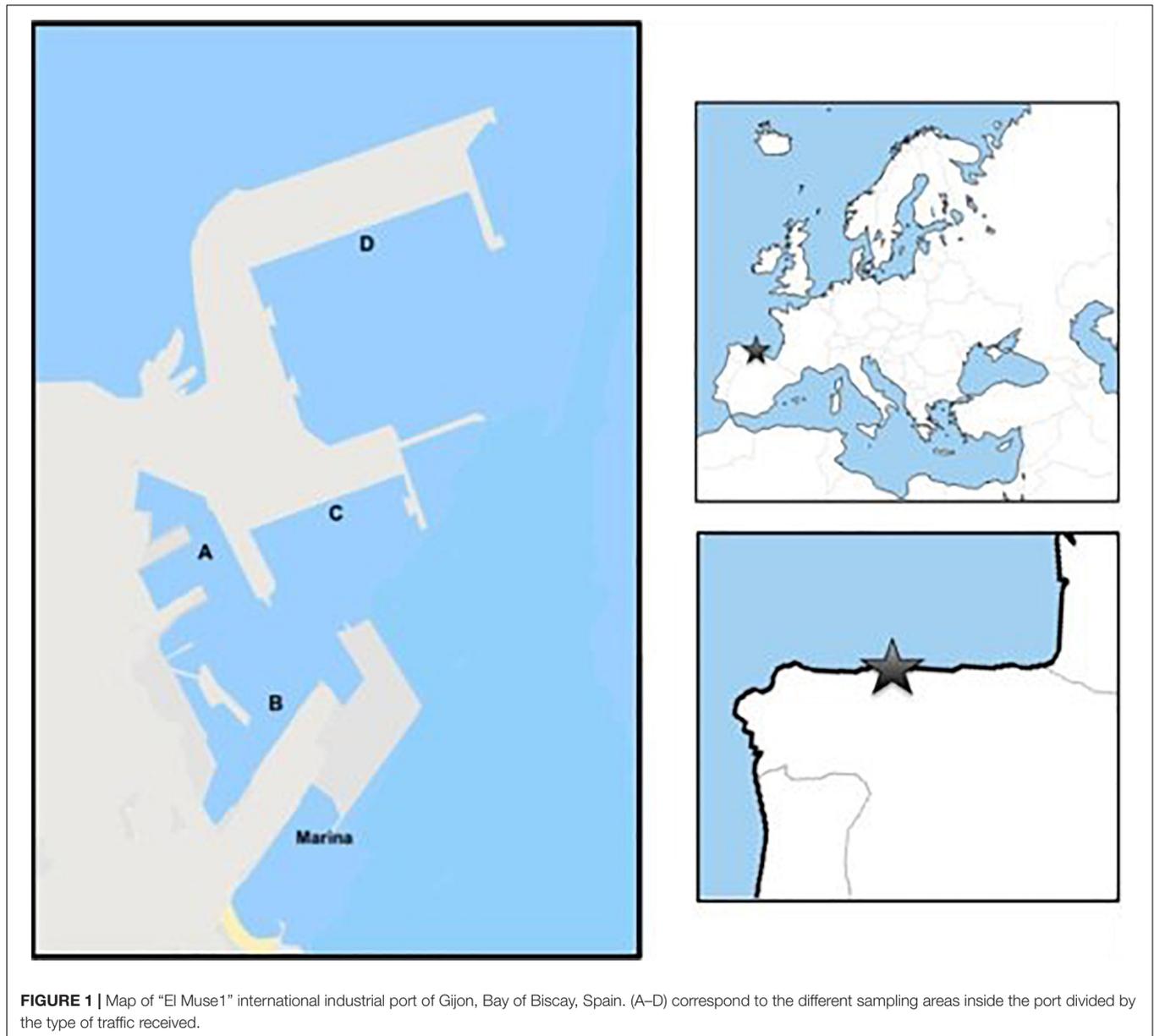
The NIS-ITS was calculated using three criteria. The first criterion was the suitability of the environmental conditions of the port of Gijon for each of the potentially arriving species. This information was obtained from the SeaLifeBase webpage (Palomares and Pauly, 2020). Values for habitat suitability (H_s) range from 0 (totally unsuitable habitat) to 1 (perfect environmental conditions for the species). These values are calculated for each species and geographic location by using estimates of environmental preferences with respect to depth, water temperature, salinity, primary productivity, sea ice concentration and distance to land (Ready et al., 2010). Each species-specific environmental preferences are matched against local environmental conditions to determine the suitability of a given area in the ocean for a particular species.

By combining the information for these parameters, an environmental envelope is generated for each species, which is essentially a response curve that describes the habitat usage of such species or its preferences with respect to certain environmental parameters. If the mean environmental conditions fall within the preferred parameter range of the species, the habitat suitability will be the highest ($H_s = 1.00$). However, when values fall outside this range, it will decrease linearly toward the species' absolute minimum or maximum parameter thresholds and it is set to zero ($H_s = 0.00$) beyond the absolute threshold values (Ready et al., 2010). The suitability is first calculated for each of the individual environmental

²<https://www.puertogijon.es/en/>

³<https://www.puertogijon.es/en/port/sustainable-port/>

⁴<http://www.issg.org/database>



parameters and the product is then used to determine the overall habitat suitability in a given location, that is:

$$Hs = S_{depth} \times S_{temperature} \times S_{salinity} \times S_{primaryproduction} \\ \times S_{iceconcentration} \times S_{landdistance}$$

This multiplicative approach allows each environmental predictor to act as a “knock-out” criterion. For instance, if the salinity in a given location exceeds the salinity preference of a particular species, the probability of occurrence with respect to salinity will be zero ($S_{salinity} = 0.00$). The overall habitat suitability for the species in that given location will then be zero ($Hs = 0.00$), even if all other environmental attributes are within the preferred range of that species. Habitat suitability data can be found for specific geographic regions and for particular species

and it is accessible at <https://www.aquamaps.org/search.php>. Georeferenced values for the mentioned parameters are provided at 0.5° resolution (Kaschner et al., 2019). For this case study, data from the point 43.75°, -5.75° were employed since the port of Gijón is located inside this area.

The second criterion was the known invasion history of the species. A numeric score is given to each species depending on the following situations: (a) the species is only present in its native habitat and there are no reports of its introduction in any other area, (b) the species has been introduced in areas outside its native range but there is no information on any relevant impact, (c) the species has invaded new areas and causes various impacts of different magnitudes. The score for each of these conditions is set based on the six EICAT (Environmental Impact Classification for Alien Taxa) impact categories from

Hawkins et al. (2015) by assigning an ascending numerical value according to the level of impact of the species (**Supplementary Table 1**). A score of zero was given to those species that are present only in their native habitat without any known introduction event. A score of 0.2 was given to those species that have been introduced in areas out of their native range but do not cause any relevant impact. This score corresponds to the “minimal concern” EICAT impact category which includes species with “small inconsequential changes,” as suggested by Roy et al. (2019). An invasion history score of 1 was given to those species that have been introduced and produce environmental or economic impacts outside their native areas. This score corresponds to the rest of categories with higher EICAT impact magnitudes (minor, moderate, major and massive) which are all merged due to the lack of specific information to differentiate between these categories with high accuracy.

Finally, the third criterion employed for the calculation of the NIS Invasion Threat Score (ITS) was the maritime traffic. In order to calculate the maritime traffic score for a given species, its native range is determined based on bibliography. Results are reported as presence or absence of the species in each of the seven biogeographic areas defined by Menéndez-Teleña (2019) (see **Supplementary Table 2**). Once the native range for a species is assessed, it is also checked if it has established populations in other biogeographic areas. If so, these areas are also considered as potential donors of species and included in the traffic score calculation. The specific dispersal method that may be employed by the species (e.g., ballast water or hull fouling) is not taken into account. If the species is present in a biogeographic zone, the maritime traffic originating from this area is analyzed and summed to other potential areas where the species is present and therefore may arrive to the port of Gijon using the ships as vectors. The traffic score is calculated by using available traffic data for the port under study. Depending on the case, this data availability may vary. However, this score can be calculated in any case without a minimum data limitation. This means that the score can be calculated, for example, annually in those cases with scarce traffic data, but it can also employ data from a longer period, such as in this case study, in which 14 years of traffic in the port of Gijon were used. Once the global distribution of the species is determined, the relative proportions of traffic from native and invaded areas (the ratio between the volumes of traffic from the area divided by the total traffic arriving at Gijon) are estimated and then summed to obtain a unique total traffic score. Areas with heavy maritime traffic will show a high traffic ratio. Then, species that are present in these areas will show high maritime traffic scores. Values for maritime traffic also range from 0 (total absence of a species in all of the biogeographic areas with marine traffic to Gijon) to 1 (a species that was present in all the areas with maritime traffic to the port of Gijon).

Following the methods used to calculate invasion risk scores by different authors (Bomford, 2008; Gallardo et al., 2016; Tsiamis et al., 2020), the NIS-ITS can be calculated by the sum of the specific criteria, which in this case are the habitat suitability, invasion history and maritime traffic. All of them are considered equally important as in Gallardo et al. (2016) and references therein. This method (unlike multiplicative ones)

allows obtaining non-zero scores when one of the parameters is null. In summary, the index can be calculated by the following formula:

$$NIS - ITS = \text{HabitatSuitability} + \text{InvasionHistory} + \text{MaritimeTraffic}$$

The maximum NIS-ITS that a species could obtain after combining all the criteria is 3 (examples of ITS calculations can be seen in **Supplementary File 1A**). A NIS-ITS = 3 (100% of the maximum ITS) indicates that the port of Gijon presents perfectly suitable environmental conditions for the specific species (habitat suitability = 1), that the species has shown previous capacity to impact ecosystems out of its native range and biologically invade elsewhere (invasion history = 1), and the species is present in all the biogeographic areas from which the port of Gijon receives marine traffic (total traffic risk = 1). On the contrary, a species could also obtain an NIS-ITS = 0 when: environmental conditions are outside the species' absolute minimum or maximum parameter thresholds on at least one of the environmental parameters (habitat suitability = 0), the species is only present on its native range, without any known introduction or invasion event (invasion history = 0) and when the species is not present in any of the biogeographic areas with maritime traffic with the port under study (maritime traffic = 0).

This method does not take into account if vessels employ techniques for the prevention of biological invasions, such as techniques to release ballast water away from ports or to avoid fouling by using paints that cover the hulls. In these cases where these preventive measures are taken, the NIS-ITS should be lower than in cases where they are not. Due to the lack of data, this method cannot be implemented in this case study, but we provide an example for its calculation that could be considered for future cases (see **Supplementary File 1B**).

The port of Gijon can also be a donor of species, so it is also important to understand which species that are present within the port can be transferred via maritime traffic. Thus, from the initial list of species containing species with the potential of using maritime traffic as a vector for expansion, those that are native to the area studied in this work (the port of Gijon) were analyzed. These species were categorized as IAS (when the species has invaded new habitats and causes impacts and environmental damages), NIS (when the species has been reported out of its native range, but there is not information about any impact in these habitats) or NA (no alien populations). It is expected that these data can help destination ports complete their own and specific NIS-ITS information with data on species habitat suitability and maritime traffic for their own conditions.

Port Biota and Biofouling Samplings

The port biota study was focused on macroinvertebrates attached to port facilities (rocky ground and port structures). Targeting macroinvertebrates is the typical sampling procedure routinely followed in most ports to estimate biotic indices for environmental quality assessments of marine ecosystems (Borja et al., 2000, 2009) and is recommended in both the EU Water Framework Directive (WFD, Directive 2000/60/EC)

and the Marine Strategy Framework Directive (MSFD, Directive 2008/56/EC) (Borja et al., 2015). Sampling was performed in two consecutive years (2016 and 2017) during the highest spring tide. Two years of sampling were selected to optimize the surveillance design, in terms of effort vs. resolution, by following the recommendations for sampling aquatic invasive species in Hoffman et al. (2016).

The sampling was divided into the structured areas of the port: A, B, C, D, and marina (**Figure 1**). In each area, two locations were studied, except for in the marina, where only one place was sampled due to its small area. At each location, sampling occurred at three different tidal levels (high, low and medium). A rapid assessment survey approach (RAS; Minchin, 2007) and the sampling protocol from Miralles et al. (2016) were adapted and conducted on the assemblages living on artificial port structures. A team of five researchers (two postdoc and three predoc) expert in marine science did the 2 years sampling. Before starting the sampling and to avoid biases due to patchy spatial distributions, a visual inspection to determine the different types of organisms present at the sampling site was conducted for approximately 10 min. To standardize the sampling effort, the surface sampled at each site within the area of the port was approximately 200 m², and the duration of the sampling was 30 min. Roughly 1% of the animals visually detected attached on that surface were collected at random. The number of individuals picked of each morphotype (group of organisms phenotypically different) was approximately proportional to the abundance of such morphotype in the area. A maximum of 50 individuals per location were selected and were placed in buckets with seawater for further morphological, taxonomical and genetic analyses.

All sampled individuals were immediately transported to the laboratory of Genetics of Natural Resources at the University of Oviedo. In the laboratory, individuals were identified based on their morphology to the species or to the lowest possible taxonomic level. Then, a piece of tissue from each individual was preserved in absolute ethanol for further genetic analysis. The taxonomic nomenclature of all identified species was verified on the World Register of Marine Species (WoRMS, 2020).

Apart from the biota on the port structures, species attached on the hulls of ships arriving at the port of Gijón were analyzed. To carry out this effort, during July 2017, 22 ships from 12 different countries were contacted directly through the Port Authority of Gijón to ask them to participate in this study. Sampling consisted of scratching all biofouling organisms from a quadrat of approximately 30 × 30 cm in three different places along the draft of the vessel: stern, mid-ship and bow. Sampling was performed under strict safety procedures with professional divers. All divers were fastened by wiring and had audio and visual connections with the supporting ship where the scientists were placed located. The general procedure of diving was based on the Collective Agreement of Professional Diving and Hyperbaric Means; the Spanish Ministerial Order of October 14, 1997; the Modification to the Collective Agreement of Professional Diving and Hyperbaric Means of October 18, 2016; and the UNE-EN 15333-1. Samples were placed in bags with seawater and transported to the University of Oviedo. In the laboratory, individuals were identified based on morphology, and

a piece of tissue from each individual was preserved in absolute ethanol for further genetic analysis. The taxonomic nomenclature of all identified species was verified on the World Register of Marine Species (WoRMS, 2020).

The list of NIS found in these samplings was enhanced by data from previous studies of macroinvertebrates and algae (such as Arias et al., 2014; Miralles et al., 2016) and from metabarcoding data (Borrell et al., 2017; Ibabe et al., 2020a,b) to generate an accurate and complete NIS scenario for the port of Gijón.

Genetic Species Identification

Genetic species identification was carried out to confirm the morphological taxonomic identification following Miralles et al. (2016). DNA was extracted from ethanol-preserved individuals using a Chelex[®] resin (Bio-Rad Laboratories Inc., United States) protocol (Estoup et al., 1996) or the EZNA Mollusc DNA Kit (Omega Bio-Tek Inc., United States). The genetic barcode employed was the mitochondrial cytochrome oxidase subunit I coding region (COI gene), with additional genes (RNAr18s or RNAr16s) to improve resolution or to determine species when the database references were insufficient (Miralles et al., 2016; Pejovic et al., 2016). The obtained DNA sequences were visually inspected and edited with BioEdit v7.2.5 (Hall, 1999). All sequences were compared with online public databases using nBLAST in NCBI⁵ and Bold Systems⁶. The sequence with a maximum score and at least a 97% nucleotide identity and an E-value <e-100 was considered the reference for genetic species identification.

All sampled port biota and ship biofouling individuals were classified as native species, NIS, IAS or cryptogenic species. The Invasive Species Specialist Group (ISSG) global database of the International Union for Conservation of Nature (IUCN), the DAISIE (Delivering Alien Invasive Species Inventories from Europe), and the AquaNIS (information system on aquatic non-indigenous and cryptogenic species database) were the references used to confirm the introduced and invasion status of each NIS when sufficient information was available. Species that are unable to be identified as native or NIS were classified as cryptogenic (Carlton, 1996). Once species identity was confirmed, species abundances for statistical analyses were calculated based on the number of individuals with confirmed identification and the total number of species identified in each sampling (2 year separately).

Statistical Analyses

Since samplings were conducted in different port areas and years, a multivariate non-metric multidimensional scaling (nmMDS) analysis was conducted to determine whether the distribution of the species (including both, native and NIS) was strongly clustered by area or year or whether they were more or less scattered. The Bray-Curtis dissimilarity index was applied, and a stress <0.2 was considered acceptable. The Shepard plot was inspected to determine the linearity of the index in the dataset. A SIMPER analysis was conducted (only for the non-indigenous biota) to assess the species that most contributed to the differences from 1 year to the other and detect any

⁵www.ncbi.nlm.nih.gov/

⁶www.boldsystems.org/

potential expansion or population growth. Comparisons between years and areas as well as diversity indexes (Shannon and Simpson), tests and plots were performed with PAST version 3.8 (Hammer et al., 2001).

RESULTS

Expected NIS Scenario

Eighty research articles recording worldwide evidences of species introductions and translocations through shipping from approximately 70 years of marine research were compiled in this work. These reports were used to create a list of species that were able to travel by ship and have the potential to be future invaders in any port. A large species taxonomical representation was obtained from that list of 380 species. From this dataset, 252 species could be considered as NIS (if found in the port of Gijón), whereas 128 were classified as native in accordance with their natural geographic distributions. NIS habitat suitability data for the port of Gijón were available only for 71 out of the 252 NIS. This information was combined with the invasiveness of each species and the corresponding port of Gijón's maritime traffic score to estimate the NIS-ITS for each individual species (Supplementary Table 3). Although NIS-ITS cannot be estimated for those species lacking habitat suitability information, scores for maritime traffic and invasion history were also estimated for these species (Supplementary Table 4).

A maximum invasion threat score (NIS-ITS = 3.00) was obtained by 3 out of the 71 species [*Bugula neritina* (Linnaeus, 1758), *Ficopomatus enigmaticus* (Fauvel, 1923), and *Botryllus schlosseri* (Pallas, 1766)]. On the contrary, some species such as *Corynactis californica* Carlgren, 1936 (NIS-ITS = 0.153) showed low scores, indicating a lower likelihood of being a successful invader in the area (Supplementary Table 3). Most of the analyzed species were crustaceans (Arthropoda), ascidians (Chordata), bivalves and gastropods (Mollusca). All species belonging to Rhodophyta and Ctenophora phyla showed invasion threat scores above 70%. Similarly, four polychaetes showed also more than a 70% score and two out of four bryozoans scored above 90% (Figure 2). A total of fifteen NIS showed more than 90% of invasion likelihood (NIS-ITS > 2.70) (Table 1). Most of them were meroplanktonic animals belonging to classes such as Malacostraca, Hexanauplia, Polychaeta, or Ascidiacea. The only exception was *Hypnea musciformis* Lamouroux, 1813, which is a seaweed belonging to the class Florideophyceae. All these species with a NIS-ITS higher than 90% showed an invasiveness score of 1, meaning that all of them had previously invaded areas out of their native ranges and produced environmental impacts. Special care and attention should be given to these species to avoid their establishment in the port of Gijón. In fact, most of the species with an invasion history score of 1 showed a NIS-ITS above 70% whilst species that are only known to be present in their native habitat (invasion history score = 0) showed low NIS-ITS values (Figure 3).

On the other hand, 128 species capable of using maritime traffic as a vector were found to be native to Gijón (Supplementary Table 5). Sixteen of these species have

already invaded other areas and produced environmental impacts outside their native range. Similarly, 32 species that are native to the area under study, despite having no records of producing environmental impacts, have already been detected outside their native habitats. These species are mainly dinophytes transferred by ballast water or bryozoans transferred by hull fouling (Supplementary Table 5).

Port Biota and Biofouling

A total of 522 individuals were genetically identified to 82 taxa in the 2 years samplings of the port of Gijón. In general, genetic analyses confirmed the previous morphological identification. The observed communities were mainly composed by polychaetes, bivalves and gastropods (GenBank accession numbers MN185333-MN185374 for COI gene, MN164033-MN164046 for 18s RNA gene, and MN164346-MN164348 for 16s RNA gene; Supplementary Table 6).

The global biodiversity in the port of Gijón was not significantly different between 2016 and 2017 (diversity *t*-test for Shannon index $t = -1.108$ and $p = 0.268$; Simpson index $t = 0.901$ and $p = 0.368$; Figure 4A). Regarding the observed NIS, IAS and cryptogenic species, significant differences between years were found (diversity *t*-test for Shannon index $t = -2.475$ and $p = 0.015$; Simpson index $t = 2.61$ and $p = 0.010$). The rarefaction plot revealed higher NIS richness in 2017 than in 2016 (Figure 4B). SIMPER analysis showed that the NIS that mainly accounted for the differences between 2016 and 2017 communities were *Magallana gigas* (Thunberg, 1793), *Mytilaster minimus* (Poli, 1795), *Watersipora subtorquata* (d'Orbigny, 1852), and *Mytilus trossulus* (Gould, 1850), which accounted for 70% of the dissimilarities. The first two species decreased, while *W. subtorquata* (7.74% of the contribution to the dissimilarities between 2016 and 2017), and *M. trossulus* (5.31% of the contribution to dissimilarities) increased in observed specimens from 2016 to 2017. In fact, *W. subtorquata* was found at all five stations in 2017 (found only at stations A and B and the marina in 2016), and *M. trossulus* was found in stations A and C and the marina in 2017 while it was present only at station C in 2016 (Supplementary Table 6).

Biodiversity comparisons among the main industrial docks (A, B, C, and D) showed that they clustered together regardless of the sampling year (Figure 5, stress = 0.080). However, the marina always appeared slightly distant from the remainder of the industrial docks both years, and notable differences were found when comparing communities inhabiting the marina in 2016 and 2017 (Figure 5).

Apart from analyzing the biota inhabiting the different areas of the port of Gijón, the species attached to the hulls of three ships docked in the port were also studied. Notably, out of the 22 ships contacted, these three ships were the only ones that agreed to participate in the study. To maintain anonymity, the sampled vessels were labeled ships 1, 2, and 3. Ship 1 was an international vessel that arrived to Gijón from Russia and its destiny was United Kingdom. Ship 2 was a dredge vessel that used to work along the Cantabrian Sea (South Bay of Biscay) but it was stopped due to maintenance in Gijón for 13 days before sampling. Finally, Ship 3 was an inactive vessel docked in Gijón for 3 months

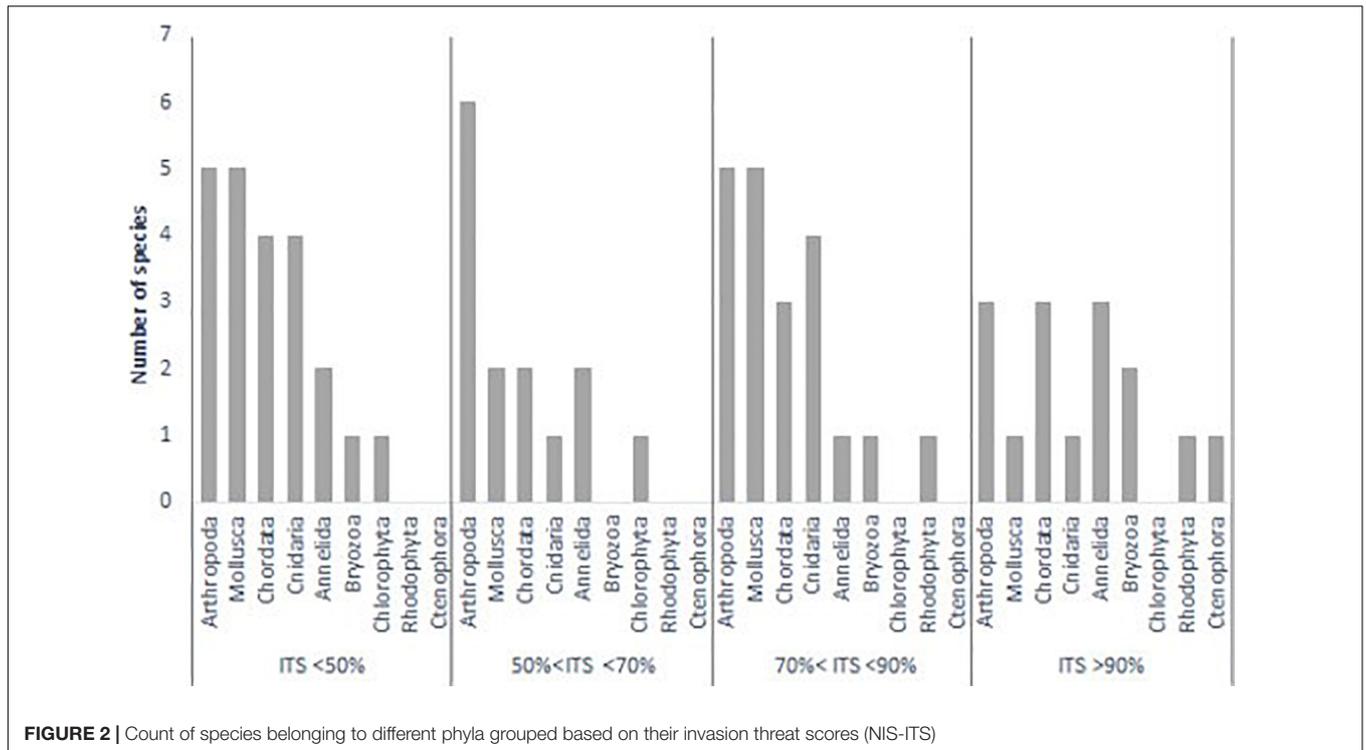


FIGURE 2 | Count of species belonging to different phyla grouped based on their invasion threat scores (NIS-ITS)

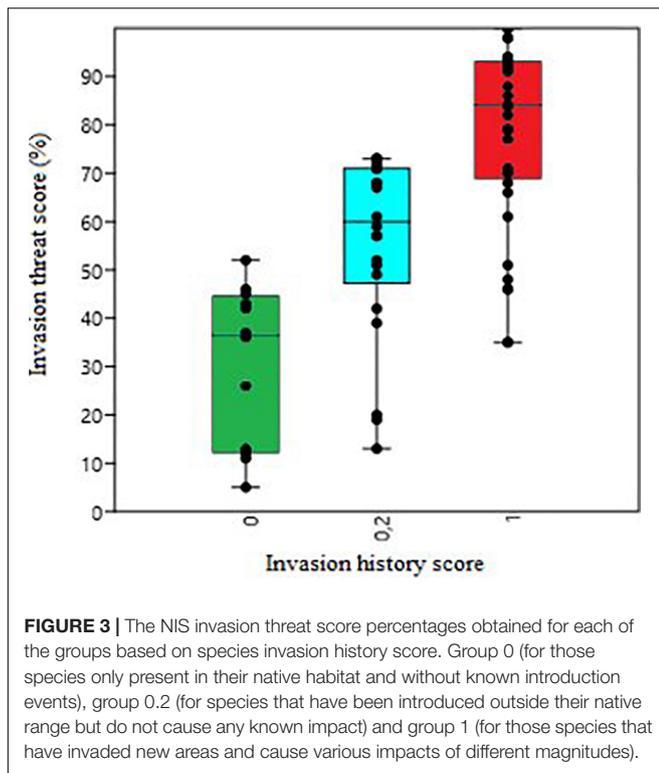
TABLE 1 | Non-indigenous species (NIS) with more than a 90% invasion threat score (NIS-ITS) in the port of Gijón.

n	Class	Species	NIS-ITS (%)	References for the species report
1	Asciacea	<i>Botryllus schlosseri</i> (Pallas, 1766)*	100	Nydam et al., 2017
2	Gymnolaemata	<i>Bugula neritina</i> (Linnaeus, 1758)*	100	This study
3	Polychaeta	<i>Ficopomatus enigmaticus</i> (Fauvel, 1923)	100	Miralles et al., 2016
4	Asciacea	<i>Diplosoma listerianum</i> (Milne Edwards, 1841)*	99.97	This study
5	Anthozoa	<i>Diadumene lineata</i> (Verrill, 1869)	97.87	Altuna, 2015
6	Polychaeta	<i>Polydora cornuta</i> Bosc, 1802	97.87	Not found
7	Hexanauplia	<i>Acartia tonsa</i> Dana, 1849	94.30	Aravena et al., 2009
8	Gymnolaemata	<i>Schizoporella errata</i> (Waters, 1878)	94.03	Not found
9	Asciacea	<i>Styela plicata</i> (Lesueur, 1823)*	92.67	Miralles et al., 2016
10	Malacostraca	<i>Callinectes sapidus</i> Rathbun, 1896	92.63	Arias et al., 2014
11	Malacostraca	<i>Rhithropanopeus harrisi</i> (Gould, 1841)	92.63	BOE, 2013
12	Bivalvia	<i>Mytilopsis leucophaeata</i> (Conrad, 1831)	92.20	BOE, 2013
13	Tentaculata	<i>Mnemiopsis leidyi</i> Agassiz, 1865	92.20	BOE, 2013
14	Florideophyceae	<i>Hypnea musciformis</i> Lamouroux, 1813	91.63	Arias et al., 2014
15	Polychaeta	<i>Marenzelleria viridis</i> (Verrill, 1873)	91.30	Not found

Those species detected in previous reports in the Bay of Biscay (the geographic area of the port) are shown in bold. Species already detected in the port of Gijón in this and previous studies are shown with an asterisk.

(details about the study vessels in **Supplementary Table 7**). A total of 73 specimens and 15 different species, including native but also 5 NIS, were found fouling the hulls of the ships (**Table 2**). Most of the species were natives already observed in the samplings from other port areas. However, some species were only found on the vessels. For instance, Ship 3 had been docked in Gijón for the longest period, and almost all the fouling species were already described in the sampled areas of the port. Nevertheless, a NIS [*Diplosoma listerianum* (Milne Edwards, 1841)] that did not appear in the port samplings was collected

from this ship. Two new potentially dangerous species (which were not reported in the 2 years previous monitoring surveys) appeared on Ship 2: *Monocorophium acherusicum* (Costa, 1853) and *Spirobranchus taeniatus* (Lamarck, 1818), in addition to other NIS, *Spirobranchus latiscapus* (Marenzeller, 1884) and *Dipolydora capensis* (Day, 1955) that had already been observed in the port samplings (**Table 2**). Thus, the risk of being a vector of NIS through biofouling was not the same for the three vessels. Ship 1 could be classified as “no invasion risk” (because there was no attached biota), while ships 2 and 3 could represent a “potential



vector of invasions” since new species were present in their hull (and not in the port). In addition, these ships could represent a “vector of dispersion” because they carried several fouling native species that could eventually spread to new areas.

Expected vs. Observed NIS Scenario

Complete data to estimate the proposed NIS-ITS were available only for approximately a quarter (28.2%) of the potential invaders. Four out of the fifteen species (26.7%) with NIS-ITS values above 90% have been already detected in the port of Gijón (Table 1). Two of them [*Bugula neritina* (Linnaeus, 1758) and *Diplosoma listerianum* (Milne Edwards, 1841)] were found in the samplings from this study, and the other two were found in samplings from a previous report for the same port [*Botryllus schlosseri* (Pallas, 1766) (Nydám et al., 2017) and *Styela plicata* (Lesueur, 1823) (Miralles et al., 2016)]. Moreover, eight other NIS (*Ficopomatus enigmaticus* (Fauvel, 1923); *Diadumene lineata* (Verrill, 1869); *Acartia tonsa* Dana, 1849; *Callinectes sapidus* Rathbun, 1896; *Rhithropanopeus harrisi* Gould, 1841; *Mytilopsis leucophaeata* (Conrad, 1831); *Mnemiopsis leidyi* Agassiz, 1865; and *Hypnea musciformis* Lamouroux, 1813) have also been reported in the Bay of Biscay in previous studies (Table 1). Therefore, 80% (12 NIS) of the potentially most dangerous species have already been detected in the Bay of Biscay area. In addition, two other NIS [*Watersipora subtorquata* (d’Orbigny, 1852) ITS = 2.593 (86%) and *Balanus trigonus* Darwin, 1854 ITS = 2.199 (73%)] that were also included in the expected NIS scenario have also been found in the samplings from this study (Table 2). Globally, from the total list of 252 NIS (including

species with and without habitat suitability information), 12 species have already been reported in Gijón (6 were species detected in the samplings from this work) and 52 species (25.6%) in the Bay of Biscay (Supplementary Tables 3, 4).

DISCUSSION

In this work, a representative case study in one of the main seaports in the Atlantic arc (Port of Gijón) was conducted as a proof of concept to be implemented in other international seaports with the aim of reaching real “blueports” (ports with an adequate management of resources and wastes that put in practice new strategies to avoid biological invasions) according to EU policy (European Commission, 2012). Prevention and early detection of NIS must be at the core of current port policies since they are the most cost-effective strategies for NIS management (CBD, 1992). However, some species may slip through prevention and detection efforts and might establish and become abundant pests (Lodge et al., 2006). Prioritization of efforts for particular “target species,” defined as species previously identified as high invasion risks within a port, is indispensable for the correct management of biological invasions (McGeoch et al., 2016). On the other hand, biological port surveys set a baseline of native and non-indigenous port biodiversity. This knowledge can help in both, early NIS detection avoiding the proliferation and spread of NIS, and in the establishing of effective IAS management strategies (e.g., Ojaveer et al., 2014, 2018; Miralles et al., 2016, 2018; Marraffini et al., 2017).

The present study represents the first step to establish a NIS threat assessment procedure in the Port of Gijón and in all ports connected with it via maritime traffic. This biodiversity baseline can be consulted in future NIS assessments for the port. As a result, a list of potential NIS ranked by their Invasion Threat Scores (NIS-ITS) is now available. The biota list is now ready to be used by stakeholders and researchers, and should be updated from further samplings, and include species detected in the area from other studies. Fixing a specific score to consider a species a threat will always be a raw approximation or a little bit subjective decision. For example, Gallardo et al. (2016) set the threshold in 12 points out of the maximum 16 that a species can obtain. Similarly, Tsiamis et al. (2020) set the limit in 38 points out of 48, so that species above this score will be considered as high-risk or high-likely invaders. Both studies use expert judgment to establish this threshold. In the case of NIS-ITS, authorities should pay special attention to those species over the 90% NIS-ITS (high likely invaders). Even more, most of the species with known invasive potential (invasion history score = 1) showed invasion threat scores over 70% (Figure 3). In fact, a *posteriori* analyses revealed that those species that according to the Spanish catalog of invasive species (BOE, 2013) are classified as IAS are all above the 70% threshold, except *Caulerpa taxifolia* Agardh, 1817, which obtained a 60% score due to very low habitat suitability in the Bay of Biscay area (Supplementary Table 3). These results indicate the importance of carefully studying the potential threat of those species with more than 70% NIS-ITS, which show specific traits that can make them especially dangerous.

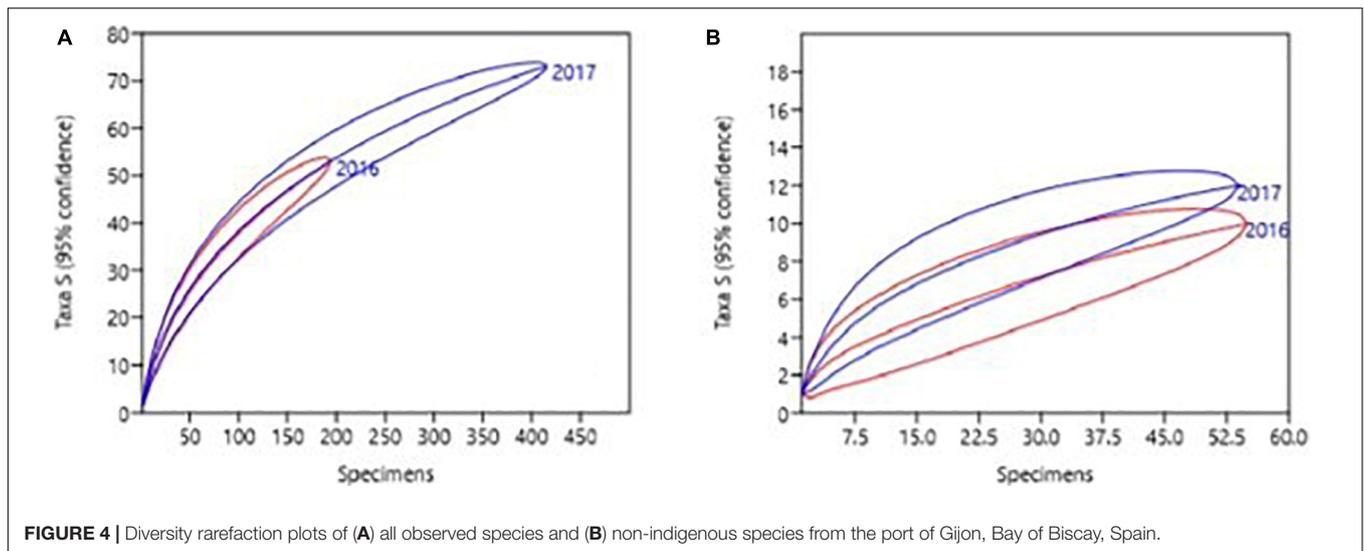


FIGURE 4 | Diversity rarefaction plots of (A) all observed species and (B) non-indigenous species from the port of Gijon, Bay of Biscay, Spain.

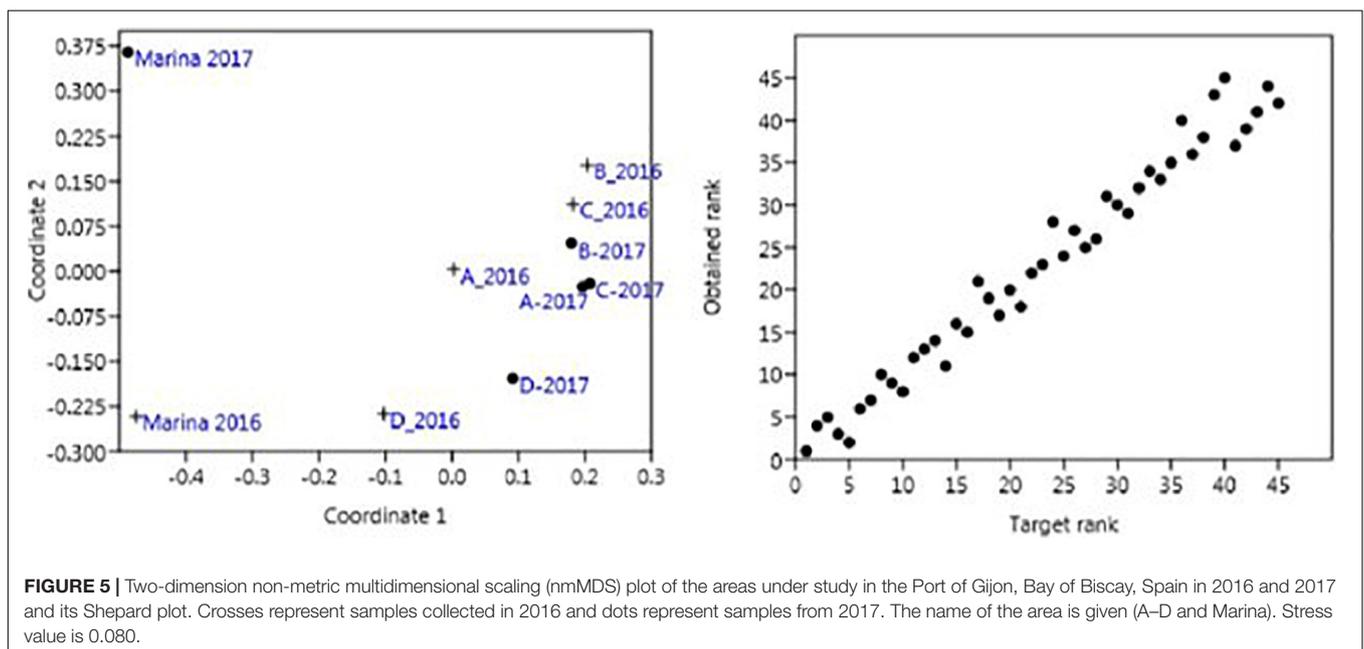


FIGURE 5 | Two-dimension non-metric multidimensional scaling (nmMDS) plot of the areas under study in the Port of Gijon, Bay of Biscay, Spain in 2016 and 2017 and its Shepard plot. Crosses represent samples collected in 2016 and dots represent samples from 2017. The name of the area is given (A–D and Marina). Stress value is 0.080.

Given this information, it is expected that specific port prevention strategies would be enriched and environmental monitoring would be developed in advance and thus improved. Currently, in addition to physical surveys, environmental DNA (eDNA) assays could also be a useful alternative for determining the presence of species based on organism vestigial particles that remain in the environment, negating the need to catch or observe the organism (Darling and Mahon, 2011; Borrell et al., 2017). It is also possible to design species-specific primers for the rapid, early detection of certain species of interest within a seawater community using eDNA (e.g., Devloo-Delva et al., 2016; Miralles et al., 2019).

The new NIS-ITS can be implemented and used by any other international port as a first line of defense against biological invasions without the need of experts. The index contains

three components, and two of them (habitat suitability and invasiveness/impact) are based on joint efforts by researchers and are currently public and available for all stakeholders concerned about this challenge. The AquaMaps⁷ and the Global Invasive Species Database⁸ websites are constantly updated and improved, and they have incorporated relevant tools such as the Hawkins et al. (2015) and Palomares and Pauly (2020) approaches, respectively. The third NIS-ITS component is a relatively easy-to-create port database that compiles all the destinations and cargo origins (traffic) within a port (e.g., 15 years for our case study at the port of Gijon). However, this score can be calculated in any case without a minimum data limitation. This last component

⁷www.aquamaps.org

⁸<http://www.iucngisd.org/gisd/search.php>

TABLE 2 | The number of non-indigenous (NIS) and invasive (IAS) specimens observed in macroinvertebrates samplings conducted in two consecutive years (2016, 2017) in the port of Gijón, Bay of Biscay, Spain.

n	Class	Species	Status in Gijón	Sampling		
				2016	2017	Vessels
1	Anthozoa	<i>Anthopleura anjunae</i> Den Hartog and Vennam, 1993	NIS	1	1	
2	Anthozoa	<i>Anthopleura elegantissima</i> (Brandt, 1835)	NIS	2	2	
3	Gymnolaemata	<i>Bugula neritina</i> (Linnaeus, 1758) ⁺	IAS	1	1	
4	Gymnolaemata	<i>Crassimarginatella papulifera</i> (MacGillivray, 1882) ⁺	NIS	1		
5	Gymnolaemata	<i>Watersipora subtorquata</i> (d'Orbigny, 1852) ⁺	IAS	3	9	
6	Asciadiacea	<i>Diplosoma listerianum</i> (Milne Edwards, 1841) ⁺	IAS			3
7	Asciadiacea	<i>Microcosmus squamiger</i> Michaelsen, 1927	NIS		1	
8	Polychaeta	<i>Dipolydora capensis</i> (Day, 1955)	NIS		1	1
9	Polychaeta	<i>Phyllodoce groenlandica</i> Örsted, 1842	NIS	1	2	
10	Polychaeta	<i>Spirobranchus laticapus</i> (Marenzeller, 1884)	NIS		2	1
11	Polychaeta	<i>Spirobranchus taeniatus</i> (Lamarck, 1818)	NIS			1
13	Bivalvia	<i>Magallana gigas</i> (Thunberg, 1793)	IAS	18	14	
15	Bivalvia	<i>Mytilus trossulus</i> Gould, 1850	NIS	1	5	
16	Bivalvia	<i>Ostrea stentina</i> Payraudeau, 1826	NIS	2		
17	Bivalvia	<i>Talochlamys multistriata</i> (Poli, 1795)	NIS		2	
18	Hexanauplia	<i>Balanus trigonus</i> Darwin, 1854	NIS		2	6
	Total			30	42	12

Those species marked with plus symbol (+) are colonial species. In these cases, the number of specimens is the number of representative units of each colony (size 2 mm³, number of individuals can vary).

could also be improved if the complete, and accurate, itinerary of the cargo were available in all port registries and if treatments of the vessels hulls and ballast water management procedures were known. The full history of ships beyond immediate cargo origin can be important since hull foulers can be entrained in all locations visited. In the case of ballast water, knowing the origin and history or the water being discharged into Gijón is important and its origin is likely not where ships sourced cargo. Unfortunately, complete history of the vessels travels and treatments/procedures are not always available. Despite this, we propose how to incorporate BW and Hull Fouling treatments (if available) in our maritime traffic element when calculating the NIS-ITS in future studies.

This work is consistent with previous studies (Cardeccia et al., 2018) that indicate the need to collect data about the biological and ecological traits of NIS. These data can be used to generate habitat suitability information, which is necessary to calculate indices such as NIS-ITS that may facilitate the prevention of biological invasions. In fact, due to the lack of this information, we obtained only approximately 30% of the complete habitat suitability data for all the potential NIS travelers identified from the literature screening. At the same time, for some species with wide and abundant distributions in the Bay of Biscay, such as *Magallana gigas* (Thunberg, 1793), very low values of habitat suitability were found (0.110) (**Supplementary Table 3**). Most likely, aquaculture practices, which could imply diverse sources and sizes of the inoculums, periodic introduction events and differential extents of bottleneck processes in the new environments (Roman and Darling, 2007), could be a modulating factor currently not considered (but relevant) when predicting habitat suitability for those species. In the same way, although

the IUCN Environmental Impact Classification for Alien Taxa (EICAT) was developed a few years ago and it is in use, data related to the impacts that are produced in the invaded areas are still lacking for many relevant IAS (Ojaveer and Kotta, 2015). One of the main recommendations that arises from this work is that more efforts and initiatives must be developed (and funded) to collect biological data for the main IAS and NIS around the world and that this information must be integrated into databases that are easily accessible for any type of stakeholder.

Another possible biological invasion risk for the port of Gijón, which was not considered in the newly proposed NIS-ITS, could originate from other surrounding marinas. There have been multiple cases where recreational vessels, yachts and boats (not industrial traffic) have been cited as either the primary introduction vector or secondary spread vector for an invasive species (Floerl and Inglis, 2005; Davidson et al., 2010; Clarke-Murray et al., 2014; Ulman et al., 2019; Iacarella et al., 2020). For large industrial ports, traffic data are often accessible and public but recreational vessel traffic data are hard to collect and compile (Le Tixerant et al., 2018). A relevant tool to integrate this information into biological invasion predictions could be tracking ship movements with an automatic identification system (AIS) (EU, 2002; Merchant et al., 2012; Ulman et al., 2017). Coastal and offshore area data from satellite-based AIS (Loretta, 2016) could provide information about the spatial and temporal distribution of shipping (Shelmerdine, 2015). However, AIS transmitters are only required on larger ships, passenger vessels and large fishing vessels. Thus, millions of recreational vessels worldwide are not required to have an AIS, which makes it difficult to develop AIS-based models (Hermannsen et al., 2019). Despite this issue, it is clear that in the future, NIS Invasion

Threat Scores should include recreational vessel traffic data from surrounding marinas when estimating a specific port's ITS and predicted scenarios.

Despite the previously mentioned limitations, in the predicted port of Gijón NIS scenario based on the new ITS, we found 15 marine species with values above the 90% ITS. After sampling macroinvertebrates on the port structures and in the hulls of vessels (as conducted during routine water quality inspections within ports; see Borja et al., 2015), 26.7% of these high-threat species were observed in the port. The outcome revealed that eight other NIS have also been reported in or near the port of Gijón (south Bay of Biscay) (Table 1). These results increase the percentage of NIS findings with regard to the previously described expected NIS scenario with an 80% coincidence level. This scenario could indicate that expected species arrivals based on the NIS-ITS could be, even with the current limitations, close to reality and should be considered at the time of developing prevention/management port strategies. In the case of Gijón, specific management plans must be developed for the species already present in the port. This is also necessary for those NIS that are not present in the port yet but were classified as high-threat species by the NIS-ITS, indicating their potential to become future invaders that are spread by port industrial activity and traffic (Colautti et al., 2006; Floerl et al., 2009; Bishop and Hutchings, 2011; Ojaveer et al., 2018).

One important discovery here was that three species were found on boat hulls and not in the port where they were sampled in. Considering that only 3 out of 22 ships authorized hull samplings, it seems that Spanish port policies should be updated in correspondence with the great threat that biological invasions are for biological diversity. When vessels move from one port to another, organisms can fall off the hulls and may become introduced into a new host environment (Dobroski et al., 2017). Thus, mandatory routine monitoring, along with periodic inspections on arriving boats should be implemented to detect new arrivals of NIS (Bishop and Hutchings, 2011), and marine stakeholder participation (authorities, ship owners, scientists and citizens) will be fundamental.

The port of Gijón is not only a destination for exotic species, it can also be the origin of new introductions to other regions. For example, native species that commonly inhabit the port structures can be transported to other areas outside their native range where they can display invasive features. This study also provides a list of native species at risk of being inoculum (port of Gijón as a donor) to spread/start new invasion processes overseas, since the results of the bibliographic search showed that they have the ability to use maritime traffic as a vector for expansion. As an example, *Carcinus maenas* (Linnaeus, 1758) is native to the European coasts (where the port of Gijón is located) but has been listed as one of the “100 worst invasive species” as it has become invasive in areas like the northwest Pacific or the southwest Atlantic because it is a voracious predator that affects epibenthic and infaunal species through predation, competition and burrowing activities (Pollard and Hutchings, 1990; Bravo et al., 2007; Briski et al., 2011; Edgell and Hollander, 2011; Lim et al., 2017). This is why this list can be very useful in other regions and these species can be considered in future risk assessments.

Globalization has increased international marine traffic, particularly the shipping industry (Tournadre, 2014; Halpern et al., 2015), by moving not only people and goods but also NIS around the world (Hulme, 2009). It is urgent to develop effective prevention strategies and management plans of biological invasions. Each international port should develop its expected NIS scenario and prepare themselves for early/rapid detection of high-risk species. Thus, the use of the new NIS-ITS index proposed here (where there is not the need of experts) and the implementation of species-specific detection tests using eDNA or species-specific physical sampling strategies within ports will be of help for that. Finally, annual biota monitoring within ports is fundamental and must be mandatory. These data will help update global data about port biodiversity (and its risks) and to obtain substantially more insights into NIS colonization dynamics and a clearer picture of the real influence of maritime traffic on marine biological invasion processes worldwide.

Together with these ecological studies, it is necessary to put into practice other strategies related to the management of shipping activities that facilitate the spread of NIS, such as the implementation of mandatory periodic hull cleanings, discharges of ballast water at least 200 nautical miles from the nearest land or the implementation of specific policies related to the boat treatments in recreational boating that can avoid the dispersal of species (Anwar, 2016; IMO, 2018). The combination of these methods can lead to a better understanding and management of biological invasions and a significant reduction of environmental risks, leading to a real blue economy.

DATA AVAILABILITY STATEMENT

The datasets generated for this study can be found in the online repositories. The names of the repository/repositories and accession number(s) can be found below: <https://www.ncbi.nlm.nih.gov/genbank/>, MN185333-MN185374; MN164033-MN164348.

AUTHOR CONTRIBUTIONS

LM: methodology, formal analysis, investigation, writing—original draft, and writing—review and editing. AI: methodology, formal analysis, investigation, visualization, and writing—review and editing. MG: project administration and funding acquisition. EG-V: conceptualization, supervision, project administration, and funding acquisition. YB: methodology, investigation, conceptualization, supervision, project administration, and funding acquisition, and writing—review and editing. All authors contributed to the article and approved the submitted version.

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