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Environmental DNA metabarcoding reveals spatial and seasonal patterns in the fish community in the Venice Lagoon

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Environmental DNA (eDNA) is an emerging tool for assessing biodiversity and understanding spatial and temporal community patterns and processes, directly from DNA sequencing of environmental samples such as air, water, and sediments. We applied eDNA methods to monitor bony fish communities, detecting as well locally allochthonous species, and to reveal seasonal patterns at two sites in the Venice Lagoon. We analyzed 17 water samples collected over 12 months at two ecologically distinct sites by using available primers for teleosts and High Throughput Illumina sequencing. We identified 1,289 amplicon sequence variants (ASVs) assigned to 62 fish taxa. Most of the species known to inhabit or to enter the Venice Lagoon were detected, with eDNA data reflecting differences in fish communities between the internal (freshwater associated) and the external (sea associated) part of the lagoon. Moreover, seasonal trends of migration have been portrayed, highlighting the most involved species and disclosing possible clashes between migration events and the temporary interruption of sea-lagoon connectivity due to MOSE (MOdulo Sperimentale Elettromeccanico). Of interest, the first-time detection of Oceanic puffer (Lagocephalus lagocephalus) DNA in the Venice Lagoon provides evidence of the further northward expansion of this species in the high Adriatic Sea. eDNA successfully profiled fish communities by season and habitat in the Venice Lagoon. Our results support routine application of eDNA to monitor potential ecological consequences of MOSE closures in this World Heritage site.

KEYWORDS

environmental DNA, metabarcoding, fish community, Tele02, alien species, oceanic pufferfish, Venice, lagoons

Introduction

Environmental DNA (eDNA) is an emerging and noninvasive technique which is considered highly promising for species monitoring studies (Bohmann et al., 2014). eDNA is a mixture of intracellular and/or extracellular nucleic acids, commonly released by organisms in the environment through feces, shred tissue, mucus, and/or blood, which can be analyzed by purification from different substrates (e.g., water, sediments, air, ice) (Taberlet et al., 2018). By combining co-amplification of small portions of barcode genes from purified eDNA with high throughput sequencing technologies (HTS), the eDNA analysis is highly effective in simultaneously identifying and differentiating multiple species (Deiner et al., 2017), thus providing "snap-shots" of different ecosystems (DiBattista et al., 2020). The analysis of environmental samples eases the detection of rare and endangered species in large systems (Bergman et al., 2016; Eva et al., 2016; Anderson et al., 2018). Besides being non-invasive, eDNA metabarcoding is also costand time-effective and has higher sensitivity compared with traditional approaches, allowing species detection at any life stage (Shu et al., 2020).

In the last decade, eDNA has been successfully applied, giving higher priority to fish macro-fauna (Tsuji et al., 2019), as support to conventional monitoring approaches, in species surveys of aquatic environments to obtain more detailed inventories of species distribution in time and space (Stoeckle et al., 2017; Priè et al., 2020; Aglieri et al., 2021). However, the studies that focused on aquatic ecosystems rarely considered transitional environments like lagoons (Suarez-Menendez et al., 2020; Oka et al., 2021; DiBattista et al., 2022). These are shallow coastal waters, intermittently subjected to variation in salinity and other parameters. As seen by their high productivity, they usually host a remarkable number of species by providing many habitat types, nurseries, and feeding grounds (Malavasi et al., 2005; Basset et al., 2006; Elliott and Whitfield, 2011; Sigovini, 2011; Tagliapietra et al., 2012; Whitfield et al., 2012; Basset et al., 2013). Due to their direct economic role in intensive aquaculture and fisheries, transitional environments also suffer from overexploitation and many other human threats, including water pollution, climate change, and introduction of alien species, raising the need for their protection (Giupponi et al., 1999; Solidoro et al., 2010; Occhipinti-Ambrogi et al., 2011).

We applied eDNA metabarcoding and set up sampling, filtration, and amplification protocols to analyze water samples in transitional environments. Short fragments of the 12S mitochondrial gene have been selected as the gold standard for eDNA detection of fishes. We analyzed samples that were collected in the Venice Lagoon, a wide coastal environment located in the northern Adriatic Sea. For the incredible cultural patrimony, Venice and its lagoon have been included in the UNESCO World Heritage Site List. The Venice Lagoon is separated from the sea by islands that can be temporarily connected by a huge complex of mobile dams (MOSE - MOdulo Sperimentale Elettromeccanico), whose function is to prevent the lagoon water level from flooding Venice during extreme high tides. After a long stalemate, the MOSE infrastructure has been nearly completed and tested over the few past years and it is currently working in cases of tides over 130 cm (https://www.mosevenezia.eu/). From a long-term perspective, an increase of the number of closures per year and an even larger and more rapid increase of the duration of closures is likely in the future (Mel et al., 2021) due to the high tide phenomenon that has strongly impacted the lagoon area for many years and has recently been increasing in frequency, which is most likely because of global climate change. If this occurs, the connectivity between the sea and the lagoon will alter, leading to ecological consequences.

The objectives of the present study were: 1) to obtain a fish biodiversity inventory in the Venice Lagoon, observing the resolution capacity of the Tele02 metabarcoding primer pair by comparing the detected amplicon sequence variants (ASVs) to what is known from traditional species surveys of the lagoon; 2) to examine fish communities, thus providing baseline information on which species are present in the lagoon (autochthonous and allochthonous) before the complete activation of MOSE, with a particular focus on the marine migratory component that is expected to be the most vulnerable to sea-lagoon connectivity perturbations, which can help anticipate ecological consequences; 3) to assess if, through eDNA metabarcoding, it is possible to portray differences in fish communities between two different areas of the same lagoon, a sea-associated one and a freshwater-associated one; and 4) to test if eDNA analysis can be useful in identifying patterns of seasonality in migratory fishes.

The present study tests a tool to enhance the lagoon species inventory and aims at providing the basis for a future systematic survey that might be crucial in understanding how healthy the functioning of the Venice Lagoon ecosystem will be after the connectivity perturbations of the MOSE infrastructure.

Materials and methods

Study area, eDNA sampling procedure, and extraction

The sampling was carried out in the Venice Lagoon (Italy), whose surface extends 500 km² into the north of the Adriatic Sea (Umgiesser et al., 2004). The lagoon environment is linked with the sea by three inlets (from north to south): Lido, Malamocco, and Chioggia (Figure 1). A total of 17 sampling campaigns were performed from November 2018 to December 2019 at two sites 31.39 km apart, one in the northern basin of the lagoon, close to Torcello (45°29.952'N 12°25.043'E), and the other in the southern part, close to the inlet of Chioggia (45°13.938'N 12°17.184'E)



(Figure 1). The northern site is approximately 7.5 km from the closest connection to the sea (Lido inlet) and is part of the Long Term Ecological Research network (LTER_EU_IT_016, station 5). The southern site (station 15) lies within proximity to the southern inlet (Chioggia inlet). During each sampling campaign, at least two replicates of superficial water were collected in plastic tanks of about 10 L.

Prior to sampling, the equipment was carefully sterilized by bleaching (1:10 solution) and thoroughly rinsing with MilliRo water. Afterwards, the samples were transported to the laboratory (Department of Biology at UNIPD) and stored at 8° C until filtration. Filtration took place within 24 hours of sampling, using a vacuum pump Edwards 5 two-stage. Given that most studies reported particles of a size up to 10 µm as the most common fish DNA molecules found in water samples (Shu et al., 2020), filtration was conducted for each water sample using both glass fiber filters GF/C (Whatman[®]) with a pore size of 1.2 µm and cellulose acetate filters (Sartorius®) with a 0.45 µm pore size, until the filters became clogged. Prior to filtration of each sample, all equipment, components of the filtration device, and surfaces were cleaned with 10% bleach. After each cleaning, a minimum of 1.5 L of pure water (MilliRo) was filtered, and these samples were used as filtration "blanks" to estimate the level of contamination, as suggested in Taberlet et al. (2018).

Two different filter types were used to identify the most appropriate for the specific protocol. However, since the results

obtained didn't show significant differences between the kind of filters (Martino, 2022), they are considered replicates for the purpose of this study. Full details on samples and filters are reported in Table 1.

A total of 81 filters, 42 from station 5 and 39 from station 15, were obtained from filtering lagoon water samples and stored at -80° C until the DNA extraction.

DNA extraction was carried out using the DNeasy Blood & Tissue kit (Qiagen[®]) in a separated pre-PCR environment equipped with positive air pressure. The amount of ATL buffer (Qiagen[®]) was increased until it completely covered the filters during the first incubation, and the volume of Proteinase K was modified accordingly. After incubation, 200 μ l of the solution containing DNA was processed following the manufacturer's protocol. At the end, the elution buffer was halved to increase the DNA concentration and the final DNA product was eluted in 100 μ l of AE buffer (Qiagen[®]). For each day of extraction, two extraction blanks were processed under the same conditions, by simulating an extraction without any filter (De Barba et al., 2014; Taberlet et al., 2018; Aglieri et al., 2021). Extracted eDNA samples were stored at -20°C.

Library amplification

For eDNA fish amplification, we used the primer pair Tele02 described in Taberlet et al. (2018), targeting, based on *in-silico*

Site	Date	No. Filters (GFC/cell)	Vol (L) - GFC	Vol. (L) - cell
5	14/03/2019	4 (2/2)	4.1	1.7
5	12/04/2019	4 (2/2)	3.0	1.0
5	16/05/2019	4 (2/2)	2.9	0.7
5	12/06/2019	4 (2/2)	3.9	0.9
5	08/07/2019	8 (4/4)	1.4	0.5
5	22/08/2019	4 (2/2)	0.7	0.2
5	23/09/2019	3 (1/2)	1.6	0.7
5	08/10/2019	4 (2/2)	2.9	0.9
5	04/11/2019	3 (1/2)	3.0	0.9
5	04/12/2019	4 (2/2)	2.6	1.0
15	13/11/2018	2 (1/1)	1.5	1.5
15	08/03/2019	18 (9/9)	3.4	2.3
15	14/03/2019	4 (2/2)	4.5	3.0
15	17/04/2019	3 (2/1)	3.8	1.5
15	16/05/2019	4 (2/2)	3.6	1.5
15	22/08/2019	4 (2/2)	2.7	0.7
15	04/11/2019	4 (2/2)	3.7	1.0

TABLE 1 Sample collection table reporting the site, the date of sampling, the total number of filters analyzed (between parentheses, the number of GF/C and cellulose acetate filters used), and the average volume (in liters) of filtered water for each filter type.

validation, a short fragment of about 130-209 bp of the 12S rRNA mtDNA gene of bony fishes.

Since 18 filters, from September 2019 to December 2019, out of the 81 starting filters, were processed twice, and one filter from March 2019 was amplified three times, the total number of eDNA libraries processed in this study was 101.

The eDNA samples were amplified using a single-step PCR protocol, as this minimized the possibility of contaminations by reducing the protocol steps, and it is reported to be the most effective in amplifying low-concentration templates (Taberlet et al., 2018; however, see Ushio et al., 2022, for potential disadvantages of this approach, such as the increase in the cost for primers and the variation among tags). To be multiplexed, Tele02 forward and reverse primers were tailed with an 8 bp-barcode and up to four random nucleotides at the 5' end (e.g., 5'NNNN-8bpBarcode-Primer-3'). This step allowed for tagging each amplicon by a unique and identifiable double-tag combination. The complete list of barcodes used in this study is given in Table S1; the 36 barcodes used were retrieved from Taberlet et al. (2018).

In addition to the eDNA samples, positive controls represented by DNA of the black goby, *Gobius niger*, were included in the PCR reaction to assess the success of amplification, sequencing, and taxonomic identification. This species was excluded from later analysis because of its use as a positive control during the PCR step. Moreover, PCR blanks were added, using water as a template, as well as sequencing blanks (one for each plate column and row, according to Taberlet et al., 2018) consisting of plate tubes with just the enzyme and primers tagged with unique tags. The latter blanks allow us to estimate the amount of "tag-jumping" (Schnell et al., 2015). PCR reactions were performed using a thermocycler (SimpliAmpTM, Applied Biosystems[®]) in triplicate to reduce amplification stochasticity. Final PCR volume was 10-20 µl, containing AmpliTaq GoldTM 360 MasterMix 1X (Life Technologies), each primer at the final concentration of 0.5 µM, and 2 µl of template (extracted eDNA or blank). The amplification thermal profile started with 10 min at 95°C, followed by 35 cycles of 30 s at 95°C, 1 min at 50°C, 30 s/ 1 min at 72°C, and finally 7 min at 72°C. Extractions and PCR reactions were set up in a pre-PCR environment, exclusively dedicated to eDNA, using filtered pipette tips.

The presence of PCR amplicons was checked on 1.8% agarose gel with a transilluminator (Gel DocTM XR+, Bio-Rad) for all reactions. PCR products were mixed to produce two different pools, the first including 96 and the second 48 samples (comprising eDNA samples, replicates, negative and positive controls). The amplicon pools were purified with the MinElute PCR Purification Kit (Qiagen[®]), with a final elution volume of 16 μ l, and the purification success was checked on 1.8% agarose gel. To reduce stochasticity, six purification replicates of each pool were carried out. Finally, after combining the six purification replicates, the two purified pools were quantified using Nanodrop 2000c (ThermoFisher), and QubitTM 4 Fluorometer (ThermoFisher). All the steps were performed in a post-PCR environment with the use of pipettes with filtered tips.

Afterwards, 20 µl of each purified pool was sent to a sequencing service (Norwegian Sequencing Center, Oslo, or BMR Genomics S.R.L., Padova) for adding Illumina-adapters, and for sequencing with Illumina Technologies 150 bp pairedend. The first pool was sequenced on an HiSeq4000 platform, starting from a sequencing library produced by the NEBNext[®] UltraTM DNA Library Prep Kit, and splitting the sequencing library on two half-lanes. The second pool was sequenced on a single MiSeq run, starting from a library obtained with the NEBNext[®] UltraTM II DNA Library Prep Kit. Two different sequencing platforms were used due to the unavailability of the HiSeq sequencing facility, and because the preliminary analysis of HiSeq data showed that the MiSeq throughput was adequate for our level of complexity, as indeed confirmed by the growing literature that used this latter platform (e.g., Miya et al., 2020).

Sequence analyses

Sequences were demultiplexed with Cutadapt (Martin, 2011), independently processed through QIIME2 pipeline (Bolyen et al., 2019), and denoised with DADA2 algorithm (Callahan et al., 2016). Chimeras were removed through the DADA2 algorithm (-p-chimera-method 'consensus').

Taxonomic assignments were performed by using the Mitohelper genetic database (Lim and Thompson, 2021), a QIIME2 compatible fish specific database. The whole pipeline is deposited in Github at the following link https://github.com/ Slide95/combinatorial-dual-indexes-metabarcoding. Taxonomic assignation of ASVs was curated manually through BLAST (Altschul et al., 1990) when ASVs were assigned to controversial species (species whose presence in the study area is not documented) or in case of clear mis-assignments (ASVs assigned to species whose presence in the study area is not possible). We removed ASVs that have been assigned to taxa with a taxonomic resolution lower than family (i.e., order, class etc.). As a general rule, if BLAST top matches had a percent identity lower than 97%, the ASV was discarded; if the top matches consisted of congeneric species with a very close percent identity, the ASV was assigned to the genus; if they were non-congeneric species but belonged to the same family, the ASV was assigned to the family. In the case of two similar congeneric species, if only one was reported in the Atlantic Ocean or the Mediterranean Sea, the ASV was assigned to this species. The ASVs that were assigned to the same taxon were at this point clustered into Operational Taxonomic Units (OTUs) for the subsequent analyses. OTUs assigned to the same taxon were merged, and their reads were summed.

To account for the different throughput of the two sequencing platforms used, read counts were normalized. To this end, we first calculated the ratio between the number of HiSeq and MiSeq retained reads, and then multiplied, for each sample and OTU, the MiSeq reads count by this correction factor. Then, OTU tables from the two sequencing outputs (HiSeq and MiSeq) were merged. Consequently, for each sample, read counts lower than 10 were removed to increase the reliability of the results such as in Adamo et al. (2020) and

Lopes et al. (2021). Finally, data were cleaned of contaminations through R (R Development Core Team, 2021). To this purpose, we first calculated, for each OTU, the distribution of the number of reads observed in the negative controls (filtration, extraction, PCR, and sequencing blanks), and then we subtracted the value of the third quartile of the distribution from the number of reads assigned to the corresponding OTU of each sample. To evaluate if our normalization and filtering procedure was successful in removing the bias associated with different sequencing depths, we performed a correlation test between the two vectors of normalized OTUs counts obtained from one technical replicate, consisting of a sample sequenced on both HiSeq and MiSeq platforms. The test was performed using the cor.test function with Pearson's method in R package version 4.1.2. In addition, as an alternative to normalization, we performed a coverage-based rarefaction (Chao and Jost, 2012) of the HiSeq and MiSeq retained reads, using the function phyloseq_coverage_raref of the metagMisc package (available at https://rdrr.io/github/ vmikk/metagMisc/man/phyloseq_coverage_raref.html), and we cross-checked the results of the ecological analyses obtained, starting from normalized and rarefied data.

Statistical and ecological analyses

For statistical and ecological analyses, the R Vegan Package (Oksanen et al., 2007) and QIIME2 were used. To account for eDNA performance, a species table was built with species names, sites in which they were detected by eDNA, IUCN status, and ecological guild. The table was then compared to fish distribution data retrieved from the literature (Cavraro et al., 2017). To investigate community dissimilarity between north and south, and seasonal clusterization, pairwise PERMANOVA analyses were performed through QIIME2 ("qiime diversity" function) starting from a Bray-Curtis dissimilarity matrix built from the refined OTU table of normalized abundances. This approach was selected because it was used in similar studies (Turon et al., 2020; DiBattista et al., 2022). However, since the Bray-Curtis distance is based on counts (number of reads in metabarcoding studies), and a quantitative interpretation of eDNA is still being debated, we checked to see if the relevant results were affected by our choice by performing PERMANOVA, using a Jaccard distance, after conversion of our matrix of normalized OTUs abundances to presence/absence data. In addition, considering that our dataset was comprised of different numbers of replicates for each site and sampling date (Table 1), we checked to see if these differences affected the PERMANOVA results by repeating this analysis on a different matrix. For this purpose, we subsampled our dataset to select only strictly comparable samples, i.e., those with the same number of replicates that were collected in the same month and year at the two sites. This resulted in keeping 38 of the original 101 samples and, indeed, also controlled for the inclusion of different sampling months at the two sites in the original dataset.

Differences between sites and seasons were plotted by nonmetric Multidimensional Scaling (nMDS), using the metaMDS function in Vegan. To identify singular patterns of species habitat selection between the two study areas, the number of reads were normalized for the two sites by dividing the read counts of each OTU of each sample by the average number of reads assigned to the corresponding site (north and south), and the ratios of normalized number (NAR) of reads corresponding to each species for both sites were compared. To identify the species which mostly contributes to the observed seasonality pattern, the "similarity percentage breakdown" procedure (SIMPER; Clarke et al., 2014) was performed using the Bray-Curtis dissimilarity matrix. Since the results can be skewed by overrepresented species (in metabarcoding, the species with the highest number of reads), the complete OTU table was normalized by dividing the read counts by the mean number of reads assigned to the OTUs in the relative sample, and species overrepresented in terms of reads were excluded before the SIMPER analysis. From the results, the first two species with a contribution higher than 5% and a p-value smaller than 0.05 were selected for ecological inferences.

Results

HiSeq and MiSeq sequencing (Illumina) produced, respectively, approximately 240 x 10^6 and 12 x 10^6 raw reads, with demultiplexing resulting in 77,487,725 (32.3%) and 8,780,314 (73%) reads. Denoising of reads, using the DADA2 pipeline, was performed independently for the two sequencing outputs. After filtration, denoising, paired-end merging, and chimera-removal, the number of reads retained for HiSeq and MiSeq outputs were, respectively, 53,409,947 (69% of demultiplexed reads) and 3,937,504 (45% of demultiplexed reads; complete denoising stats for every sample available in supplementary material, Table S2). Since two different sequencing platforms (HiSeq and MiSeq) were used at different times before proceeding to the ecological analyses, we estimated the stability of the data across different sequencing depths comparing, for a sample sequenced twice, its vector of OTU abundances obtained with HiSeq with the vector of its normalized MiSeq OTU abundances, as described in the Materials and Methods section. The high and significant correlation (rho = 0.989; cor.test p-value < 2.2e-16) supports the soundness of our approach and also, indirectly, the robustness of our results as, indeed, does the concordance of the ecological results obtained from the normalized data and from the coverage-based rarefied data (see below and Figures S1, S2).

eDNA detection of fish biodiversity

The analysis of 17 water samples from the Venice Lagoon allowed us to detect 42,888 ASVs, of which 1,289 were assigned to 62 taxa (2 families, 6 genera, 54 species) that are potentially present in this environment (Cavraro et al., 2017); the distribution of each species from south to north is reported in Table S3. Each taxon was assigned, based on its behavioral habits, to a given ecological guild, following Potter et al. (2015) and Cavraro et al. (2017). A graphic representation of the overall proportions of ecological guilds detected in our samples is reported in Figure 2.

In the whole dataset (Table S4), the most abundant OTUs with a taxonomic resolution at the species level, based on the number of reads assigned, were the big-scale sand smelt, Atherina boyeri, (10% of total reads assigned to OTUs), and the grass goby, Zosterisessor ophiocephalus (48%). This result was expected because the two species belong to the guild of the resident fishes in the Venice Lagoon. Overall, we identified seven species of lagoon residents (R) and ten species of marine migratory fishes (MM), the latter corresponded to fish species that tend to migrate from the sea to the lagoon during a specific time of year seeking food or that aggregate for the breeding season. The catadromous guild, here represented exclusively by Anguilla anguilla, was assimilated to marine migrants for these kinds of analyses. In addition, twelve species belonging to the freshwater guild (FW), twelve occasional migratory species (MO), as well as seventeen exclusively marine species (EM) have been detected. Based on the Italian Committee IUCN (International Union for Conservation of Nature) Red List (Battistoni et al., 2013), critically endangered species were found, such as Anguilla anguilla and Squalus acanthias, together with endangered or vulnerable species such as Mustelus mustelus and Alosa fallax (Table S3). Most of the other detected species were not currently in danger.

Detection of locally allochthonous species

Besides the alien species that are already known to be present in the Venice Lagoon, such as *Silurus glanis* (Corro, 2020), *Oncorhynchus mykiss* (Sicuro et al., 2016), and *Gambusia holbrooki* (Monti et al., 2021), eDNA analysis found traces of the presence of the oceanic pufferfish, *Lagocephalus lagocephalus*, a species that has never been detected in the lagoon or in the northern portion of the Adriatic Sea. The strong signal represented by the number of reads (more than 7000, see Table S4), and the absence of this species in the list of possible contaminants obtained by negative controls indicate that the detection of the eDNA of this species may be an early signal of its presence. Indeed, the presence of eDNA of the pufferfish was detected at the South site in March 2019 and at the North site in September and October of the same year.

Differences in northern and southern lagoon biodiversity

Retained reads, totals after the normalization and merging of the HiSeq and MiSeq data, were then assigned to the sites.



Respectively, 9,134,931 reads were attributed to Torcello and 10,080,623 to Chioggia. Through eDNA metabarcoding, 54 taxa were assigned to the southern and 48 to the northern lagoon.

Prior to the ecological analyses, we excluded from the dataset species that had never been documented before in the Mediterranean Sea Clupea harengus, Pleuronectes platessa, Salmo salar, and Sebastes mentella because their detection was probably the result of contamination. A pairwise PERMANOVA analysis between the two sampling sites was performed on a Bray-Curtis matrix built from a refined OTU table, obtained from the ASV table with realistic taxonomic assignments (see sequence analyses for details), and it returned highly significant results (Pseudo-F = 7.97; p = 0.001). Consequently, a nMDS ordination was performed on the distance matrix, and its results (k = 3, stress = 0.171) were plotted in a two-dimensional scatterplot, showing a clear differentiation among the northern and southern sites (Figure 3, see also Figure S1 for coveragebased rarefied data). The existence of differences between these sites was confirmed by using presence/absence data with a Jaccard distance matrix (PERMANOVA Pseudo-F = 9.44; p = 0.001) and by the analysis of a reduced dataset comprising of only samples with the same number of replicates collected in the same month and year at the two different sites (PERMANOVA Pseudo-F = 2.43; p = 0.005).

Eight species were detected exclusively in the northern lagoon, while fourteen were unique to the southern site. Of the species uniquely assigned to Torcello, five (62.5%) were freshwater fishes ("FW" ecological guild), while eight (57.1%) species, which were uniquely detected in Chioggia, were exclusively marine fishes ("EM" ecological guild). Moreover, to account not only for exclusivity but also for a clear disproportionality of the read numbers between the two sites in some OTUs, a ratio between normalized abundances (NAR) was calculated and used to produce a graphic representation of site selection for species. Even though most of the species were detected in both sites of the study, some of those were clearly more abundant in one site compared to the other one in terms of actual number of sequences detected (Figure 4). Species of the freshwater guild were mostly present in the internal site, the one closer to freshwater inputs. For example, *Cyprinus carpio* and *Rhodeus sericeus*, two freshwater species, were clearly more present in the northern site of the lagoon (Figure 4). On the other hand, species such as *Conger conger* and *Sardinella aurita*, exclusively marine, were more abundant in the site of Chioggia, which is the nearest site to the sea.

Seasonality

PERMANOVA pairwise analysis results are illustrated in Table 2. As before, Atlantic species known to be absent in the Venice Lagoon were excluded from this analysis. Significant differences were detected between all seasons, with a particularly clear distinction between Fall and Winter samples with respect to the other seasons, as also shown by nMDS (Figure 5, see also Figure S2 for coverage-based rarefied data). Similar to what was reported for the differences between sites, PERMANOVA also confirmed the differences between seasons when using presence/ absence data (Table S5) and when analyzing only strictly comparable samples (Table S6), though in this latter case some of the pairwise differences were smaller due to the use of a reduced dataset.

SIMPER analysis was performed after also excluding, in addition to Atlantic species, OTUs assigned to *Atherina boyeri* and to *Zosterisessor ophiocephalus* because these species are overrepresented (Table S4) and possible sources of bias.



Species that contributed the most to the differences between Summer and Fall were *Dicentrarchus labrax* (10.5%, p = 0.010) and *Chelon auratus* (9.7%, p = 0.033). Fish that primarily contributed to the Winter-Summer dissimilarity were *Chelon auratus* (10.7%, p = 0.009) and *Diplodus sargus* (5.0%, p = 0.001), while the ones that contributed to the Spring-Summer were *Chelon auratus* (18.3%, p = 0.001) and *Pomatoschistus minutus* (6.5%, p = 0.001). Finally, when including in the analysis only the samples of the northern site, also *Anguilla anguilla* resulted as a significant driver of the Summer-Winter diversity (6.6%, p = 0.031). In general, species found to influence the most seasonality patterns belonged mostly to the migratory ecological guilds, such as migratory fishes *Dicentrarchus labrax*, *Diplodus sargus* and *Chelon auratus*.

Discussion

Like in other eDNA studies metabarcoding fish diversity in the Mediterranean Sea by using Tele02 primers (Aglieri et al., 2021; Maiello et al., 2022), a high number of taxa was detected, and an elevated taxonomic resolution was achieved. In addition to bony fishes, three species of cartilaginous fishes were identified, indicating the applicability of this marker to simultaneously detect Actinopterygii and Chondrichthyes. The species richness retrieved in this study using eDNA metabarcoding was comparable to that reported in species checklists of the Venice Lagoon (Franzoi et al., 2010; Cavraro et al., 2017) and confirmed the high power of metabarcoding reported by Schroeder et al. (2020), who analyzed zooplankton samples collected at five sites in the Venice Lagoon, including LTER_EU_IT_016, station 5, one of the two chosen for our study.

Using eDNA, we detected species belonging to different ecological guilds (defined following Potter et al., 2015): occasional migrants (MO), marine migrants (MM), lagoon resident fishes (RL), freshwater species (FW), exclusive marine species (EM), and anadromes (AN). This variety of ecological guilds shows that surface water samples, taken from the Venice Lagoon, can produce a realistic picture of the functional ecological diversity along with the quantitative diversity represented by the number of species detected. The freshwater species detected (Abramis brama, Alburnus alborella, Blicca bjoerkna, Cyprinus carpio, Gambusia holbrooki, Padogobius martensii, Pseudorasbora parva, Rhodeus sericeus, Sander lucioperca, Silurus glanis, and Squalius cephalus) have different tolerances to salinity. Consequently, the detection of species less tolerant to low salinity could be due to transport of eDNA with river discharge, while the detection of tolerant species can be, more confidently, attributed to the presence of a local community. Interestingly, the RL guild, composed of species



Overrepresented species at the North and South sites. On the x-axis, the ratio between North and South normalized abundances is reported in Log2 scale; positive values indicate species overrepresented in the northern site, negative values indicate species overrepresented in the southern site. Only species both present in north and south sites have been considered in this graph and reported on the y-axis.

Group 1	Group 2	Sample Size	Pseudo-F	p-Value	q-Value	
Fall	Spring	57	6.46	0.001	0.0012	
Fall	Summer	54	3.59	0.001	0.0012	
Fall	Winter	66	5.11	0.001	0.0012	
Spring	Summer	35	2.46	0.020	0.0200	
Spring	Winter	47	6.45	0.001	0.0012	
Summer	Winter	44	5.11	0.001	0.0012	

TABLE 2 Summary of pairwise PERMANOVA results for seasonality, based on Bray-Curtis dissimilarity and 101 samples.

Each row reports the results of the pairwise comparison between groups of samples for different seasons. The number of samples used in each comparison, a measure of inter-group differences (Pseudo-F), and the corresponding p- and q- values are reported. Number of Permutations = 999.

adapted to live in transitional environments, was represented by the highest number of reads. The MM and MO guilds were also remarkable, and they deserve particular attention due to the predictable impact that the MOSE system can have on these kinds of species, limiting or interrupting their movement between the lagoon and the sea. In this perspective, the fact that migrators were detected in this study in the same periods when MOSE was first activated (https://www.mosevenezia.eu/ilmose-in-funzione/#mvbtab_61894462f0019-1) highlights the potential impact of this infrastructure on these species. Regarding unexpected species, this study is the first, to the best of our knowledge, to detect the oceanic puffer *Lagocephalus lagocephalus* along the upper Adriatic coasts (http://www.iucn.it/ scheda.php?id=-2081245894), though caution is needed since only eDNA was detected. The literature on the oceanic puffer mostly focuses on its strong poison (Saoudi et al., 2008; Pinto et al., 2019), whereas studies on the congeneric species *Lagocephalus sceleratus* also focus on ecological factors and the ongoing invasion (Coro et al., 2018). The oceanic puffer is diffused throughout the oceanic tropical and subtropical waters, and it was Cananzi et al.



first reported in the Mediterranean Sea, around Sicily, by Doderlain (1878). The species is considered present in the Mediterranean Sea by Tortonese (1986), except for the Adriatic and Black seas. Several recent "first detections" have been recently reported for areas in the Mediterranean Sea, where L. lagocephalus was never recorded before (see for instance: Erguden et al., 2017; Alshawy et al., 2019), suggesting that the species is currently expanding its range. In addition, according to Zava et al. (2005), starting from 1999, the number of detections of the oceanic puffer near the Sicilian coasts has increased. In the Adriatic Sea, the oceanic puffer was first spotted in 2004, with one specimen captured near the southern Adriatic Croatian coast (Dulcic and Pallaoro, 2006), and then, in 2015, two new records were reported further north along the Croatian coasts, with one specimen from the Kornati archipelago and one near to the island of Rab (Tsiamis et al., 2015). The detection of the oceanic pufferfish DNA in the Venice Lagoon adds a further support for an ongoing range expansion of this species in the Mediterranean Sea and raises concerns for the Adriatic Italian coast.

The northern and internal site (Torcello) and the southern seaassociated one (Chioggia), surveyed in this study, are both inside the Venice Lagoon but 31.39 km apart. They represent two different habitats inside a transitional ecosystem: Torcello is about 7.5 km from the closest lagoon inlet, while the Chioggia site is directly next to the Chioggia inlet. Based on this, our strategy has been to retrieve, through eDNA metabarcoding in both sites, the portrait of the ichthyological communities in both high salinity and low salinity environments (see https://issos.ve.ismar.cnr.it/for salinity maps). Our results indicate the presence of statistically significant differences and a clear clustering between samples belonging to the two sites. Moreover, the list of species (Table S3) and the proportion of the normalized reads abundance (Figure 4) show that exclusive marine species are predominant in Chioggia and freshwater fishes in Torcello. This, as other studies such as DiBattista et al. (2022), shows that eDNA metabarcoding can portray changes in fish communities caused by salinity in a lagoon environment. As a note of caution, differences between these two sites could be due to the proximity to bias factors. In fact, eDNA samples from Chioggia contained species known to be absent in the lagoon, such as the swordfish Xiphias gladius or the tuna Thunnus sp., whose DNA could have been detected due to the fact that they are traded, as a consequence of fish market discards. Such a hypothesis is supported by the presence of one of the largest fish-markets in Italy in the city of Chioggia, about 1.5 km from the Chioggia sampling site. Moreover, it cannot be excluded that the detection of eDNA belonging to these species at the Chioggia site is due to transport from off-shore because of hydrodynamics.

In addition to investigating differences between sites, our study aimed at picturing seasonal trends of marine migratory fishes through eDNA, in line with other works (Sigsgaard et al., 2017; Djurhuus et al., 2020; DiBattista et al., 2022). In our case, these trends are important because the MOSE infrastructure, a huge complex of dams that will alter the sea-lagoon connectivity, will inevitably condition the marine migratory ichthyological guild. Statistical analyses detected significant differences between seasons, particularly when comparing winter and fall with the other seasons. Migrators such as *Chelon auratus*, *Diplodus sargus*, and *Dicentrarchus labrax* contributed most to these differences, as evidenced by the SIMPER analysis. These species are characterized by marine reproduction and spend the early stages, as well as other periods of their lives, in shallow coastal waters or lagoons (Rossi, 1986; Franzoi et al., 1989; Franzoi and Trisolini, 1991; Franco et al., 2006). Abundance peaks of these species in the Venice Lagoon have been documented before in late winter/early spring and at the end of the summer/start of fall (Rossi, 1986), coherent to the SIMPER results obtained in this study.

Importantly, also the catadromous species Anguilla anguilla was detected in the Venice Lagoon, almost exclusively in eDNA for the fall and winter seasons (Table S4). Although it was not one of the species that contributed most to the seasonality patterns when the SIMPER analysis was performed on the whole dataset, it resulted as a significant driver of the summerwinter dissimilarity when considering only the more freshwater associated northern site. Individuals of this species are known to leave rivers to migrate to the Sargasso Sea for reproduction (Cresci, 2020), and this species is temporarily abundant in the Venice Lagoon in fall, according to catches of fyke net surveys (Scapin et al., 2022). The MOSE infrastructure, according to official data, have been activated 33 times in the period of October 2020 - October 2022, exclusively in fall and winter (https://www.mosevenezia.eu/il-mose-in-funzione/#mvbtab 61894462f0019-1). A correct management of this alteration in the future will be important as Anguilla anguilla is classified by the IUCN as a critically endangered species, and its severe demographic decrease is attributed to many causes, among which is the migratory barrier establishment (Starkie, 2003).

Conclusions

In this work, we obtained an ichthyological list of the Venice Lagoon species by using eDNA metabarcoding. The list of species is comparable to available checklists obtained by traditional monitoring. Beyond the qualitative result, functional and ecological diversity was observed, through the detection of six different ecological guilds. Two distinct communities were observed, Torcello with a strong component of freshwater fishes' eDNA, and Chioggia with a dominant component of exclusive marine fishes. In addition, eDNA metabarcoding unveiled seasonality trends, and species most involved in this pattern were catadromous or migratory fishes. Migration periods of several species do coincide with the functioning of MOSE, the huge complex of dams that is currently working to preserve Venice from high tides. This structure, despite its importance for the protection of the historical and cultural patrimony that Venice represents, could function as a barrier for migrators, thus impacting the life strategy of these species, many of which are of strong commercial importance at the local scale. We highlight the urgence to account for this phenomenon and the need to perform extensive biodiversity monitoring, particularly for migratory species. Considering that the expected frequency of extreme events in the future, together with the projected sea level rise, will increase the number and duration of MOSE closures, we believe that eDNA should be routinely used to document ongoing changes in lagoon biodiversity and to provide input to stakeholders useful to move toward a "regulated lagoon".

Data availability statement

The datasets presented in this study can be found in online repositories. The name of the repository and accession numbers can be found below: NCBI; PRJNA878955, BioSamples SAMN30788935 and SAMN30788936.

Author contributions

GC, IG, FM, EB, IAMM, and LZ conceived and designed the experiments. GC, IG, and FM performed the experiments. GC, IG, FM, and TL analyzed the data. EC, MP, AS, and LC contributed sampling/reagents/materials/analysis tools. GC, IG, FM, TL, EB, EC, LC, IAMM, MP, AS, and LZ critically discussed results; GC, IG, FM, TL, and LZ wrote the paper, which was approved by all the coauthors.

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Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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Supplementary material

The Supplementary Material for this article can be found online at: https://www.frontiersin.org/articles/10.3389/ fmars.2022.1009490/full#supplementary-material

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