



Gorgonians Are Foundation Species on Sponge-Dominated Mesophotic Coral Reefs in the Caribbean

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Foundation species (FS) regulate ecological processes within communities often facilitating biodiversity and habitat complexity. Typically FS are dominant structure-forming taxa; but less dominant taxa having disproportionate ecological impacts to the community can also be FS. Mesophotic coral ecosystems (MCEs) are deep coral reef (~30–150 m) communities, often dominated by emergent sponges in the Caribbean Basin. Despite the potential competitive advantage of sponges on MCEs, gorgonians are also common constituents of these reefs. Data from the Bahamas demonstrate increased biodiversity and densities of sponges on mesophotic reefs with gorgonians relative to reefs without these species. Drawing upon fifteen years of field surveys at five sites in the Caribbean Basin we assessed *in situ* interactions between gorgonians and sponges to quantify outcomes consistent with competition (i.e., tissue necrosis and overgrowth). Gorgonians were effective competitors against a variety of sponges, and two allelochemicals produced by *Ellisella elongata* were mechanistically important in interactions with *Agelas clathrodes*. We also examined invertebrate recruitment patterns near gorgonians to assess their role in facilitating MCE biodiversity. Our results indicate that live gorgonians, *Antillogorgia bipinnata* and *E. elongata*, facilitate biodiverse recruitment into MCEs, indicating that this process is governed by more than passive hydrodynamics. Collectively, these data indicate that these gorgonians exhibit both positive and negative ecological interactions (i.e., facilitation and competition, respectively) with sponges, and other taxa. Thus, these gorgonians are FS of MCE communities within the Caribbean Basin that display several traits contributing to the ecological structure of these understudied communities.

Keywords: allelopathy, biodiversity, competition, facilitation, foundation species, gorgonians, mesophotic coral ecosystems, recruitment

INTRODUCTION

Foundation species (FS) are critical constituents of a community that promote the success of other species by regulating ecological processes (Dayton, 1972); they are typically dominant structure-forming taxa that occupy low trophic positions within an ecosystem (e.g., red mangroves: Ellison et al., 1996; giant kelp: Lamy et al., 2020). However, less dominant and undersized species can

also be FS (ark cockle: Gribben et al., 2009). These FS can be further parsed into primary FS responsible for the initial habitat modification that facilitates community structure and function (Dayton, 1972), and secondary FS (i.e., obligate FS: require primary habitat modification) that add important biocomplexity and/or biodiversity to the habitat (Thomsen et al., 2018). Multiple FS can coexist within the same habitat, or they can live in contiguous zones (“nested” and “adjacent” assemblages, respectively: Angelini et al., 2011). In marine communities, FS provide important ecosystem services that augment biodiversity through substrate stabilization (Wulff, 1984), water filtration (Altieri and Witman, 2006), and/or nutrient transfer (Fiore et al., 2013), as well as defense of other species (Hay, 1986). However, corals are often treated as a single functional FS regardless of the number of species, and their roles, within the community (Knowlton and Jackson, 2001). Recently, gorgonian corals have been identified as FS that enhance community diversity and ecological functioning (McLean and Yoshioka, 2007; Verdura et al., 2019). They can facilitate larval settlement (Ponti et al., 2014; Privitera-Johnson et al., 2015), and enhance sponge growth (Mercado-Molina et al., 2018), possibly by altering their hydrodynamic environment (Gili and Coma, 1998). Moreover, gorgonians provide significant structure and function to shallow (1–30 m) and deeper (30–100 m) coral reefs throughout the Caribbean Basin (Tsounis et al., 2018; Sánchez et al., 2019; Lasker et al., 2020).

The worldwide decline of corals, due to natural and anthropogenic stressors (Hughes et al., 2017), has resulted in regime shifts from hard coral dominance to algal, sponge, or gorgonian dominance (Norstrom et al., 2009). The roles of increasing algal cover on coral reefs has been well documented (McCook et al., 2001; Hay and Rasher, 2010), but all three structural groups are active competitors with corals, and with each other (Chadwick and Morrow, 2011). Although ecological processes such as overgrowth and/or shading are common amongst these taxa, contact-mediated or water-borne delivery of allelopathic compounds is another competitive option for some species of algae, sponges, and gorgonians (Granéli and Pavia, 2006; Slattery and Gochfeld, 2012). For example, allelochemicals from the Brazilian gorgonian, *Phyllogorgia dilatata*, reduced growth rates in the zoanthid, *Palythoa caribaeorum* (Ribeiro et al., 2017). Likewise, 29 gorgonian species exhibited competitive interactions that effected spatial distributions of this taxa in St. John USVI (Gambrel and Lasker, 2016). In addition, taxonomic relatives (i.e., Indo-Pacific soft corals) also demonstrate significant allelopathic activities against hard and soft corals (Maida et al., 2001; Slattery et al., 2008).

Mesophotic coral ecosystems (MCEs) are deep reef communities (~30–150 m depth) structured by strong gradients of solar irradiance and food availability (Lesser et al., 2009, 2018). They are typically less affected by land-based anthropogenic stressors and are often below the depth limits of natural stressors affecting shallow reefs (but see: Lesser and Slattery, 2011; Smith et al., 2016; Slattery et al., 2018). This has led to speculation that MCEs, as a whole, are relatively stable habitats that might serve as important refuges or nursery grounds for depth-generalists that could reseed damaged shallow reefs through larval connectivity

(Bongaerts et al., 2010; Slattery et al., 2011) although recent data indicate this “deep reef refuge hypothesis” is clearly context-dependent (e.g., Bongaerts et al., 2017). Throughout the Caribbean Basin MCEs are often dominated by habitat-forming sponges. Sponge biomass and density increase dramatically within these MCEs (Slattery and Lesser, 2012; Lesser and Slattery, 2018), and there is a clear demarcation between the transitional communities of the upper mesophotic (~30–60 m) and the endemic communities of the lower mesophotic (~61–150 m) (Lesser et al., 2019). While sponges dominate many MCE communities (Pomponi et al., 2019), other taxa, including gorgonian corals, also co-occur at high densities at these depths (Sánchez et al., 2019). Here we examine the ecological roles of the dominant species of Caribbean MCEs (i.e., sponges and gorgonians), and assess the relationships between these species through strong positive and negative interactions (i.e., facilitation and competition, respectively). Specifically, we test the hypotheses that: (1) gorgonians are the primary FS of MCEs, and (2) the gorgonian *Ellisella elongata* structures MCEs with allelopathic compounds.

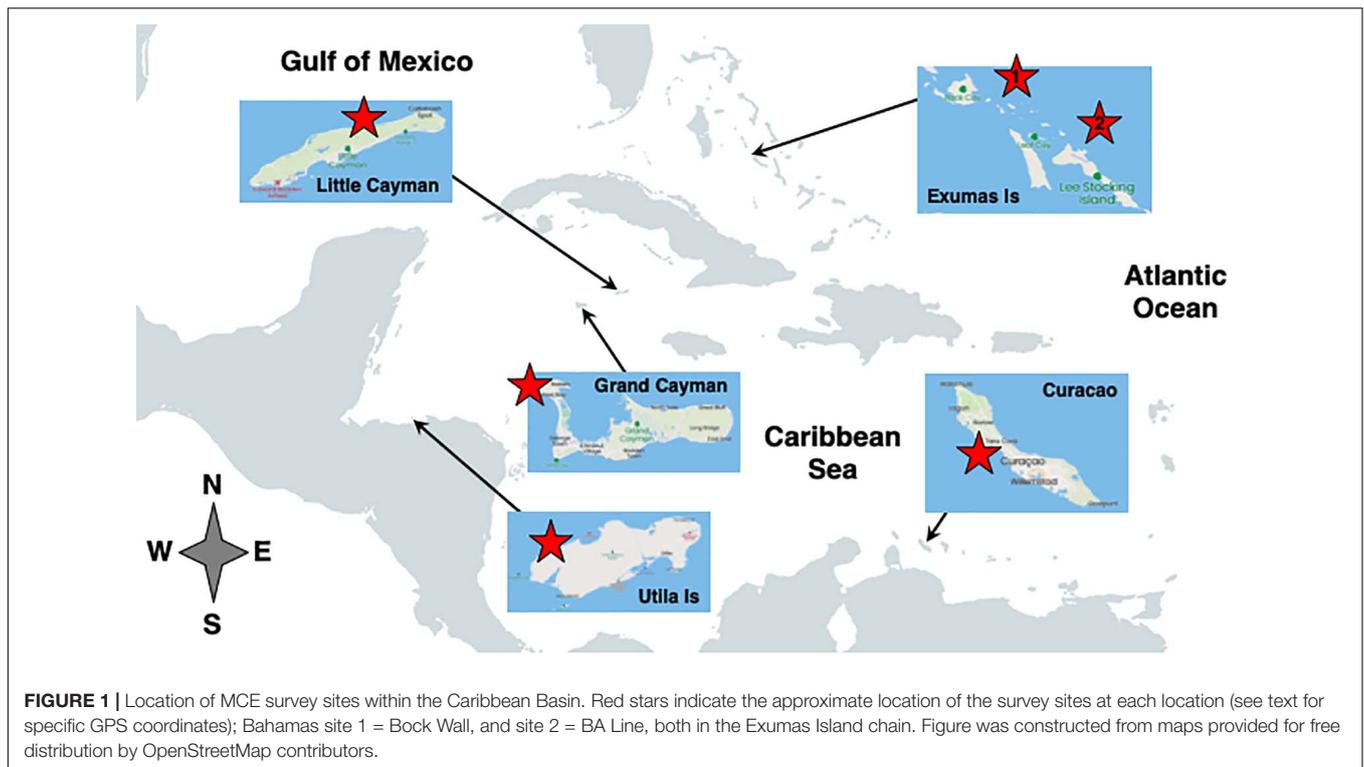
MATERIALS AND METHODS

Study Sites

This study was conducted on MCEs in the Bahamas (Exumas Archipelago, Bock Wall: 23°46.5' N, 76°05.5' W; BA Line: 23°46' N, 76° 06' W), the Cayman Islands (Grand Cayman, Sand Chute: 19°21.43' N, 81°24.44' W; Little Cayman, Rock Bottom Wall: 19°42.03' N, 80°03.25' W), Curaçao (Buoy 1: 12°7.33' N, 69°03.8' W), and Honduras (Utila, CJ's Drop-off: 16°4.59' N, 86°55.1' W) over the period from 2003 to 2018 (Figure 1). Research in the Bahamas included field surveys, as well as manipulative laboratory and field experiments, over multiple years, while research at the other locations represent a few weeks of opportunistic field surveys. Data were collected via open-circuit and closed-circuit (i.e., rebreather) trimix technical dives at depths of 30–91 m. Given the extensive decompression schedules associated with the deeper dives, fewer replicate quadrats and transects were conducted at those depths; this resulted in unequal sample sizes relative to gorgonian and sponge density assessments.

Gorgonian and Sponge Densities at Mesophotic Depths

Gorgonian and sponge diversity and abundance were surveyed in replicate 1 m² quadrats ($n = 5–10$) at random points along replicate 30 m transects ($n = 3–9$) at depths of 30, 46, 61, 76, and 91 m at all of the aforementioned sites, except Curaçao where surveys were conducted to a maximum depth of 61 m (Slattery and Lesser, 2019; and see **Supplementary Table 1** for total replicates at each depth). Our sponge density data have been reported previously (Lesser and Slattery, 2018) for all sites except the BA Line, Bahamas, which we compared to Bock Wall, since the former site lacks gorgonians, and the latter site has a dense gorgonian population. These data were assessed using a two-way analysis of variance (ANOVA) with site and depth as the



fixed factors. Significance in any of the main effects were further parsed using Tukey's HSD *post hoc* test. Prior to all analyses, the data were tested for normality and homogeneity of variance and transformed as needed. All data within are presented as the mean \pm SE.

Gorgonian abundance throughout the Caribbean Basin was analyzed using a one-way ANOVA, with site as the fixed factor. Due to the patchy distribution of gorgonians on these MCEs, all gorgonians were combined into "upper mesophotic" (30 and 46 m) and "lower mesophotic" (61, 76, and 91 m) groupings, consistent with the ecological zones described in Lesser et al. (2019), and the depth differences were tested using an unpaired Student's *t*-test. The abundance of the dominant gorgonian species in the upper and lower mesophotic zones, *Antillologorgia bipinnata* and *Ellisella elongata* (formerly *Pseudopterogorgia bipinnata* and *E. barbadensis*; Castro et al., 2010; Williams and Chen, 2012), were also compared using an unpaired Student's *t*-test.

Facilitation of Invertebrate Recruitment by Gorgonians

At each of our five sites within the Caribbean Basin we quantified invertebrate recruitment near the bases of *A. bipinnata* and *E. elongata* colonies, as well as to bare substrate further away, in the upper and lower mesophotic zones. The larvae of sponges, hard corals, and gorgonians typically recruit to rocky substrata and/or components of that substrate such as crustose coralline red algae (Maldonado, 2006; Ritson-Williams et al., 2009) that are common on many MCE reefs throughout the Caribbean

(Slattery and Lesser, pers. obs.). Replicate 5 \times 5 gridded m² quadrats ($n = 4$) were positioned around conspecific patches of gorgonians ($n = 1-4$ colonies) at 45 \pm 3 m and 75 \pm 3 m, except in Curaçao where the lower mesophotic recruitment surveys occurred at 61 \pm 3 m. We planned to assess recruitment near several dominant sponges (e.g., *Agelas conifera*, *Plakortis* spp., and *Xestospongia muta*) but there was either no recruitment or limited conspecific recruitment surrounding these taxa (Slattery and Lesser, pers. obs.). Replicate quadrats ($n = 4$) were positioned on "hardpan" substrate at each depth as a control for larval settlement cues (i.e., chemical and physical cues; Pawlik, 1992), and at least 5 m from any other conspecific or heterospecific gorgonians. However, a few of these plots ($n = 11$ of 80) included some turf and/or foliose algae, and/or small sponges (<5% cover of these structural groups). Within each quadrat, all invertebrate recruits were identified to the lowest possible taxon, counted, and their position relative to the central gorgonian(s) was noted. The recruit diversity (H') and richness (S) were compared using a three-way ANOVA with site, depth, and treatment (i.e., quadrats: *A. bipinnata*, *E. elongata*, and hardpan) as fixed factors.

To address the influence of inherent quadrat-specific variables within our recruitment survey data, we conducted a field experiment into cleared 1 m² plots at depths of 45 m and 75 m in the Bahamas, near the Caribbean Marine Research Center (CMRC). Quadrats ($n = 12$ at each depth) were affixed to the reef using stainless steel screws, and the substrate within was scoured to hardpan, using wire brushes and paint scrapers (Slattery and Lesser, 2014). *Ellisella elongata* colonies (~30–60 cm tall) were collected from populations at 45 \pm 3 m

and 75 ± 3 m and transplanted into a “common garden” at 35 m, where they were affixed with Z-Spar epoxy compound into numbered PVC caps (2.5 cm diameter; Kim and Lasker, 1997), and allowed to acclimatize for 14 d. Dead gorgonian skeletons were collected from the beaches near CMRC, and the branchlets were trimmed from the main stalk to mimic the whip-like structure of *E. elongata*; these were also affixed into numbered PVC caps and transported to the common garden site. A 3×3 rack of nine “gorgonian caps” spaced 15 cm apart was positioned in the center of each quadrat and cable-tied to stainless steel “eye” bolts anchored in the substrate. These experimental racks contained either living *E. elongata* ($n = 6$) or dead gorgonian skeletons ($n = 6$) at each depth. This density exceeds averages of *E. elongata* recorded in our field surveys (see below), but it is representative of the patchy abundances of gorgonians on our transects (Slattery and Lesser, pers. obs.). Quadrats were examined 12 mo later, and within each quadrat all recruits were identified to the lowest possible taxon and enumerated. These data were analyzed using a two-way ANOVA with depth and treatment (i.e., live vs dead gorgonians) as fixed factors.

Gorgonian–Sponge Interactions

All contact interactions between gorgonians and sponges in the quadrats were quantified, as well as responses for each species pair (categorized as tissue necrosis and/or overgrowth). Although mortality was a potential outcome, we did not include this endpoint since we could not definitively attribute cause of death during single time-point observations. Given the limited number of contact interactions, we consolidated observations from all locations and collapsed the five depths into upper and lower mesophotic. These data were analyzed using a one sample Wilcoxon signed rank test with two potential outcomes (i.e., “win” or “lose”) compared to the null hypothesis (i.e., “draw”; Ladd et al., 2019).

Competition Between Gorgonians, Sponges, and *Ellisella elongata*

The success of *E. elongata* within the sponge-dominated lower mesophotic community was addressed in field and laboratory experiments. Live and dead gorgonian caps were cable-tied in contact with one of three sponge species (*Agelas clathrodes*, *Plakortis angulospiculatus*, or *Xestospongia muta*) at the common garden site. This provided contact treatments (live gorgonians; $n = 12$ replicates per sponge species) and contact controls (dead gorgonians; $n = 12$ replicates per sponge species). In addition, replicate non-contact sponges ($n = 12$ replicates per sponge species) were identified by attaching numbered cow-tags adjacent to each individual. The condition of these contact pairs and non-contact individuals were examined at three time points post-acclimatization (1, 3, and 12 mo) for evidence of overgrowth, tissue necrosis, or mortality. This allowed documentation of species responses through time, as well as competitive reversals (Chadwick and Morrow, 2011). Data were analyzed using the non-parametric equivalent of a repeated measures ANOVA

(= Friedman’s Test) to address the number of interactions with the outcomes: win, lose, or draw. In addition to gorgonian-sponge competition, we also assessed competition between *A. bipinnata* and *E. elongata*. At each of two depths (45 and 75 m), we cabled-tied acclimatized gorgonian caps containing either *A. bipinnata* or *E. elongata* into contact with replicate conspecifics or heterospecifics ($n = 15$ of each combination). The responses of each gorgonian in these contact pairs were recorded after 3 mo and analyzed using a 3×2 contingency table that compared expected (= no allelopathic effect; healthy tissue) to observed rates of necrotic and/or overgrown tissue.

The allelopathic potential of *E. elongata* was assessed in samples from the Bahamas. Approximately 1,500 g wet mass of this gorgonian was collected from each of two depths (45 ± 3 m and 75 ± 3 m). In addition, 60 cc syringe arrays were used to collect replicate water samples at fixed distances of 0.5, 1.0, 1.5, and 2.0 cm from the surface of *E. elongata* ($n = 4$ each; Slattery et al., 1997) at each depth. Samples were returned to CMRC on ice and stored at -40°C for up to 2 mo prior to transport to the University of Mississippi. Tissue samples were freeze dried, weighed, extracted in 2:1 dichloromethane/methanol, and evaporated to dryness in pre-weighed scintillation vials. These crude extracts were separated on a Waters Alliance High Performance Liquid Chromatograph (HPLC) configured with a Photo Diode Array detector and a fraction collector. HPLC peaks, often representing pure compounds, were collected into pre-weighed vials to determine mass relative to the dry weight of the gorgonian. The three most abundant compounds were further characterized using Nuclear Magnetic Resonance (^1H and ^{13}C NMR) and High-Resolution Mass Spectrometry (HRMS) and compared to data from known compounds archived in MarinLit (Jones et al., 2020). These compounds were tested in a sponge cell-based allelopathy assay described in Slattery and Lesser (2014). Briefly, the sponge *A. clathrodes* ($n = 4$ replicates) was disassociated into cell suspensions through sterile gauze, filtered, layered with Percoll, and added to replicate Sedgwick Rafter counting chambers in the presence of the gorgonian crude extracts or pure compounds, with 3-([3-cholamidopropyl]-dimethylammonio) propanesulfonate (CHAPS) as a positive lysis control, and calcium-magnesium-free filtered seawater (CMFSW) as a negative control. The counting chambers were maintained at 25°C under a 12:12 L:D cycle for 48 h, after which the first 100 cells within the microscopic field of view were scored as intact or lysed. Cellular lysis mimics responses of allelopathic interactions in the field (Slattery and Lesser, 2014). Data were analyzed using a two-way ANOVA with depth and treatment (i.e., extract, compounds, or controls) as fixed factors. Water sampled near *E. elongata* was also assessed for allelochemicals using procedures described in Slattery et al. (1997). Briefly, the samples were partitioned across dichloromethane in 1 L separatory funnels, dried under vacuum, weighed, and quantified/identified using the analytical HPLC, NMR, and HRMS techniques described above. The concentrations of the water-borne compounds at four distances from the surface of the gorgonians were plotted, and differences

in the slopes of each dilution curve were analyzed using linear regression.

RESULTS

Gorgonian Densities at Mesophotic Depths

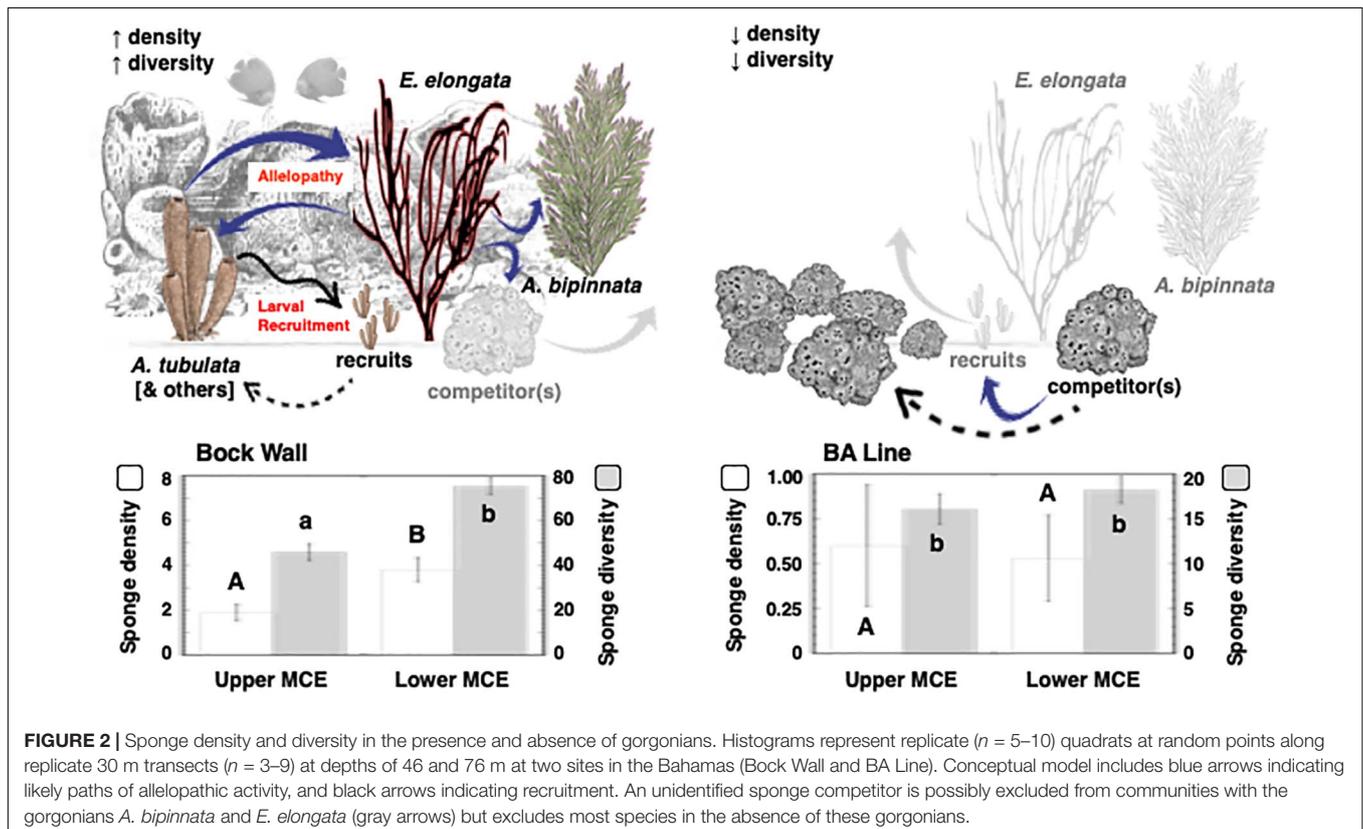
A total of 3,185 gorgonians were observed within 1,000 censused quadrats (Supplementary Table 1). Gorgonian abundance was not significantly different across the Caribbean Basin (ANOVA: $F = 1.756$, $P = 0.1819$; Bahamas: $1.3 \pm 0.3 \text{ m}^{-2}$; Grand Cayman: $1.9 \pm 0.6 \text{ m}^{-2}$; Little Cayman: $2.8 \pm 0.5 \text{ m}^{-2}$; Curaçao: $2.5 \pm 0.5 \text{ m}^{-2}$; Honduras: $4.0 \pm 1.4 \text{ m}^{-2}$) and averaged $2.5 \pm 0.4 \text{ gorgonians m}^{-2}$. However, there were significant differences in gorgonian abundances between the upper ($3.8 \pm 0.6 \text{ m}^{-2}$) and lower ($1.5 \pm 0.2 \text{ m}^{-2}$) mesophotic depths (t -test: $t = 3.713$, $P = 0.0013$). There were also site-specific significant differences between the dominant gorgonians on these reefs (*A. bipinnata* and *E. elongata*) at all depths in the upper mesophotic (Supplementary Table 1, t -tests: $P < 0.05$), but not in the lower mesophotic (Supplementary Table 1, t -tests: $P > 0.05$). When these gorgonians were merged into a Caribbean Basin location, there were significant differences between *A. bipinnata* and *E. elongata* in the upper and lower mesophotic zones (Supplementary Table 1, t -tests: $P = 0.0211$ and $P < 0.0001$, respectively). There were also greater sponge densities, and an increase in sponge species, in the presence

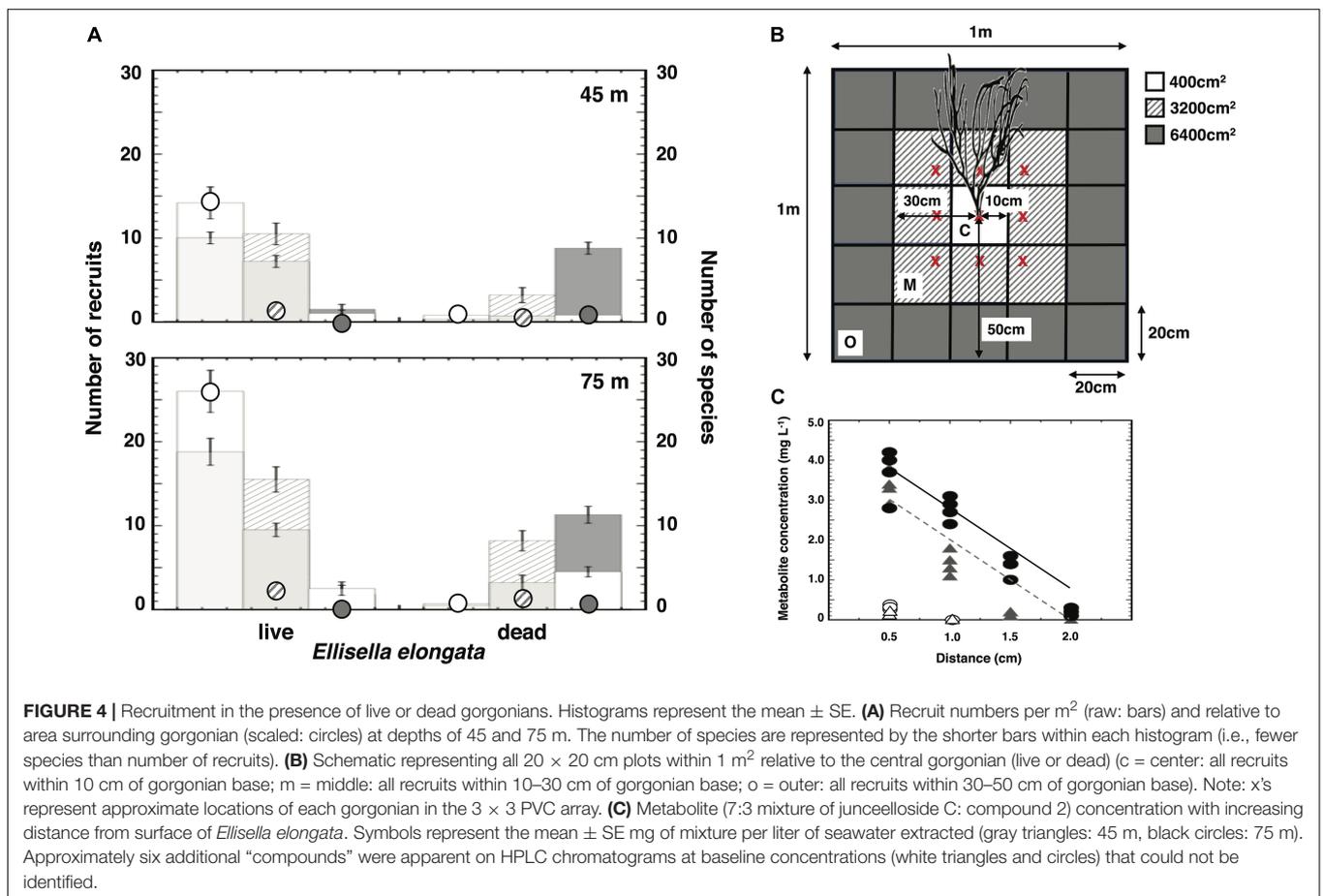
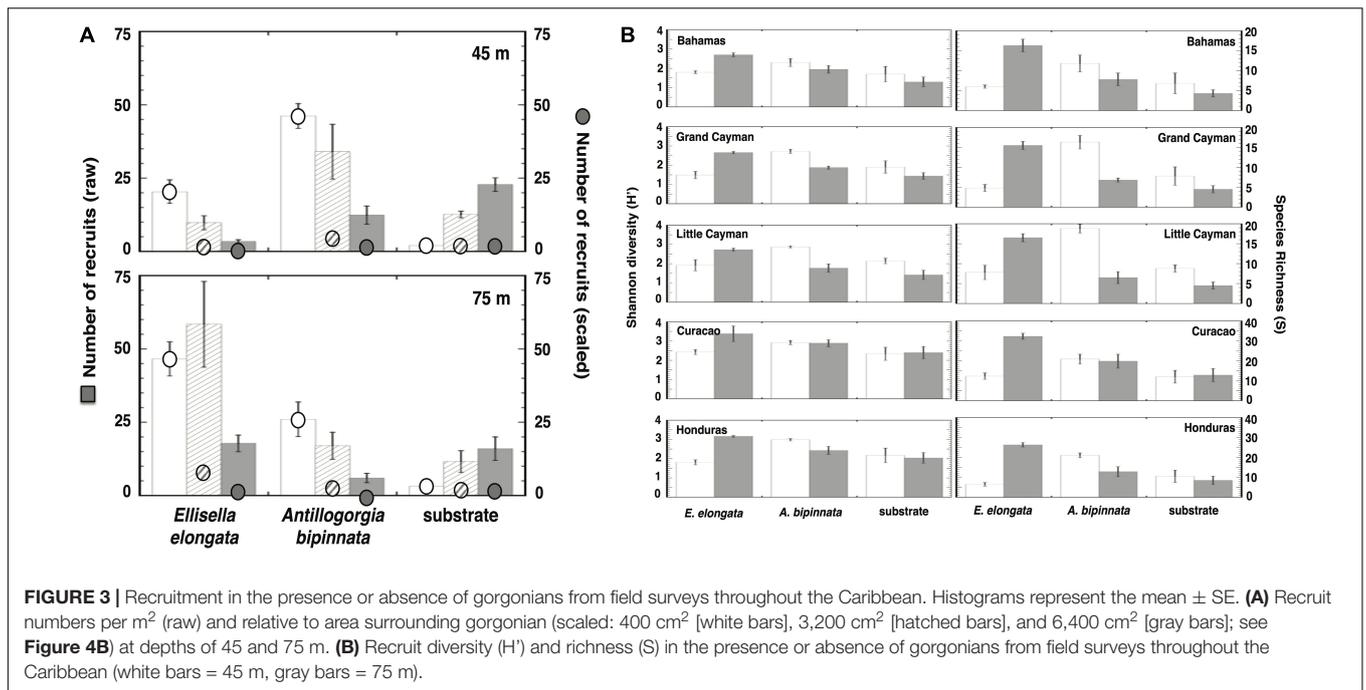
of gorgonians relative to their absence (i.e., Bock Wall vs BA Line, respectively; Figure 2), and the interaction between site and depth was significant (Figure 2; 2-way ANOVAs: $F = 5.593$, $P = 0.0223$ and $F = 21.339$, $P < 0.0001$, respectively).

Gorgonian Facilitation of Invertebrate Recruitment

The number of recruits observed near the bases of live *A. bipinnata* and *E. elongata* during our field surveys were significantly higher than recruits onto bare substrate further away (Supplementary Table 2), and these results were particularly pronounced when data were scaled to the surface area available for recruitment (Figure 3A). At each of our locations within the Caribbean, the Depth \times Distance \times Treatment interaction was significant (three-way ANOVAs: $P < 0.0001$; Supplementary Tables 3–7). In general, diversity and richness were higher around *A. bipinnata* in the upper mesophotic, while diversity and richness were higher around *E. elongata* in the lower mesophotic (Figure 3B and Supplementary Tables 8, 9).

Invertebrate recruitment into cleared quadrats increased in the presence of live gorgonians (Figure 4A). Specifically, there were more recruits and greater diversity near living *E. elongata* than near dead gorgonian skeletons on upper and lower mesophotic reefs. In addition, these metrics declined with increasing distance from live gorgonians (Figure 4B), and the results were particularly pronounced when the data were scaled to the surface area available for recruitment (Figure 4A). There





was a significant Depth \times Distance \times Treatment interaction (3-way ANOVA: $P < 0.0001$; **Supplementary Table 10**) that likely reflected differences in recruitment to the furthest plots relative to live vs dead skeleton treatments.

Gorgonian–Sponge Interactions

Contact interactions between gorgonians and sponges were relatively common across the mesophotic depth gradient and interactions between *Antillogorgia bipinnata*, *E. elongata*, *M. laxa*, or *N. schmitti*, and eight sponges were further examined (749 of 3,185 gorgonians observed; **Table 1**). Sponges were typically the competitive dominant in gorgonian-sponge interactions, with 20 wins vs 6 wins for gorgonians, in all interactions that resulted in an obvious outcome of tissue necrosis (85% of interactions) or overgrowth (30% of interactions). All gorgonian wins resulted in sponge tissue necrosis, and many interactions ($n = 20$) resulted in a draw. Three of the four gorgonians examined (i.e., all but *M. laxa*) exhibited some degree of competitive dominance over at least one sponge species (**Table 1**), and only the sponge *Verongula rigida* was a competitive dominant to all four gorgonians. In 15 of the 64 gorgonian-sponge interactions, there was at least one example of a competitive reversal (i.e., a winning individual within the population of a losing taxa). Finally, 14 of 32 gorgonian-sponge interactions included paired upper and lower mesophotic data; in 10 of those cases competition was significantly greater within the lower mesophotic samples (unpaired t -tests: $P < 0.05$).

Ellisella elongata Competition With Sponges and Other Gorgonians

The contact interactions between transplanted *E. elongata* and the sponges, *A. clathrodes*, *P. angulospiculatus*, and *X. muta*, exhibited species-specific outcomes. Specifically, *E. elongata* was the winner when in contact with *A. clathrodes*, and this outcome was enhanced with increasing time and depth (**Figure 5A-A**; Friedman's Test: tie-corrected Chi Square = 24.111, $P < 0.0001$). In contrast, *P. angulospiculatus* was the winner when in contact with *E. elongata*, and this outcome was also enhanced with increasing time and depth (**Figure 5A-B**; Friedman's Test: tie-corrected Chi Square = 10.330, $P = 0.0057$). However, the interaction between *E. elongata* and *X. muta* was a draw (**Figure 5A-C**; Friedman's Test: tie-corrected Chi Square = 1.059, $P = 0.5890$). Although competitive interactions increased in the lower mesophotic for both *E. elongata* (**Figure 5A-B**) and *P. angulospiculatus* (**Figure 5A-C**), the negative effects on the sponge declined over 12 mo, while effects on the gorgonian did not. Nonetheless, the gorgonian was unable to exert a competitive reversal in this time period. Gorgonians placed in contact with conspecifics for 3 mo exhibited minimal tissue necrosis (= 1 out of 15 pairings between *A. bipinnata*, and 0 out of 15 pairings between *E. elongata*, Chi Square: $P > 0.05$), while heterospecific contact pairings exhibited significant tissue necrosis in *A. bipinnata* (= 12 out of 15 pairings, Chi Square = 28.774, $P < 0.0001$) but no impact to the *E. elongata* colonies.

Extracts and pure compounds from *E. elongata* exhibited allelopathic activity against cells of *A. clathrodes* (**Figure 5B**) that were highly significant (ANOVA: $F = 832.871$, $P < 0.0001$). The crude extracts from 45 to 75 m resulted in 58.5 ± 1.9 and 80.0 ± 2.0 percent cell lysis, respectively, and included three pure compounds that occurred at 2.9–7.6% of the extract mass: 1) funicolide E (Guerrero et al., 1995), 2) (22E,24S)-5a,8a-epidioxy-24-methyl-cholesta-6,22-dien-3b-ol (Ioannou et al., 2009), and 3) juncecelloside C (Jiang et al., 2013) (**Figure 5C**). *Post hoc* multiple comparison testing revealed that funicolide E and juncecelloside C caused significant sponge cell lysis compared to the seawater controls (Tukey's HSD: $P < 0.05$), while compound 2 had no effect on sponge cells (Tukey's HSD: $P > 0.05$, **Figure 5C**). Moreover, these two compounds represented most of the allelopathic effect of the gorgonian crude extract. While the metabolomic profiles of the gorgonians collected at 45 and 75 m were qualitatively identical, there was a significant, and independent, increase in the concentrations of funicolide E (3.6 to 6.4%) with increasing depth that resulted in a significant increase in allelopathic activity (ANOVA: $F = 66.502$, $P < 0.0001$). Sponge cell lysis almost doubled (33.8 ± 2.3 vs. $61.0 \pm 3.5\%$) in samples from 75 m compared to samples from 45 m (**Figure 5C**). The positive lysis control, CHAPS, caused about 90% cell lysis in *A. clathrodes* consistent with its measured bioactivity (Slattery and Lesser, 2014). There were no interactions between treatment and depth for the allelopathic assays (ANOVA: $P > 0.05$).

A 7:3 mixture of juncecelloside C and compound 2 was isolated from seawater collected at distances of 0.5 to 2.0 cm from the surface of *E. elongata* (**Figure 4C** and **Supplementary Table 11**). The concentration of this mixture in seawater exhibited a significant decline with distance from the gorgonian (2-way ANCOVA: $P < 0.0001$), as well as a significant increase with increasing depth (i.e., 45–75 m; 2-way ANCOVA: $P < 0.0001$). At least six additional HPLC peaks, likely representing specific gorgonian compounds, were noted on the HPLC chromatograms, but the baseline concentrations (**Figure 4C**) could not be further purified for MS characterization.

DISCUSSION

Gorgonians Are Primary Foundation Species of MCEs

Gorgonian abundance and diversity on Caribbean MCEs are lower than nearby shallow coral reefs, in contrast to the Red Sea where these metrics are higher at mesophotic depths (Shoham and Benayahu, 2017). Sponge abundance is similarly high at MCE depths in the Red Sea and eastern Mediterranean (Idan et al., 2018; Eyal et al., 2019), although their morphology is encrusting like most sponges on Indo-Pacific MCEs (Slattery and Lesser, 2012). The lower gorgonian abundance in the Caribbean may be due to competitive interactions with the space-dominant sponge fauna (Lesser and Slattery, 2018; Ladd et al., 2019), potentially mediated by sponge allelochemicals (Thacker et al., 1998; Slattery and Gochfeld, 2012). Nonetheless, gorgonians span the upper to lower mesophotic zones at densities that are second only to those of sponges (Pomponi et al., 2019;

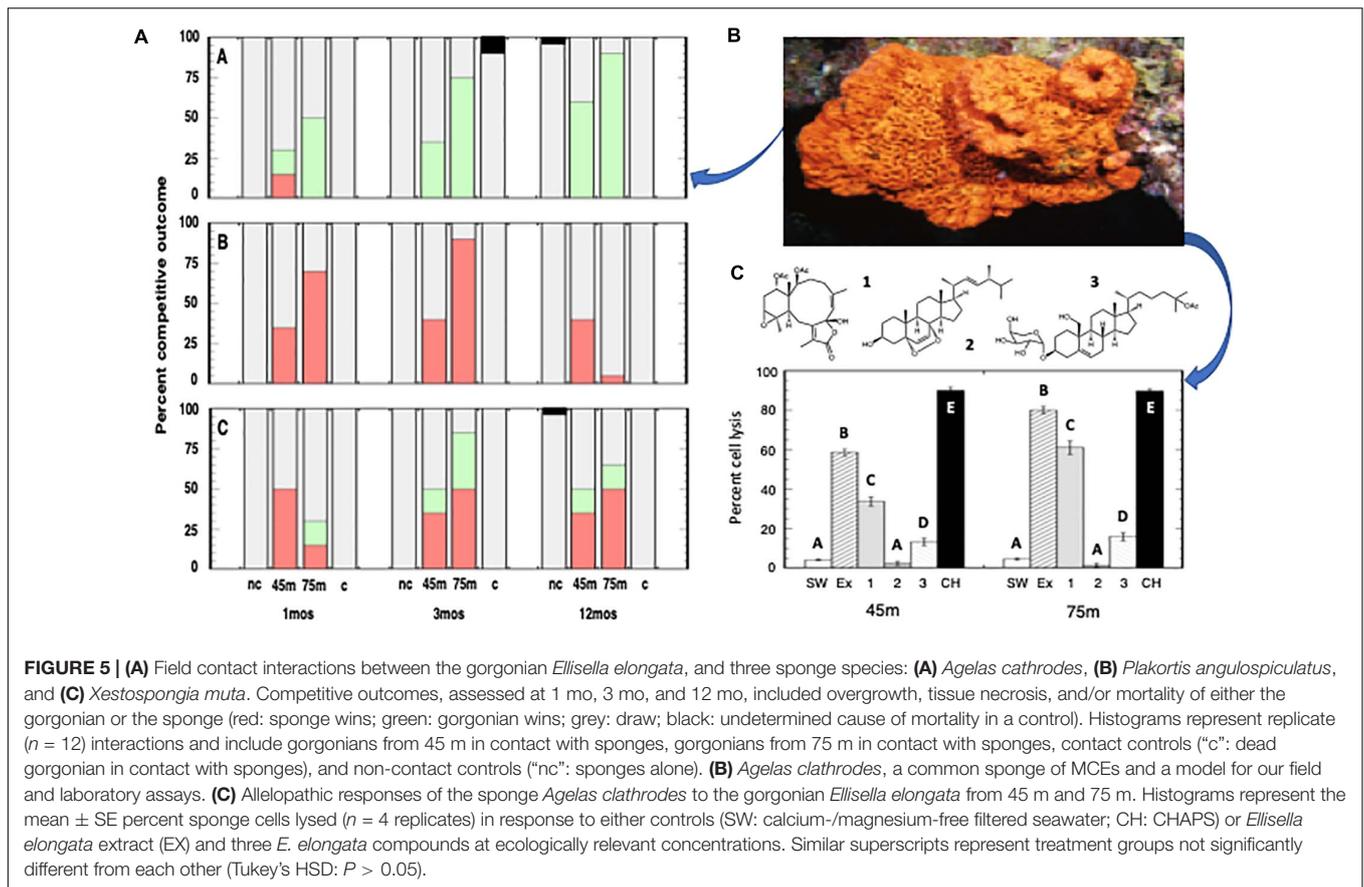
Sánchez et al., 2019). Given the abundance and biomass of sponges on Caribbean mesophotic reefs (Slattery and Lesser, 2012; Lesser and Slattery, 2018), it is not surprising that contact interactions with gorgonians are common. McLean and Yoshioka (2007) noted numerous contact interactions between gorgonians and sponges on the shallow reefs of Puerto Rico, with most interactions resulting in gorgonian mortality. The majority of the interactions documented here on Caribbean MCEs (= 70%) resulted in sub-lethal effects to gorgonians (i.e., limited tissue necrosis and/or overgrowth). Sponges are competitive dominants on coral reefs worldwide (Slattery and Gochfeld, 2012; Wulff, 2012) where they often out-compete corals (Porter and Targett, 1988; Aerts, 1998) and other sponges (Engel and Pawlik, 2000; Wulff, 2005), but not algae (Easson et al., 2014; Slattery and Lesser, 2014). However, competitive reversals by gorgonians were observed in many pairings during this study where sponges were the initial winner. Therefore, gorgonians can coexist with sponges in the MCE sponge belt, as they do in some shallow reef communities (Gili and Coma, 1998; Cúrdia et al., 2013).

While gorgonians do not exhibit the density and biomass of sponges on lower MCEs (Slattery and Lesser, 2012), the most common species on Caribbean MCEs (i.e., *A. bipinnata* and *E. elongata*) play a critical role in the facilitation of benthic invertebrate recruitment. An established MCE community, that lacked these gorgonians (= BA Line) exhibited 3–5 times lower sponge densities and diversity relative to a nearby site with these gorgonians (= Bock Wall). It is possible that site-specific abiotic factors (e.g., sedimentation on BA Line: Liddell et al., 1997) are partially responsible for differences in recruitment. However, our survey data of recruits near these gorgonians, at five locations throughout the Caribbean Basin, validated this observation of gorgonians as facilitators of recruitment which was higher near gorgonians than on bare substrate. Gorgonian forests in the Mediterranean and the USVI have also been shown to enhance recruitment of algae and invertebrates (Ponti et al., 2014, 2018; Privitera-Johnson et al., 2015), probably due to hydrodynamic processes and/or entrainment of larvae within the arborecent landscape (Gili and Coma, 1998). However, we observed higher recruitment into plots containing live *E. elongata*, compared

TABLE 1 | Contact interactions between gorgonians (top row) and sponges (left column) from five sites in the Caribbean Basin.

	<i>Antillogorgia bipinnata</i>	<i>Ellisella elongata</i>	<i>Muricea laxa</i>	<i>Nicella schmitti</i>
<i>Agelas clathrodes</i> upper	[4] 0; n/a	[12] 3G; Z = -1.732; P = 0.0833	[11] 0; n/a	[5] 1G; n/a
<i>Agelas clathrodes</i> lower	[13] 0; n/a	[19] 12G; Z = -3.464; P = 0.0005	[26] 0; n/a	[14] 6G; Z = -2.449; P = 0.0140
<i>Agelas conifera</i> upper	[7] 0; n/a	[17] 6S; Z = -2.646; P = 0.0196	[21] 10S; Z = -3.162; P = 0.0016	[3] 1S; n/a
<i>Agelas conifera</i> lower	[10] 0; n/a	[11] 8S + 1G; Z = -2.333; P = 0.0196	[30] 21S + 1G; Z = -4.264; P < 0.0001	[10] 7S; Z = -1.732; P = 0.0828
<i>Aplysina fulva</i> upper	[8] 2S + 1G; Z = -1.414; P = 0.1573	[10] 0; n/a	[9] 5S; Z = -2.000; P = 0.0455	[16] 0; n/a
<i>Aplysina fulva</i> lower	[15] 9S; Z = -2.530; P = 0.0110	[7] 2G; Z = -1.414; P = 0.1573	[7] 6S; Z = -1.449; P = 0.0143	[4] 0; n/a
<i>Iotrochota birotulata</i> upper	[6] 0; n/a	[9] 4S; Z = -2.000; P = 0.0455	[11] 3S + 1G; Z = -1.000; P = 0.3173	[2] 1G; n/a
<i>Iotrochota birotulata</i> lower	[7] 0; n/a	[15] 11S; Z = -2.250; P = 0.0126	[23] 11S; Z = -3.317; P = 0.0009	[15] 5G + 1S; Z = -2.236; P = 0.0249
<i>Neofibularia nolitangere</i> upper	[3] 2G; n/a	[15] 0; n/a	[7] 0; n/a	[8] 0; n/a
<i>Neofibularia nolitangere</i> lower	[7] 5G; Z = -1.414; P = 0.1573	[11] 2S; Z = -1.414; P = 0.1573	[15] 0; n/a	[17] 0; n/a
<i>Plakortis angulospiculatus</i> upper	[9] 4S; Z = -2.000; P = 0.0460	[9] 4S; Z = -2.000; P = 0.0455	[7] 3S + 1G; Z = -1.000; P = 0.3173	[6] 1G; n/a
<i>Plakortis angulospiculatus</i> lower	[9] 8S; Z = -1.449 P = 0.0143	[28] 16S + 2G; Z = -3.300; P = 0.0010	[15] 9S + 1G; Z = -2.530; P = 0.0114	[14] 7G + 1S; Z = -2.646; P = 0.0079
<i>Verongula rigida</i> upper	[4] 1S + 1G; n/a	[15] 7S; Z = -2.646; P = 0.0082	[6] 3S + 1G; Z = -1.000; P = 0.3173	[11] 3S; Z = -1.732; P = 0.0827
<i>Verongula rigida</i> lower	[8] 4S; Z = -2.000; P = 0.0460	[20] 16S; Z = -4.000; P < 0.0001	[17] 8S + 1G; Z = -2.333; P = 0.0196	[18] 8S; Z = -2.828; P = 0.0050
<i>Xestospongia muta</i> upper	[4] 0; n/a	[18] 7G; Z = -2.646; P = 0.0082	[8] 0; n/a	[8] 2G; Z = -1.414; P = 0.1573
<i>Xestospongia muta</i> lower	[3] 0; n/a	[23] 15G + 1S; Z = -3.500; P = 0.0005	[18] 0; n/a	[11] 8G + 1S; Z = -1.890; P = 0.0580

Upper and lower mesophotic samples represent combined data from 30 to 45 m, and 61, 76, and 91 m, respectively. Parenthetic numbers are the total interactions; the subsequent numbers represent the wins for gorgonians (G) or sponges (S), and an additional number represents a potential competitive reversal (i.e., at least one of the losing taxa is a winner). The Wilcoxon Signed Rank test statistic and P value are included for all interactions that could be analyzed statistically. Note: n/a = "draw" (i.e., no difference in ranks), or no analysis due to limited replicates.



to plots containing dead gorgonian skeletons, indicating that hydrodynamics was not solely responsible for the recruitment patterns we documented. Chemical settlement cues and/or substrate characteristics might also select for the success of recruits (Smith and Witman, 1999; Whalan et al., 2012). Alternatively, the presence of algae in a few of the plots might have inhibited recruitment to control plots via allelopathic mechanisms (Slattery and Lesser, 2014). In contrast to our data from the Caribbean, the presence of gorgonians within the Mediterranean had little effect on the percent cover of recruits, but they did select against certain taxa, and thereby lowered biodiversity, possibly due to size-selective feeding on larvae (Ponti et al., 2014).

Our survey data also demonstrated depth-specific differences in recruitment, with more recruits near *A. bipinnata* at 45 m than at 75 m, and more recruits near *E. elongata* at 75 m than at 45 m. This may be due to the higher population densities of *A. bipinnata* and *E. elongata* at 45 and 75 m, respectively. Yoshioka (1996) reported density-dependent recruitment of 14 gorgonian species into conspecific adult forests in Puerto Rico, implying species-specific "aggregative" settlement cues might be important in recruitment patterns (Pawlik, 1992). While similar trends were observed in 10 genera of gorgonians in St. John USVI, *Antillogorgia* spp. did not exhibit a positive relationship between recruits and adults (Privitera-Johnson et al., 2015) suggesting *A. bipinnata* may not respond to conspecific

settlement cues. Instead, larval settlement patterns might be supported by characteristics of the gorgonian mucus and/or their microbial biofilms (Slattery, 1992; Webster et al., 2004; Whalan and Webster, 2014). Alternatively, depth-specific differences may be a function of species richness and diversity (i.e., S and H, respectively) across the broader Caribbean Basin (*sensu* Witman et al., 2004). However, it is important to note that abiotic factors such as reef slope angle between the upper and lower mesophotic zones, with its influence on irradiances, water flow, and nutrients (Lesser et al., 2018), also varies at our sites.

Taken together, *A. bipinnata* and *E. elongata* facilitate the success of multiple species within MCEs by regulating recruitment processes consistent with the definition of a FS (Dayton, 1972). Moreover, these gorgonian species exhibit inter-taxa competitive interactions resulting in physical segregation and varying dominance hierarchies, throughout the upper and lower mesophotic zones, that are consistent with adjacent FS assemblages (Angelini et al., 2011). Therefore, it is likely that these gorgonians, as primary FS, structure their respective MCE zones to allow additional obligate FS (e.g., sponges) to be successful (Altieri et al., 2007; Sousa et al., 2007).

***Ellisella elongata* Structures MCEs With Allelopathic Compounds**

Given that gorgonians and sponges are both suspension feeders one might assume that would result in a network of resource

competitors, but these species often co-occur without evidence of competitive feeding interactions (Gili and Coma, 1998). In fact, gorgonians are passive suspension feeders that consume relatively large plankton (~4–700 μm in size: e.g., Ribes et al., 1998, 1999), whereas sponges are active suspension feeders that primarily consume the abundantly available bacterioplankton (<1 μm in size: e.g., Lesser, 2006; Hadas et al., 2009; Lesser and Slattery, 2013), resulting in a significant increase in sponge growth, biomass, and density at mesophotic depths (Slattery and Lesser, 2012; Lesser and Slattery, 2018). Thus, food limitation and/or niche overlap seems unlikely to provide any pressure for competition between gorgonians and sponges on MCEs within the Caribbean Basin. Instead, it is possible that the space itself, and the access it provides to flow-mediated trophic subsidies (Vogel, 1977; Sebens, 1984), might be the cause of the observed competitive outcomes (i.e., tissue necrosis and/or overgrowth). In contrast, habitat-partitioning between *A. bipinnata* and *E. elongata* (upper and lower mesophotic zones, respectively) may be due, in part, to differences in their light requirements (e.g., Dattolo et al., 2017). Specifically, the former species hosts photosymbionts, while the latter species is photosymbiont free (Sánchez et al., 2019). However, it is also clear that *E. elongata* is competitively dominant over *A. bipinnata* which might provide the ecological pressure that constrains the latter species to the upper mesophotic zone.

Tissue necrosis often involves the production of allelochemicals that are active through surficial contact and/or water-borne routes of transmission (Granéli and Pavia, 2006), and both gorgonians and sponges produce a diverse array of compounds with known or presumed allelopathic roles (Slattery and Gochfeld, 2012). Tissue necrosis was the most common result of contact interactions for both gorgonians and sponges, suggesting that allelopathy is an important competitive mechanism at mesophotic depths (e.g., Slattery and Lesser, 2014). Although water-borne allelopathic effects are less common, due to dilution of compounds within the surrounding seawater (Slattery et al., 1997), we did note a few isolated sponges with evidence of tissue necrosis. This necrosis may have been due to prior contact during current-mediated abrasion (Box and Mumby, 2007), “retreat” from contact with a dominant species (Hoeksema and De Voogd, 2012), and/or to a true water-borne allelochemical (de Nys et al., 1991). In contrast, Ladd et al. (2019) noted that the gorgonian *Erythropodium caribaeorum* was the most aggressive species on the shallow reefs of Florida, but its competitive success was due to overgrowth rather than allelopathic interactions. That said, most gorgonians have a very small “footprint” relative to the substrate, and horizontal overgrowth is a limited competitive strategy for these species.

We observed competitive dominance by the gorgonian, *E. elongata*, over the sponge, *A. clathrodes*, in field surveys and contact transplants, as well as in a laboratory experiment. Our results indicate that the additive effects of two compounds, funicolide E and juncecelloside C, are responsible for 80–96% of the activity of the gorgonian extract. Funicolide E is briarene diterpenoid, a class of compound that is broadly distributed geographically and taxonomically (Guerriero et al., 1995), and Juncecelloside C is a steroid commonly associated with gorgonians

from the family Ellisellidae throughout their geographic range (Sun et al., 2010). This study is the first report of ecological activity for either compound, although certain briarenes exhibit antifouling activity against barnacles (Qi et al., 2006) and apparently have some feeding deterrent activity (Harvell et al., 1993), and steroids can deter predators (Slattery et al., 1997). The presence of two allelopathic compounds in *E. elongata* is not surprising given the potential importance of competition on MCEs, and this may represent a form of “bet-hedging” against different species and/or species-specific acclimatization to individual compounds (Rasher et al., 2011). It is interesting that lower MCE gorgonians exhibited higher concentrations of funicolide E than their shallow MCE counterparts which resulted in greater allelopathic activity of the lower MCE gorgonians at depths where sponge abundance and biomass were significantly greater (Slattery and Lesser, 2012, 2019; Lesser and Slattery, 2018). Depth-dependent variation may indicate that funicolide E represents an inducible defense (Karban and Baldwin, 1997) to increasing competition with sponges at lower MCE depths (*sensu* Anstett et al., 2016).

Despite the presence of a potential inducible chemical defense, the gorgonian *E. elongata* actively facilitated the recruitment of a diverse benthic community to nearby substrata, including many species of sponges, scleractinian corals, gorgonians, and a tunicate. In the Mediterranean garrigue scrub community, a monoterpene produced by the herbaceous shrub *Thymus vulgaris* repressed the abundance of a space-dominant competitive grass, allowing other rarer species to persist (Ehlers et al., 2014). It is possible that the gorgonian allelochemicals we isolated were selected for their allelopathic effects on a particular space-dominant sponge species (e.g., *A. clathrodes*), and a result of that competitive success was the recruitment of less competitive species into the benthos. These results are consistent with associational defense systems (e.g., algae: Hay, 1986; soft corals: Kerr and Paul, 1995; Ribeiro et al., 2017; sponges: Farren and Donovan, 2007) often attributable to FS (Stachowicz, 2001). Alternatively, these gorgonians may make the substrate more suitable for recruits (sponges, hard and soft corals, and/or tunicates), akin to the role of the Mesquite in semi-arid savannas of the southwest United States (Archer et al., 1988). For example, gorgonians regularly slough mucus and mucus-associated bacteria (Ruble et al., 1980) that might induce larval settlement of multiple invertebrate taxa (e.g., Pawlik, 1992; Slattery, 1992). It is also possible that greater stress occurs on upper MCE reefs from disease or storm activity (e.g., Woodley et al., 1981; Smith et al., 2010, 2016; Frade et al., 2018), compared to the reduced stress on lower MCE reefs (Slattery et al., 2011), and this provides the selective environment for facilitation to override the ecological costs of competition (i.e., the Stress Gradient Hypothesis: Bertness and Callaway, 1994).

In conclusion, common gorgonians of the Caribbean Basin structure the upper and lower mesophotic zones through counterbalanced positive (i.e., facilitation) and negative (i.e., competition) interactions. Although gorgonians are less common than the space dominant MCE sponges on Caribbean reefs, *A. bipinnata* and *E. elongata* are FS that play critical roles in the recruitment and allelochemical-mediated survival

(= associational defense) of diverse benthic communities in the adjacent upper and lower MCE zones (Lesser et al., 2019).

DATA AVAILABILITY STATEMENT

The original contributions presented in the study are included in the article/**Supplementary Material**, further inquiries can be directed to the corresponding author/s.

AUTHOR CONTRIBUTIONS

MS and ML designed the study, conducted the field studies, conducted the statistical analyses, and contributed to revisions. MS conducted the chemical analyses and wrote the first draft. Both authors contributed to the article and approved the submitted version.

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SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fmars.2021.654268/full#supplementary-material>

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Conflict of Interest: The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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