



Spring Ichthyoplankton Assemblage Structure in the Yangtze Estuary Under Environmental Factors

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Specialty section:

This article was submitted to Marine Fisheries, Aquaculture and Living Resources, a section of the journal *Frontiers in Marine Science*

Received: 31 October 2021

Accepted: 29 November 2021

Published: 16 December 2021

Citation:

Wang Y, Liang C, Chen Z, Liu S, Zhang H and Xian W (2021) Spring Ichthyoplankton Assemblage Structure in the Yangtze Estuary Under Environmental Factors. *Front. Mar. Sci.* 8:806096. doi: 10.3389/fmars.2021.806096

Estuaries, where fresh and salty water converge, provide abundant nutrients for ichthyoplankton. Ichthyoplankton, including fish eggs, larvae, and juveniles, are important fishery recruitment resources. The Yangtze Estuary and its adjacent waters comprise a typical large-scale estuary and supply many important fish spawning, feeding, and breeding areas. In this study, 1,291 ichthyoplankton individuals were collected in the Yangtze Estuary in spring, from 2013 to 2020. The aims of the study were to provide detailed information on characteristics of the ichthyoplankton assemblage, explore interannual variation, and evaluate the effects of environmental variables on the temporal variation in assemblage structure. Twenty-six species in seventeen families were identified. The dominant species were *Coilia mystus*, *Chelidonichthys spinosus*, *Engraulis japonicus*, *Hypoatherina valenciennei*, *Larimichthys polyactis*, *Salanx ariakensis*, *Stolephorus commersonii*, and *Trachidermus fasciatus*. The ichthyoplankton assemblage changed significantly over time, and *Chelidonichthys spinosus* became one of the dominant species. Canonical correspondence analysis showed that temperature and chlorophyll a were the key factors affecting the assemblage structure in the Yangtze Estuary in spring.

Keywords: fish eggs, larvae and juveniles, environmental factors, assemblage structure, interannual variability

INTRODUCTION

Estuaries are semi-closed coastal water bodies that are freely connected to the open sea, in which the seawater is diluted by fresh water produced by land drainage (Day, 1981; Potter et al., 1990). The interactions of physical, chemical, biological, and geological processes lead to complex change processes and a sensitive and unique estuarine environment that differs from adjacent fresh water and marine environments (Luo, 1994; Shan et al., 2004). Estuaries are considered to be among the most productive and nutrient-rich ecosystems on earth, and provide important habitat for a variety of species (Hodgson et al., 2020). Many species complete some or all of their life histories in estuaries, and these areas are particularly important foraging and breeding habitats for many fish species (Sheaves et al., 2015; Xian et al., 2016; Lefcheck et al., 2019). As a result of their unique geographical location, estuaries have been sites of human development for thousands of years

(Limburg, 1999; Lotze, 2010). Currently, about 40% of the world's population lives in coastal areas (Barragán and de Andrés, 2015). More than 90% of economical species, including fish, have been depleted, and biomass is lower than 50% of its historical abundance in some estuaries. The main drivers of this change are resource overexploitation and habitat loss (Lotze et al., 2006).

Ichthyoplankton, including eggs, larvae, and juveniles, are significant stages during the growth and development of fish (Butler et al., 2003; Shan et al., 2004) and are the basis for the sustainable utilization of fishery resources (Miller and Kendall, 2009; Zhang et al., 2015). Although these planktonic stages can be very short, they are the most vulnerable in a fish's life history (Shao et al., 2001; Jiang et al., 2006) and can be strongly affected by environmental variability. The abundance and survival of ichthyoplankton are indicators of available biomass and interannual dynamics of fish populations in the future (Butler et al., 2003; Song et al., 2019). In addition, ichthyoplankton play a significant role in energy transfer in estuarine ecosystems (Wan and Sun, 2006), consuming and converting bioenergy as a key link in aquatic food webs (Wan and Jiang, 2000).

Ichthyoplankton assemblages, species composition, and distribution are affected by the interactions of a range of biotic and abiotic processes (Zhang et al., 2015). Biological factors include the location, timing, and modes of spawning; larval life history and behavior; rates of predation; and feeding (Azeiteiro et al., 2006). Physical factors include salinity (Wang and Zhang, 1998; Wooldridge, 1999; Campfield and Houde, 2011); temperature (Neira and Potter, 1992; Primo et al., 2011; Wang et al., 2017); turbidity (Harris and Cyrus, 1995); dissolved oxygen (Rakocinski et al., 1996); depth (Zhang et al., 2015); light (Shoji et al., 2006); and hydrographic events such as river flow (Taylor, 1982), tides (Chen et al., 1997), currents (Wu, 1989), upwelling (Ramos et al., 2006; Valencia et al., 2019), and winds (Tzeng et al., 2002). Studies of ichthyoplankton provide a foundation for understanding trends in fishery resources, determining spawning sites and cycles, and clarifying fishing mechanisms; they can also be used to monitor ecosystem health.

The Yangtze Estuary is the link between the Yangtze River and the East China Sea. The estuary is influenced by fresh water from the Yangtze River, as well as the Yellow Sea cold water mass and the Taiwan warm current. The estuary is generally considered to be important nursery habitat used by many fishes to reproduce, forage, and grow (Luo and Shen, 1994). Although researchers have previously investigated ichthyoplankton composition and distribution in the Yangtze Estuary and the relationships between ichthyoplankton assemblages and environmental factors, the duration of these studies was short (Yang et al., 1990; Zhu et al., 2002; Jiang et al., 2006; Liu et al., 2008; Zhang et al., 2019) or discontinuous (Zhang et al., 2015, 2016). A clear understanding of how the physical and chemical characteristics of the Yangtze Estuary affect ichthyoplankton assemblages is still lacking.

The present study was based on consecutive survey data from spring 2013 to 2020. The aims of the study were to provide detailed information on the characteristics of the ichthyoplankton assemblage in the Yangtze Estuary in spring, explore interannual variability in ichthyoplankton assemblages,

and evaluate the effects of environmental variables on temporal variations in assemblage structure.

MATERIALS AND METHODS

Data Source

Forty stations were established in the Yangtze Estuary (30°45'–32°00' N, 121°00'–123°20' E), where the biological and oceanographic data were collected in eight springs (05/2013, 05/2014, 05/2015, 05/2016, 05/2017, 05/2018, 05/2019, 05/2020) (Figure 1).

In total, 320 ichthyoplankton samples were collected by surface tows using a plankton net (0.8 m mouth diameter, 2.8 m length, 0.505 mm mesh size) equipped with a flow meter. At each station, the net was towed against the tidal flow for 10 min at a depth of 0.5 m from the surface, with a towing speed of approximately 2–3 knots. After completion of each tow, the nets were cleaned and the samples were fixed and preserved in 5% buffered formaldehyde–seawater solution.

The geographical locations of sampling stations were determined using GPS (Magellan 315) (Zhang et al., 2016). A conductivity, temperature, and depth device (CTD) was used to measure depth (D), salinity (S), and temperature (T). Other environmental indicators, including dissolved oxygen (DO), pH, total phosphorus (TP), total nitrogen (TN), suspended particulate matter (SPM), chemical oxygen demand (COD), and chlorophyll *a* (Chla), were determined in the study according to GB/T 12763-2007 specifications for oceanographic surveys (Standardization Administration of China (SAC), 2007).

Data Analysis

Ichthyoplankton were identified in the laboratory to the lowest possible taxonomic level according to their morphological characteristics. Numerical density for each species was standardized to catch per unit effort (CPUE) as abundance per tow every 10 min.

The dominant species were determined using the index of relative importance (IRI) (Zhu et al., 2002), calculated as

$$IRI = N \times 100\% \times F \times 100\%,$$

where N 100% and F 100% are the relative abundance and frequency of occurrence, respectively. The IRI of the dominant species should be greater than 100.

To determine the significance of interannual trends in assemblage structure from 2013 to 2020, non-parametric ANOSIM analysis was performed based on a Bray–Curtis similarity matrix calculated using $\log(x + 1)$ transformed data (Clarke and Warwick, 2001). R-statistic values of paired comparisons provided by ANOSIM were used to determine differences between groups. R values close to 1 indicate that the differences between groups are greater than those within groups, and R values close to 0 indicate little difference between groups.

The ichthyoplankton assemblage structure and its relationship with environmental factors was analyzed by canonical correspondence analysis (CCA) (CANOCO

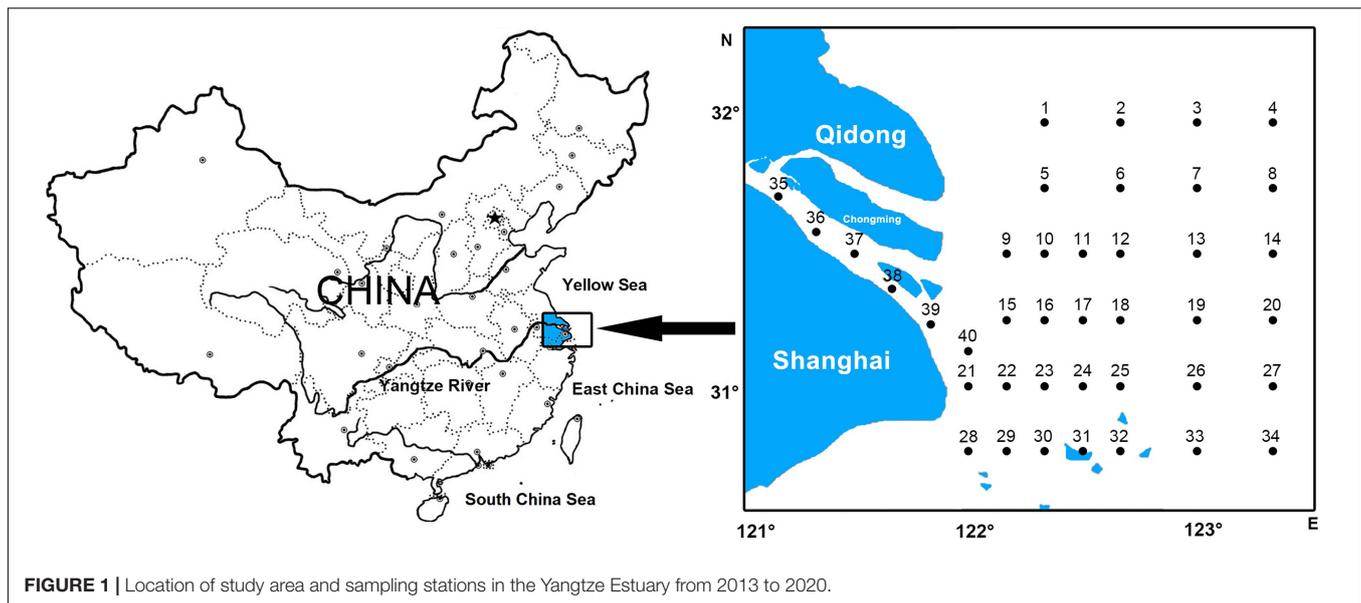


FIGURE 1 | Location of study area and sampling stations in the Yangtze Estuary from 2013 to 2020.

Software, version 5.0). Only species that occurred in $>0.5\%$ of the catches were included in the analysis. Fourteen species found between 2013 and 2020 were included in the analyses. The species abundance data were $\lg(x + 1)$ transformed to reduce the dominance effects of some species (Clarke and Warwick, 2001; Zhang et al., 2015). These 14 species and 10 environmental factors were used in the CCA.

RESULTS

Ichthyoplankton Composition

In total, 1,291 ichthyoplankton individuals, belonging to 26 species in 17 families, were captured in the Yangtze Estuary in spring 2013–2020. The scientific names, family, ecological guilds, and IRI values for the assemblages are shown in **Table 1**. The dominant species were *Coilia mystus*, *Chelidonichthys spinosus*, *Engraulis japonicus*, *Hypoatherina valenciennesi*, *Larimichthys polyactis*, *Salanx ariakensis*, *Stolephorus commersonii*, and *Trachidermus fasciatus*. Most of the species with low abundance were collected occasionally.

Interannual Variation of Assemblage Structure

The assemblage structure changed significantly in the Yangtze Estuary according to the IRI values for spring 2013–2020 (**Table 2**). Although *Engraulis japonicus* was always the dominant species in the Yangtze Estuary and adjacent waters, the composition of the other dominant species changed significantly, highlighted by the shift in the relative abundances of *Coilia mystus* and *Chelidonichthys spinosus*. The assemblage composition based on presence–absence methods varied among years and the average dissimilarity among years was very high (**Table 3**).

Ichthyoplankton Assemblages and Environmental Variables

The CCA results are shown in **Table 4**. Eigenvalues (0–1) indicate the importance of the CCA axes. The sum of all canonical eigenvalues and unconstrained eigenvalues were 1.338 and 7.729, respectively. The former only accounted for 17.31% of the latter, indicating the restrictive effect of building environmental relationships into the CCA model. The eigenvalues of the first four axes were 0.362 (CCA 1), 0.334 (CCA 2), 0.233 (CCA 3), and 0.167 (CCA 4) in the present study. The eigenvalues of the first two axes were moderately high, with correlation coefficients of 0.716 and 0.646, whereas those of the latter two were relatively low (<0.3). Consequently, the results for the first two axes were plotted (**Figure 2**), which explained 9.01 and 64.62% of the cumulative percentage variance of species and species–environment, respectively (**Table 4**).

On the basis of the Monte Carlo tests of F-ratios ($P < 0.05$), T and Chl_a were the most significant environmental variables affecting the ichthyoplankton assemblages (**Table 5**). In addition to these variables, the other eight environmental variables were evaluated by their inter-set correlations. There are two types of scores in this plot. (1) Environmental variable arrows, where each arrow points in the direction in which the value of the variable increases most sharply. The length of the arrow represents the strength of the environmental variable on the axis. The angle between arrows represents the correlation between the variables. More accurately, the approximate correlation between two environmental variables can be found by projecting the arrow of one variable onto an imaginary line running along the arrow direction of the other. (2) Species symbols: the distance between symbols is similar to the difference in the distribution of the relative abundance of these species in the sample, measured by chi-square distance. Adjacent points correspond to species that often occur at the same time. The species symbol can be projected vertically onto the line above the arrow of a

TABLE 1 | The basic ichthyoplankton information in the Yangtze Estuary in spring 2013–2020.

Family	Species	Abbreviation	Ecological guilds	Index of relative importance (IRI)									
				2013	2014	2015	2016	2017	2018	2019	2020		
Atherinidae	<i>Hypoatherina valenciennei</i>	Hv	Coastal assemblage	100.86									
Clupeidae	<i>Sardinella zunasi</i>	Sz	Offshore assemblage			10.00			10.00				
Cottidae	<i>Trachidermus fasciatus</i>	Tf	Estuary assemblage	2.16	127.36								
Engraulidae	<i>Coilia mystus</i>	Cm	Estuary assemblage	378.24	1745.28	2.50			2.50				40.00
Engraulidae	<i>Coilia nasus</i>	Cn	Estuary assemblage				3.08	92.59				1.97	
Engraulidae	<i>Engraulis japonicus</i>	Ej	Offshore assemblage	1793.59		70.00	370.37	583.33	350.00	1970.47	665.00		
Engraulidae	<i>Stolephorus commersonnii</i>	Sc	Coastal assemblage	18.01		35.00		27.78	35.00	236.22			
Engraulidae	<i>Setipinna tenuifilis</i>	St	Estuary assemblage			10.00		18.52	10.00				
Engraulidae	<i>Thryssa kammalensis</i>	Tk	Coastal assemblage					4.63					
Haplogenyidae	<i>Haplogenyis analis</i>	Ha	Offshore assemblage	0.36									
Hemiramphidae	<i>Hyporhamphus sajori</i>	Hs	Estuary assemblage								15.75	5.00	
Liparidae	<i>Liparis tanakae</i>	Lt	Offshore assemblage				3.09				1.97		
Lobotidae	<i>Haplogenyis kishinouyei</i>	Hk	Offshore assemblage										20.00
Salangidae	<i>Salanx ariakensis</i>	Sar	Estuary assemblage			150.00	870.37	9.26	187.50				15.00
Sciaenidae	<i>Larimichthys polyactis</i>	Lp	Offshore assemblage	90.06	7.08	250.00		74.07	250.00				5.00
Sciaenidae	<i>Pennahia argentata</i>	Pa	Offshore assemblage										5.00
Scombridae	<i>Scomber japonicus</i>	Sj	Offshore assemblage	1.44									
Sebastidae	<i>Sebastes marmoratus</i>	Sm	Offshore assemblage									1.97	
Sebastidae	<i>Sebastes schlegelii</i>	Ss	Offshore assemblage					4.63					
Synanceiidae	<i>Minous monodactylus</i>	Mm	Offshore assemblage			2.50			2.50				70.00
Syngnathidae	<i>Syngnathus acus</i>	Sac	Coastal assemblage					4.63					5.00
Synodontidae	<i>Harpadon nehereus</i>	Hn	Coastal assemblage			7.50	6.17	13.89	7.50				
Tetraodontidae	<i>Takifugu rubripes</i>	Tr	Estuary assemblage									11.81	
Triglidae	<i>Chelidonichthys spinosus</i>	Cs	Coastal assemblage		9.43	180.00	74.07	111.11	180.00	251.97	275.00		
Xenocyprididae	<i>Pseudolaubuca engraulis</i>	Pe	Estuary assemblage			2.50			2.50				
Zoarcidae	<i>Zoarcis elongatus</i>	Ze	Coastal assemblage			2.50			2.50				

TABLE 2 | Dominant species determined by index of relative importance (IRI) in the Yangtze Estuary in spring 2013–2020.

Species	2013	2014	2015	2016	2017	2018	2019	2020
<i>Coilia mystus</i>	378.24	1745.28						
<i>Chelidonichthys spinosus</i>			180.00		111.11	180.00	251.97	275.00
<i>Engraulis japonicus</i>	1793.59			370.37	583.33	350.00	1970.47	665.00
<i>Hypoatherina valenciennei</i>	100.86							
<i>Larimichthys polyactis</i>			250.00			250.00		
<i>Salanx ariakensis</i>			150.00	870.37		187.50		
<i>Stolephorus commersonnii</i>							236.22	
<i>Trachidermus fasciatus</i>		127.36						

specific environmental variable. These predictions can be used to approximate the optimal value of an environmental variable for a single species. The projection points for the species are arranged in the order of increasing value of the predictive variable. According to these rules, some environmental variables were inferred to be important correlates of one or both of the first two CCA axes. pH, Chla, S, and T were highly correlated with the first axis (CCA 1), while TN, SPM, TP, D, DO, and COD were strongly correlated with the second axis (CCA2). Although there was no obvious correlation between temperature and chlorophyll *a*, they both affected the interannual variation in ichthyoplankton (Figure 2).

Ichthyoplankton were divided into three species groups according to their ecological guilds (Table 1 and Figure 2). Offshore assemblage: six species (*Chelidonichthys spinosus*, *Sardinella zunasi*, *Minous monodactylus*, *Salanx ariakensis*, *Larimichthys polyactis*, and *Engraulis japonicus*) distributed close to CCA axis 2 and mainly correlated with T. All species, except for *Sardinella zunasi* and *Minous monodactylus*, were dominant. Coastal assemblage: three species (*Hypoatherina valenciennei*, *Stolephorus commersonnii*, and *Harpadon nehereus*) distributed in brackish waters. *Hypoatherina valenciennei* and *Stolephorus commersonnii* were dominant species. Estuarine assemblage: five species (*Hyporhamphus sajori*, *Setipinna tenuifilis*, *Trachidermus*

TABLE 3 | Inter-annual comparison of the assemblage structure according to one-way ANOSIM (R value and significance level) and SIMPER analysis.

Groups	ANOSIM		SIMPER
	R	P	Average dissimilarity
2013 and 2014	0.404	0.001	91.09
2013 and 2015	0.183	0.001	89.24
2013 and 2016	0.265	0.001	87.91
2013 and 2017	0.137	0.005	85.38
2013 and 2018	0.183	0.001	89.24
2013 and 2019	0.123	0.004	82.08
2013 and 2020	0.141	0.006	85.13
2014 and 2015	0.308	0.001	97.06
2014 and 2016	0.577	0.001	98.61
2014 and 2017	0.393	0.001	98.36
2014 and 2018	0.308	0.001	97.06
2014 and 2019	0.499	0.001	98.24
2014 and 2020	0.296	0.001	92.06
2015 and 2016	-0.059	0.884	83.46
2015 and 2017	-0.019	0.673	86.52
2015 and 2018	-0.056	0.999	84.00
2015 and 2019	0.089	0.016	86.42
2015 and 2020	0.014	0.279	87.79
2016 and 2017	0.049	0.145	85.88
2016 and 2018	-0.059	0.899	83.46
2016 and 2019	0.103	0.07	82.55
2016 and 2020	0.034	0.211	84.54
2017 and 2018	-0.019	0.694	86.52
2017 and 2019	0.041	0.11	81.68
2017 and 2020	-0.002	0.441	84.21
2018 and 2019	0.089	0.019	86.42
2018 and 2020	0.014	0.256	87.79
2019 and 2020	0.013	0.253	80.01

fasciatus, *Coilia mystus*, and *Coilia nasus*) with dispersed distribution and mainly correlated with S, pH, S, and Chla. *Trachidermus fasciatus* and *Coilia mystus* were dominant species.

DISCUSSION

Variation on Ichthyoplankton Composition

The ichthyoplankton assemblage in the Yangtze Estuary is changing, as reflected in the reduction in species number and shifts in species composition. In the 1980s, 53 families and 94 species of ichthyoplankton were recorded in the Yangtze Estuary (Yang et al., 1990). By 2000–2003, the number of species collected had decreased sharply, with only 30 families and 45 species (Jiang et al., 2006). From 2012 to 2014, the number of ichthyoplankton species decreased again, with only 38 species in 18 families (Kindong et al., 2020). In this study (2013–2020), 17 families and 26 species of ichthyoplankton were collected, which is only 27.66% of the diversity observed in the 1980s. These results show that the ichthyoplankton species composition in the Yangtze Estuary has changed significantly in the last 35 years.

The abundance of ichthyoplankton in estuaries is determined by the environmental characteristics and nutritional status of the estuarine ecosystem. The community structure of ichthyoplankton in the estuary was composed of a few high abundance species and a large number of rare species, which is a common feature of estuaries (Gaughan et al., 1990; Harrison and Whitfield, 1990; Drake and Arias, 1991). For example, ichthyoplankton communities in warm water estuaries were found to be dominated by Gobiidae, which live in estuaries, and Clupeidae and Engraulidae, which lay eggs seasonally in estuaries (Talbot and Able, 1984; Drake and Arias, 1991). Our results are in accordance with those of previous studies. The eight dominant species accounted for 94.58% of all ichthyoplankton sampled. Interestingly, *Engraulis japonicus* was the dominant species in the Yangtze Estuary from 2013 to 2020 in our study, accounting for 44.15% of the total catch, which is consistent with previous reports in which this species represented 67.91% of the total catch (Shen et al., 2011; Zhang et al., 2015). In addition, *Chelidonichthys spinosus* gradually became one of the dominant species after 2015, although it accounted for only 4.96% of the total catch. As far as we know, this finding is the first report of this shift.

Interannual Variability of Assemblage Structure

Estuaries have complex and variable physical, chemical, and hydrological conditions, with environmental factors that change dramatically in time and space (Whitfield, 1994; Harris and Cyrus, 1995; Hettler and Hare, 1998). Ichthyoplankton are very sensitive to environmental conditions (Shao et al., 2001). Thus, ichthyoplankton assemblages usually show high interannual variability in abundance and composition. By comparing the shifts in dominant species between this study (2013–2020) and previous ones (1999, 2001, 2004, and 2007), it was found that the dominant species of ichthyoplankton had decreased significantly, and the species composition of dominant species had changed (Figure 3). During the study period, there were obvious changes in the structure of the ichthyoplankton community in the Yangtze Estuary, indicating that environmental variables have a significant effect on interannual changes in ichthyoplankton community structure. However, there were no significant differences in the structure of the ichthyoplankton community in the spring in 1999, 2001, 2004, and 2007 (Zhang et al., 2015). Distribution patterns of ichthyoplankton are closely related to the reproductive strategy and life history of the adult populations, which are often associated with oceanography and meteorology (Hernández-Miranda et al., 2003). Therefore, the ability of juvenile fish to survive in the pelagic environment and migrate to suitable adult habitats may depend on their ability to regulate their diffusion or migration (Muhling et al., 2007).

Stochastic climatic events (cyclones) can alter the community composition of ichthyoplankton in estuaries by bringing intense rainfall, which significantly changes the distribution and abundance of estuary residents and marine migratory species (Martin et al., 1992). In addition, long-term climate trends, such as climate change and El Niño events, can also drive changes in community structure (Franco-Gordo et al., 2004;

TABLE 4 | Results of canonical correspondence analysis (CCA) relating ichthyoplankton abundance data to environmental factors in the Yangtze Estuary in spring 2013–2020.

Eigenvalues	CCA axes				Total inertia
	1	2	3	4	
Eigenvalues	0.3623	0.3339	0.2334	0.1668	7.72918
Species-environment correlations	4.69	9.01	12.03	14.18	
Cumulative percentage variance of species data	0.7157	0.6462	0.64	0.5049	
of species-environment relation	4.69	9.01	12.03	14.18	
Sum of all canonical eigenvalues	27.09	52.05	69.5	81.97	1.3375
Sum of all unconstrained eigenvalues					7.72918

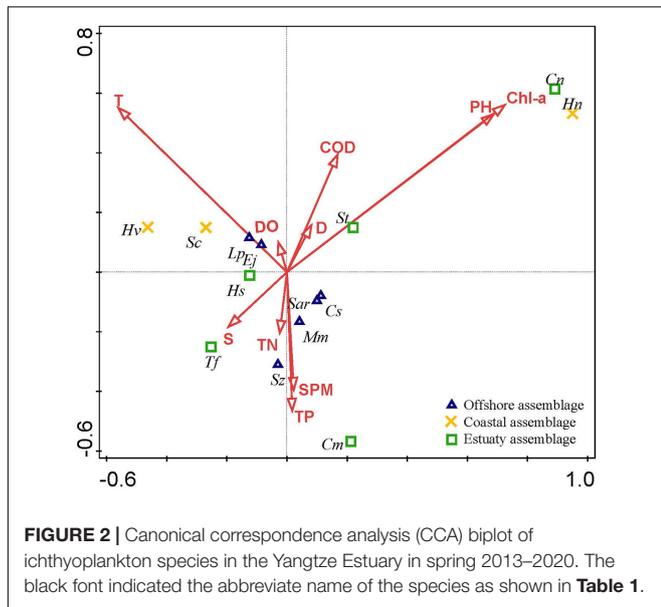


TABLE 5 | Conditional effects and correlations of environmental variables with the canonical correspondence analysis (CCA) axes.

Variables	Explains (%)	Contribution (%)	pseudo-F	P
Chl-a	4	22.9	3.9	0.002
T	3.3	19.2	3.4	0.002
TP	1.9	10.8	1.9	0.048
S	2	11.5	2.1	0.012
D	1.7	10	1.8	0.052
DO	1.1	6.2	1.1	0.304
COD	0.9	5.2	0.9	0.46
TN	0.8	4.8	0.9	0.556
PH	0.8	4.5	0.8	0.586
SPM	0.8	4.9	0.9	0.476

(Song et al., 2019). Under the joint influences of overfishing and environmental change, the population and community structure of fish resources in estuarine ecosystems have changed significantly (Bian et al., 2010), leading to loss of ecosystem resilience and integrity (Ryder et al., 1981).

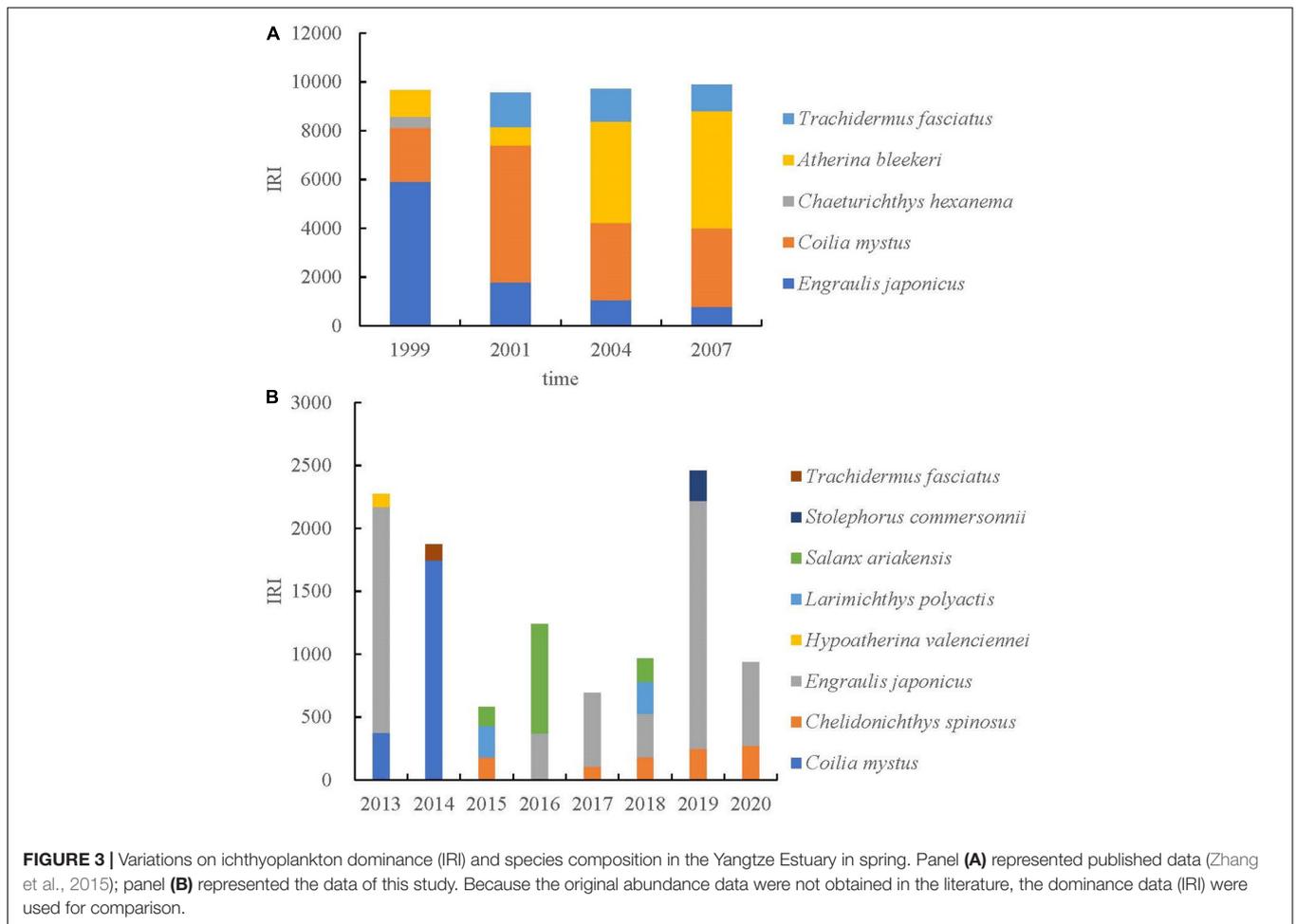
Acha et al., 2012). Climate change is altering the temperature, salinity, dissolved oxygen, and other estuarine conditions, resulting in changes to the temporal and spatial distribution of ichthyoplankton (Sloterdijk et al., 2017). Increasing temperatures could lead to earlier fish spawning (Thaxton et al., 2020), extension of the spawning time (Primo et al., 2011; Acha et al., 2012), shortening of egg hatching time (Hassell et al., 2008), early gonadal maturation (Thaxton et al., 2020), and northward shift of spawning areas (Auth et al., 2018; Zhang et al., 2019), resulting in changes in ichthyoplankton community structure.

In recent years, the modernization of fishing gear and improvements in fishing technology have led to increases in fishing intensity, which now far exceeds the capacity of resource replenishment. Some important economic species have become rare, and even endangered (Zhang et al., 2009). In the Yangtze Estuary, overfishing of important economic species, such as *Trichiurus japonicus* and *Larimichthys polyactis*, has led to fish miniaturization, early maturity, and juvenile structure (Shan et al., 2004). The main reason for the miniaturization of fish communities is the uncontrolled overuse of marine resources in coastal waters (Zhu et al., 2002). Overfishing is the main reason for the significant reductions in fishery resources

Ichthyoplankton Assemblage Structure and Its Influencing Factors

The ichthyoplankton community in the Yangtze Estuary was divided into three groups: estuarine, coastal, and offshore species (Zhang et al., 2015). It is generally believed that the distribution patterns of ichthyoplankton communities is affected by both biological and environmental factors (Zhu et al., 2002). In this study, CCA was used to determine the correlations between ichthyoplankton abundance (time–abundance data) and environmental variables (time–environment data). In general, T and Chl-a were the most significant factors affecting community structure. The other factors, such as pH, D, COD, DO, TN, and SPM, had no significant effect on ichthyoplankton assemblage structure.

Temperature has a significant and direct effect on fish reproduction, physiology, growth, and behavior (Sabates et al., 2006; Santos and Severi, 2019), which is very important for the abundance and spatial distribution of ichthyoplankton (Bruno et al., 2014; Chermahini et al., 2021). Estuaries are generally shallow, with large temperature changes, and diurnal and seasonal changes, therefore estuarine species adapted to a wide temperature range often dominate (Yang et al., 2006).



In temperate and subtropical estuaries, the peak number of species usually occurs in the spring and summer because of the annual changes in temperature (Neira and Potter, 1992; Primo et al., 2011; Wang et al., 2017). Therefore, during the growth and developmental stages of larvae and juveniles, temperature could have an important influence on growth, density, and reproduction (Ramos et al., 2006). Here, interannual changes in the abundance of *Engraulis japonicus* and *Larimichthys polyactis* were closely correlated with increasing temperature, and interannual changes in *Salanx ariakensis* and *Chelidonichthys spinosus* abundance were closely associated with decreasing temperatures.

Chla is an important environmental variable that has a significant influence on ichthyoplankton community structure, and was positively correlated with pH and negatively correlated with S. Chla is a relatively accurate reflection of phytoplankton standing stock; the larger the phytoplankton biomass, the higher the primary productivity (Whitfield, 1999). Chla is closely associated with fresh water nutrient levels entering the estuary. When nutrient levels are high, the Chla level in the estuary is also high, and food resources are rich, which is necessary for the development and growth of larvae and juveniles. Abundant food resources also improved the survival rate and successful foraging rate (Zhang et al., 2015). In this

study, three estuarine species (*Trachidermus fasciatus*, *Coilia nasus*, and *Setipinna tenuifilis*) were closely associated with the content of Chla (Figure 2). Zhang et al. (2016) also reported that Chla was closely related to the distribution of three freshwater species.

In addition to T and Chla, other environmental variables also affected the community structure of ichthyoplankton. TP and TN represent nutritional status, S, DO, COD, SPM, and pH represent water quality. Salinity is a key abiotic factor determining the structure of the ichthyoplankton community in estuaries (Wooldridge, 1999; Campfield and Houde, 2011). A decrease in salinity level can lead to peaks in fish eggs and larvae, and low salinity coastal systems have a significant influence on the hatching and development of fish eggs and larvae (Jiang et al., 2006). SPM is one of the most important factors affecting the density of estuarine species (Harris and Cyrus, 1995). Rakocinski et al. (1996) used canonical correlation analysis to analyze the changes in ichthyoplankton community structure in the Mississippi River and found that dissolved oxygen had a strong influence, as did temperature and salinity.

The variation in species distribution explained by the first four axes of the CCA was only 14.18%, which indicates that other factors are affecting the ichthyoplankton assemblage structure

in the Yangtze Estuary. Biological and human factors were not included in this study and the influence of these variables should be investigated in the Yangtze Estuary.

CONCLUSION

In this study, 1,291 ichthyoplankton individuals were collected in the Yangtze Estuary from 2013 to 2020, and identified as 26 species in 17 families. The dominant species were *Coilia mystus*, *Chelidonichthys spinosus*, *Engraulis japonicus*, *Hypoatherina valenciennae*, *Larimichthys polyactis*, *Salanx ariakensis*, *Stolephorus commersonnii*, and *Trachidermus fasciatus*. The interannual ichthyoplankton assemblage has changed significantly, reflected in the replacement of some dominant species and shifts in species composition. In particular, *Chelidonichthys spinosus* became a dominant species. Temperature and chlorophyll *a* were the key factors affecting the ichthyoplankton assemblage in the Yangtze Estuary in spring. Temperature and chlorophyll *a* were closely related to the abundance of offshore and estuarine species assemblages, respectively.

DATA AVAILABILITY STATEMENT

The original contributions presented in the study are included in the article/supplementary material, further inquiries can be directed to the corresponding authors.

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ETHICS STATEMENT

Ethical review and approval was not required for the animal study because this study did not involve animal ethics.

AUTHOR CONTRIBUTIONS

YW analyzed the data and completed the first draft. HZ and WX provided guidance on the structure of the manuscript. ZC and SL provided the suggestions on this manuscript. YW, CL, and HZ modified the manuscript. All authors are contributed to revise the manuscript.

FUNDING

The present work was supported by the National Natural Science Foundation of China (Nos. 41976094 and 31872568), the Key Deployment Project of Center for Ocean Mega-Science, Chinese Academy of Sciences (COMS2019Q14), and the Youth Innovation Promotion Association CAS (No. 2020211).

ACKNOWLEDGMENTS

We are grateful to the editors and reviewers for their constructive feedback and concerning on our work.

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