



Single-Turnover Variable Chlorophyll Fluorescence as a Tool for Assessing Phytoplankton Photosynthesis and Primary Productivity: Opportunities, Caveats and Recommendations

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Phytoplankton photosynthetic physiology can be investigated through single-turnover variable chlorophyll fluorescence (ST-ChlF) approaches, which carry unique potential to autonomously collect data at high spatial and temporal resolution. Over the past decades, significant progress has been made in the development and application of ST-ChlF methods in aquatic ecosystems, and in the interpretation of the resulting observations. At the same time, however, an increasing number of sensor types, sampling protocols, and data processing algorithms have created confusion and uncertainty among potential users, with a growing divergence of practice among different research groups. In this review, we assist the existing and upcoming user community by providing an overview of current approaches and consensus recommendations for the use of ST-ChlF measurements to examine *in-situ* phytoplankton productivity and photo-physiology. We argue that a consistency of

practice and adherence to basic operational and quality control standards is critical to ensuring data inter-comparability. Large datasets of inter-comparable and globally coherent ST-ChlF observations hold the potential to reveal large-scale patterns and trends in phytoplankton photo-physiology, photosynthetic rates and bottom-up controls on primary productivity. As such, they hold great potential to provide invaluable physiological observations on the scales relevant for the development and validation of ecosystem models and remote sensing algorithms.

Keywords: variable chlorophyll fluorescence, phytoplankton, photo-physiology, photosynthesis, primary productivity, data synthesis, FRRF

INTRODUCTION

The immense size and inaccessibility of many oceanic regions has historically rendered them under-sampled with respect to key biogeochemical variables, requiring extrapolation of sparse measurements over large areas. In recent years, however, rapid advancement of technologies for data collection and processing has begun to drastically change the notion of the chronically under-sampled ocean; more oceanographic data are now typically acquired in a single year than over the entire preceding century (Tanhua et al., 2019; Brett et al., 2020). The collection of high-resolution *in-situ* data has been led by physical and chemical variables that are amenable to measurement by autonomous sensors, including salinity, temperature, light, and certain nutrients and dissolved gasses. More recently, autonomous measurement systems have shown great potential for providing standardized and inter-comparable *in-situ* observations of plankton standing stocks and diversity on a global scale (Lombard et al., 2019). In contrast, acquisition of globally consistent, high-resolution measurements of phytoplankton physiology and biomass turnover remains challenging. This limits our understanding of phytoplankton productivity, which is a critical component of global biogeochemical cycles, and ultimately controls the carrying capacity of marine ecosystems (Falkowski et al., 1998). Characterizing the potential response of primary productivity to perturbations over a range of scales is one of the key objectives of oceanographic research. Achieving this requires autonomous methods to monitor *in-situ* variability in phytoplankton physiology and productivity at a spatial and temporal resolution comparable to that obtainable for other key oceanographic variables.

Single-turnover variable chlorophyll fluorescence (ST-ChlF) approaches, such as fast repetition rate fluorometry (FRRF), are unique in providing autonomous, instantaneous, non-destructive, and sensitive observations of phytoplankton photosynthetic physiology. Measurements of ST-ChlF can be used to derive insight into the fate of absorbed photons, which, in turn, can be related to photosynthetic capacity. Early oceanographic application of ST-ChlF instruments demonstrated *in-situ* phytoplankton responses to physical forcing (Kolber et al., 1990; Falkowski et al., 1991) and iron limitation (Kolber et al., 1994; Behrenfeld et al., 1996; Behrenfeld and Kolber, 1999), and revealed strong proportionality to estimates of primary productivity derived from ST-ChlF and ^{14}C -uptake

(Kolber and Falkowski, 1993). More recent work indicates that the relation between carbon-based productivity and the photochemical flux in photosystem II (PSII) derived from ST-ChlF measurements (J_{PII} , see Table 1 for abbreviations and units), is modulated by a number of environmental factors that vary regionally (e.g., Lawrenz et al., 2013; Hughes et al., 2020), across seasonal and diel cycles (e.g., Ryan-Keogh et al., 2018; Schuback and Tortell, 2019), and among phytoplankton taxa (e.g., Hughes et al., 2021). This complicates the application of ST-ChlF measurements as a metric of carbon fixation, but also opens important insights into the plasticity of the photosynthetic process in response to environmental and taxonomic variability.

Over the past decades, significant progress has been made in the use of ST-ChlF methods in aquatic ecosystems. Developments in sensor technology have greatly improved measurement sensitivity, with current instruments able to collect robust data in the most oligotrophic waters and from autonomous platforms (Lin et al., 2016; Carvalho et al., 2020). At the same time, new approaches to interpret ST-ChlF data in terms of phytoplankton photo-physiology and taxonomic composition allow for a better understanding of the environmental and taxonomic factors driving variability in derived parameters. With maturing technology and a strengthening theoretical framework, ST-ChlF measurements are poised to contribute significant new insights into the variability of phytoplankton photosynthesis over a range of spatial and temporal scales, enabling us to address organismal and ecosystem-level responses to global change. However, the field now sits at a crossroads, as operational, computational, and conceptual approaches to extract and interpret ST-ChlF derived parameters are rapidly diverging. An increasing number of sensors (both commercial and custom-made), sampling protocols, and processing algorithms for ST-ChlF measurements are being developed, yet no standards for best practice have been formally adopted by the international research community. Rapidly growing data sets may thus become increasingly difficult (if not impossible) to reconcile, leading to a “Tower of Babel” scenario, which would limit our ability to build globally coherent ST-ChlF observations and examine large-scale patterns and long-term trends in phytoplankton physiology.

To address the challenges outlined above, SCOR-WG156 was established to assemble minimum standards of best practice for the acquisition and archiving of aquatic ST-ChlF data. Bringing together instrument manufacturers and users from 10 countries and five continents, our group seeks

TABLE 1 | Notations and terminology.

Parameter	Synonym(s)	Meaning	Derivation	Units
Primary ChlF parameter				
F_o	F_o F_{min}	Minimum ST-ChlF in the dark-regulated state.	Minimum ST-ChlF at beginning of ST-ChlF transient in the dark-regulated state.	Relative units
F_m	F_{max}	Maximum ST-ChlF in the dark-regulated state.	Maximum ST-ChlF from ST-ChlF transient in the dark-regulated state.	
F'	F F_t F_s	Steady-state ST-ChlF in the light-regulated state.	Measured as minimum ST-ChlF from ST-ChlF transient in the light-regulated state. (note that as the fit parameter is F_o , many instruments report the biological parameter F' as F_o)	
F_m'		Maximum ST-ChlF in the light-regulated state.	Maximum ST-ChlF from ST-ChlF transient in the light-regulated state. (note that as the fit parameter is F_m , many instruments report the biological parameter F_m' as F_m)	
F_o'		Minimum ST-ChlF in the light-regulated state.	Minimum ST-ChlF at beginning of a ST-ChlF transient measured after a brief (~1 s) period of darkness to promote opening of all RCII. Alternatively estimated as: $F_o/(F_v/F_m + F_o/F_m')$.	
σ_{PII}	σ_{PSII}	Absorption cross-section for PSII photochemistry in the dark-regulated state.	Derived from initial ST-ChlF rise during the saturation phase of a ST-ChlF transient.	$m^2 \text{ photon}^{-1}$ $m^2 \text{ PSII}^{-1}$
σ_{PII}'	σ_{PSII}'	σ_{PII} in the light-regulated state.		
ρ	ρ	"Connectivity factor", defining the probability of excitation transfer from the pigment antenna serving a closed RCII to that of an open RCII.	Derived from the sigmoidicity of the ST-ChlF rise during the saturation phase of a ST-ChlF transient.	Dimension-less
ρ'	ρ' J	ρ in the light-regulated state.		
τ_{QA}	τ_1	Time constant for PSII (Q_A) re-oxidation in the dark-regulated state.	Derived from the relaxation phase of ST ChlF transients by fitting a multi-component exponential decay curve to the data.	s
τ_{QA}'	τ_1'	τ_1 in the light-regulated state.		
Secondary ChlF parameter				
F_v		Variable ST-ChlF in the dark-regulated state.	$F_m - F_o$	Relative units
F_v/F_m		Estimate of the maximum quantum yield of photochemistry in PSII (Φ_{PII}). The maximum fraction of light energy absorbed by PSII, which can be used for photochemistry under given environmental conditions. Note that a strict interpretation as the maximum quantum yield of PSII explicitly assumes all measured fluorescence comes from photochemically active PSII (i.e., $F_{o,c}$ and $F_{m,c}$, see section "Blanks and baseline correction" and Figure 2).	$(F_m - F_o) / F_m$	Dimension-less
F_q'	$\Delta F'$	Variable ST-ChlF in the light-regulated state.	$F_m' - F'$	Relative units
F_v'		Maximum variable ST-ChlF in the light-regulated state.	$F_m' - F_o'$	
F_q'/F_m'		Estimate of the quantum yield of photochemistry in PSII (Φ_{PII}'). The fraction of photons absorbed by PSII used for photochemistry under given background irradiance. Note that a strict interpretation as the maximum quantum yield of PSII explicitly assumes all measured fluorescence comes from photochemically active PSII. Note that $F_q'/F_m' = F_q'/F_v' \cdot F_v'/F_m'$.	$(F_m' - F') / F_m'$	Dimension-less

(Continued)

TABLE 1 | Continued

Parameter	Synonym(s)	Meaning	Derivation	Units
F_q'/F_v'		Coefficient of photochemical quenching, qP . Quantifies the drop in F' below F_m' attributable to photochemistry and, under certain assumptions, may therefore be interpreted as an estimate of the fraction of open RCII (strictly assuming no connectivity between PSII units). By definition, the coefficient is 1 in the dark-regulated state and decreases with increasing background light.	$(F_m' - F') / (F_m' - F_o')$	
F_v'/F_m'		The quantum yield of photochemistry at open RCII in the light-regulated state. Quantifies the extent to which PSII photochemistry is limited by competition with thermal dissipation processes.	$(F_m' - F_o') / F_m'$	
JV_{PII}	ETR LET PET P_e	PSII photochemical flux per unit volume. Commonly reported as electron transport rate, the rate of charge separation in PSII per unit volume.	See section 1.2.1	Photon $m^{-3} s^{-1}$ or electron $m^{-3} s^{-1}$
J_{PII}	ETR_{PSII} ETR_{RCII} $PSII_{ETR}$	Photochemical flux per PSII. Commonly reported as electron transport rate, the rate of charge separation in individual photochemically active PSII.		Photon $PSII^{-1} s^{-1}$ or electron $PSII^{-1} s^{-1}$
NPQ_{SV}	NPQ	Regulated (i.e., light-induced) energy dissipation quantified as Stern–Volmer type ChlF quenching.	$(F_m - F_m') / F_m'$	
Φ_{NPQ}	$Y(NPQ)$	Quantum yield of regulated energy dissipation processes (NPQ).	$F'/F_m' - F'/F_m$	
Φ_{NO}	$Y(NO)$	Quantum yield of non-regulated energy pathways through ChlF and non-regulated heat dissipation (both, F and D_{NR} in Figure 1). Note that $\Phi_{NPQ} + \Phi_{NO} + \Phi_{PII} = 1$.	F'/F_m	
NSV	NPQ_{NSV}	A measure of non-photochemical energy dissipation including regulated (i.e., light-induced) processes and increases in energy dissipation in the dark-regulated state (e.g., photo-inhibition). Referred to as normalized Stern–Volmer quenching.	F_o'/F_v'	

The primary ST-ChlF parameters shown below are derived by fitting biophysical models to measurements of ST-ChlF transients. Secondary ST-ChlF parameters are calculated from primary ST-ChlF parameters and provide an interpretation of changes in ChlF in terms of the reduction state of RCII, the fraction of RCII in different states, and the efficiency of different processes within PSII (i.e., NPQ, J_{PII}). Note that numerous additional secondary ST-ChlF parameters exist.

to provide consensus recommendations on the use of ST-ChlF instruments to examine *in-situ* phytoplankton photo-physiology and productivity. We also seek to broaden the application of ST-ChlF approaches among the aquatic research community, and to support the development of a global synthesis of existing and future data. Importantly, it is not our intention to favor any one particular approach, instrument or conceptual model. Rather, we aim to facilitate the sharing of datasets collected by different researchers and instrument types, establishing protocols to promote inter-comparability of observations at a fundamental level. To this end, our goal is to provide consensus recommendations on instrument deployment, data retrieval, and data archiving. While recognizing that any given scientific application may require context-specific methods and protocols, the generation of a globally consistent data archive would enable researchers to apply

existing and emerging approaches of data interpretation, and assess local and global patterns of phytoplankton photo-physiology, photosynthetic rates, and bottom-up controls on primary productivity. Coherent datasets of this kind are invaluable for providing a large-scale and long-term view, and to inform the development of ecosystem models and remote sensing algorithms.

This article represents a collective effort by members of SCOR-WG156 to address key challenges and opportunities for successful global integration of ST-ChlF measurements. We begin with a brief overview of foundational concepts and focus on recommendations toward the acquisition of inter-comparable primary ST-ChlF parameters. We then present a short description of different current approaches to deriving secondary ST-ChlF parameters, emphasizing advantages, and caveats of each approach. In section “Operational and practical

considerations” we discuss the need for standards-of-best-practice for field data acquisition. Stressing the importance of data inter-comparability and collaborative community efforts, we also provide recommendations on data reporting and archiving needed to produce globally consistent datasets (section “Data reporting and archiving”). We conclude by discussing the wide scope of potential ST-ChlF applications in the context of phytoplankton primary productivity, community composition, and the refinement of ecosystem models and remote sensing algorithms (section “Integration and application”). Our goal is to highlight key developments in ST-ChlF methodologies, and stimulate broader interest in the application of these powerful approaches to a range of research questions. This discussion forms the starting point of a community-led and evolving Community-Best-Practice document (SCOR Working Group 156, 2021) for, which will provide detailed guidelines and recommendations.

THEORETICAL FOUNDATIONS AND CONCEPTS

Chlorophyll-a (chl_a), the primary light-harvesting pigment of photosynthetic organisms, re-emits a fraction of absorbed photons at longer wavelengths as fluorescence (ChlF, e.g., Harbinson and Rosenqvist, 2003). This property provides an optical signal that has been widely used to study photosynthetic organisms under field and laboratory conditions (e.g., Kautsky and Hirsch, 1931; Lichtenthaler, 1988; Papageorgiu and Govindjee, 2004; Suggett et al., 2010a). In aquatic systems, ChlF has long been used to infer total chl_a concentrations as a proxy for phytoplankton biomass. When measured *in-vitro* (i.e., after sample extraction in an organic solvent), ChlF is, indeed, proportional to the total chl_a concentration. However, *in-vivo* ChlF measurements, such as those derived from chl_a-fluorometers deployed on depth profiling systems or connected to shipboard continuous flow systems, are subject to variable amounts of so-called “quenching” that cause changes in the ChlF:chl_a ratio. Quenching mechanisms represent the redirection of variable proportions of the absorbed photons to pathways other than ChlF. The variable ratio between *in-vivo* ChlF and chl_a concentration represents an unwanted complication during routine surveys of biomass, necessitating correction procedures (e.g., Thomalla et al., 2018). On the other hand, the variable nature of ChlF provides valuable insights into underlying photo-physiological processes. It is precisely this variability that is examined using variable ChlF methods such as ST-ChlF.

All variable ChlF approaches are based on a fundamental generalized concept, illustrated in **Figure 1**. Light energy absorbed by the photosynthetic pigments serving PSII follows one of three pathways: (1) photochemistry (P); (2) dissipation as heat (D); or (3) re-emission as fluorescence (F; e.g., Butler, 1978). The distribution of excitation energy among the three pathways is variable. Photochemistry and a component of heat-dissipation (D_{NPQ}) are actively regulated, and changes in these two processes modulate the remaining fraction of absorbed

energy re-emitted as ChlF. When more absorbed light energy is directed to either P or D_{NPQ} , less energy is re-emitted as ChlF. For this reason, P and the D_{NPQ} are typically referred to as photochemical and non-photochemical quenching of ChlF, respectively. It follows that changes in ChlF can be used to assess variations in P, as long as changes in D_{NPQ} can be accounted for. This general concept has been applied and refined for over a century of photosynthetic research (e.g., Govindjee, 1995), leading to many important insights. At the same time, there has been considerable conceptual and methodological confusion associated with varying approaches and nomenclatures employed by different investigators across various photosynthetic taxa and measurement techniques.

Numerous books, reviews, and manuals have explained the details of ST-ChlF techniques and the derivation of ChlF parameters (e.g., Roháček and Barták, 1999; Roháček et al., 2008; Huot and Babin, 2010; Kolber, 2021; Oxborough, 2021). Here, we summarize the essentials in **Figure 1**, and then focus on aspects most relevant to aquatic field deployments of ST-ChlF instruments and the interpretation of the resulting data. In our discussion, we make a distinction between primary ST-ChlF parameters, which are those properties that are directly derived from induced changes in ChlF (ST-ChlF transients), and secondary ST-ChlF parameters, which are subsequently computed from primary ST-ChlF parameters. Importantly, it is not our goal (nor in the interest of scientific progress) to favor any one particular approach for the acquisition or interpretation of ST-ChlF data. Rather, we aim to present foundational concepts and procedures applicable to any hardware and analysis routine, enabling datasets collected by different research groups, instrument types and field campaigns to remain inter-comparable at a fundamental level. Such inter-comparability is needed for the construction of sustainable global datasets and their robust interpretation using existing and emerging approaches.

Primary ChlF Parameters

Variable ChlF approaches use intense light pulses on microsecond timescales to transiently saturate the photochemical pathway, thereby inducing measurable changes in ChlF (**Figure 1**). Such induced changes in ChlF, referred to as ChlF transients, involve the rapid increase in ChlF up to a maximum value (i.e., the saturation phase), followed by a return to a basal level (i.e., the relaxation phase; **Figure 1C**). The ChlF signal measured by variable ChlF approaches is assumed to derive exclusively from PSII (but see section “Blanks and baseline correction”). Consequently, the technique is most suited to study reactions and processes taking place at or close to PSII reaction centers II (RCII). However, given tight coupling of reductant and energy fluxes across the entire photosynthetic system and beyond, information well beyond PSII function can be inferred from variable ChlF measurements.

Biophysical models have been developed to interpret ChlF transients and derive primary ChlF parameters (e.g., Dau, 1994; Trissl and Lavergne, 1995). During a saturating single-turnover (ST) flash, light energy sufficient to reduce all primary electron acceptors, Q_A (i.e., to “close” all RCII), is delivered over a short period (<200 μ s), before significant electron transport

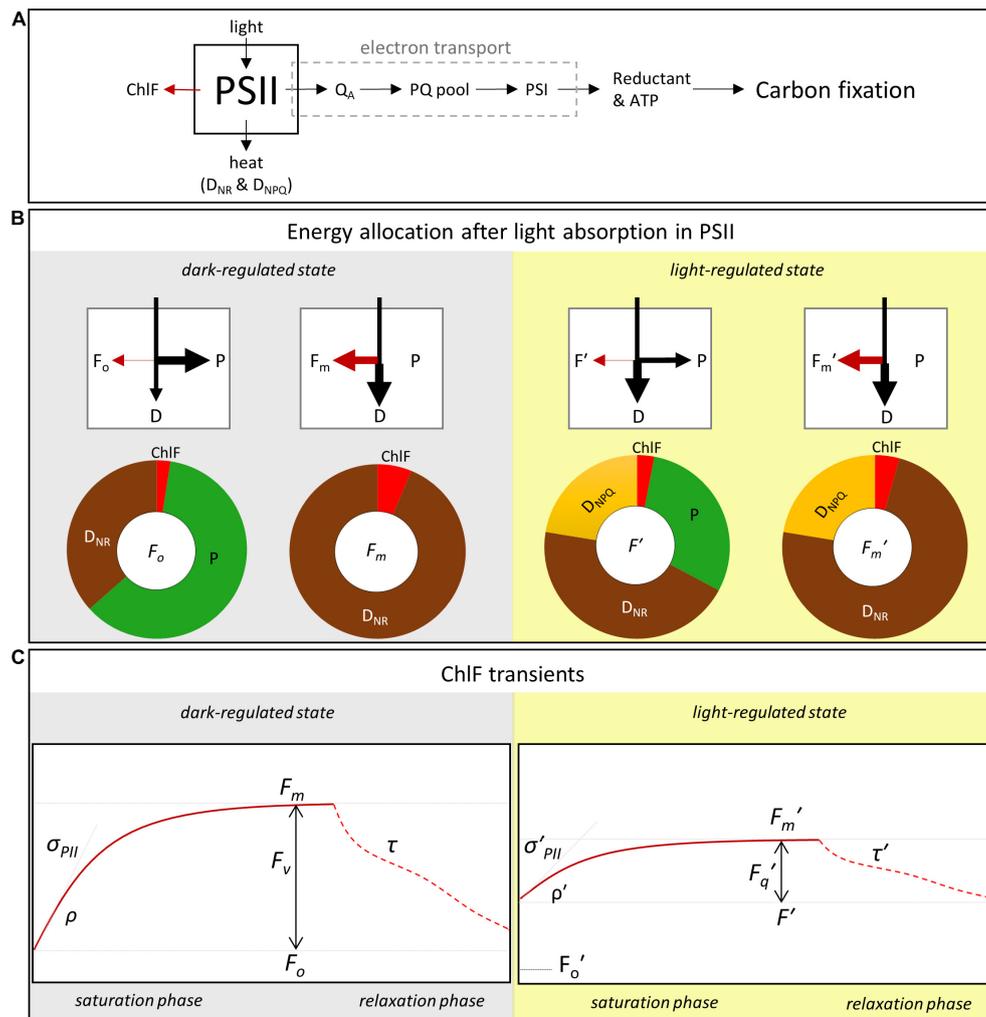


FIGURE 1 | The three energy pathway concept and ChIF transients from typical ST protocols. ChIF induced and detected by ST-ChIF instruments is typically assumed to originate primarily from PSII (**A**). (**B**) provides a conceptual overview of energy allocation to the three competing pathways of photochemistry (P), re-emission as heat (D), or ChIF (F). The heat-dissipation pathway is composed of a non-regulated (D_{NR}) and an actively regulated (D_{NPG}) part. Changes in both P and D_{NPG} modulate ChIF. During ST-ChIF transients (**C**), energy allocation to P is selectively modulated, leading to changes in ChIF (see main text). Left hand panels in (**B,C**) (gray shading) represent the dark-regulated state, while right hand panels (yellow shading) represent the light-regulated state. In the dark-regulated state, it is assumed that D_{NPG} is zero, and that all electron transport-chains are fully oxidized at the beginning of the saturation phase (i.e., all RCII are open), leading to the maximum potential for absorbed light energy to be used for photochemistry and thus maximal photochemical quenching of ChIF (ChIF = F_0). During the “saturation phase,” all primary electron acceptors Q_A are progressively reduced (i.e., all RCII are closed, $P = 0$), thus decreasing photochemical quenching and increasing the energy re-emitted as ChIF (ChIF = F_m). As shown in (**C**), by fitting the ST-ChIF saturation phase in the dark-regulated state we can derive: minimum (F_0) and maximum (F_m) ChIF, the absorption cross-section for photochemistry (σ_{PII}), and the connectivity factor (ρ). The decrease of ST-ChIF during the relaxation phase can be interpreted in terms of electron transport rates downstream of charge separation in PSII (τ). In the light-regulated state, the ST-ChIF level at the beginning of saturation phase increases to a steady-state fluorescence, F' . This increase in ChIF reflects the fact that some PSII are engaged in electron transport (Q_A reduced, RCII closed), such that the fraction of absorbed energy potentially allocated to photochemistry is no longer maximal. The maximum ST-ChIF decreases from F_m in the dark-regulated state to F_m' in the light-regulated state, as a result of ChIF quenching by regulated heat-dissipation pathways (D_{NPG}). Further, σ'_{PII} , ρ' , and τ' can be derived from the light-regulated ST-ChIF transient. The parameter F_0' represents the minimum ST-ChIF measured immediately after the transition from light to dark. It is the ST-ChIF level under maximal photochemical quenching (all Q_A oxidized, all RCII open), while D_{NPG} is still active at the level induced during the light-regulated state. The conceptual model shown is a simplified and idealized representation and that its applicability to different phytoplankton with a range of photosynthetic architectures and mechanisms will vary.

downstream of PSII can re-open RCII (**Figure 1A**). In contrast, multiple turnover protocols are designed to more gradually reduce the entire electron transport chain over ~ 100 – $1,000$ ms, usually leading to higher levels of maximum ChIF (e.g., Kolber et al., 1998; Kromkamp and Forster, 2003; Brown et al., 2019).

We focus our discussion here on ST instruments and analysis protocols only, as these are more commonly applied for research on phytoplankton. The rapid reduction of Q_A in ST instruments can be achieved by a series of light “flashlets” in FRRE, or by a single light pulse in fluorescence induction

and relaxation (FiRe) and single turnover active fluorometry (STAF) instruments.

Primary ChlF Parameters From the Saturation Phase

In the “dark-regulated” state, measurements are made without any background illumination and after relaxation of any NPQ processes (section “The dark-regulated states and NPQ-relaxation”). Under this condition (left panel in **Figures 1B,C**), it is assumed that D_{NPQ} is minimal. All RCII are open at the beginning of the ST-ChlF transient, allowing for a maximum fraction of absorbed photon energy to be partitioned to photochemistry (P is maximal and ChlF minimal). In the dark-regulated state, the amplitude of a ChlF transient (F_v) can be interpreted in terms of the maximum photochemical efficiency for a given population of PSII under a given environmental condition. As described in **Figure 1**, and in much detail elsewhere (e.g., Kolber et al., 1998; Huot and Babin, 2010), the primary ST-ChlF parameters derived from the saturation phase of dark-regulated ST-ChlF transients are the minimum (F_o) and maximum (F_m) ChlF, the absorption cross-section for PSII photochemistry (σ_{PII}) and the “connectivity” among PSII units (ρ). F_o and F_m are typically measured in arbitrary units, although they can be calibrated to a reference signal, providing useful additional quantitative information (Oxborough, 2021). Values of σ_{PII} are derived from the initial ST-ChlF transient rise and are reported in units of area photon⁻¹ or area PSII⁻¹. Note that σ_{PII} has frequently been reported in Å², but the use of such non-SI units is not recommended. Connectivity among PSII units (ρ) is a unitless value, derived from the sigmoidicity of the ST-ChlF rise from F_o to F_m (**Table 1**, Lavorel and Joliot, 1972; Lavergne and Trissl, 1995; Kolber et al., 1998).

In the light-regulated state, where the sample is exposed to background illumination during measurements (right panel in **Figures 1B,C**), a fraction of the RCII pool is already closed at the beginning of the ST-ChlF transient. As a result, the minimum ChlF level derived from the ChlF transient (F') is generally increased relative to the minimum ChlF for fully open RCII (F_o). Depending on the intensity and duration of the background irradiance, the fraction of absorbed photon energy dissipated as heat may be increased relative to the dark-regulated state, resulting in a drop (i.e., quenching) of ChlF ($F_m' < F_m$, $F_o' < F_o$, see section “Non-photochemical quenching”). As described in **Figure 1**, the primary ST-ChlF parameters derived from the saturation phase in the light-regulated state are F' , F_m' , σ_{PII}' , and ρ' . We note here that, in contrast to higher plants and green algae, the light-dependent decrease in F_m' relative to F_m in many phytoplankton samples is frequently preceded by a transient increase in F_m' upon moderate illumination (e.g., Gorbunov et al., 2011). The underlying causes of such transient increases in F_m' are still debated, and readers are referred to SCOR Working Group 156 (2021) for a more detailed discussion of this phenomenon.

Primary ST-ChlF Parameters Derived From the Relaxation Phase

Following the transient closure of RCII during the saturation phase, ChlF decreases back to its minimal level (**Figure 1C**).

The time-course of this ChlF decrease largely reflects Q_A re-oxidation kinetics through downstream photosynthetic electron transport (**Figure 1A**). With the FRRF method, this ST-ChlF “relaxation phase” is typically resolved through a series of low frequency “probing flashlets” often applied at gradually increasing intervals (Kolber et al., 1998). An alternative approach is to apply a small number of more widely spaced ST saturation phases. A “multi-flash” (comprising five ST saturation phases) protocol, with an increasing interval between adjacent ST phases was implemented in a single-cell FRRF (Gorbunov et al., 1999). A “dual-pulse” protocol with a variable gap between two ST saturating phases has been incorporated within “single-turnover active fluorescence” (STAF) instruments (Oxborough, 2021). In all cases, the time-dependent decrease in ChlF after saturation is fit to a multi-component exponential decay curve to resolve the time constant(s; τ) for Q_A re-oxidation (Kolber et al., 1998; Gorbunov and Falkowski, 2020; Oxborough, 2021, **Figure 1**). The use of a three-component kinetic analysis is critical for the most accurate description of Q_A re-oxidation kinetics (Gorbunov and Falkowski, 2020). However, the fitting of three (or more) components requires a high signal-to-noise ratio, which is not always achievable in oligotrophic regions or with older and less sensitive instrument types.

Uncertainty and Error

Multiple factors can limit the accuracy with which primary ST-ChlF parameters can be retrieved. However, despite previous attempts to draw attention to some of these issues in specific instruments (e.g., Laney, 2003; Laney and Letelier, 2008), uncertainty and error associated with primary ST-ChlF parameters are not routinely described in the literature, nor reported in published datasets. This limits our ability to gage data quality, and the strength of any derived observations and interpretations. To address this limitation, we outline several considerations specific to the derivation and reporting of primary ChlF parameters. While the exact approach may be instrument specific, an explicit consideration of data quality and confidence is nonetheless critical to support globally coherent and inter-comparable observations.

Single-turnover variable chlorophyll fluorescence instruments are now capable of acquiring data even in very low biomass regions, but the low signal typical for oligotrophic waters often requires considerable data averaging from repeated rounds of ST-ChlF transients to achieve fits of reasonable quality (e.g., Ryan-Keogh and Robinson, 2021). A minimum level of fit quality for the derivation of primary ST-ChlF parameters should preferably be assessed during real-time data acquisition. Using this information, appropriate instrument settings and signal averaging can ensure minimum data quality standards. In addition to measurements taken in oligotrophic waters, the assessment of quality of ST-ChlF transient fits is particularly important for measurements taken at high background light levels, where the amplitude of the ST-ChlF transient decreases and retrieved parameters, σ_{PII}' and τ' in particular, become imprecise. As discussed in section “Data reporting and archiving,” information regarding the statistical goodness of the fit of

ST-ChlF transients should be archived alongside primary ST-ChlF parameter data.

In addition to the general statistical issues described above, other factors need to be considered in the derivation of individual primary ST-ChlF parameters. For example, both $F_m(')$ and $\sigma_{PII}(')$ rely, in principle, upon closure of all RCII on time-scales shorter than RCII reopening ($<200 \mu\text{s}$), which results in a clear plateau of the ST-ChlF transient toward the end of the saturation phase (**Figure 1C**). On a practical level, this requirement is generally easily achieved when ChlF is excited in the 410–500 nm spectral range, which is strongly absorbed by most eukaryotic phytoplankton species, and for which LEDs with high photon flux are readily available. However, when excitation power is delivered at wavelengths poorly absorbed by the present phytoplankton taxa, or at wavelengths served by less effective LEDs, the photon flux achievable during the short ST saturation phase may be insufficient for near complete Q_A reduction. Such “under-saturation” of the ChlF transient is taken into account during data fitting (Kolber et al., 1998), but very low (50%) saturation combined with a low signal-to-noise ratio, can make it challenging to derive accurate primary ST-ChlF parameters. Given these challenges, users should confirm sufficient Q_A reduction (ST-ChlF transient saturation) in their measurements (**Table 2**).

Special considerations are required for the light-regulated ChlF parameter F_o' . This parameter represents the minimum ChlF expected when the photochemical potential is maximal (i.e., all photochemically active RCII in the open state), but with light-dependent regulation of the heat dissipation pathway (D_{NPQ}) still active (e.g., Genty et al., 1989). In principle, F_o' can be measured by acquiring a ST-ChlF transient immediately after turning off the background light, under the assumption that re-oxidation of the electron transport chain will occur on time-scales much shorter than those required for the relaxation of NPQ (Ni et al., 2017). In practice, some NPQ processes may begin relaxing on very short timescales (Roháček et al., 2014; Ni et al., 2017). Measuring an accurate F_o' thus becomes problematic, particularly in low biomass regions where averaging of many sequential ChlF transients may be necessary to obtain good quality data. In a second approach, introduced by Oxborough and Baker (1997), F_o' is estimated as $F_o' = F_o/(F_v/F_m + F_o/F_m')$. The derivation is based on the widely accepted concept of competing energy pathways in PSII, but is susceptible to distortion by baseline fluorescence (section “Blanks and baseline correction”) and relies on measurements in the dark regulated state (section “The dark-regulated states and NPQ-relaxation”). As discussed in section “PSII photochemical flux, J_{PII} ,” incorrect values for F_o' can introduce systematic error in the derivation of some secondary ST-ChlF parameters.

In addition to the sources of error and uncertainty described above, other sources of variability in the derivation and interpretation of primary ST-ChlF parameters include uncertainty in conceptual models assessing connectivity between PSII units (ρ ; e.g., Stirbet, 2013; Oxborough, 2021); the appropriate number of exponential decay lifetimes (τ) used to model the relaxation of ST-ChlF (e.g., Gorbunov and Falkowski, 2020) and the effects of carotenoid quenching on $F_m(')$ (e.g.,

Kolber et al., 1998; Schreiber et al., 2019). Specific details of these effects are beyond the scope of this article, but readers are referred to SCOR Working Group 156 (2021) for further details.

Finally, it is important to highlight taxonomic diversity as a factor that complicates the derivation and interpretation of primary ST-ChlF parameters from models developed for homogeneous populations of PSII. For example, values of σ_{PII} measured on mixed phytoplankton assemblages are unlikely to scale linearly with the proportional contributions of σ_{PII} from the individual species present (Suggett et al., 2004; Laney, 2010). In addition, taxonomic variability in baseline fluorescence levels (see section “Blanks and baseline correction”) has the potential to significantly increase non-variable ChlF, and disproportionately increase apparent minimum (F_o or F') relative to maximum (F_m or F_m') ChlF. For example, fluorescence from phycobilins can be falsely attributed to PSII, and such contributions can vary significantly between dark and light-regulated states. It has furthermore been shown that taxonomic trends in PSII:PSI ratios can lead to differential contributions of PSI-derived ChlF to signals usually interpreted in terms of PSII (Campbell et al., 1998). Such taxonomic influences, and other sources of baseline fluorescence not emanating from the PSII pigment pool, will complicate the physiological interpretation of widely used primary and secondary ST-ChlF parameters (section “Blanks and baseline correction”).

Secondary ST-ChlF Parameters

The primary ST-ChlF parameters described above are those derived directly from applying photo-physiological models to ST ChlF transients (**Figure 1**). These primary parameters can, in turn, be used to derive secondary parameters of physiological interest (**Table 1**). Here, we focus on J_{PII} and NPQ, reviewing the most common algorithms used to estimate these parameters and describing their respective advantages and disadvantages with respect to field data. In this discussion, it is important to understand that secondary ST-ChlF parameters such as J_{PII} are not directly measured, but rather derived from primary ST-ChlF parameters using conceptual models based on current understanding of the photosynthetic process. Under field conditions, where taxonomic and environmental variability are the norm, the applicability of different models to derive J_{PII} (and other secondary ST-ChlF parameters) may vary, and results could diverge. While this can create uncertainty, the goal here is not necessarily to identify one “correct” approach, but rather to understand how and why different models may differentially capture underlying physiological processes under various conditions (**Table 3**). The compilation of globally consistent and inter-comparable primary ST-ChlF parameter data will greatly facilitate the comparisons of different models to derive J_{PII} (and other parameters). Important insights will likely be found in situations where results from different modeling approaches diverge.

PSII Photochemical Flux, J_{PII}

Primary ST-ChlF parameters can be used to quantify the photochemical pathway (**Figure 1**) in terms of the PSII photochemical flux. We use the term PSII photochemical

TABLE 2 | Consensus recommendations for the field deployment of ST-ChIF instruments, aimed at supporting the development a globally coherence ST-ChIF dataset.

Operational considerations	
Blanks (section “Blanks and baseline correction”, Figure 2)	<p>For benchtop measurement of discrete samples:</p> <p>Run B_{inst} at the beginning of each measurement series, monitor at regular intervals to detect, e.g., biofouling in cuvette. If B_{inst} is low (<5% of F_m in oligotrophic waters), subtract B_{inst} from all measurements automatically during sample acquisition, such that the “filtrate blank” represents F_{diss}.</p> <p>Run filtrate blank with every sample.</p> <p>For benchtop applications, in particular for depth profiles and in low biomass regions, sample-specific F_{diss} should always be subtracted from ST-ChIF measurements. Values of F_{diss} contain valuable information and should be recorded and archived as % of F_m.</p> <p>For underway flow-through deployments:</p> <p>Run B_{inst} at the beginning of each field campaign, monitor at regular intervals to detect, e.g., biofouling in cuvette. If B_{inst} is low (<5% of F_m in oligotrophic waters), subtract B_{inst} from all measurements automatically during sample acquisition, such that the “filtrate blank” represents F_{diss}.</p> <p>Run filtrate blank at regular intervals (e.g., daily) and interpolate this value to the frequency of underway measurements. For continuous data acquisition from surface waters, F_{diss} can be considered negligible if it remains consistently <5% of F_m.</p> <p>For in-situ deployments:</p> <p>Run B_{inst} at the beginning and end of each field campaign to detect, e.g., biofouling in cuvette. Subtract B_{inst} from all measurements.</p> <p>Where F_{diss} cannot be measured regularly during autonomous data acquisition (floats, moorings, etc.), systematic error can be introduced in regions of low biomass or below the chlorophyll max. We therefore recommend caution when interpreting data from such regions and that values of F_{diss} from on-board measurements are routinely measured and archived (as % of F_m), in order to compile global, instrument-independent data compilations to systematically characterize F_{diss} with respect to, e.g., region, bloom-stage, etc.</p> <p>Use low light (<5–10 $\mu\text{mol photon m}^{-2} \text{s}^{-1}$) at sampling temperature. If the goal is to relax NPQ but not photo-inhibition, a period of 10–20 min is generally sufficient, but note that no “ideal” NPQ-relaxation time exists, as it is highly dependent on the sampling situation and scientific question addressed.</p> <p>Always report the time and light intensity used during NPQ-relaxation.</p> <p>For benchtop application:</p> <p>It is recommended to run trials to determine the optimal NPQ-relaxation time.</p> <p>For underway deployments:</p> <p>A “NPQ-relaxation” step should be added to the measurement protocol (e.g., before each automated light-response curve acquisitions).</p> <p>For in-situ deployments:</p> <p>NPQ-relaxation can be achieved by enclosure of a sample into a sample chamber before measurements of, e.g., light-response curves.</p>
NPQ-relaxation (section “The dark-regulated state and NPQ-relaxation”)	<p>Use low light (<5–10 $\mu\text{mol photon m}^{-2} \text{s}^{-1}$) at sampling temperature. If the goal is to relax NPQ but not photo-inhibition, a period of 10–20 min is generally sufficient, but note that no “ideal” NPQ-relaxation time exists, as it is highly dependent on the sampling situation and scientific question addressed.</p> <p>Always report the time and light intensity used during NPQ-relaxation.</p> <p>For benchtop application:</p> <p>It is recommended to run trials to determine the optimal NPQ-relaxation time.</p> <p>For underway deployments:</p> <p>A “NPQ-relaxation” step should be added to the measurement protocol (e.g., before each automated light-response curve acquisitions).</p> <p>For in-situ deployments:</p> <p>NPQ-relaxation can be achieved by enclosure of a sample into a sample chamber before measurements of, e.g., light-response curves.</p>
Light-response curves (section “Light-response curves”)	<p>Always report details of the protocol used (e.g., lengths of NPQ-relaxation, length, number and sequence of light steps, and fit used to derive parameters, etc.).</p> <p>The length of light steps should be adjusted such that primary ST-ChIF parameters reach steady-state during each light level.</p> <p>Ensure temperature within the sampling chamber remains close to <i>in-situ</i> temperature throughout the light-response curve.</p>
Spectral correction (section “Spectral correction”, Figure 3)	<p>Always report spectral distribution of excitation and background LEDs alongside data.</p> <p>If information on spectral light absorption is available, all data should be spectrally corrected.</p>
Assessment and reporting of error and uncertainty in primary ChIF parameters	
Quality of fit and retrieved primary ST-ChIF parameters (section “Uncertainty and error”)	<p>The quality of fit of the ST-ChIF transient should be statistically assessed, ideally in real-time such that data quality can be improved through increasing the number of data acquisitions used for each curve fit, if necessary.</p> <p>The quality of the ST-ChIF transient fit and derived primary ST-ChIF parameters can be assessed, for example, as a Signal-to-Noise Ratio (Kolber, 2021) or RMSE (Ryan-Keogh and Robinson, 2021).</p>
ST-ChIF transient saturation (section “Uncertainty and error”)	<p>The user should verify that the photon flux delivered during the saturation phase of a ST-ChIF transient is sufficient to achieve ~80% reduction of Q_A, in particular if less well absorbed excitation wavelength are used.</p> <p>During data acquisition, this is possible by verifying that the ST-ChIF transient reaches a clear plateau at the F_m level, or that $1/2 F_m$ is reached during the first 50 μs of the saturation phase.</p>
F_o' (section “Uncertainty and error”)	<p>If the F_o' parameter is used for the calculation of secondary ST-ChIF parameters, details of how it was measured or derived must be reported.</p>
Instrument calibration	
Excitation and background light (section “Calibration of light sources”)	<p>Regularly confirm LED output following manufacturers’ instructions.</p>

See referenced sections in main text for nomenclature.

flux (J_{PII}) rather than the commonly used term electron transport rate to emphasize that the parameter quantifies the flux of solar photons toward metabolically useful biochemical

energy in the form of redox potential in the photosynthetic electron transport chain at the level of PSII. Units of J_{PII} are (absorbed) photon $\text{PSII}^{-1} \text{s}^{-1}$. However, given that

TABLE 3 | Sources of error in different J_{PII} estimates.

Source of error	Explanation	Affects
Baseline and blank fluorescence	Non-variable fluorescence from sources other than PSII, which is misinterpreted as originating from PSII (Figure 2 , section “Blanks and baseline correction”).	Eqs 1, 2, and 5
Need for ST-ChlF transient in dark and light-regulated states from the same sample	A fully dark-regulated state is difficult to achieve for phytoplankton in field samples, and is necessarily offset in time from a light-regulated state (Figure 1 , section “The dark-regulated states and NPQ-relaxation”).	Eqs 1 and 5 Eq. 2 (if F_o' is not measured directly)
F_o'	Difficult to measure directly in low biomass field samples (section “Uncertainty and error”). Calculation relies on measurement in dark-regulated and light-regulated states (Figure 1 , section “Uncertainty and error”).	Eq. 2
σ_{PII} (°)	Relies upon accurate calibration of excitation photon flux density and spectra (section “Uncertainty and error”). Uncertainty increases under higher background irradiance, and in mixed phytoplankton assemblages (section “Uncertainty and error”).	Eqs 1–4
τ_{QA} (°)	Relies upon accurate calibration of excitation photon flux density and spectra (section “Calibration of light sources”). Uncertainty in the appropriate number of τ constants used to fit relaxation kinetics with a multi-exponential decay function and in the mechanistic interpretation of derived values (section “Uncertainty and error”). Measurement uncertainty increases under higher background irradiance (Figure 1 , section “Uncertainty and error”).	Eqs 3 and 4

Uncertainties inherent to different J_{PII} derivations result from errors associated with either ST-ChlF transient amplitudes [F_o' (°), F_m' (°), F'] or kinetics [σ_{PII} (°), τ (°)]. Amplitude-based approaches will be affected by baseline fluorescence, which can be difficult to correct under the range of environmental and taxonomic variability encountered in the field. Approaches based on the kinetics of the ST-ChlF transients (σ_{PII} and τ) rely on precise instrument calibration (section “Calibration of light sources”), are prone to error under low biomass and high background light conditions, and can be ambiguous in mixed phytoplankton assemblages (section “Uncertainty and error”).

each photon absorbed and delivered to RCII leads to one charge separation, the parameter is widely reported in units of electrons $\text{PSII}^{-1} \text{ s}^{-1}$. A proportion of the biochemical energy available through J_{PII} is ultimately captured in the form of reduced organic carbon, and it is this connection to carbon-based primary productivity that often motivates the measurement of J_{PII} in aquatic environments (section “Deriving carbon-based primary productivity,” Hughes et al., 2018).

Equations 1 and 2 show different versions of the so-called “sigma-algorithm,” commonly used to derive J_{PII} in aquatic systems. Both equations follow the simple rationale that J_{PII} can be calculated from estimates of incident photon irradiance, the fraction of photons absorbed by PSII and the distribution of absorbed photon energy among the three energy dissipation pathways (**Figure 1**). Equations 1 and 2 are algebraically equivalent, but differ operationally in their approach to estimating light-dependent changes in absorbed energy allocation among the three pathways (e.g., Gorbunov et al., 2001; Suggett et al., 2010b). In Eq. 1, light absorption of PSII-associated pigments specific to all three pathways is estimated as the product of scalar irradiance (E), σ_{PII} and $(F_v/F_m)^{-1}$. This estimate of light absorption is then multiplied by the quantum efficiency of photochemistry (i.e., changes in the distribution of energy between the three pathways) under a given background light intensity, F_q'/F_m' .

$$J_{PII} = E \cdot \sigma_{PII} \cdot (F_v/F_m)^{-1} \cdot (F_q'/F_m') \quad (1)$$

In Eq. 2, light absorption directed to the photochemical pathway at a given irradiance only is quantified as the product of E and σ_{PII}' . The parameter F_q'/F_v' , calculated as $(F_m' - F')/(F_m' - F_o')$, is

used as an estimate of the fraction of RCII in the open state (**Table 1**, Kolber et al., 1998; Kramer et al., 2004).

$$J_{PII} = E \cdot \sigma_{PII}' \cdot (F_q'/F_v') \quad (2)$$

Equations 1 and 2 are equivalent when the ratio of light to dark regulated absorption cross-section of PSII photochemistry, $\sigma_{PII}'/\sigma_{PII}$, is equal to the ratio of light to dark regulated quantum yield of photochemistry, $(F_v'/F_m')/(F_v/F_m)$; (Gorbunov et al., 2001; Suggett et al., 2010b).

The approach represented by Eqs 1 and 2 has several limitations (**Table 3**). First, it relies on the measurement of ST-ChlF amplitudes (i.e., changing levels of fluorescence, F), which can be affected by baseline fluorescence (section “Blanks and baseline correction”), although this has a larger influence on Eq. 1 than Eq. 2. Further, Eq. 1 requires measurements from separate ST-ChlF transients offset in time (dark- and light-regulated state), and the need to achieve a fully dark-regulated state, which can be challenging under field conditions (section “The dark-regulated states and NPQ-relaxation”). For Eq. 2, measurements in the fully dark-regulated state are required if the F_o' value needed to calculate F_v' is derived following the approach by Oxborough and Baker (1997; section “Uncertainty and error”). To address these challenges, Eq. 3 has been proposed as an alternative to Eq. 2:

$$J_{PII} = E \cdot \sigma_{PSII}' \cdot [1/(1 + (\sigma_{PSII}' \cdot E \cdot \tau'))] \quad (3)$$

Here, the calculation of the fraction of RCII in the open state (in square brackets) uses a mechanistic model depending on σ_{PII}' (energy distributed to the photochemical pathway, i.e., closing RCII) and $1/\tau'$ (rate of re-opening of RCII). All parameters used in Eq. 3 can be derived from a single saturation/relaxation profile measured in the light-regulated state.

Recently, Gorbunov and Falkowski (2020) introduced an approach for the estimation of J_{PII} which relies almost exclusively on the kinetics of the relaxation phase of the ST-ChlF transient.

$$J_{PII} = 1/\tau \cdot [(E \cdot \frac{F'_q}{F'_m}) / (E_{max} \cdot \frac{F'_q}{F'_m(E_{max})})] \quad (4)$$

As explained in more detail in Gorbunov and Falkowski (2020), J_{PII} in this approach is derived from the time constant of Q_A re-oxidation, τ_{QA} , derived from the first of a three-component decay function fit to the relaxation phase of a ST-ChlF transient (section “Primary ST-ChlF parameters derived from the relaxation phase”), at saturating background light. τ_{QA} was confirmed in independent experiments to closely approximate the maximum rate (P_{max}) of short-term ^{14}C -uptake (here referred to as the photosynthetic turnover rate, τ). The second term (in square brackets) of Eq. 4 characterizes the shape of a light response curve, and thus scales the maximum J_{PII} to light availability. In Eq. 4, E_{max} is taken as a value three times higher than the light saturation parameter, E_k .

All of the approaches described above estimate PSII photochemical flux per photochemically active PSII (J_{PII} , photon $PSII^{-1} s^{-1}$ or electron $PSII^{-1} s^{-1}$). In order to derive a volume-specific PSII photochemical flux (JV_{PII} , photons $m^{-3} s^{-1}$ or electron $m^{-3} s^{-1}$), information is required on the concentration of functional PSII. This can be done either directly on a volumetric basis ($[PSII] m^{-3}$), or indirectly through a normalization to chl *a* concentrations ($[PSII] chl a^{-1}$; n_{PSII}) and determination of $[chl a]$. Different approaches to quantify $[PSII]$ or n_{PSII} exist (see, e.g., Silsbe et al., 2015), but none of these are practical for high-resolution autonomous field measurements. At the same time, the use of an assumed constant value for n_{PSII} can result in considerable error. Consequently, Oxborough et al. (2012) have developed an approach, subsequently verified by Silsbe et al. (2015) and extended by Boatman et al. (2019), to estimate $[PSII]$ from ST-ChlF measurements. In simple terms, the approach recognizes that values of F_o should scale with the number of active PSII within a sample, while σ_{PII} provides a measure of the size of a single PSII. From this it follows that the number of PSII within a sample can be estimated by using an appropriate instrument-specific scaling factor, K_a (m^{-1}), such that $[PSII] = F_o/\sigma_{PII} \cdot K_a$. Incorporation of $[PSII]$ into the J_{PII} -algorithm shown in Eq. 1 provides the means to calculate JV_{PII} , where the term in square brackets estimates light absorption by all PSII within a volume of water (a_{LHII} , m^{-1}) and F'_q/F'_m represents the efficiency with which this energy is used for photochemistry (Oxborough, 2021).

$$JV_{PII} = E \cdot F'_q/F'_m \cdot [K_a \cdot (F_m \cdot F_o)/F_v] \quad (5)$$

This approach, referred to as the “absorption algorithm” enables the calculation of volume-specific, rather than PSII-specific photochemical fluxes, and thus represents an important step forward in our ability to quantify phytoplankton primary productivity from ST-ChlF measurements. Importantly, it allows the quantification of light absorbed by all PSII within a volume of water (a_{LHII} , m^{-1}), and does not rely on estimates of absorption cross-sections for PSII photochemistry (σ_{PII}), which are

ambiguous in heterogeneous populations of PSII and therefore difficult to interpret in mixed phytoplankton assemblages in the field. However, as with any amplitude-based technique, estimates of $[PSII]$ and the associated absorption algorithm are prone to error introduced by baseline fluorescence (section “Blanks and baseline correction”) and requires measurements in a fully dark-regulated state (section “The dark-regulated states and NPQ-relaxation”). Furthermore, the effect of reabsorption of ChlF in large or highly pigmented cells (“pigment packaging”) must also be considered (Boatman et al., 2019), and the approach relies on precise calibration of the absolute ChlF yields through procedures not routinely applied for all instrument types. The application of the absorption algorithm to a wide range of environmental conditions and phytoplankton taxa will allow to determine how the necessary corrections can be confidently and routinely applied.

Non-Photochemical Quenching

All plants and most algae possess a range of mechanisms that can be rapidly activated to dissipate potentially harmful excess excitation energy in the pigment antenna under conditions of transient increases in incident light. Collectively these are known as non-photochemical quenching (NPQ) mechanisms, due to their ability to physiologically quench excitation energy. As described above, activation of NPQ will lead to a decrease in ChlF, such that different metrics of NPQ can be derived from ST-ChlF measurements. In this respect, “NPQ” describes the phenomenon of non-photochemical quenching of ChlF rather than excitation energy. The effect of NPQ on ChlF complicates the interpretation of ChlF as a biomass proxy and the interpretation of ST-ChlF measurements in terms of photochemistry. On the other hand, NPQ holds untapped potential as an optical signal reflecting the physiological state of phytoplankton (Campbell et al., 1998; Schuback et al., 2020; section “Exploring environmental controls on primary productivity”), which can, in turn, constrain key parameters for estimating productivity (Schuback et al., 2015).

Current definitions and terminologies for NPQ are particularly confusing, with the existing literature and assumptions derived primarily from higher plant research. As we are only beginning to understand the diversity of NPQ mechanisms and capabilities across different phytoplankton species (e.g., Goss and Lepetit, 2015; Magdaong and Blankenship, 2018; Lacour et al., 2020), great caution should be taken when interpreting patterns of different NPQ metrics in phytoplankton based on models or mechanisms extrapolated from higher plants.

Multiple metrics of NPQ can be derived from ST-ChlF measurements. Most commonly, NPQ is quantified according to Stern–Volmer quenching principles (NPQ_{SV} , Eq. 6), as the difference in maximum ChlF between the dark-regulated (F_m) and light-regulated state (F'_m ; see also **Figure 1**).

$$NPQ_{SV} = (F_m - F'_m)/F'_m \quad (6)$$

NPQ_{SV} , which was first introduced by Bilger and Björkman (1991), provides a mechanistic metric tracking the accumulation of a ChlF quencher within a sample exposed to increasing photon flux. It is an unbound parameter, with values above ~ 2

showing poor correlation with other photo-physiological metrics of excitation dissipation (e.g., Xu et al., 2018).

The parameter F_v'/F_m' arguably most closely tracks the impact of NPQ on PSII photochemical efficiency under different conditions.

$$F_v'/F_m' = (F_m' - F_o')/F_m' \quad (7)$$

A different approach can be used to derive the fractional yields of regulated NPQ (Φ_{NPQ} , Eq. 8) and non-regulated energy dissipation processes (Φ_{NO} , including both un-regulated energy re-emission as heat and ChlF, **Table 1**) akin to the yield of photochemistry ($F_q'/F_m' = \Phi_{PSII}$, **Table 1**), such that $\Phi_{NPQ} + \Phi_{NO} + \Phi_{PSII} = 1$ (**Table 1**, Hendrickson et al., 2004; Kramer et al., 2004; Klughammer and Schreiber, 2008), where:

$$\Phi_{NPQ} = F'/F_m' - F'/F_m \quad (8)$$

More recently, the so called normalized Stern–Volmer quenching metric (NSV, Eq. 9), has been applied to aquatic ST-ChlF measurements (McKew et al., 2013; Oxborough, 2021).

$$NSV = F_o'/F_v' \quad (9)$$

In this approach, changes in the heat dissipation pathway in both the dark and light-regulated state (D_{NPQ} and D_{NR} in **Figure 1**) are considered, which is useful in comparing samples acclimated to different light intensities or nutrient regimes.

All the above metrics of NPQ can be affected by baseline fluorescence (section “Blanks and baseline correction”), and it is worth pointing out that an observed decrease in F_m to F_m' can be caused irrespective of a change in the concentration of a quencher (e.g., state transition, Krause and Weis, 1991). Notwithstanding the different metrics used to quantify NPQ, current ST-ChlF instruments enable us to link our understanding of the physiological processes of NPQ in phytoplankton across molecular and eco-physiological levels. Controlled single-species laboratory experiments will provide additional insight into the photo-physiological plasticity underlying phytoplankton regulation of photosynthesis and photo-protection. At the same time, globally consistent and inter-comparable datasets of primary ST-ChlF parameters will reveal patterns in NPQ across ecologically relevant scales. Temporal or spatial patterns in photo-protection, observed through apparent variability in NPQ, could provide an optical proxy of nutrient limitation – a major determinant of aquatic productivity. Intriguingly, large scale patterns of NPQ proxies can also be obtained from (existing archives of) *in-vivo* ChlF measurements (section “Exploring environmental controls on primary productivity”), and potentially from changes in spectral absorption indices detectable by remote sensing approaches (Mélédér et al., 2018). A foundational understanding of NPQ mechanisms, which can be advanced by careful deployment of ST-ChlF instruments, will help to advance these high-level objectives.

OPERATIONAL AND PRACTICAL CONSIDERATIONS

The conceptual foundations described above provide the basis for interpreting variable ChlF data. In practice, a range of operational factors can significantly influence the quality and interpretability of observations, particularly during field deployments. We briefly discuss several important aspects below and refer readers to SCOR Working Group 156 (2021) for more detailed information.

A range of options are available for field deployment of ST-ChlF instruments, each with their own strengths and challenges. In ship-board laboratories, discrete samples can be analyzed individually, or data can be continuously acquired from a flow-through seawater supply (e.g., Behrenfeld et al., 2006). Truly *in-situ* data acquisition can also be achieved using instruments deployed on depth-profiling packages, towed platforms (e.g., Moore et al., 2003) or autonomous platforms, including moorings, floats, and gliders (e.g., Fujiki et al., 2008; Carvalho et al., 2020). Whereas continuous data collection provides advantages of high frequency measurement, discrete analysis allows for better control and optimization of experimental protocols for individual samples (e.g., light-response curves, NPQ-relaxation, tuning of ST protocols, and blank correction). For all deployment approaches, it is important to systematically identify the key operational factors with the greatest effect on ST-ChlF measurements, and to provide practical guidance on how these might be controlled and quantified (see also Laney, 2010). In the following, we focus our discussion on those operational aspects that should most influence the inter-comparability of ST-ChlF data from different researchers and instrument types. It is important to emphasize that no single approach will be optimal for all research needs across all systems. Rather, the goal is to understand and document the effects of different operational decisions on the resulting measurements.

Instrument Calibration and Standards

Calibration of ST-ChlF instruments is fundamental to ensure the collection of accurate and inter-comparable data by the global user community. The need for robust calibration approaches is particularly important given the increasing availability of custom-made instruments (e.g., Fujiki et al., 2008; Hoadley and Warner, 2017), for which data quality targets need to be defined. A comprehensive discussion of instrument calibration, taking into account particularities of specific instruments, will be provided in the Community-Best-Practice document (SCOR Working Group 156, 2021).

Calibration of Light Sources

Two categories of light sources are used in ST-ChlF instruments. In all instruments, strong “excitation light” is used to induce ST-ChlF transients. In addition, within many instruments, background (“actinic”) light is used to drive variable rates of photochemistry during light-regulated states (e.g., during light-response curves). The LEDs used as light sources in current instruments are typically very stable. Nonetheless, users should be aware of the need for routine monitoring and calibration of the photon flux density of these light sources to ensure quality

and consistency of produced data. In the case of excitation light, proper calibration is crucial for the derivation of σ_{PII} in absolute units and, by extension, derivation of J_{PII} (section “PSII photochemical flux, J_{PII} ”) using equations that involve σ_{PII} . Calibration of the background light is also important for robust data interpretation, for example to obtain inter-comparable values of light-response curve fit parameters (α and E_k ; see section “Light-response curves”). Finally, as discussed further in section “Spectral correction,” it is important to characterize and document the spectral quality of light provided by both excitation and background light, since LED spectra can vary significantly even from unit to unit within a manufacturing run.

At present, most end-users are not aware of the importance and difficulty of accurate light source calibration, and no standard protocols or procedures exist. The stability of LED light sources means that factory calibrations can remain valid for periods of months to years, but methods for “in-field” verification would clearly be desirable. Hand-held PAR sensors can be used to measure incident light fields within the sampling cuvette, but these calibrations can be rather finicky, sometimes pushing PAR sensors past their limits of dynamic range and response times. A major design challenge is achieving even illumination of all phytoplankton cells within the detected volume of the instrument. Going forward, we recommend that commercial instrument manufacturers provide guidelines, and perhaps ancillary hardware (e.g., cuvette inserts for light meter probes), for routine light source calibration.

Blanks and Baseline Correction

As described in section “Theoretical foundations and concepts,” measurements of ST-ChlF transients are used to quantify the variable ChlF between a minimum (F_o or F') and maximum (F_m or F_m') value. This variable ChlF is superimposed on a non-inducible (i.e., invariant) background fluorescence signal, which includes both a non-physiological component, the analytical blank, and a physiological component, the non-variable baseline fluorescence (**Figure 2**). The analytical blank represents background instrument noise (B_{inst}), and fluorescence originating from the dissolved phase of a sample (F_{diss}). The analytical blank can be significant relative to the measured values in oligotrophic regions (Cullen and Davis, 2003; Moore et al., 2008), and correction procedures are thus important. The baseline fluorescence is a physiological signal, often interpreted in the context of nutrient limitation (Behrenfeld et al., 2006; Macey et al., 2014), but not always fully characterized or understood. While baseline ChlF is not an operational issue, *per se*, we discuss it together with the analytical blank, as understanding the different sources of non-inducible fluorescence is crucial for correct data interpretation. The prevalence of baseline fluorescence, in particular in phytoplankton compared to higher plants, and the uncertainty around the sources and correct interpretation of this signal has contributed to confusion surrounding the application of ST-ChlF instruments in aquatic environments.

As shown in **Figure 2**, changes in the non-variable background fluorescence signal will affect F_o proportionally more than F_m , thereby affecting derived secondary ST-ChlF parameters (i.e.,

resulting in a drop in derived F_v/F_m). Consequently, correction is needed to collect the highest quality data possible (**Table 2**). Here, we separate the analytical blank into two components, B_{inst} and F_{diss} , both of which should be monitored regularly during field deployments of ST-ChlF instruments. For example, routine monitoring of B_{inst} (derived from measurements of ultra-pure water) is important to verify the absence of fouling in the sampling cuvette. Values of B_{inst} , originating from background luminescence of optical components (lenses, filters, and optical windows) induced by direct or elastically scattered excitation light can vary significantly between instrument types. Much progress has been made in lowering B_{inst} in newer instruments, resulting in blank ($=B_{inst} + F_{diss}$) values consistently $<5\%$ of F_m values even in the most oligotrophic open ocean waters. For such low B_{inst} values, it may be acceptable to subtract a constant B_{inst} from all measurements. Higher B_{inst} values may be difficult to correct because the amount of scattered excitation light inevitably varies among samples and will, for example, increase dramatically when highly scattering cells (e.g., calcified coccolithophores or chain-forming diatoms) are present in the sample.

Non-inducible fluorescence in the dissolved phase, F_{diss} , can be of particular importance in low biomass regions, and at depths below the chlorophyll maximum, where high background values can cause significant systematic error in the interpretation of F_v/F_m . During some deployments (e.g., unattended ship-board operation or extended *in-situ* deployments), routine F_{diss} correction can be challenging. Under such conditions, particularly for measurements in high biomass surface regions where the value of F_{diss} is frequently negligible relative to ChlF, it may be sufficient to subtract an instrument and wavelength-specific B_{inst} value from all measurements. Where there is a need to conduct regular F_{diss} measurements, it should be possible to develop simple fluidic systems to periodically introduce filtered water into the measurement chamber using automated micropumps and valves.

In contrast to the analytical blank, variations in baseline fluorescence can be rather complex, reflecting differences in the taxonomic composition and physiological state of phytoplankton assemblages. It is crucial to understand and resolve this term for correct interpretation of ST-ChlF data, as the models used for their interpretation explicitly assume that all measured ChlF is specific to the pigment pool of active PSII ($F_{o,c}$ and $F_{m,c}$ in **Figure 2**). Therefore, correction for baseline fluorescence is necessary for a strict interpretation of F_v/F_m as the quantum yield of photochemistry in PSII and when calculating [PSII] or J_{PII} following the absorption algorithm of Oxborough et al. (2012) and Boatman et al. (2019), or J_{PII} using Eqs 1, 2.

Low measured F_v/F_m values, not corrected for baseline fluorescence, are well established as an indicator of iron limitation and have been explained by the presence of energetically-decoupled light harvesting complexes (edLHC), which absorb light and emit ChlF, but do not transfer energy toward photochemistry (e.g., Behrenfeld and Milligan, 2013; Macey et al., 2014). These complexes contribute to the baseline fluorescence signal and thus increase the measured values of F_o and F_m by equal amounts, leaving F_v unchanged and hence lowering F_v/F_m .

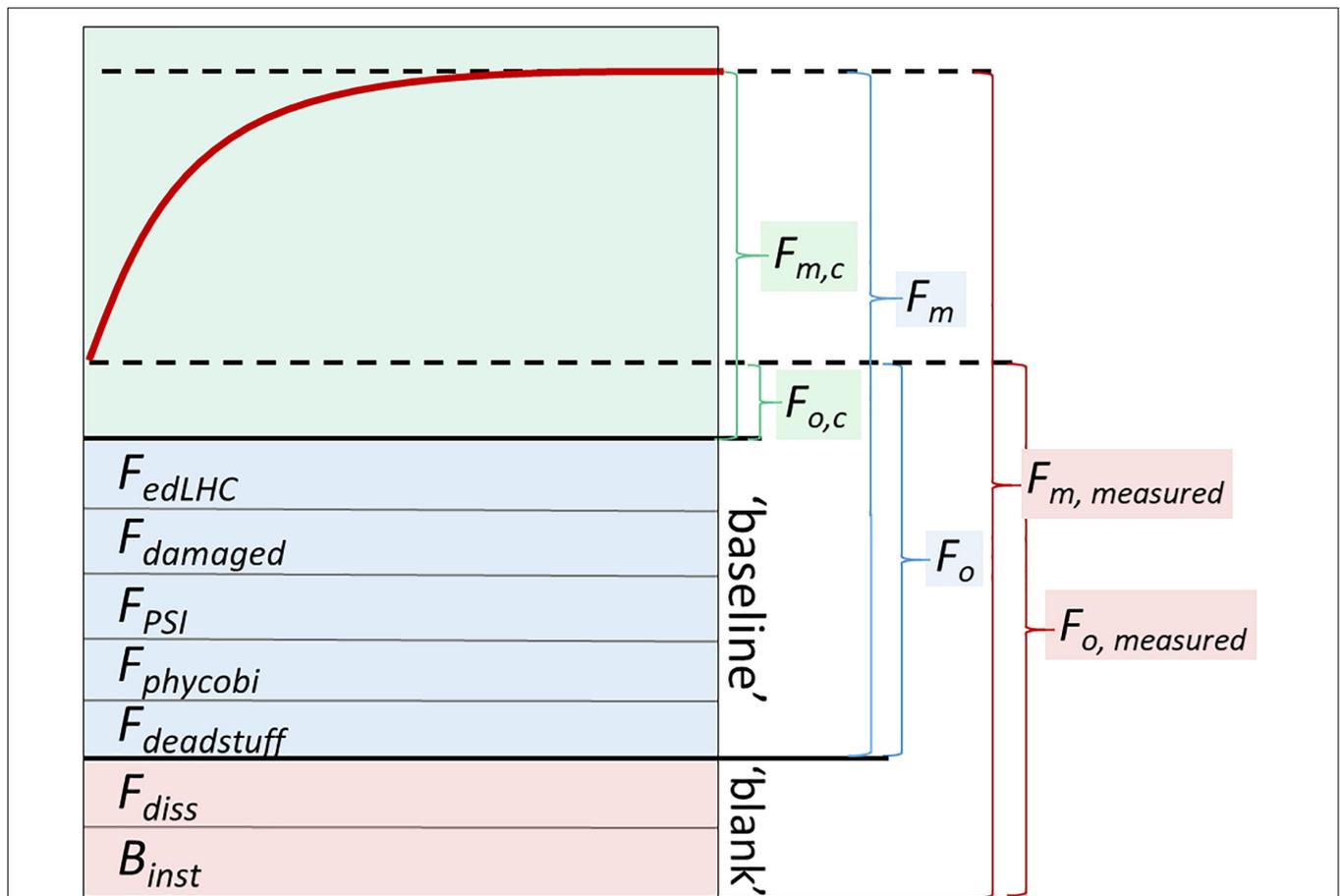


FIGURE 2 | Conceptual diagram of the components of the ChlF signal. The top part of the figure (green shading) shows a hypothetical ST-ChlF transient with ChlF from photochemically active PSII increasing from minimum ($F_{o,c}$) to maximum ($F_{m,c}$) values. This physiological signal is superimposed on several additional sources of non-inducible fluorescence. The analytical blank (pink shading) generally consists of instrument-specific blank (B_{inst}) and fluorescence from dissolved fluorophores in the sample (F_{diss}). Operationally, the analytical blank is defined as the apparent ChlF signal recorded in a 0.2 μm filtrate. B_{inst} largely reflects optical cross-talk between the instrument excitation and emission channels and can be operationally determined through measurement of ultra-pure water. The baseline fluorescence (blue shading) is composed of $F_{deadstuff}$ (non-living but still fluorescent phytoplankton and associated detritus), and a number of fluorescence sources in living phytoplankton. These latter sources include contributions from phycobiliproteins in some taxa (collectively $F_{phycobi}$), and fluorescence from PSI (F_{PSI}), which is not always negligible. Non-inducible fluorescence can also originate from damaged or inactive PSII complexes ($F_{damaged}$) or energetically decoupled light harvesting complexes (F_{edLHC}). Well-established correction procedures for baseline fluorescence do not exist, such that the interpretation of secondary ST-ChlF parameters based on models which explicitly assume ChlF to be specific to photochemically active PSII need to be treated with caution, particularly in phytoplankton assemblages in the field, where baseline fluorescence can be substantial. Conventionally, blank-corrected minimum and maximum ST-ChlF values, or values for which the blank was considered negligible, are reported as F_o and F_m in the literature, while baseline-corrected values have been referred to as $F_{o,c}$ and $F_{m,c}$. Note that the contributions of non-inducible fluorescence components in this conceptual diagram are not to scale, and would vary in amplitude in natural samples.

A decrease in F_v/F_m due to increased baseline fluorescence can also result from processes other than F_{edLHC} (Figure 2). However, limitation by nutrients other than iron does not always result in decreased F_v/F_m (e.g., Parkhill et al., 2001; Kruskopf and Flynn, 2006), likely because only severe starvation would lead to an accumulation of damaged PSII or dead cells ($F_{damaged}$ and $F_{deadstuff}$ in Figure 2), thereby increasing baseline fluorescence and decreasing F_v/F_m (Figure 2). Moreover, taxonomic effects, including the presence of phycobilin-containing species, can significantly increase baseline fluorescence ($F_{phycobi}$) as measured by ST-ChlF instruments. Finally, the fluorescence contribution from Photosystem I (F_{PSI}) is not always negligible, complicating algorithms that assume measured ChlF is solely contributed by

PSII. Going forward, it will be important to develop approaches to distinguish different sources of baseline fluorescence, and to interpret these signatures in terms of phytoplankton taxonomy and physiology. Furthermore, robust approaches for the correction of baseline fluorescence must be developed in order to improve amplitude-based J_{PII} algorithms (section 1.2.1, Table 3).

The Dark-Regulated States and NPQ-Relaxation

As discussed in section “Primary ChlF parameters” and Figure 1, interpretation of ST-ChlF measurements in the dark-regulated

state assumes that all RC are open and light-induced NPQ processes have been fully reversed. Measurements in the dark-regulated state are necessary, for example, to interpret changes in F_v/F_m in the context of iron limitation, or to use F_m as a proxy for chl biomass. Several approaches used to calculate J_{PII} from primary ChlF parameters also require measurements in the dark-regulated state (Table 2, section “Secondary ST-ChlF parameters”). A basic assumption is that all RCII are in the open state at the beginning of the ST-ChlF transient (Figure 1). Opening of RCII occurs on time-scales of milliseconds, and is thus easy to achieve with a short period of darkness. However, the time required to fully relax all light-induced NPQ (section “Non-photochemical quenching”) can be much longer and highly variable among species and environmental conditions (e.g., Goss and Lepetit, 2015). For this reason, it is challenging to achieve a fully dark-regulated state across mixed phytoplankton assemblages using standardized protocols.

One important recommendation, increasingly implemented in aquatic ST-ChlF instrument deployments, is the use of a low light treatment ($<5\text{--}10\ \mu\text{mol photons m}^{-2}\ \text{s}^{-1}$) rather than complete darkness, to induce relaxation of NPQ in samples. Low light availability will keep electron transport engaged at a basal level, minimizing so-called “dark-quenching” caused by respiratory reduction of the electron transport chain in prokaryotes (e.g., Campbell et al., 1998) or chlororespiration in diatoms (Jakob et al., 1999) and potentially in other species. Moreover, low light conditions are more appropriate for the relaxation of multiple NPQ components, often requiring energy provided by photosynthetic electron transport, particularly in diatoms, and cyanobacteria (e.g., Lavaud and Goss, 2014; Lacour et al., 2018).

A further complication is the presence of photoinactivated PSII complexes in samples from high light or otherwise stressed conditions. Repair of photoinactivated PSII can contribute to the slow relaxation of NPQ, with kinetics potentially overlapping the relaxation of other forms of NPQ (e.g., Li et al., 2016). There is no current consensus on whether or not PSII repair should be explicitly considered in NPQ relaxation or not. We note, however, that for estimations of community level productivity, the goal should be measurements that reflect the *in-situ* performance of the community rather than a hypothetical optimal performance. Thus recovery from photoinactivation is usually not the goal of the NPQ-relaxation period.

Although no “ideal” time-scale exists for NPQ-relaxation of field samples, we provide guidance for the measurement of dark-regulated ST-ChlF parameters in natural phytoplankton assemblages (Table 2). For bench-top measurements of discrete samples, users are encouraged to test different low-light exposure times whenever possible, and report results of such tests. For underway flow-through deployments, there are currently no consensus values, though results suggest 10–20 min of low-light exposure at *in-situ* temperature is needed to consistently relax most of the non-inhibitory phases of NPQ. For *in-situ* deployments, enclosure of a sample in a low-light chamber for a similar time will likely lead to the best and most comparable results. Importantly, NPQ-relaxation times should always be reported alongside published datasets.

Light-Response Curves

Photosynthetic light-response curves have been widely used to characterize environmental controls on the light dependency of photosynthesis, and to derive the photosynthetic parameters α (initial slope of light-dependent increase in photosynthetic rate), P_{max} (the maximum photosynthetic rate), and E_k (the light-saturation parameter; Platt and Gallegos, 1980; Farquhar et al., 2001). Traditionally, the rate of photosynthesis for such light-response curves has been derived from measurements of O_2 evolution or carbon uptake. However, ST-ChlF instruments also allow for easy and rapid acquisition of light response curves of primary ST-ChlF parameters, as well as derived properties such as J_{PII} and NPQ. The acquisition procedure for ChlF-based light response curves in different studies has varied widely in terms of lengths of light steps, inclusion of dark steps between light steps, spectral quality of background light, and order of light steps (low to high vs. high to low vs. non-sequential), making it challenging to compare photosynthetic parameters from different studies. Furthermore, several different model fits are used to derive photosynthetic parameters from such light-response curves (e.g., Silsbe and Kromkamp, 2012; Boatman et al., 2019). Noting that many of these issues also apply to O_2 or carbon-uptake derived light response curves (Bouman et al., 2018), these sources of variability should be systematically addressed if we are to assemble globally consistent datasets. As a minimum requirement, it is essential that photosynthetic parameters derived from ST-ChlF instruments are always reported alongside details of the acquisition protocol. In the case of different fitting routines currently available, open source software will be useful in providing end-users with a means to re-fit their data and assess differences in derived photosynthetic parameters using different models (section “Data reporting and archiving,” Ryan-Keogh and Robinson, 2021).

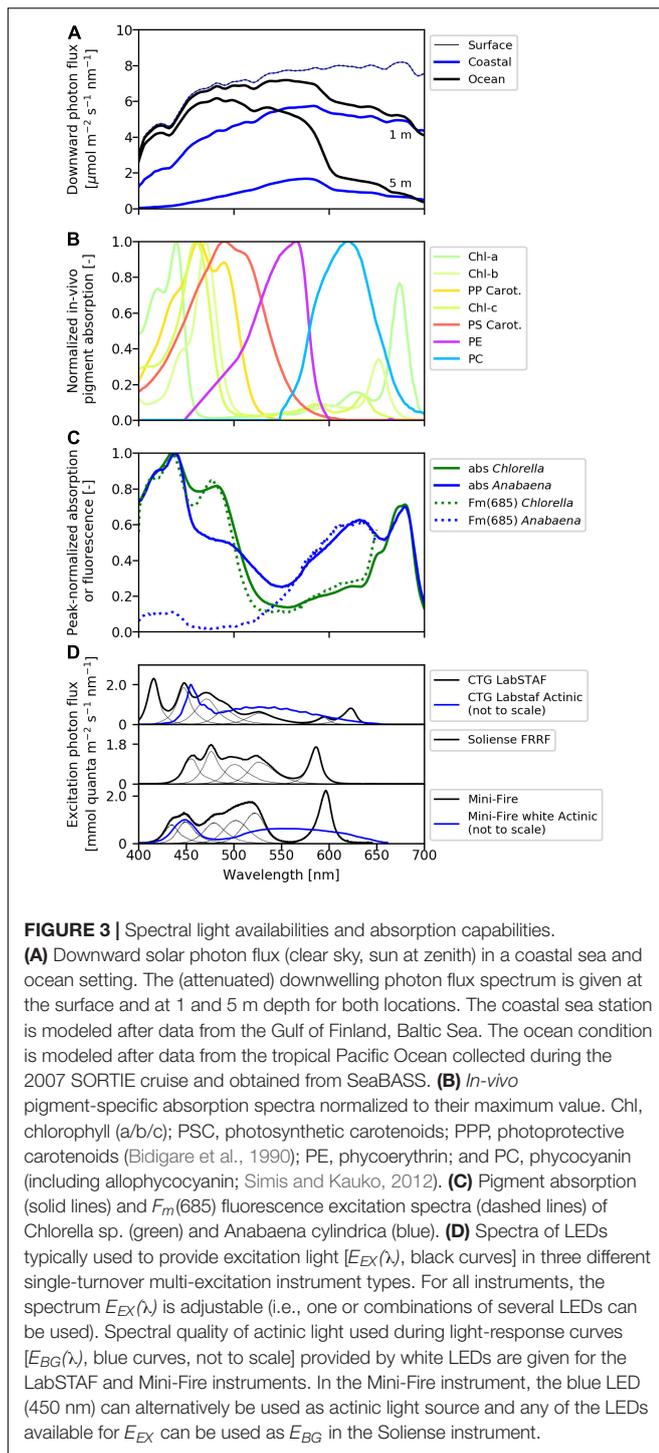
Spectral Correction

Biological oceanographers typically report light intensity in units of $\mu\text{mol photons m}^{-2}\ \text{s}^{-1}$ and integrated from 400 to 700 nm (so-called photosynthetically available radiation, PAR). Integration to a single number simplifies calculations and is justified because once absorbed, all light energy within the PAR spectrum can equally drive photochemistry. However, significant variability exists in the spectral distribution of incident light in various aquatic systems (e.g., Kirk, 2010; Johnsen, 2012; Figure 3A), in the light absorption capabilities of phytoplankton (Figure 3B), and the spectral properties of light sources used in different ST-ChlF instruments (Figure 3D). In order to collect environmentally relevant and inter-comparable ST-ChlF data, it is thus critical to consider these spectral differences, and apply corrections when necessary.

Fundamental to spectral correction procedures is the ability to calculate total PAR absorbed by phytoplankton assemblages (E_{abs}^{PAR}) as:

$$E_{abs}^{PAR} = \int_{400}^{700} a(\lambda) E_{source}(\lambda) d\lambda \quad (10)$$

Where $a(\lambda)$ is the phytoplankton absorption spectrum and $E_{source}(\lambda)$ is the spectrum of the light source. For example,



$E_{source}(\lambda)$ might correspond to the background light during light response curves ($E_{BG}(\lambda)$), available irradiance *in-situ* ($E_{IS}(\lambda)$), or a spectrally flat reference spectrum ($E_{ref}(\lambda)$; **Figure 3**). As long as measurements or estimates of these spectra are available, spectral correction procedures can be applied (e.g., Moore et al., 2006).

Not surprisingly, the primary ChlF parameters most affected by the spectral quality of the excitation light source are those

related to light capture, including σ_{PII} and the absolute value of F_o when the latter is being used to quantify light absorption by PSII (Eq. 5; Oxborough et al., 2012). Because the majority of eukaryotic phytoplankton absorb light most strongly in the blue part of the spectrum (**Figure 3C**), the use of a blue excitation source will usually result in a much larger value of σ_{PII} than excitation at other wavelengths (e.g., Gorbunov et al., 2020). In marked contrast, cyanobacteria often show a small response to blue excitation because their σ_{PII} is small in the blue waveband (e.g., Suggett et al., 2004). It is therefore important to report absolute values of σ_{PII} as a function of excitation light wavelength, $\sigma_{PII}(\lambda)$, and spectrally correct if inter-comparable or ecologically relevant absolute values are desired.

$$\sigma_{PII, BG} = \sigma_{PII, Ex} \cdot \frac{\int_{400}^{700} a(\lambda) E_{BG}(\lambda) d\lambda \cdot \int_{400}^{700} E_{EX}(\lambda) d\lambda}{\int_{400}^{700} a(\lambda) E_{EX}(\lambda) d\lambda \cdot \int_{400}^{700} E_{BG}(\lambda) d\lambda} \quad (11)$$

In Eq. 11, σ_{PII} is corrected to match the spectral quality of background light used during light response curves ($E_{BG}(\lambda)$), which is necessary if spectral quality of excitation light ($E_{EX}(\lambda)$) differs from that of background light (**Figure 3D**). The same approach can be used to calculate values of σ_{PII} relevant to $E_{IS}(\lambda)$ or $E_{ref}(\lambda)$. In order to obtain inter-comparable data, derived J_{PII} values (section “PSII photochemical flux, J_{PII} ”) require spectral correction, or should be reported as a wavelength specific value.

Spectral correction is also necessary to derive light response curve parameters (α , E_k) that are relevant to *in-situ* light availability or if comparing values from simultaneous experiments (e.g., ^{14}C -uptake) conducted with light of different spectral quality.

$$E_{k, IS} = E_{k, BG} \cdot \frac{\int_{400}^{700} a(\lambda) E_{BG}(\lambda) d\lambda \cdot \int_{400}^{700} E_{IS}(\lambda) d\lambda}{\int_{400}^{700} a(\lambda) E_{IS}(\lambda) d\lambda \cdot \int_{400}^{700} E_{BG}(\lambda) d\lambda} \quad (12)$$

In Eq. 12, the E_k value measured in a ST-ChlF instrument using a given background light spectrum ($E_{BG}(\lambda)$) is corrected to *in-situ* light availability ($E_{IS}(\lambda)$). Alternatively, all light levels used can be corrected prior to fitting the light response curve.

Inter-comparison among datasets requires that all ST-ChlF measurements are reported alongside spectral information of $E_{EX}(\lambda)$ and $E_{BG}(\lambda)$. Furthermore, it would be useful for all instruments to have at least one common excitation wavelength. Ideally, $E_{EX}(\lambda)$ and $E_{BG}(\lambda)$ would be of the same spectral quality, however, this can be difficult to achieve in practice due to engineering constraints. *In-situ* spectral light distribution should ideally be recorded simultaneously with the ST-ChlF measurement, although this can also be modeled with relatively high accuracy (Mobley, 1994; Lee et al., 2015).

Estimates of spectral light absorption $a(\lambda)$ required for spectral correction of ST-ChlF data have, until recently, relied on discrete measurements of phytoplankton specific absorption spectra using the filter-pad approach, reconstruction of absorption specific to photosynthetic pigments from pigment concentrations, or PSII fluorescence excitation spectra (e.g., Moore et al., 2006; Silsbe et al., 2015). In recent years, however, the development of multi-excitation wavelength ST-ChlF instruments has provided an approach to characterize

spectrally resolved photosynthetic responses of phytoplankton. With these instruments, relative light absorption profiles specific to PSII photochemistry, akin to PSII fluorescence excitation spectra, can now be derived with a high sampling resolution. Such multi-wavelength capabilities of next-generation instruments will greatly simplify spectral correction of ST-ChlF data.

Beyond excitation and background light sources, different ST-ChlF instruments deploy different spectral bands for detection of fluorescence emission, which will differentially bias their responses to chlorophyll fluorescence (from PSII) vs. fluorescence from other sources (including PSI, phycobiliproteins or organic matter). These instrument properties must similarly be reported and archived alongside derived measurements.

DATA REPORTING AND ARCHIVING

The development of best practices for the acquisition of ST-ChlF measurements is a critical step in the collection of coherent and inter-comparable datasets. However, in order to fully leverage the power of global data compilations, it is also necessary that raw data are freely available in a non-proprietary format, and archived with sufficient ancillary information to evaluate data quality and apply any corrections and/or future re-analysis. With this in mind, data collection and archiving should be guided by the “FAIR” principle, making information findable, accessible, interoperable, and reusable (Wilkinson et al., 2016). Implicit in this definition is a commitment to archive all data in consistent and accessible formats facilitating (re-)analysis with open-source computing tools in a manner that is platform agnostic.

Archiving of data (including raw ST-ChlF transients and primary and secondary ST-ChlF parameters) alongside all necessary ancillary information is a key requirement for efficient exchange and communication among research groups using different ST-ChlF instruments. Accessibility of raw ST-ChlF transient data will facilitate coherent re-analysis and quality control of large global datasets, helping to ensure backward and forward compatibility of measurements. Ancillary information (e.g., details of the ST protocols, wavelengths of LEDs and spectral bands of detectors, NPQ-relaxation time and light level, calibrations) is needed to evaluate potential biases in the reported data. Ideally, self-describing data formats allowing for efficient multi-dimensional storage and extraction, such as the netCDF standard, can be broadly adopted alongside a curated data ontology. For more holistic data interpretation, archived ST-ChlF measurements should be linked to additional supporting datasets that include key environmental (nutrients, temperature, surface PAR, mixed layer depth, time of day, sampling depth) and taxonomic (phytoplankton assemblage composition) variables. Guidance on how and where to archive ST-ChlF data will be provided to the user community as part of the Community-Best-Practice document (SCOR Working Group 156, 2021).

Once a robust database framework is established, end-users must be able to access and re-process raw data in a consistent and traceable way, choosing from a range of existing (and evolving) model fits. Toward this end, our group is developing a series of Python-based Jupyter notebooks allowing

users with various levels of experience and expertise to re-analyze ST-ChlF data collected with any instrument (Ryan-Keogh and Robinson, 2021). The software will continue to evolve, as new analysis approaches are developed, allowing direct comparison among different approaches to calculate J_{PII} or fit light-response curves. Drawing inspiration from the CO2SYS program used for thermodynamic calculations of the seawater carbonate system (Lewis and Wallace, 1998), we believe that the ability to re-process raw ST-ChlF data from different sources will maximize consistency and inter-comparability across research groups.

INTEGRATION AND APPLICATION

Deriving Carbon-Based Primary Productivity

Over the past two decades, there has been strong interest in deriving carbon-based primary productivity from ST-ChlF measurements (Hughes et al., 2018). This approach is based on the premise that a significant fraction of J_{PII} is used for the generation of ATP and NADPH, which in turn is utilized for carbon fixation (Figure 1). In practice, however, the measured stoichiometry between J_{PII} and carbon fixation (often referred to as the electron requirement for carbon fixation, $\Phi_{e,C}$, $\text{mol e}^- [\text{mol C}]^{-1}$) varies significantly. Many studies have experimentally determined $\Phi_{e,C}$ from parallel J_{PII} estimations and ^{14}C -uptake experiments (see Hughes et al., 2018 and references within). Collectively, such studies are beginning to reveal some coherent trends, such as increases in $\Phi_{e,C}$ above its reference minimum of ~ 5 under conditions of environmental stress, including excess light or limiting nutrients. Additionally, $\Phi_{e,C}$ appears to be significantly influenced by taxonomic variability, due to differences in metabolic strategies across different phytoplankton groups (Suggett et al., 2009a,b; Hughes et al., 2021). Moving forward, it is important to better understand the underlying mechanistic factors driving variability in $\Phi_{e,C}$. Ultimately, this variability represents the adjustable coupling of primary photosynthetic energy production and growth due to (taxon-specific) metabolic plasticity. An understanding of such plasticity will provide crucial insights into the environmental controls on photosynthetic energy use, carbon fixation in aquatic ecosystems, and the response of aquatic photosynthesis to environmental change.

To date, meta-analyses of $\Phi_{e,C}$ data from the literature have clearly demonstrated the challenge of separating methodological biases from true physiological variability in this parameter (e.g., Lawrenz et al., 2013). In many cases, procedural differences in J_{PII} or carbon fixation measurements across studies (e.g., different wavelengths of light used or varying ^{14}C -uptake incubation times) can introduce a level of variability comparable to that expected from taxonomic and/or environmental influences. For this reason, consistent approaches of data acquisition, reporting, and archiving are fundamental to characterize physiological and taxonomic variability in $\Phi_{e,C}$. Recommended standards of best practice in the deployment of ST-ChlF instruments (Table 2) and in the subsequent analysis of resulting data are critical steps toward this goal. At the same time, it will be important for users

to adopt consistent approaches for ^{14}C -uptake measurements, taking into consideration key factors including, for example, the incubation duration and the time of day that the incubation was initiated (e.g., Halsey et al., 2011; Milligan et al., 2015; Schuback et al., 2016).

Exploring Environmental Controls on Primary Productivity

Beyond the potential for high resolution estimates of JV_{PII} and associated carbon fixation rates (with the caveats discussed above), ST-ChlF approaches also provide a means of directly observing bottom-up controls on phytoplankton physiological ecology. In response to environmental variability on various time-scales (e.g., changing light fields and nutrient concentrations), phytoplankton can rapidly adjust the fraction of light energy absorbed and used to generate chemical energy, and the fraction of the chemical energy invested directly in carbon fixation and growth (Halsey and Jones, 2015). ST-ChlF approaches can provide insight into such physiological processes, enabling real-time observations of changing photosynthetic light utilization in response to interacting effects of multiple environmental and metabolic (e.g., cell cycle) factors.

As an example, ST-ChlF-based studies have been instrumental in examining the effect of iron limitation on phytoplankton physiology and productivity in the global oceans. Chronic iron limitation – affecting phytoplankton in >30% of the global ocean (Moore et al., 2013) – directly influences photosynthetic rates, as iron is a vital component of the photosynthetic electron transport chain (Raven et al., 1999; Yruela, 2013). PSII content can be lowered under iron limitation, and energetically disconnected light-harvesting complexes increase the non-inducible baseline fluorescence (section “Blanks and baseline correction”), resulting in the diagnostic decrease in measured F_v/F_m (Figure 2; Behrenfeld and Milligan, 2013). Furthermore, the absorption cross-section of PSII photochemistry (σ_{PII}) often increases under iron-limiting conditions, necessitating rapidly inducible NPQ under fluctuating light availability (e.g., Schuback and Tortell, 2019). Increased NPQ levels under iron-limited conditions can be extrapolated to observations of diurnal quenching of ChlF from “standard” *in-vivo* ChlF fluorometers deployed in continuous flow through and *in-situ* systems (Roesler and Barnard, 2013; Ryan-Keogh and Thomalla, 2020; Schallenberg et al., 2020).

The deployment of ST-ChlF instruments in continuous shipboard underway mode, or on moorings, floats and gliders provides a means of monitoring phytoplankton photo-physiology at a resolution comparable to that obtainable for other key oceanographic variables (light, temperature, etc.). Integration of ST-ChlF instruments with complementary high-resolution bio-optical measurements (e.g., absorption line height for chl concentrations, particulate carbon and cell size from backscatter, etc.) thus provide the means to observe how interacting physical and chemical conditions drive physiological changes resulting in changes in standing stocks. In this respect, it will be particularly powerful to integrate multi-instrument measurements in the analysis of observed diurnal periodicity in various productivity and biomass tracers, including oxygen (O_2/Ar), particulate

carbon (from optical backscatter and attenuation), and ST-ChlF derived photo-physiology in the surface ocean. Furthermore, integration of ST-ChlF data with high-resolution “omics” data from established and emerging approaches can be used to link phytoplankton cellular metabolism with optical indices, including those retrieved from remote sensing. These combined datasets will provide a mechanistic framework for the interpretation and continued refinement of various empirical modeling approaches and algorithms.

Special attention should also be given to parallel field deployments of ST-ChlF instruments with pico-second lifetime fluorometers, recently developed for absolute measurements of the quantum yield of ChlF in natural seawater (Lin et al., 2016; Falkowski et al., 2017). These highly complementary measurements can provide valuable insight into the controls of photosynthetic energy conversion in aquatic environments.

Application of Multi-Wavelengths ST-ChlF Instruments

In addition to the “bulk” phytoplankton photo-physiological properties discussed above, multi-spectral ST-ChlF data can provide information on the taxonomic composition of mixed phytoplankton assemblages (e.g., Gorbunov et al., 2020). The use of spectral fluorescence as a continuous monitoring tool of phytoplankton taxonomic composition is not new. Indeed, more than 40 years ago Yentsch and Yentsch (1979) wrote: “For continuous monitoring of phytoplankton in the open ocean, one anticipates the use of a series of fluorometers or a new fluorometric instrument for detecting emission and inducing excitation at several wavelengths.” Today, high-sensitivity, autonomous ST-ChlF instruments can collect data at >5 excitation and several emission wavebands, and these will provide large datasets to be interpreted in the context of community composition.

The presence of characteristic pigment compliments in different phytoplankton taxa (“spectral groups”) has long been exploited to infer information about phytoplankton taxonomic composition from discrete pigment samples (e.g., Mackey et al., 1996; Kramer et al., 2018) or deconvolution of light absorption and reflectance spectra (Bracher et al., 2017). The variable ChlF at ~ 680 nm, as measured by ST-ChlF instruments, stems primarily from chl associated with PSII. However, other pigments present in the PSII light harvesting antenna can pass absorbed energy on to chl such that the light absorbed by accessory photosynthetic pigments can induce ChlF. For this reason, fluorescence excitation spectra can vary significantly between phytoplankton groups (e.g., Johnsen and Sakshaug, 2007; Silsbe et al., 2015; Gorbunov et al., 2020).

Single-turnover variable chlorophyll fluorescence instruments are not yet able to fully resolve entire ChlF emission spectra, as is possible with fluorescence spectrometers (e.g., Seppälä and Balode, 1998; Seppälä and Olli, 2008). However, ChlF excitation at five to eight wavebands is sufficient to resolve characteristic “spectral profiles.” Importantly, such profiles do not merely track ChlF emission, but provide – for each waveband – all primary and secondary ST-ChlF parameters described in

section “Theoretical foundations and concepts.” Additionally, with the use of different bandpass filters, ST-ChlF instruments can be easily adapted to detect fluorescence emission at multiple wavebands. For example, fluorescence emission at 650–670 nm arising from phycobiliproteins and phycobilisomes found in cyanobacteria and cryptomonads can be used to monitor the occurrence and distribution of these groups (e.g., Seppälä et al., 2007). Looking forward, many exciting possibilities exist for the interpretation of multi-wavelength ST-ChlF data in the context of phytoplankton community composition. However, some words of caution are also necessary.

The main caveat inherent to all approaches deriving phytoplankton community composition from pigment complements (based on pigment concentrations, spectral light absorption, reflectance, or ChlF) is the fact that specific pigments are rarely unique to individual taxa, and their relative concentrations within a group will change in response to light and nutrient availability. Thus, while the photosynthetic efficiency could perhaps be derived from ST-ChlF data for “functional pigment groups,” the specificity of such analysis to any one particular taxa will likely be limited. Another limitation is the effect of “pigment packaging,” where variations in intracellular pigment concentration affect absorption, reflectance, or ChlF spectra. Variability in spectral shape due to pigment packaging is likely greater among species of the same phytoplankton group under different environmental conditions, than among different spectral groups. In the specific case of ST-ChlF measurements, systematic error can also be introduced into spectrally resolved ST-ChlF parameters if less well-absorbed wavelengths do not provide sufficient energy for full saturation of a ChlF transient during the ST saturation phase (section “Uncertainty and error”).

Notwithstanding these caveats, ST-ChlF instruments have great potential for fine-scale monitoring of phytoplankton taxonomic succession or early detection of harmful algal blooms. In the future, it will be particularly important for the research community to identify the most useful excitation and emission wavelengths for the detection and discrimination of specific phytoplankton groups. In the interest of compiling globally coherent and inter-comparable datasets, an effort should be made to match the wavelengths used in different instrument types.

The development of multi-spectral ST-ChlF instruments coincides with the increasing use of spectrally resolved approaches in satellite remote sensing and biogeochemical modeling, which depend on globally coherent datasets. Further development and application of these instruments should consider, for example, the wavebands measured by satellites, or target key taxa as identified by biogeochemical modeling.

ST-ChlF and the Interpretation of Sun Induced ChlF

In ST-ChlF instruments and *in-vivo* chl_a fluorometers, phytoplankton ChlF is controlled through application of an artificial light source. Natural sunlight also induces ChlF in phytoplankton, and the resulting emission at ~680 nm

(known as sun-induced fluorescence, SIF) can be detected by spectroradiometers, including those installed on satellites. The SIF signal detected by satellites has most commonly been measured as a normalized fluorescence line height, quantifying the remote sensing reflectance at the available waveband most affected by ChlF against a baseline from two adjacent, unaffected wavebands (e.g., Letelier and Abbott, 1996; Huot et al., 2005; Gupana et al., 2021).

Not surprisingly, the strongest determinant of the SIF signal is the underlying chl_a biomass in a given water mass. As discussed above, however, fluorescence per chl_a (or more correctly per absorbed photon) is not constant. For this reason, a number of studies have explored the use of SIF signatures as a diagnostic tool to assess phytoplankton physiological state (e.g., Letelier et al., 1997; Morrison, 2003; Schallenberg et al., 2008; Behrenfeld et al., 2009; O’Malley et al., 2014; Gilerson and Huot, 2017). Remote sensing of phytoplankton physiology over entire ocean basins would significantly advance the development of global primary productivity models. Yet, large uncertainties in the accurate retrieval of the faint SIF signal and in the physiological interpretation of this signal has so far limited the application of this approach in aquatic ecosystems. By comparison, there has been considerably more research exploring the empirical and mechanistic links between SIF, plant physiology, and primary productivity in the terrestrial realm where foliage with high chlorophyll concentrations generate larger signals (e.g., reviews by Porcar-Castell et al., 2014; Frankenberg and Berry, 2017; Mohammed et al., 2019).

The physiological information inherent to the SIF signal are variations in the quantum yield of fluorescence (Φ_f). The theoretical concepts underlying the physiological interpretation of Φ_f variability are the same as those outlined above for ST-ChlF approaches. Namely, the fraction of absorbed light energy re-emitted as ChlF is controlled by physiological changes in the fraction of energy used for photosynthesis and dissipated as heat (**Figure 1**). Despite this existing theoretical framework, extracting Φ_f from the SIF signal and interpreting variations in Φ_f in terms of physiology has proven to be far from trivial (Huot et al., 2005; Behrenfeld et al., 2009; Browning et al., 2014; Lin et al., 2016).

Great care must be taken when comparing active (ST-ChlF) and passive (SIF) fluorescence, as the approaches (and derived parameters with similar abbreviations) differ in crucial details. Nevertheless, globally coherent high resolution *in-situ* data from (autonomous) ST-ChlF instruments hold significant promise to improve the ground-truthing and interpretation of remotely sensed SIF. For example, current modeling approaches to derive Φ_f from SIF must account for changes in [chl_a], light absorption due to pigment packaging, and NPQ. ST-ChlF instruments are capable of collecting high-resolution *in-situ* data relevant to all these processes, and global data compilations will allow to detect spatial and temporal patterns needed for model parameterization. Furthermore, modern ST-ChlF instruments deployed *in-situ* or connected to continuous underway systems, have the ability to collect data that can be used to ground-truth the remotely sensed SIF signal itself.

Coordinated effort should thus be initiated to maximize the usefulness of *in-situ* ST-ChlF data to validate remotely

sensed SIF and improve model representation of Φ_f and other physiologically useful parameters. This work is particularly timely, given anticipated improvements of SIF detection with new generation hyperspectral sensors (e.g., Erickson et al., 2019; Köhler et al., 2020; Tenjo et al., 2021) and the increased use of ST-ChlF instruments on autonomous, *in-situ* platforms. Guidance and inspiration can be taken from the terrestrial remote sensing community, where paired measurements of active and passive ChlF in laboratory and field have been extensively utilized to examine fluorescence-photosynthesis linkages and explain observed diurnal, seasonal, and stress-induced variations in remotely sensed SIF (e.g., Porcar-Castell et al., 2014; Magney et al., 2017; Wyber et al., 2017; Maguire et al., 2020; Choudhary et al., 2021).

The use of ST-ChlF instruments to reveal global-scale patterns in photo-physiological metrics will contribute to the validation and refinement of ecosystem models and remote sensing algorithms, providing an exciting, and thus far under-utilized, opportunity to connect molecular-scale photosynthetic processes with global-scale biogeochemical cycles.

CONCLUSION AND OUTLOOK

Single-turnover variable chlorophyll fluorescence methods provide a powerful tool for high resolution photo-physiological measurements, with significant potential to examine aquatic productivity and its environmental controls over a range of spatial and temporal scales. Recent advances in instrumentation and data analysis are now beginning to significantly expand the application of ST-ChlF methods to a range of research questions. As the field continues to expand, it is essential to promote global coordination in the development of best practice, using flexible, open-source tools to disseminate information, software, and data products. Through the application of consensus recommendations (**Table 2**), and a robust system of documenting user-specific protocols, inter-comparison among emerging datasets will be greatly facilitated. This, in turn, will enable the synthesis of synoptic ST-ChlF observations at global scales, providing new insights into the response of marine productivity to a range of perturbations.

The UN Decade of Ocean Science for Sustainable Development (2021–2030) alongside the UN Sustainable Development Goals 6 and 14 (dealing with clean and productive inland and marine waters, respectively) will provide the opportunity to revolutionize the collection, storage, and analysis of ocean data, leading to better understanding of global-scale patterns in key ocean properties and their response to various environmental factors. ST-ChlF-derived observations are important supplements to existing observations that represent phytoplankton standing stocks. Current advances in

automation of measurements will enable data compilations at unprecedented resolution in both time and space. Such data will provide information on physical-biological coupling, including the impacts of localized hydrographic fronts, river plumes, and glacial discharge. At larger scales, regional patterns of phytoplankton physiology can be examined in relation to climate forcing, providing empirical correlations and mechanistic understanding for the improvement of ecosystem models, and remote sensing algorithms. Through concerted international cooperation, we are confident that the expansion of ST-ChlF measurements will significantly advance our understanding of global aquatic ecosystems.

AUTHOR'S NOTE

SCOR WG156 does not advocate for the use of any one particular instrument or protocol. The aim of this review is to outline necessary best-practices needed to ensure data inter-comparability across all ST-ChlF instruments, regardless of manufacturer.

AUTHOR CONTRIBUTIONS

NS and PT wrote the manuscript. All other authors contributed through discussions and comments, and editing of the manuscript.

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DEDICATION

We dedicate this article to the memory of our dear friend and colleague, Jacco Kromkamp, who passed away on Oct. 5, 2020. Jacco was a pioneer in the development and application of variable ChlF measurements to assess phytoplankton productivity in aquatic systems, and he believed passionately in the potential of this method to inform our understanding of ecosystem health. His ideas and insights contributed greatly to the development of our field, and his warmth and friendship helped bring our community together. He positively impacted our lives in numerous ways and will be deeply missed.

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