



Chaotic Genetic Patchiness in the Highly Valued Atlantic Stalked Barnacle *Pollicipes pollicipes* From the Iberian Peninsula: Implications for Fisheries Management

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The stalked barnacle *Pollicipes pollicipes* inhabits rocky shores from the Atlantic coasts of Brittany (France) to Senegal. Because of the culinary traditions of southern Europe, stalked barnacles represent an important target species for local fisheries on the Iberian Peninsula. To manage this fishery sustainably, it is therefore important to assess the dynamics of local populations over the Iberian coast, and how they are interconnected at a wider scale using finely tuned genetic markers. In this work, a new enriched library of GT microsatellites for *P. pollicipes* was prepared and sequenced using Ion Torrent™ Next Gen-Sequencing Technology. 1,423 adults and juveniles were sampled in 15 localities of three geographic regions: southern Portugal, Galicia and Asturias (both in northern Spain). Twenty polymorphic loci arranged in five multiplex PCRs were then tested and validated as new molecular tools to address the spatial and temporal genetic patterns of *P. pollicipes*. Our results revealed high genetic diversity among adults. However, juveniles were genetically more structured than their adult counterparts, which alternatively displayed much more connectivity among the three studied regions. The lack of spatial genetic heterogeneity in adults may be due to the overlapping of several generations of settlers coming from different geographic origins, which mainly depends on the orientation of residual currents along the coast during reproduction. The genetic differentiation of juveniles may indeed be congruent with Iberian Peninsula hydrodynamics, which can produce chaotic genetic patchiness (CGP) at small temporal scales due to sweepstake reproductive success, collective

dispersal and/or self-recruitment. Remarkably, most of the genetic heterogeneity of juveniles found in this work was located in Galicia, which could represent an admixture between distinct metapopulations or an old refuge for the most northern populations. To conclude, high genetic variation in *P. pollicipes* can lead to the false impression of population panmixia at the Iberian scale by masking more restricted and current-driven larval exchanges between regions. This possibility should be taken into consideration for further specific management and conservation plans for the species over the Iberian Peninsula.

Keywords: stalked barnacle, multiplex PCR, microsatellite, small-scale fisheries, recruitment, stock management, connectivity

INTRODUCTION

The percentage of stocks exploited at biologically unsustainable levels increased from 10% in 1974 to 34.2% in 2017 (FAO, 2020) after decades of management strategies based on catch-rate limitations (i.e., the EU Common Fisheries Policy). As an alternative or complementary approach, management practices are increasingly incorporating the spatial allocation of fishing intensity through marine protected areas, marine zoning, or spatial user rights, particularly for sessile or low-motility species (Lorenzen et al., 2010; Rassweiler et al., 2012). Optimization of these processes depends on the accurate estimation of the connectivity among management units, mediated by the dispersal of the planktonic larval stages (Silva et al., 2019). In this regard, a fundamental issue concerns whether the dispersal scales are consistent with the management scales (Ouréns et al., 2015). Although advection by ocean currents should lead to long dispersal distances exceeding the scale of management, there is increasing evidence that long-distance dispersal may be rare on ecological time scales (Palumbi, 2003; Selkoe et al., 2010; D'Aloia et al., 2015). This phenomenon can be explained by a combination of seascape characteristics such as eddies, gyres or upwellings of deep water bodies and specific larval behavior that would favor local retention and reduced dispersal (Morgan et al., 2009, 2018; Barshis et al., 2011; Kough et al., 2013). An additional line of evidence reveals surprising patterns of spatial and temporal genetic structure observed in some marine species at a scale where genetic variation should be efficiently homogenized by gene flow via larval dispersal, collectively coined chaotic genetic patchiness (CGP) (Johnson and Black, 1982; Hedgecock and Pudovkin, 2011; Eldon et al., 2016).

The stalked barnacle (*Pollicipes pollicipes*) is a pollicipedomorph cirriped (Chan et al., 2021) inhabiting rocky coasts that are highly exposed to waves in the northeast Atlantic. Its range extends from southwestern England through the coasts of Brittany (France), Spain, Portugal, and West Africa to Dakar (Senegal) (Barnes, 1996, 2008; Southward, 2008; Fernandes et al., 2010). In the Iberian Peninsula, it represents a highly valued resource that reaches very high market prices due to an old gastronomic tradition (Molares and Freire, 2003; Jacinto et al., 2011; Rivera et al., 2014). Remains of its consumption

have been found in early Holocene archeological sites, mainly associated with Mesolithic and Neolithic shell-middens on both the Atlantic and Mediterranean coasts (Álvarez-Fernández et al., 2010, 2013; Álvarez-Fernández, 2011). Between 2013 and 2016, the European stalked barnacle fisheries have an annual economic value of EUR 10 million, involving approximately 500 t of landings and 2,100 professional fishers (Aguión et al., 2021). At some localities, the pressure exerted by poachers can be extremely high (more than 60% of their catches), especially in banned areas or periods (Jacinto et al., 2010; Rivera et al., 2014; Ruiz-Díaz et al., 2020).

Management of the stalked barnacle fishery in the Iberian Peninsula is highly heterogeneous (Aguión et al., 2021). In Galicia (NW Spain) since 1992, the regional government has developed a co-management system between fishers' guilds ("cofradías") and the fisheries authority through territorial user rights for fishing (TURFs) (Molares and Freire, 2003; Macho et al., 2013), where exclusive right of access are granted to fishing communities (Costello et al., 2010; Rivera et al., 2014). Similarly, in the West coast of Asturias (N Spain), the barnacle fishery has been managed through a co-management system with TURFs since 1994 (Rivera et al., 2014, 2017). Both Galicia and western Asturias present adaptive spatial management with nested scales at regional, local and patch/rock levels; recognized to promote fisheries sustainability (Aguión et al., 2021). However, on the eastern coast of Cape Peñas (eastern Asturias) and Portugal, the fishery is managed at a regional scale through general regulations without management plans (Aguión et al., 2021). In Portugal, however, there are two protected areas subjected to specific regulations for harvesting *P. pollicipes*: the Reserva Natural das Berlengas (RNB) and the Parque Natural do Sudoeste Alentejano e Costa Vicentina (PNSACV) (Sousa et al., 2013; Cruz et al., 2015; Carvalho et al., 2017). The first one (RNB) is subjected to local management, resembling a TURF in many aspects (Aguión et al., 2021). Currently, there is interest and potential for developing co-management systems similar to the one in Galicia and western Asturias in both Portuguese protected areas (Cruz et al., 2015; Sousa et al., 2020). Among different management approaches, TURFs represent the best option for the sustainable management of small-scale sessile fisheries (Gutiérrez et al., 2011; Rivera et al., 2017; Aguión et al., 2021). However, the design of management areas mandates a good understanding of population renewals for which estimates of connectivity

Abbreviations: CGP, chaotic genetic patchiness; TURE, territorial use rights for fishing.

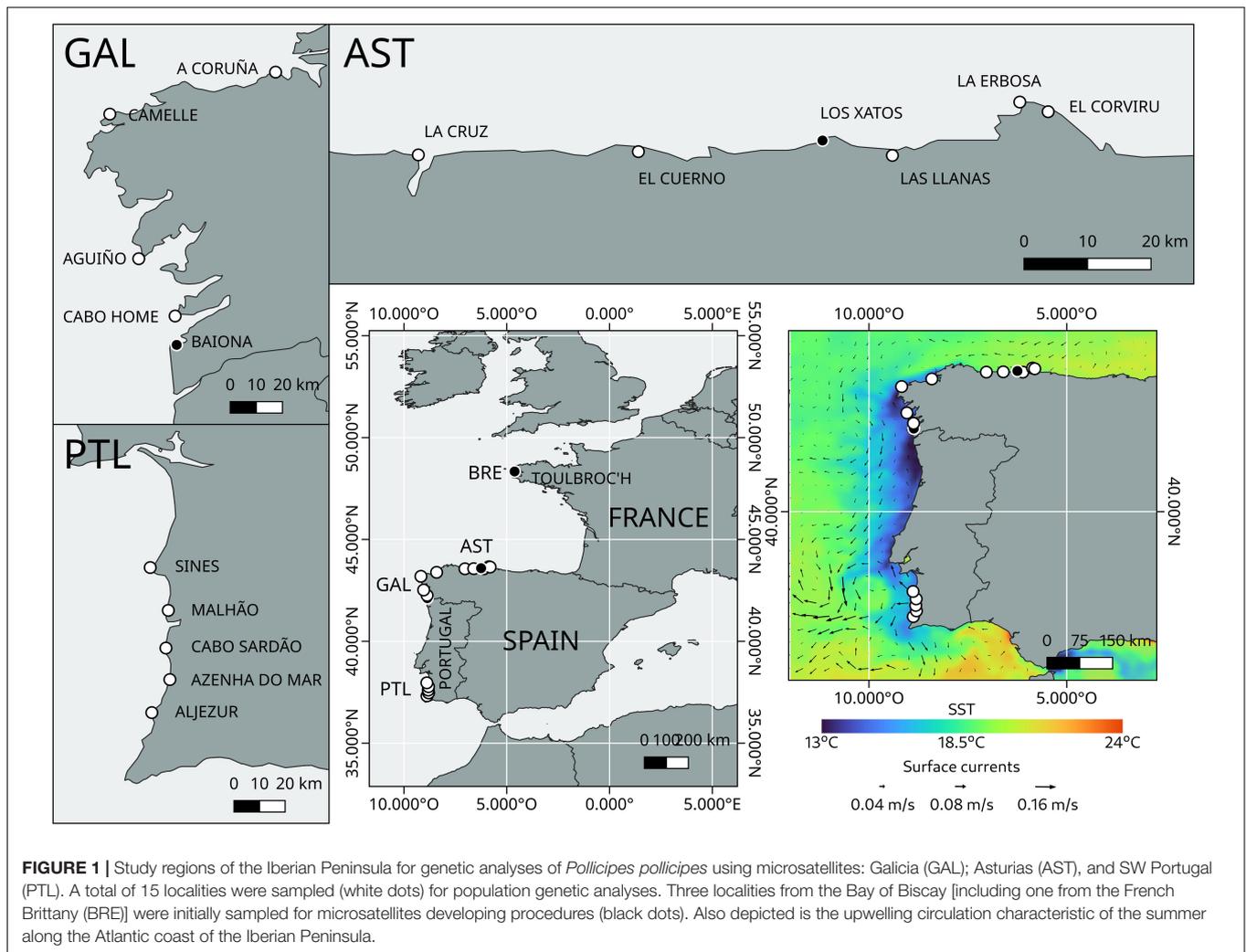
are crucial (Aceves-Bueno et al., 2017; Silva et al., 2019). Dispersal, settlement, and subsequent recruitment are decisive processes in the population dynamics of marine invertebrates with planktonic larval stages, allowing the connection between remote populations and leading to meta-populations that are globally viable (Cowen and Sponaugle, 2009).

Pollicipes pollicipes larvae go through six planktotrophic nauplius stages before turning into a lecithotrophic stage, called *cypris*. According to Molares et al. (1994) and Franco et al. (2016, 2017), the pelagic larval development is finalized after 15 days to 1 month under optimal conditions in the laboratory, whereas in the natural environment, the total pelagic larval duration is estimated to last 2 months (Cruz, 2000; Macho, 2006). The presence of stalked barnacles on the shore might favor the settlement of *cyprids*, because recruitment is intense on conspecifics (e.g., Cruz et al., 2010; Fernandes et al., 2021). The dynamics of ocean circulation are recognized as important aspects in shaping connectivity patterns among marine populations (Tremblé et al., 2008). In this situation, significant effort is required to study population dynamics locally to adequately manage the resource (Molares and Freire, 2003). For *P. pollicipes*, a minimum potential passive migration distance of 600 km during the planktonic stage has been suggested (Quinteiro et al., 2007); nevertheless, reanalysis of genetic data and basic biophysical modeling point to modest dispersal distances in the range of tens of kilometers in the Asturian region (Rivera et al., 2013). At a large spatial scale, it has been suggested that *P. pollicipes* displays a metapopulation structure, where disconnected adult populations share a common larval pool (the n-islands model hypothesis) (Molares and Freire, 2003). However, the metapopulation structure has not yet been addressed. Alternatively, species with long larval dispersal potential, such as *P. pollicipes*, may exhibit surprising patterns of spatial and temporal genetic structure. CGP (Johnson and Black, 1982; Hedgecock and Pudovkin, 2011; Eldon et al., 2016) has been consistently reported in marine species that broadcast larvae at a scale where genetic variation should be efficiently homogenized by gene flow via larval dispersal. Eldon et al. (2016) reviewed and discussed how selection, sweepstakes reproductive success, collective dispersal, and temporal shifts in local population dynamics may play a crucial role in generating such unexpected patterns. Moreover, Pineda et al. (2006) reported the existence of “recruitment windows” in a close barnacle species (*Semibalanus balanoides*), in which after a recruitment period of approximately 3 months, only recruits able to settle in just a couple of weeks survive after settlement and mature into adults. In spite of its interest for the management of this species, the processes that shape the genetic structure of *P. pollicipes* in the Atlantic Ocean have not been studied.

Genetic markers are a powerful tool for fisheries management because they present an array of very useful applications: they can address the correct identification of species, delimit distinct fish stocks (Borrell et al., 2012; Papa et al., 2020), assess relatedness levels within populations (Veliz et al., 2006; Plough et al., 2014), expose population connectivity (Pascual et al., 2017; Muñoz-Ramírez et al., 2020), estimate larval dispersal (Van Wyngaarden et al., 2017) and larval diversity

(Chen et al., 2013; Wong et al., 2014; Alshari et al., 2021) or the source-sink dynamics within the population structure (Pineda et al., 2007; Brault et al., 2013; Lindegren et al., 2014). Genetic data, however, integrate information on the past demographic history of populations and are not always easily applicable for the present-days management for marine species with high fecundity and dispersal capabilities (Gagnaire et al., 2015). Estimating some of the population parameters that are crucial for stock management imposes the need to develop numerous highly polymorphic markers. These will help to discriminate between past and present-day processes that shape populations of species with highly effective population sizes (e.g., Hongjamrassilp et al., 2020). Despite the economic relevance of the *P. pollicipes* fishery, only a few articles have been published on the genetics of the stalked barnacles, most of which are based on mitochondrial markers (Quinteiro et al., 2007; Campo et al., 2010; Rivera et al., 2013). According to Quinteiro et al. (2007), the panmixia hypothesis is rejected, and 5 population groups are established: (1) Brittany; (2) Asturias-East; (3) Galicia, Portugal and Morocco; (4) Canary Islands; and (5) Cape Verde Islands, with the latter being extremely divergent. The Cape Verde population was later considered a new species (Van Syoc et al., 2010) and described as *Pollicipes caboverdensis* (Fernandes et al., 2010). Campo et al. (2010) revealed genetic differences among populations between Brittany (France) and the rest of the species distribution range, while Rivera et al. (2013) described small-scale, asymmetric connectivity in gooseneck barnacle populations, when reanalyzing data from Campo et al. (2010) for the Cantabrian coast. Microsatellites usually display high levels of genetic variation and can detect subtle genetic differentiation among populations separated by only a few hundred kilometers (Borrell et al., 2012). Moreover, they seem to be very useful to detect parentage/familial structures, when assessing the origin of recruits (St-Onge et al., 2015; Couvray and Coupé, 2018; Dubé et al., 2020). Microsatellite markers have been previously developed and, in some cases, used to infer the population genetic structure for several closely related acorn barnacles, such as *S. balanoides* (Dufresne et al., 1999; Flight et al., 2012); *Chthamalus montagui* (Pannacciulli et al., 2005; Fontani, 2009); *Tetraclita* spp. (Dawson et al., 2010; Chen et al., 2015); *Megabalanus coccopoma* (Reigel et al., 2015); *Chelonibia testudinaria* (Ewers-Saucedo et al., 2016, 2017); *Notochthamalus scabrosus* (Barahona et al., 2019) and two stalked barnacle species: *Pollicipes elegans* (Plough and Marko, 2014) and *P. pollicipes* (Seoane-Miraz et al., 2015; Fernandes et al., in prep.). The latter appear to have shown positive results with specific cross-amplifications in the congeners *P. elegans*, *Pollicipes polymerus*, and *P. caboverdensis*.

The aim of the present study was to revisit and test the previously described genetic homogeneity of *P. pollicipes* at the scale of the Iberian Peninsula with highly polymorphic microsatellite markers. The final goal is to provide support for the design of adequate and sustainable fishery management plans, using an in-depth analysis of genetic patterns inferred from a hierarchical geographic sampling of the barnacle populations along the Iberian coastline. However, preliminary tests using published microsatellite markers have provided inconsistent and



non-reproducible PCR results in two different and independent genetic labs, necessitating the development of new highly variable genetic markers for the species *P. pollicipes* (this study).

MATERIALS AND METHODS

Study Area and Sampling

A total of 1,423 individuals from 15 different localities belonging to three Atlantic regions of the Iberian Peninsula covering the most important spots of the barnacle fishery were sampled. These three regions were SW Portugal, Galicia (NW Spain), and W Asturias (N Spain). Thus, a hierarchical sampling of populations was performed in which five distinct localities were sampled within each region (**Figure 1**). The five localities belonging to Portugal are Aljezur (AL), Azenha do Mar (AZ), Cabo Sardão (CO), Malhão (MA) and Sines (CS). The five localities belonging to Galicia are Baiona (BA), Cabo Home (CH), Aguiño (AG), Camelle (CA) and A Coruña (AC). The five localities belonging to Asturias are La Cruz (PC), El Cuerno (CU), Las Llanas (LM), La Erbosa (ER), and El Corviru (EC) (**Figure 1**). Samples were

transferred to the laboratory and frozen on the same day of collection until further individualization and labeling.

Within each of the targeted localities, one hundred individuals collected in September and October 2017 were randomly sampled within two distinct developmental cohorts according to their rostrum-carinal (RC) length [see **Figure 4** in Parada et al. (2013)] (50 adults of commercial size greater than > 18 mm; 50 juveniles between 2 and 4 mm). As barnacles are usually found in groups of sessile individuals, fixed on primary rocky substrates with small juveniles attached to adult peduncles (Cruz et al., 2010), juveniles were first removed from the adults, avoiding the collection of more than one juvenile by adult and then treated secondarily. Each barnacle was put individually in a tube previously labeled and preserved in absolute ethanol at room temperature. In the laboratory, a small portion of the peduncle muscle was dissected from each individual for genomic DNA extraction. In the case of adults, special care was taken to dissect the tissue from the inner part of the peduncle to avoid possible contamination by attached post-larvae (*cyprids*) and juveniles.

To characterize the typical upwelling circulation during the stalked barnacle larval season in summer/autumn 2017 along the

coasts of northern and western Iberia, sea surface temperature (SST) along with modeled sea surface currents datasets were retrieved during the peak of meridional Ekman transport at central Portugal on 11-08-2017. Daily 4 km SST data were obtained from the all-satellites combined Copernicus' product¹. The 5-days averaged meridional and zonal components of the surface currents were obtained from the OSCAR model with a spatial resolution² of 0.33°.

Microsatellite Markers and Multiplex PCR Development

Genomic DNA from five adult individuals was extracted using the EZNA[®] Mollusk Kit (Omega Bio-Tek Inc., Norcross, GA, United States). An enriched biotin-labeled CT/GT library for dinucleotides was obtained using the methodology described by Bloor et al. (2001) and Sotelo et al. (2007), where DNA was digested with *Hae*III (NEB). Digestions were run in 1.5% agarose gels stained with ethidium bromide. Fragments between 400 and 800 bp were excised from gels and purified using a QIAquick Gel Extraction Kit (Qiagen). Fragments were ligated to a double-stranded adaptor using ligase (NEB) and enriched by PCR using oligoA. Purified PCR products were denatured and incubated with 200 pmol of 5' biotinylated (CT)₁₂ and (GT)₁₂ probes (Invitrogen) attached to streptavidin-coated magnetic beads (Streptavidin MagneSphere Paramagnetic Particles, Promega). Hybridization was carried out in 6 SSC for 30 min at 60°C in a thermocycler. Specific fragments were recovered after washing the bead suspension with solutions progressively desalted at 60°C, and subsequently amplified using Oligo A. A DNA library was prepared using an Ion Plus Fragment Library Kit (Thermo Fisher Scientific, Austin, TX, United States) according to the manufacturer's protocol. Next-generation Ion Torrent sequencing of the library was conducted using the Ion Torrent platform on an Ion PGM System (Life Technologies, Foster City, CA, United States) using Ion PGM 400 sequencing reagents and Ion 318v2 chips following the manufacturer's instructions at the University of Vigo Central Services (CACTI). Quality control procedures and filtering of the resulting reads were afforded using PRINSEQ software (Schmieder and Edwards, 2011). Tag Sequence Check and Sequence Duplication routines were used to trim adapters and eliminate duplicates. Sequences shorter than 100 bp with a mean quality Phred score lower than 20 were removed. Tandem Repeats Finder (Benson, 1999) was used with all the parameters by default for locating and displaying tandem repetitions in DNA sequences. Forward and reverse primers were designed for effective microsatellite amplifications using FastPCR 6.5 software following Kalendar et al.'s (2009) recommendations. Finally, primers were proposed and tested by individual PCR on 30 individuals of *P. pollicipes* from 3 distinct geographic populations: 10 individuals from Baiona (Galicia, Spain), 10 individuals from Los Xatos (Asturias, Spain) and 10 individuals from Toulbroc'h (Brittany, France) (**Figure 1**). PCR tests were equally subdivided between different research laboratories with

41 primers tested per laboratory at the University of Vigo, the University of Oviedo and the Roscoff Marine Station. In this way, microsatellite markers were calibrated between geographic regions and the three institutes involved in the project.

Individual PCRs were conducted in a 20 µL total volume with Green GoTaq[®] Flexi Buffer (1×) (Promega Corporation, Madison, WI, United States), MgCl₂ (2.5 mM), dNTPs (0.5 mM), 0.2 µM of each primer, 0.1 U of GoTaq[®] G2 Flexi Polymerase (Promega Corporation, Madison, WI, United States), and 50 ng of DNA in sterile distilled water. The PCR program included an initial 5 min denaturation step at 95°C, 35 cycles of denaturation at 95°C for 30 s, annealing at 57°C for 30 s and elongation at 72°C for 30 s. PCR products were visualized using electrophoresis on a 2% agarose gel stained with SimplySafe[™] (EURx, Gdańsk, Poland). Primer pairs without amplification, leading to a multiband pattern or a band size differing from its expected size, were discarded. Twelve microsatellite loci were amplified reliably and arranged in three multiplex PCRs (M1, M2, and M3 with four microsatellite markers per multiplex each) using Multiplex Manager 1.2 software (Holleley and Geerts, 2009) according to the dye colors and expected amplicon sizes. In addition to the twelve microsatellite markers retained with this screening, eight microsatellite markers previously developed in a parallel study (Fernandes et al., in prep) were tested, calibrated, and added in two supplementary multiplexes (M4 and M5) following the previously detailed methodology. This process resulted in a total of 5 multiplex PCRs. Forward primers were labeled using fluorescent dyes: 6-FAM[™], NED[™], VIC[™], and PET[®] (Applied Biosystems, Foster City, CA, United States) (**Table 1**). PCR products were sequenced at the Genomer platform of the Roscoff Marine Station and at Servicios Científico-Técnicos of the University of Oviedo. Allele sizes were manually scored using GeneMapper v.4.0 (Applied Biosystems, Foster City, CA, United States).

Multiplex PCR and Microsatellite Genotyping

As explained above, all adult DNA was extracted with the EZNA[®] Mollusk Kit (Omega Bio-Tek Inc., Norcross, GA, United States). Juvenile DNA was extracted using the Chelex[®] 100 (Bio-Rad Laboratories Inc., Hercules, CA, United States) method (Estoup et al., 1996). PCRs were carried out following a unidirectional workflow that started in a pre-PCR room to prepare PCR plates. Amplification by PCR and processing of the subsequent PCR products always took place in a post-PCR area to avoid any possible contamination. M1, M2, and M3 multiplex PCRs were conducted using the QIAGEN Multiplex PCR Kit (QIAGEN Inc., Venlo, Netherlands) in a final reaction volume of 13 µL with the following components: 1× QIAGEN Multiplex PCR Master Mix, 1× Q-Solution, 50 ng of DNA template and 0.2–0.5 µM of each primer (**Table 1**). PCR conditions consisted of an initial denaturation step at 95°C for 15 min, followed by 40 cycles at 94°C for 30 s, an annealing temperature of 60°C (M1 and M2) or 64°C (M3) for 1:30 min and 72°C for 1 min, with a final extension at 60°C for 30 min. M4 and M5 multiplex PCRs were incorporated and tested in a later stage and they were conducted

¹<https://cds.climate.copernicus.eu/cdsapp#!/dataset/satellite-sea-surface-temperature?tab=overview>

²https://coastwatch.pfeg.noaa.gov/erddap/griddap/jplOscar_LonPM180.html

TABLE 1 | Overall microsatellites information based on multiplex PCRs typifying *P. pollicipes* populations coming from 15 different localities of the Iberian Peninsula.

M	Locus/Genbank accession number	Dye	Repeat motif	C _F	T _A	Primer sequence (5'–3')	ASR	N	k	A _R (n (26))	HO	H _E	F _{IS} (p-value)	F _{ST} (p-value)	F _{ST} ENA	F _{IT}	B
M1	RF12 MW443104	6-FAM	CGCA	0.4	60°C	F: ATTGGATACCCCGTCTAGCTGA R: GTGCTAAGCTCGCCTTATCA	131 – 221	1397	30	12.929	0.685	0.889	0.231 (0.0001)	0.003 (0.0001)	0.002	0.234	0.110*
	OV100 MW443110	VIC	AC	0.2		F: AACGATCCACAAGCATGCAACACG R: CATAATTGCAAAATTAAGCCGGTG	177 – 323	1408	59	23.881	0.862	0.954	0.096 (0.0001)	0.006 (0.0001)	0.005	0.101	0.047
	RF17 MW443105	NED	CGTG	0.2		F: GGCGTTGGTCACCACCTGA R: AGTTAATCTGCGTGTCCAGGAT	135 – 239	1393	12	3.877	0.582	0.602	0.032 (0.0544)	0.001 (0.0652)	0.001	0.033	0.013
	OV113 MW443112	PET	GT	0.4		F: GTGGACTACATGTCCCACTGC R: GATTCTCTGCAACTCAGCGAT	107 – 245	1396	62	23.202	0.505	0.944	0.466 (0.0001)	0.002 (0.1068)	0.002	0.467	0.226*
	VG49 MW443106	6-FAM	TGAG	0.4	60°C	F: AGGTAATCGTCTGATAGTCAGCTCGC R: TGTGGACACGCATGTGTGCTGGC	331 – 439	1389	37	15.707	0.893	0.921	0.030 (0.0001)	0.000 (0.8637)	0.000	0.029	0.013
M2	OV89 MW443109	VIC	CA	0.2		F: CACCTTTTGTGCTCCCAATGGA R: GACTAACACCAGCTGTCCGT	127 – 185	1416	13	5.358	0.277	0.404	0.312 (0.0001)	0.011 (0.1251)	0.013	0.320	0.092*
	VG55 MW443107	NED	CA	0.2		F: GCAACTATCAGCGCTTGACCAT R: AGGGGAATCCTAATACCGTCGT	161 – 209	1419	18	8.512	0.557	0.600	0.070 (0.0001)	0.003 (0.3934)	0.001	0.073	0.027
	OV122 MW443114	PET	CACG	0.2		F: GACGCCATATAGCCTCAGCA R: GTCAAAAAGTGTGCCACAGAA	111 – 169	1417	25	12.072	0.713	0.773	0.077 (0.0001)	–0.001 (0.2484)	0.000	0.077	0.033
	OV121 MW443113	6-FAM	TG	0.2	64°C	F: GATCCGGTCTGTCAGACAC R: TGCTATCACTTGGCACCGTC	95 – 155	1405	29	13.489	0.704	0.888	0.209 (0.0001)	0.003 (0.0115)	0.003	0.211	0.097*
	OV81 MW443108	VIC	GA	0.2		F: GGCTGTGGAGCATTAGACGT R: CCAATGTGGTAGCATCGTTACC	341 – 423	1356	42	20.294	0.850	0.945	0.096 (0.0001)	0.002 (0.2331)	0.002	0.097	0.047
M3	OV103 MW443111	NED	ATGT	0.5		F: CACGTGTGCCGATTTGTA R: GGCAGAAATAGCCACGCTC	199 – 296	1340	19	7.749	0.308	0.544	0.437 (0.0001)	0.000 (0.0151)	0.004	0.437	0.154*
	RF03 MW443103	PET	TG	0.2		F: TCTTGATTGTGGACCCATGTT R: GGACTAACTCGTCTGCACC	207 – 367	1260	63	12.882	0.392	0.792	0.494 (0.0001)	0.010 (0.0148)	0.006	0.500	0.224*
	Ppol_01 MZ576446	6-FAM	CTGT	0.06	60°C/55°C	F: GTGGGTCTTCCCTGTCAAAC R: GATCGTATCAGCACGAAGCTC	210 – 254	1356	11	3.801	0.602	0.598	–0.012 (0.7583)	–0.003 (0.9429)	–0.002	–0.015	–0.006
	Ppol_03 MZ576448	NED	CACG	0.06		F: GTTGTGTATCCCAGGCTTGC R: GATATTTGGCAGCCATAGCC	86 – 142	1381	16	8.185	0.446	0.607	0.270 (0.0001)	0.000 (0.5606)	0.001	0.270	0.101*
	Ppol_05 MZ576450	PET	GCGT	0.06		F: CGCGCACGTGTGATTTAAAC R: ATCTTCGCGTTGCTGAC	166 – 194	1372	11	5.690	0.559	0.559	0.002 (0.4446)	–0.001 (0.8485)	–0.001	0.002	0.000
M4	Ppol_09 MZ576453	VIC	TAG	0.06		F: CAAAACACCGTATGACGTTTAC R: ACCCGTACTACTGCTTTTACCG	146 – 247	1344	34	21.147	0.905	0.947	0.042 (0.0001)	–0.001 (0.6503)	–0.001	0.041	0.020

(Continued)

TABLE 1 | (Continued)

M	Locus/Genbank accession number	Dye	Repeat motif	C _F	T _A	Primer sequence (5'–3')	ASR	N	k	A _R (n = 26)	H _O	H _E	F _{IS} (p-value)	F _{ST} (p-value)	F _{IT}	B
M5	Pp01_08 MZ576452	NED	CGCA	0.1	60°C/55°C	F: TTCTGACCCGTTAAGCTTGC R: AACTGCACCACCAATTCTCC	156 – 276	1366	51	24.716	0.898	0.960	0.063 (0.0001)	0.000 (0.6328)	0.063	0.030
	Pp01_02 MZ576447	6-FAM	GTCT	0.1		F: CGTTGCATTCTATGCCTATC R: CGCTGACCCGACAAAGTTAC	176 – 232	1371	16	9.136	0.761	0.793	0.034 (0.0041)	–0.001 (0.1719)	0.033	0.014
	Pp01_04 MZ576449	PET	CACG	0.13		F: TGCACAAATCAAGATGCACAG R: TCTCTCCAGCCGTCCTTG	102 – 178	1165	24	13.196	0.366	0.893	0.594 (0.0001)	0.001 (0.1494)	0.594	0.280*
	Pp01_07 MZ576451	VIC	(TAC) ₈ (TGC)(TAC) ₇	0.06		F: CCACTCAGACATTACACCAC R: GAGCATCGGGCTTCAGGAC	104 – 155	1382	10	5.815	0.683	0.673	–0.016 (0.8298)	0.000 (0.2741)	–0.017	–0.007
						Average	1366.650	29.100	12.582	0.627	0.764	0.176	0.002	0.002	0.178	0.076

M, multiplex. C_F, PCR final concentration. T_A, annealing temperature. ASR, allele size range in bp. N, sample sizes. k, number of alleles per locus. A_R, allelic richness for the minimum possible number of diploid individuals per sample (n = 26). H_O, observed heterozygosity. H_E, expected heterozygosity. Weir and Cockerham (1984) F statistics: F_{IS} (*p < 0.05 evaluated using 10,000 permutations in FSTAT software) F_{ST}, F_{ST} ENA (excluding null alleles following Chapuis and Estoup, 2007) and F_{IT}. B, brookfield 1 statistic for null allele's inferences using the Microchecker software (*q > 0.05). Bold is used to highlight marker names and averages.

using the TouchDown PCR technique (Hecker and Roux, 1996). TouchDown PCRs were conducted in a 15 µL total volume with Colorless GoTaq[®] Flexi Buffer (1×) (Promega Corporation, Madison, WI, United States), MgCl₂ (1.5 mM for M4 and 1.16 for M5), dNTPs (0.1 mM), 0.06–0.13 µM of each primer (Table 1), 0.4 U of GoTaq[®] G2 Flexi Polymerase (Promega Corporation, Madison, WI, United States), 200 ng/µL bovine serum albumin (BSA), and 5–10 ng of DNA in distilled water. The samples were initially heated at 95°C for 5 min, followed by 10 cycles consisting of 95°C for 30 s, 60°C (decreasing incrementally by 0.5°C per cycle) for 40 s, and 72°C for 40 s, followed by 25 cycles at 95°C for 30 s, 55°C for 40 s, and 72°C for 40 s, culminating in a final cycle at 72°C for 10 min. PCR results were checked on a 2% agarose gel. For each multiplex amplification, 2 µL of reaction product (diluted 1/40 with Milli-Q water for M1, M2, and M3) was mixed with 9.5 µL of Hi-Di formamide (Applied Biosystems, Foster City, CA, United States) and 0.50 µL of SM594 molecular weight marker (Mauger et al., 2012). The mixture was heated at 94°C for 5 min, immediately chilled on ice for 2 min, loaded in an ABI Prism[®] 3130XL automatic sequencer (Applied Biosystems, Foster City, CA, United States) of 16 capillaries using POP-7 polymer and run at 60°C, 15 kV, 1,200 s using the sequencing platform Plateforme Genomer (Station Biologique de Roscoff). To ensure that the allele spread calibration held between the set of samples analyzed, controls were included in each plate to be genotyped as reference genotypes. Each genotype was then scored after analyzing the amplification products with Genemapper 4.0 (Applied Biosystems, Foster City, CA, United States).

Population Genetic Analyses

The allele frequencies, number of alleles per locus (*k*), observed heterozygosity (*H_O*) and unbiased expected heterozygosity (*H_E*) were calculated with GENETIX 4.05 (Belkhir et al., 2004). Moreover, possible genotyping errors and null allele frequency estimation were conducted using MICRO-CHECKER 2.2.3 (Van Oosterhout et al., 2004) and FreeNa (Chapuis and Estoup, 2007) with a number of replicates fixed to 10,000. Moreover, to explore the influence of null alleles on data we assessed *F_{IS}* and *F_{ST}* correlation, *F_{IS}* and the number of missing data (putative null homozygotes) correlation and estimated the StrdErr*F_{IS}* and StrdErr*F_{ST}* values following the De Meeùs (2018) and Manangwa et al. (2019) [but see Waples (2018)]. The significance of correlations was tested with a unilateral ($\rho > 0$) Spearman's rank correlation test with Rcmdr package (Fox, 2005, 2007) for R. Furthermore, for each population, the number of private alleles was calculated with GENALEX 6.5.03 (Peakall and Smouse, 2012).

Possible deviations from expected proportions in Hardy Weinberg's equilibrium and linkage disequilibrium for each locus and population were assessed using FSTAT 2.94 software (Goudet, 1995). FSTAT 2.94 software (Goudet, 1995) was used to calculate the allelic richness (*A_R*) and to determine the fixation indices (*F*-statistics) within and across populations using the method described by Weir and Cockerham (1984). Significance levels of *F_{IS}* were estimated by permutating alleles between genotypes within samples 2,000 times and adjusted following Bonferroni correction (Rice, 1989) from all tested juvenile and

adult samples. To test self-recruitment, the relatedness between individuals (R_{XY}) was estimated with the “related” package in R (Pew et al., 2015). The relative performance of seven different relatedness estimators was examined (dyadml, lynchli, lynchr, queller, ritland, trioml, and wang) through comparison of the observed values to expected values generated from a simulated sample set of 400 individuals of known relatedness [with one hundred individuals from 4 categories: parent-offspring ($R_{XY} = 0.500$), full-sib ($R_{XY} = 0.500$), half-sib ($R_{XY} = 0.250$), and unrelated pairs ($R_{XY} = 0.000$)]. The results showed that the dyadic likelihood relatedness estimator (dyadml) provided the most consistent estimates through all possible levels of kinship; therefore, it was performed with 500 iterations. The bottleneck hypothesis was tested using the software BOTTLENECK 1.2.02 (Piry et al., 1999) under the two-phased model of mutation (TPM), taking into account 90% single stepwise mutations with a variance of 12.

Comparisons between regions and between cohorts (adults and juveniles) were conducted using a two-sided statistical analysis included in the FSTAT software for several statistics [A_R , H_O , H_E , F_{IS} , F_{ST} , relatedness (R), and corrected relatedness]. In addition, F_{ST} values were estimated using FreeNA, which estimates unbiased F_{ST} following the ENA method (Chapuis and Estoup, 2007). The F_{ST} values and associated p -values between cohorts and within and between regions were also calculated using FSTAT 2.94 (Goudet, 1995) to test for the regional and local structure. To assess the significance levels of F_{ST} , multilocus genotypes were permuted 2,000 times between pairs of samples, and the significance threshold was obtained by applying a false discovery rate (FDR) over samples (Benjamini and Hochberg, 1995). Partial Mantel tests to estimate the correlation between genetic and geographical distance were performed with FSTAT 2.94 (Goudet, 1995) using the INA correction method for the chord distance (Cavalli-Sforza and Edwards, 1967) (D_{CSE}) provided by FreeNA (Chapuis and Estoup, 2007) and combining a \ln transformation of Haversine geographic distances following Séré et al. (2017) and Rousset's $\theta/(1-\theta)$ and a log transformation of Haversine geographic distances with 10,000 permutations (Rousset, 1997).

The software BayeScan v2.1. (Foll and Gaggiotti, 2008) was used to identify candidate loci deviating from neutral expectations from genetic data using differences in allele frequencies between populations. Twenty pilot runs of 5,000 iterations each, followed by an additional burn-in of 50,000 iterations and then 5,000 samplings with a thinning interval of 10, were conducted. To correct for multiple testing, the program computes q -values based on the posterior probability for each locus. Loci with α -values significantly > 0 and q -values < 0.05 were defined as “outliers” –, i.e., loci putatively under directional selection. Loci with α -values significantly < 0 were considered putatively under balancing selection. The remaining loci were classified as neutral.

An analysis of molecular variance (AMOVA) implemented in Arlequin 3.5.1.3 (Excoffier et al., 2005) to partition genetic variation across nested levels, regions and sites within regions was used. For the AMOVA, the number of different alleles was used as a measure of genetic variation (F_{ST} -like option in Arlequin), and

10 000 permutations were used to test for statistical significance. Moreover, the “adeget” package in R was used to estimate the genetic differentiation and visualize individual clustering with principal component analysis (DAPC, Jombart, 2008; Jolliffe, 2011) among adults and juveniles from each of the three regions separately and both among adults and among juveniles for all three regions pooled together. A neighbor-joining (NJ) tree based on the pairwise Nei's genetic distance D_A (Nei and Takezaki, 1983) for all microsatellites and localities (15 localities; adults and juveniles grouped together) and then adding temporal cohorts as independent samples (i.e., 15 localities and 2 cohorts, 30 samples) was constructed with the software POPTREEW (Takezaki et al., 2014) using 10 000 bootstraps and visualized in The Interactive Tree of Life³ (Letunic and Bork, 2019). Finally, STRUCTURE 2.3.4 (Pritchard et al., 2000) was also run to explore the population structure with Bayesian clustering. STRUCTURE was run using the 15 localities and 30 samples using admixture (Gilbert et al., 2012; Novembre, 2016) and also using adults and juveniles taken separately from the three regions (Portugal, Galicia and Asturias) in the same conditions to explore putative genetic units. The settings used were an admixture model from $K = 1$ to $K = 30$ in 20 runs following Evanno et al. (2005) and (Gilbert et al., 2012). Assignment clusters were made with burn-in periods of 20,000 and 200,000 Markov chain Monte Carlo repetitions. The most likely value of K was chosen using the delta K statistic (Evanno et al., 2005) using STRUCTURE HARVESTER software (Earl and VonHoldt, 2012), and visualization and grouping of the individual STRUCTURE runs was performed using CLUMPAK (Kopelman et al., 2015).

RESULTS

The typical upwelling circulation during the stalked barnacle larval season in summer/autumn 2017 along the coasts of northern and western Iberia, sea surface temperature (SST) along with modeled sea surface currents datasets revealed that the SST patterns showed strong onshore advection of cold waters (13–15°C) on the Galician and Portuguese shelves with upwelling filaments extending further offshore especially at the upwelling centers of Fisterra, A Guarda, and Cape da Roca (Figure 1). Slightly onshore cooling indicative of upwelling was also observed along the western Cantabrian coast. Westward and southward currents in the order of few cm/s off the Cantabrian and Atlantic shores, respectively, clearly pointed to upwelling circulation (Figure 1). These flows are weaker close to the coast probably due to friction with the coastal boundary layer. Off southern/central Portugal in between Cape da Roca and Cape San Vicente, an anticlockwise cyclonic eddy was apparent with strong southward currents (> 10 cm/s) along its western side. The dynamic structure of this feature matched SST patterns remarkably well, with a warm core (18°C) surrounded by colder upwelled waters (14°C).

The microsatellites markers development process produced libraries with a total amount of 42,860 reads showing a mean

³<https://itol.embl.de>

sequence length of 91.61 ± 103.29 bp (minimum length: 25 bp – maximum length: 517 bp) and a mean GC content of $63.66 \pm 18.90\%$. A total of 10 781 sequences with a mean sequence length of 244.48 ± 97.61 bp, a length range of 418 bp and a mean GC content of $50.30 \pm 5.61\%$, resulted after quality control procedures. A total of 1,140 sequences containing di, tri, tetra, and pentanucleotides were selected after locating and displaying tandem repetitions in DNA sequences. Finally, 123 pairs of primers were proposed and tested in three different research laboratories (University of Vigo, University of Oviedo, and the Roscoff Marine Station). A new set of twelve microsatellite loci currently arranged into three multiplex PCRs (M1, M2, and M3) was developed for the stalked barnacle *P. pollicipes* in this work (Genbank accession numbers: MW443103–MW443114). Moreover, eight previously developed microsatellite loci by Fernandes et al. (in prep) were also tested and included in another two multiplexes (M4 and M5, Genbank accession numbers: MZ576446–MZ576456). This procedure resulted in a total of 5 multiplex PCRs (Table 1) leading to scorable and reproducible genotypes for all 20 microsatellite loci. None of these loci showed evidence of linkage disequilibrium between alleles ($p > 0.05$). These loci were highly polymorphic and exhibited approximately 15% private alleles ($n = 87$) only present at one locality (Tables 1, 2).

The number of alleles per locus (k) varied greatly from 10 to 63 between loci, with an average of 29.10, and yielded an average (min–max) allelic richness (A_R) of 12.170 (11.074–12.918) per locality. The observed and expected heterozygosities across loci ranged from $H_O = 0.277$ (M2; OV89) and $H_E = 0.404$ (M2; OV89) to $H_O = 0.905$ (M4; Ppol_09) and $H_E = 0.960$ (M5; Ppol_08), with observed and expected multilocus mean heterozygosities equal to 0.627 (0.561–0.667) and 0.764 (0.746–0.785), respectively (Table 1). All markers and all populations showed significant deviations from Hardy–Weinberg equilibrium (mean $F_{IS} = 0.179$) due to heterozygote deficiencies (Table 2). When testing these markers for null alleles with MICRO-CHECKER 2.2.3. (Van Oosterhout et al., 2004) and FreeNA (Chapuis and Estoup, 2007), we found that heterozygote deficiency could be due to null alleles for at least 8 highly polymorphic loci: RF12 locus (Brookfield 1 Statistic = 0.110); OV113 locus ($B = 0.226$); OV89 locus ($B = 0.092$); OV121 locus ($B = 0.097$); OV103 locus ($B = 0.154$); RF03 locus ($B = 0.224$); Ppol_03 locus ($B = 0.101$); and Ppol_04 locus ($B = 0.280$) (Table 1). The correlation between F_{IS} and F_{ST} appeared to be significant (Spearman's $\rho = 0.606$, p -value = 0.005). However, F_{IS} and the number of missing data (putative null homozygotes) were not correlated (Spearman's $\rho = 0.098$, p -value = 0.3402) and the standard error for F_{IS} ($\text{StrdErr}F_{IS} = 0.044$) was higher than for F_{ST} ($\text{StrdErr}F_{ST} = 0.001$). The mean overall F_{ST} value for the 20 microsatellites was $F_{ST} = 0.002$ ($P = 0.0001$), and three loci clearly showed higher F_{ST} values [OV100 ($F_{ST} = 0.006$), OV89 ($F_{ST} = 0.011$), and RF03 ($F_{ST} = 0.010$)] (Table 1).

The comparative analysis for levels of genetic variation between regions [Portugal (PTL), Galicia (GAL), and Asturias (AST)] revealed no significant differences for expected heterozygosities (H_S) (Table 2). Slight differences in genetic diversities were, however, observed depending on the population

parameter estimated. Galicia showed the highest values for allelic richness and observed and expected heterozygosity in adults and juvenile populations (Table 2). Portugal showed the highest number of private alleles (mean $A_{PPTL} = 3.3$), which was mainly attributable to adults (mean $A_P = 4.4$) (Table 2). Significant differences in allelic richness were also observed for juveniles (A_{RPTL} : 11.569; A_{RGAL} : 12.418; A_{RAST} : 11.900; $p < 0.01$) and in observed heterozygosity for adults (H_{OPTL} : 0.640; H_{OGAL} : 0.656; H_OAST : 0.615; $p < 0.05$) and juveniles (H_{OPTL} : 0.605; H_{OGAL} : 0.648; H_OAST : 0.601; $p < 0.01$) (Table 2). This phenomenon was especially obvious in Portuguese samples, where the average number of private alleles decreased by 50% (A_{PAD} : 4.4 to A_{PJV} : 2.2) (Table 2). In this later region, significant differences were found between adults and juveniles both in terms of allele richness (A_{RAD} : 12.318; A_{RJV} : 11.569; $p < 0.01$) and observed heterozygosity (H_{OAD} : 0.640; H_{OJV} : 0.605; $p < 0.05$) or expected heterozygosity (H_{SAD} : 0.773; H_{SJV} : 0.757; $p < 0.01$) (Table 2). Globally, juveniles were also more related to each other ($R_{XY} = 0.067$) than their adult ($R_{XY} = 0.058$) counterparts, as indicated by relatedness analyses. Juveniles from Portugal (R_{XY} value = 0.073, $p < 0.002$) and Asturias ($R_{XY} = 0.070$, $p < 0.002$) were much more related than expected from panmixia. Bottleneck software showed that none of the 30 samples tested (15 localities \times 2 cohorts) exhibited a significant excess of predicted heterozygotes under the TPM model and could not be considered to have experienced a recent genetic bottleneck (Table 2). When the bottleneck hypothesis was tested with all juveniles and adults together (15 samples) and at the regional scale (15 samples grouped in 3 regions, for 2 cohorts), the statistics remained non-significant (Table 2).

According to the overall F_{ST} , there was no significant genetic differentiation of adults between and within regions (Figure 2A). Only 20 out of the 75 possible pairwise F_{ST} values between adult samples from different regions (25.3%) showed p -values lower than the 0.05 cutoff value, and these critical values were more often encountered between Galicia and Asturias (12/25 = 48%) (Figure 2A). However, no p -values remained significant after FDR correction (Figure 2A). In contrast, the overall F_{ST} statistics estimated for the juveniles between and within regions indicated notable regional and spatial structuring (Figure 2B). Pairwise F_{ST} estimated between juvenile samples from Galicia and Portugal (12/25 = 48% before and 6/25 = 24% after FDR) and between Galicia and Asturias (13/25 = 52% before and 6/25 = 24% after FDR) confirmed this trend and showed clear regional structuring (Figure 2B). Asturias and Portugal were, however, less differentiated from each other, with fewer significant pairwise F_{ST} values (3/25 = 12% before and 2/25 = 8% after FDR) (Figure 2B). Some spatial structuring within regions was detected for juveniles using pairwise F_{ST} analyses but only in the case of Portugal (2/15 = 13% before and 1/15 = 6% after FDR) (Figure 2B). The pairwise F_{ST} analyses between adults and juveniles within regions revealed that only 8 out of the 75 possible comparisons (11%) had p -values lower than the 0.05 cutoff threshold for Portugal and Asturias (but not in Galicia), which, however, did not remain significant after FDR correction (Figures 2C–E).

TABLE 2 | Genetic variability of *P. pollicipes* populations (adults vs. juveniles) coming from 15 distinct localities along the Atlantic Iberian Peninsula coastline.

Country	Region	Locality	Coordinates	Sampling date	Life stage	Code	N	N _A	A _P	A _R	H _O	H _E	F _{IS}	R _{XY}	TPM p
Portugal	SW Portugal	Aljezur	37.32141, -8.879	5-9/10/2017	Adult (AD)	PTL_AL_AD	48	14.800	1	12.455	0.631	0.764	0.176*	0.045	0.959
Portugal	SW Portugal	Azenha do Mar	37.46747, -8.79988	11/10/2017	Adult (AD)	PTL_AZ_AD	48	13.800	4	11.964	0.624	0.762	0.184*	0.048	0.861
Portugal	SW Portugal	Cabo Sardão	37.6068, -8.81716	21/09/2017	Adult (AD)	PTL_CO_AD	45	14.650	6	12.152	0.627	0.770	0.186*	0.058	0.983
Portugal	SW Portugal	Malhão	37.77324, -8.8068	6/10/2017	Adult (AD)	PTL_MA_AD	48	15.000	8	12.514	0.667	0.783	0.150*	0.043	0.968
Portugal	SW Portugal	Sines	37.96286, -8.88591	23/09/2017	Adult (AD)	PTL_CS_AD	48	14.800	3	12.508	0.650	0.779	0.168*	0.045	0.980
<i>Average PTL_AD</i>							<i>47.400</i>	<i>14.610</i>	<i>4.4</i>	<i>12.318</i>	<i>0.640</i>	<i>0.772</i>	<i>0.173*</i>	<i>0.057</i>	<i>0.999</i>
Portugal	SW Portugal	Aljezur		5-9/10/2017	Juvenile (JV)	PTL_AL_JV	45	12.350	1	11.228	0.602	0.750	0.201*	0.085	0.982
Portugal	SW Portugal	Azenha do Mar		11/10/2017	Juvenile (JV)	PTL_AZ_JV	48	13.750	3	11.868	0.617	0.772	0.203*	0.055	0.995
Portugal	SW Portugal	Cabo Sardão		21/09/2017	Juvenile (JV)	PTL_CO_JV	46	13.450	2	11.513	0.606	0.749	0.193*	0.055	0.980
Portugal	SW Portugal	Malhão		6/10/2017	Juvenile (JV)	PTL_MA_JV	46	13.600	3	11.776	0.596	0.747	0.205*	0.058	0.980
Portugal	SW Portugal	Sines		23/09/2017	Juvenile (JV)	PTL_CS_JV	48	13.500	2	11.459	0.602	0.756	0.205*	0.070	0.985
<i>Average PTL_JV</i>							<i>46.600</i>	<i>13.330</i>	<i>2.2</i>	<i>11.569</i>	<i>0.605</i>	<i>0.755</i>	<i>0.201*</i>	<i>0.073**</i>	<i>0.999</i>
Average PTL							47.000	13.970	3.3	11.944	0.622	0.763	0.187*		1.000
Spain	Galicia	Baiona	42.11847, -8.86672	09/10/2017	Adult (AD)	GAL_BA_AD	48	14.700	1	12.273	0.663	0.776	0.147*	0.049	0.988
Spain	Galicia	Cabo Home	42.25244, -8.87372	05/10/2017	Adult (AD)	GAL_CH_AD	48	14.600	1	12.253	0.655	0.765	0.145*	0.047	0.997
Spain	Galicia	Aguiño	42.51861, -9.04111	09/10/2017	Adult (AD)	GAL_AG_AD	48	15.700	5	12.874	0.662	0.764	0.136*	0.042	0.998
Spain	Galicia	Camelle	43.19, -9.1743	09/10/2017	Adult (AD)	GAL_CA_AD	48	14.850	4	12.433	0.635	0.758	0.165*	0.049	0.993
Spain	Galicia	A Coruña	43.38502, -8.41133	06/10/2017	Adult (AD)	GAL_AC_AD	48	15.650	5	12.918	0.666	0.776	0.143*	0.042	0.959
<i>Average GAL_AD</i>							<i>48.000</i>	<i>15.100</i>	<i>3.2</i>	<i>12.550</i>	<i>0.656</i>	<i>0.768</i>	<i>0.147*</i>	<i>0.055</i>	<i>0.999</i>
Spain	Galicia	Baiona		09/10/2017	Juvenile (JV)	GAL_BA_JV	48	14.250	1	12.025	0.628	0.753	0.146*	0.051	0.996
Spain	Galicia	Cabo Home		05/10/2017	Juvenile (JV)	GAL_CH_JV	48	14.950	1	12.383	0.659	0.770	0.168*	0.046	0.983
Spain	Galicia	Aguiño		09/10/2017	Juvenile (JV)	GAL_AG_JV	42	14.850	3	12.669	0.647	0.759	0.149*	0.043	0.990
Spain	Galicia	Camelle		09/10/2017	Juvenile (JV)	GAL_CA_JV	44	13.700	0	12.372	0.643	0.785	0.183*	0.059	0.938
Spain	Galicia	A Coruña		06/10/2017	Juvenile (JV)	GAL_AC_JV	48	15.350	5	12.640	0.663	0.768	0.137*	0.046	0.997
<i>Average GAL_JV</i>							<i>46.000</i>	<i>14.620</i>	<i>2</i>	<i>12.418</i>	<i>0.648</i>	<i>0.767</i>	<i>0.157*</i>	<i>0.058</i>	<i>0.999</i>
Average GAL							47.000	14.860	2.6	12.484	0.652	0.767	0.152*		0.999
Spain	Asturias	La Cruz	43.55691, -7.02893	20/09/2017	Adult (AD)	AST_PC_AD	50	15.050	7	12.511	0.607	0.754	0.197*	0.043	0.987
Spain	Asturias	El Cuerno	43.56585, -6.60318	19/09/2017	Adult (AD)	AST_CU_AD	48	15.250	4	12.568	0.610	0.757	0.196*	0.047	0.998
Spain	Asturias	Las Llanas	43.56212, -6.10582	20/09/2017	Adult (AD)	AST_LM_AD	49	14.500	2	12.103	0.625	0.757	0.176*	0.050	0.955
Spain	Asturias	La Erbosa	43.6631, -5.86407	20/09/2017	Adult (AD)	AST_ER_AD	50	14.900	4	12.319	0.620	0.760	0.186*	0.047	0.968
Spain	Asturias	El Corviru	43.64414, -5.80895	20/09/2017	Adult (AD)	AST_EC_AD	44	13.500	1	11.834	0.614	0.762	0.196*	0.056	0.971
<i>Average AST_AD</i>							<i>48.200</i>	<i>14.640</i>	<i>3</i>	<i>12.267</i>	<i>0.615</i>	<i>0.758</i>	<i>0.190*</i>	<i>0.059</i>	<i>0.999</i>

(Continued)

TABLE 2 | (Continued)

Country	Region	Locality	Coordinates	Sampling date	Life stage	Code	N	N _A	A _P	A _R	H _O	H _E	F _{IS}	R _{XY}	TPM p
Spain	Asturias	La Cruz		20/09/2017	Juvenile (JV)	AST_PC_JV	47	14.750	3	12.505	0.606	0.768	0.213*	0.055	0.993
Spain	Asturias	El Cuerno		19/09/2017	Juvenile (JV)	AST_CU_JV	50	14.750	3	12.067	0.616	0.760	0.191*	0.053	0.995
Spain	Asturias	Las Llanas		20/09/2017	Juvenile (JV)	AST_LM_JV	50	14.200	3	11.804	0.561	0.753	0.257*	0.060	0.995
Spain	Asturias	La Erbosa		20/09/2017	Juvenile (JV)	AST_ER_JV	50	14.450	3	12.052	0.611	0.752	0.190*	0.053	0.998
Spain	Asturias	El Convitu		20/09/2017	Juvenile (JV)	AST_EC_JV	45	12.150	1	11.074	0.613	0.752	0.186*	0.088	0.978
						Average AST_JV	48.400	14.060	2.6	11.900	0.601	0.757	0.207*	0.070**	0.999
						Average AST	48.300	14.350	2.8	12.084	0.608	0.76	0.199*		0.999
						Average Iberian Peninsula	47.433	14.393	2.9	12.170	0.627	0.76	0.179*		

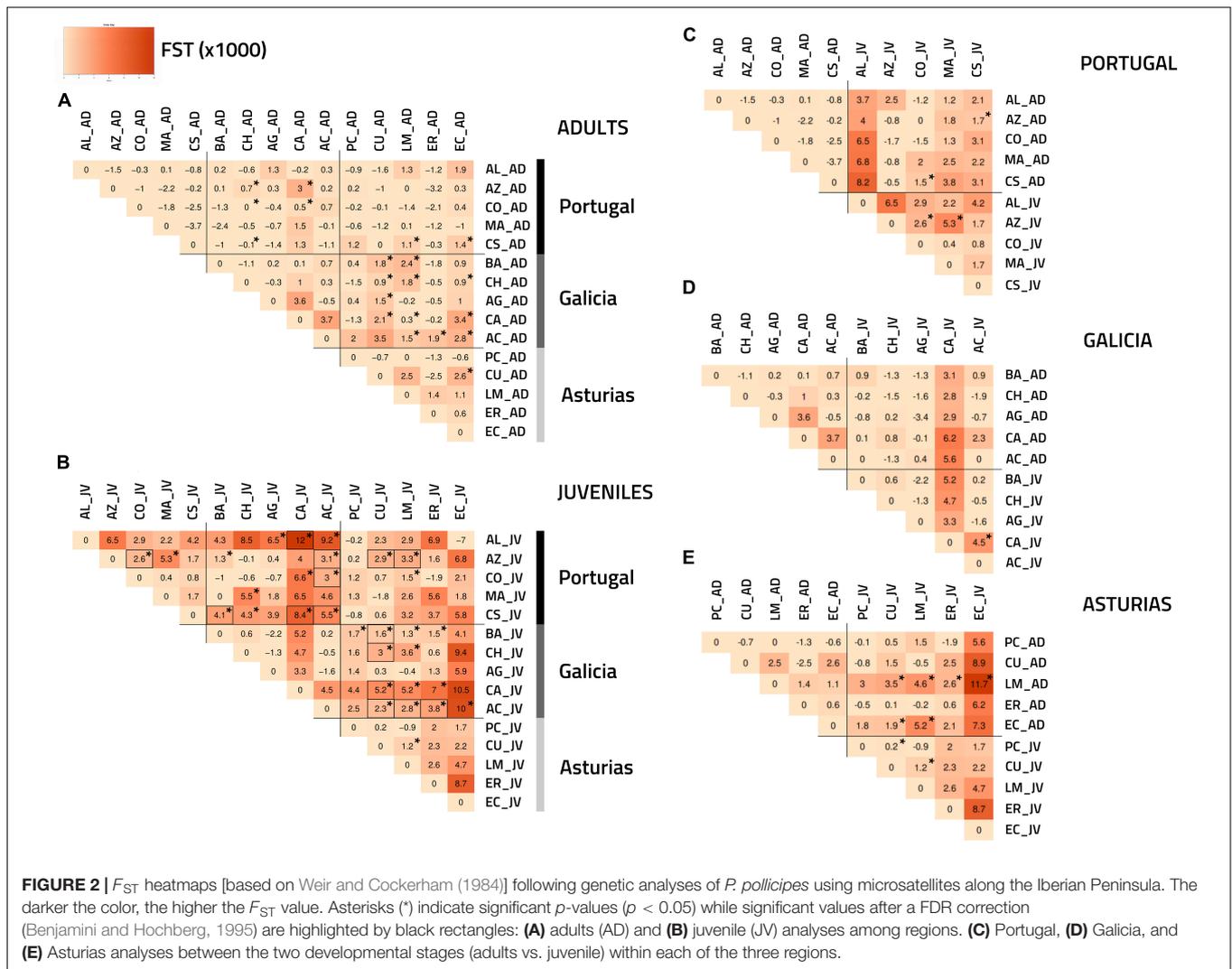
N, sample sizes. N_A, mean number of alleles by locus. A_P, private alleles. A_R, allelic richness for the minimum possible number of diploid individuals per sample. H_O, observed heterozygosity. H_E, expected heterozygosity. F_{IS}, degree of departure from expected Hardy–Weinberg proportions within samples. R_{XY}, average relatedness within each of the specified groups. TPM p, Wilcoxon test probability under TPM method. *P < 0.05, **p < 0.01. Bold is used to highlight marker names and averages. Italics are used as a way to highlight the average in different stages of development.

The analysis conducted with BayeScan v2.1 for outlier detection resulted in no loci under selection or biased by species admixture and hybridization which have the same expectations in terms of outliers; the twenty loci showed signatures of balanced or purifying selection with negative alpha values. The results of the partial Mantel tests indicated no correlation between genetic and geographic distances, with R² = 1.61 and p-value = 0.1916 for adults and R² = 3.16 and p-value = 0.0698 for juveniles using the INA correction method for D_{CSE}, and R² = 0.06 and p-value = 0.8002 for adults and R² = 0.22 and p-value = 0.6241 for juveniles using the Rousset method. The population structure was therefore closer to an n-island model than a stepping stone model, and the pairwise F_{ST} between adjacent sites often exceeded those obtained between geographically distant locations.

The DAPC analyses and the hierarchical analysis of molecular variance (AMOVA) using φ_{ST} statistics showed no significant genetic differentiation of adults among and within regions [AD: φ_{CT(among)} = 0.00032, p > 0.05; φ_{SC(within)} = 0.00013, p > 0.05] (Figure 3A). However, a globally significant genetic differentiation for juveniles among and within regions was found [JUV: φ_{CT(among)} = 0.00093, p < 0.05; φ_{SC(within)} = 0.00217 p < 0.001] (Figure 3B). The analyses also revealed significant genetic heterogeneity between *P. pollicipes* generations in Portugal [φ_{CT(among)} = 0.00127, p < 0.01] and Asturias [φ_{CT(among)} = 0.00120, p < 0.01], but not in Galicia (Figures 3C–E). The neighbor-joining tree using adults and juveniles grouped together by localities clearly separated Galicia with high bootstrapping values (i.e., 90%), where Camelle and Baiona fall apart from the rest of the Galician localities, after which two other different Portuguese and Asturian clades appeared (Figure 4A). When all populations (15 localities and 2 cohorts, 30 samples) were analyzed, the neighbor-joining tree again showed Galicia samples falling apart and becoming heterogeneous, whereas the Portuguese and Asturian samples were mixed together, with aggregations showing low bootstrap values (Figure 4B). The STRUCTURE runs using admixture suggested 3 genetic clusters [Evanno’s k = 3, L(K) = −117589.9100] when all populations (30 samples) were analyzed (Figure 5A). The STRUCTURE results also indicated the co-occurrence of 2 genetic clusters [Evanno’s k = 2, L(K) = −60822.8150] for adults (Figure 5B) and 3 clusters [Evanno’s k = 3, L(K) = −56241.9750] for juveniles when run separately (Figure 5C).

DISCUSSION

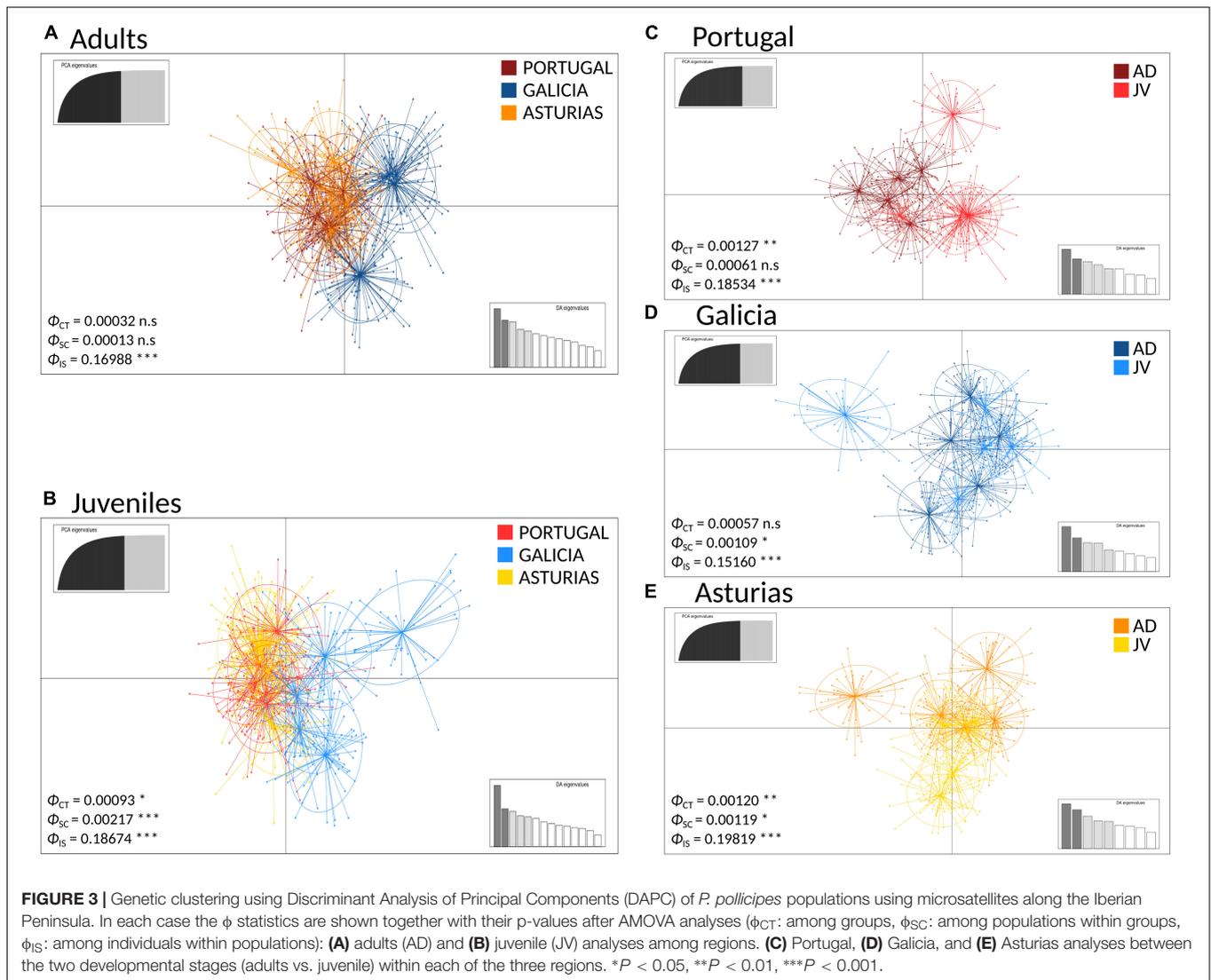
The analyses using twenty new microsatellite loci aimed to define, more accurately, the temporal and spatial evolution of the genetic structure of stalked barnacle *P. pollicipes*. This species is highly appreciated in the Spanish and Portuguese markets, and its management must be based on reliable scientific data. Previous studies have suggested that larval dispersal driven by ocean currents, in particular, the Iberian Poleward Current have played a crucial role in determining the population structure, and two distinct regional configurations



have been established using mitochondrial DNA for *P. pollicipes* within its distribution range along the northeastern Atlantic. Quinteiro et al. (2007) suggested that *P. pollicipes* is structured into four genetically differentiated groups: French populations, eastern Asturian populations, Galician-Portuguese populations, and Canarian populations. Conversely, Campo et al. (2010) suggested the presence of only two groups, among which French populations were highlighted as a peculiar and differentiated genetic entity, as a result of a past population fragmentation during Pleistocene glacial/interglacial periods. Regardless, later studies based on estimates of population migration rates have suggested that barnacle population connectivity occurred on a small scale and in an asymmetric manner in the Cantabrian coast (Rivera et al., 2013). Information based on highly variable nuclear molecular markers can provide crucial information on both population connectivity and stock renewal for this species within the Iberian Peninsula. This information is needed for the delimitation of conservation/management units in this fishery and the improvement of the management plans and the performance of TURFs.

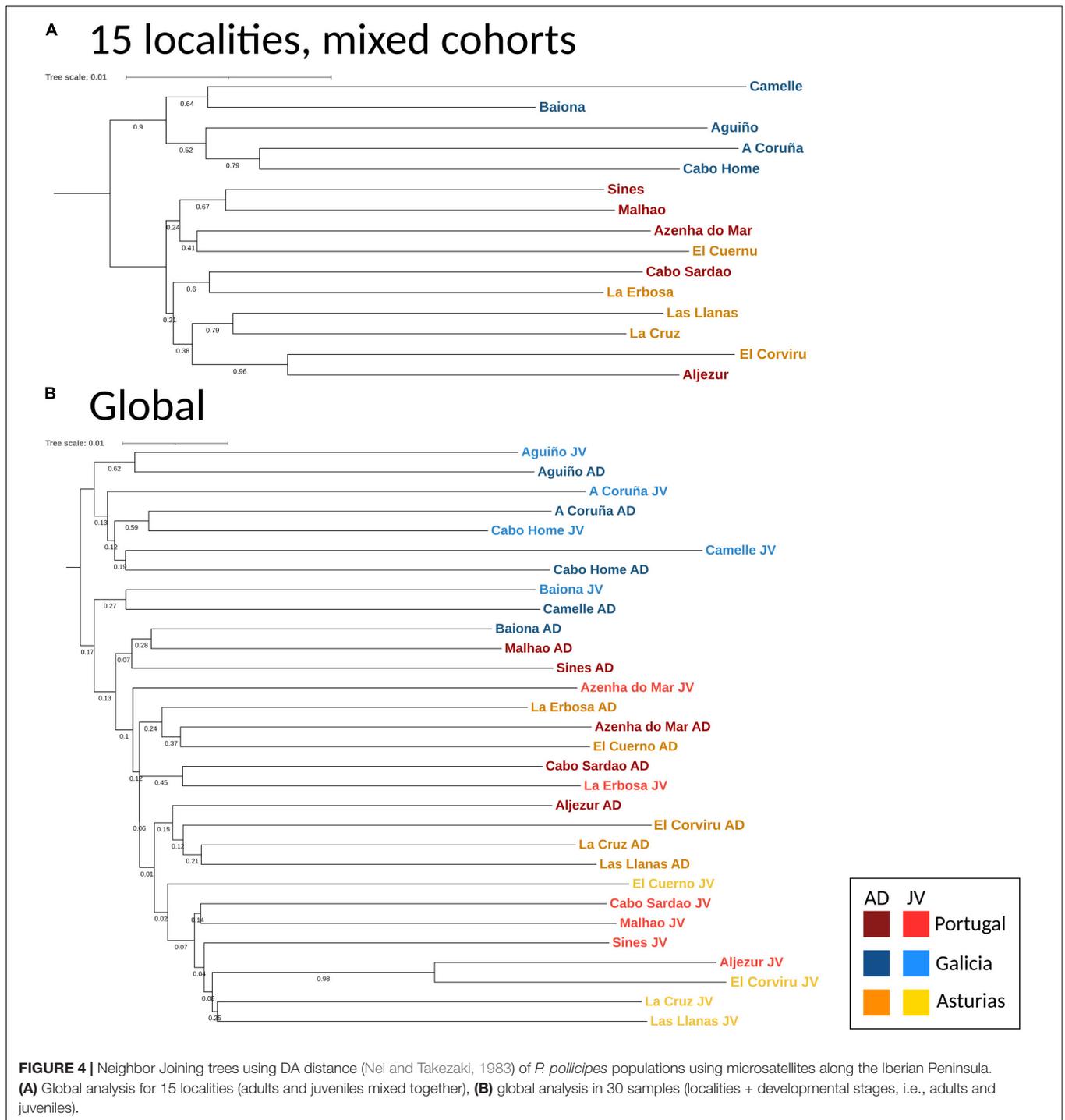
Genetic diversity contributes to the ability of a species to respond to environmental changes, and highly fecund species that release high numbers of small eggs into the environment (the so-called *r*-strategists) are much more polymorphic than species that produce a small number of relatively large offspring and provide parental care (called *K*-strategists) (Ellegren and Galtier, 2016). Recent studies in *S. balanoides* have confirmed that barnacles harbor high levels of genome-wide genetic variation (Nunez et al., 2021). The level of genetic diversity of *P. pollicipes* found in this work was particularly high. We observed higher levels of genetic variation in *P. pollicipes* than in other barnacles of the same genus, such as *P. elegans* (Plough and Marko, 2014). Our results showed that Galicia exhibited the highest values for allelic richness and observed and expected heterozygosity in adult and juvenile populations. Conversely, newly settled cohorts (juvenile) had a lower genetic diversity than adults across all the studied regions, particularly when examining both allelic richness and private alleles.

The main principal assumption of the Hardy-Weinberg principle is that the sample comes from a single, randomly mating



population where perturbing forces (such as selection, genetic drift, mutation, and migration) are absent or balanced (Waples, 2014). All loci and populations showed significant deviations from Hardy-Weinberg equilibrium in this work due to, more or less pronounced, heterozygote deficiencies. This phenomenon could be the consequence of local admixtures of genetically differentiated populations (Wahlund effect), assortative mating, inbreeding, selection (Palumbi, 2003) and finally null alleles. The presence of null alleles has been reported in the vast majority of previous microsatellite studies in barnacles (Dufresne et al., 1999; Pannacciulli et al., 2005; Plough and Marko, 2014; Reigel et al., 2015; Abreu et al., 2016; Ewers-Saucedo et al., 2016) as well as in other marine invertebrate species such as clams (Borrell et al., 2014; Chiesa et al., 2016; Rico et al., 2017), octopus (Greatorex et al., 2000; De Luca et al., 2016), sea urchins (McCartney et al., 2004; Carlon and Lippé, 2007), jellyfish (Aglieri et al., 2014), and polychetes (Jolly et al., 2003, 2009, 2014). The presence of null alleles is an inherent trait of microsatellite loci and is caused by mutations in the primer sequences, leading to the lack of

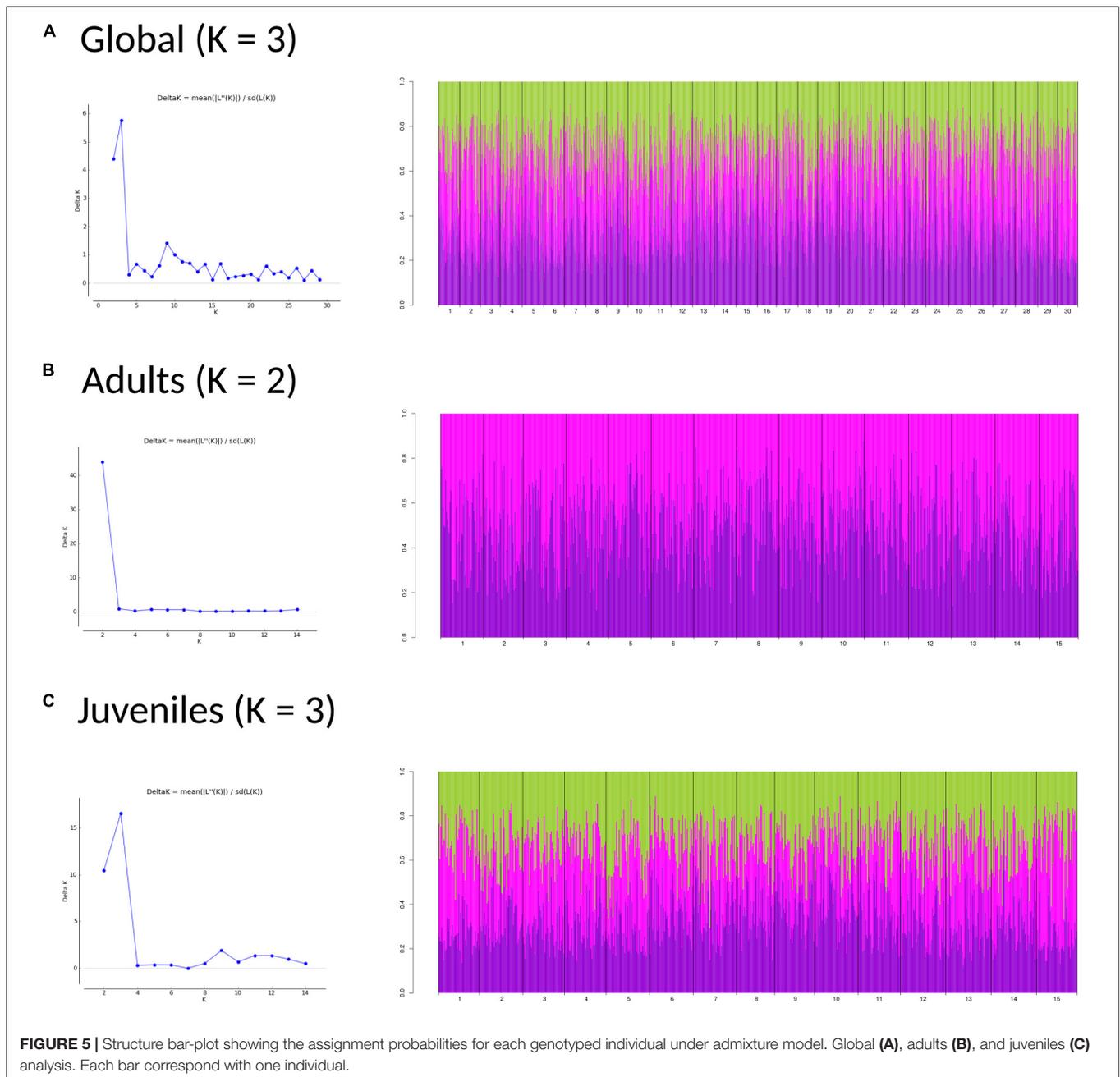
amplification and the dropout of alleles (Selkoe and Toonen, 2006). In addition, an increase of the null allele frequency would be expected with the increase of alleles per locus and previous studies have indicated that the presence of null alleles seems to be particularly common in populations with high effective population sizes (Chapuis and Estoup, 2007). Although the presence of null alleles leads to an overestimation of both F_{ST} and genetic distances in cases of significant population differentiation (Chapuis and Estoup, 2007), our results showed no differences worth considering for both the F_{ST} or F_{ST} ENA values. It has been argued that the conservative approach of discarding loci deviating from Hardy-Weinberg equilibrium expectations could rob us of our most informative markers, weakening our ability to interpret biological phenomena (Dharmarajan et al., 2013). Moreover, De Meeùs (2018) stated that in case of null alleles, F_{IS} and F_{ST} are augmented and a positive correlation is expected between F_{IS} and F_{ST} as is expected a positive correlation between F_{IS} and the number of missing data (putative null homozygotes), and $StrdErrF_{IS}$ being at least twice $StrdErrF_{ST}$. If



such correlations do not exist and if $\text{StrdErr}F_{IS} > \text{StrdErr}F_{ST}$, then a Wahlund effect better explains the data (De Meeüs, 2018; Manangwa et al., 2019). Waples (2018) had also argued about this and simulated 10% of null alleles suggesting that caution in interpreting $F_{IS} \times F_{ST}$ correlations under conditions where null alleles might be common it is indeed necessary and more efforts will be needed for a comprehensive evaluation of this complex topic. In this work, panmixia is rarely met for

any locus (**Table 1**), we found positive $F_{IS} \times F_{ST}$ correlations, $\text{StrdErr}F_{IS} > \text{StrdErr}F_{ST}$ and we did not find positive correlations between F_{IS} and the number of missing data (putative null homozygotes) pointing out to the fact that, even when null alleles are present, other biological factors also play a fundamental role to explain significant heterozygote deficits in our data.

Heterozygote deficiencies can as well be the result of local admixtures of genetically differentiated cohorts in populations, or



due to sweepstake reproductive effort (Waples, 1998; Hedgecock and Pudovkin, 2011). Growth of individuals in *P. pollicipes* populations is highly heterogeneous (Cruz et al., 2010; Jacinto et al., 2015), so that individuals of similar size may differ greatly in age. Our adult samples likely contained a mixture of cohorts from different reproductive and dispersal events, potentially leading to significant departures from Hardy-Weinberg equilibrium, locally. Genetic heterogeneity of cohorts can potentially blur the genetic signal in adults and may decrease the genetic differences over time, given that the geographic origin of migrants might change throughout the breeding/dispersal seasons depending on prevailing local hydrodynamics during these periods. However,

it should be noted that a special care was taken in this work to sample only one cohort of juveniles with a specific size (2–4 mm RC). If the deficiencies of heterozygotes were due the superimposition of cohorts, juveniles should not show such deficiencies. This was clearly not the case here as our results demonstrated that juvenile mean F_{IS} values were higher than those for adults in all the three regions (Table 2). It has been stated that the surf zone and its surrounding nearshore waters are known to act as selective barriers to the onshore transport of many larval invertebrates on the local scale (Porri et al., 2006; Rilov et al., 2008). The permeability of such barrier is modulated by small scale topography that generates retentive oceanographic

features like coastal fronts (Pineda, 1999; Shanks et al., 2003). In fact, the larvae of *P. pollicipes* and other barnacles have been shown to accumulate in great numbers at internal waves and river plume fronts off the Cantabrian coast only at some specific locations (Weidberg et al., 2014; Höfer et al., 2017). In this topic the available evidence are indeed scarce, however, genetic data seems to confirm it.

Pollicipes pollicipes has asynchronous broods during the reproductive season which usually occurs from March to September (e.g., Cardoso and Yule, 1995; Cruz and Hawkins, 1998; Pavón, 2003; Macho, 2006), where several batches of larvae are produced, and potentially lead to the co-occurrence of different settlement events. Juveniles sampled in this study might, however, come from one to few settlement events. Despite the possibility of several discrete settlement events, post larval mortality might favor one specific batch of survivors, and in the end, the 2–4 mm RC juveniles might become more related than what would have been expected from the mixing of several reproductive events. Pineda et al. (2006) found that recruitment to the reproductive stage of acorn barnacles (*S. balanoides*) was composed of survivors that settled in a recruitment window. The recruitment window (to reproduction in the case of the Pineda study, to 2–4 mm in our study) might be narrower than the recruitment season. If by some reason these survivors correspond to larvae that are genetically more related, then a pattern of genetic differentiation could occur among recruits. The concept of a “recruitment window” proposed by Pineda et al. (2006) matches quite well with Hedgecock’s “sweepstakes-chance matching hypothesis” also known as “sweepstakes reproductive success hypothesis,” which is based in part on the observation of reduced genetic variability in young-of-the-year populations relative to adult populations. This reduced genetic variability among recruits suggests that the surviving young of the year are the products of spawning by only a small fraction of the adult population, which, according to Hedgecock’s hypothesis, happened to produce their offspring at a place and time that was suitable for survival (Hedgecock, 1994). Moreover, barnacles rear embryos in bags before hatching and there is also the possibility that the larval release is only efficient for a small proportion of the reproductive adults depending on the local hydrodynamics. In this work, we found evidence indicative of reproductive sweepstakes in adult and juvenile samples. Although globally, the relatedness coefficients estimated for *P. pollicipes* were in the same range as those from other studies previously conducted with barnacles (Veliz et al., 2006; Plough et al., 2014), they were significantly slightly greater in juveniles (i.e., Asturias and Portugal) compared with adults. Juveniles were significantly more related to each other than expected from random mixing despite their larval entrainment in the water column during the planktonic phase.

Unexpected genetic differentiation in marine invertebrates can occur due to three neutral processes: sweepstake reproductive success (Hedgecock, 1994), collective dispersal (Johnson et al., 1993; Li and Hedgecock, 1998) and asynchronous population dynamics (Eldon et al., 2016), but also selective processes during the settlement process. According to the Hedgecock’s “sweepstakes-chance matching hypothesis” or

selective sweepstakes (Hedgecock, 1994), only a fortunate combination (hence sweepstakes) of reproductive traits and oceanographic conditions would allow an individual to complete the long mobile phase from spawning and fertilization through larval survival to recruitment back to the adult habitat. In a highly fecund species and a locally heterogeneous oceanographic setting, this would involve strong selection favoring just a handful of genotypes at each locality, leading to a local-scale genetic mosaic but a relatively large-scale uniformity. Post-larval settlement selection under different environmental conditions has been argued to create CGP in coastal areas of temperate regions over a mosaic of contrasting habitats able to impose a strong differential selective sieve or a target for habitat choice in larvae (Eldon et al., 2016). We detected significant genetic differentiation for juveniles among and within regions (but not for adults), together with significant genetic heterogeneity between *P. pollicipes* generations. However, we did not find evidence of such selective processes for the assayed microsatellites. There seemed to be a genome-wide pattern that was more parsimoniously explained by neutral processes such as sweepstake reproductive success, which may greatly reduce the genetic diversity of a given cohort while provoking unexpected heterozygote deficiencies, as seen previously, by mimicking local bottlenecks (genetic diversity drawn from a small subset of parents). In addition to this phenomenon, genetic differentiation may persist in recruits when dispersal is limited in space, when larvae from different cohorts do not mix completely during dispersal (collective dispersal), or when local conditions may promote self-recruitment (Eldon et al., 2016).

The genetic data obtained in this work, after applying dissimilar approaches (F and ϕ_{ST} statistics, Discriminant Principal Component and Bayesian analyses), pointed all out to the existence of significant genetic heterogeneity in the Iberian coasts rejecting previous findings using mitochondrial DNA. The results herein highlighted Galicia as a peculiar genetic entity possibly representing a superimposition of two distinct metapopulations or potentially an old refuge for the most northern populations from France (not sampled in this study). Among Galician northernmost populations, Camelle (CA) and A Coruña (AC) are also the most differentiated from Portugal and Asturias and may have a specific demographic history. The sampled *P. pollicipes* populations are located along the Atlantic Iberian coast, whose hydrodynamic patterns have been well studied. The western peninsular coast (SW Portugal and Galicia) is characterized by a complex current system subject to strong seasonality and mesoscale variability, showing inverse patterns between summer and winter in the upper layers of the shelf and slope. During spring and summer (coinciding with *P. pollicipes* breeding season), northerly winds along the coast are dominant, causing coastal upwelling and producing a southward current on the surface and a northward undercurrent on the slope. In the Cantabrian Sea (Asturias) the surface currents flow generally eastward in winter and early spring and shift westward in late spring and summer following the wind force with intermittent summer upwelling events west of Cape Peñas (ICES, 2021). Different aspects of the oceanographic circulation in Iberia were reviewed by Relvas et al. (2007).

Casabella et al. (2014) divided the upwelling affecting the coasts of Galicia into three regions: Rías Baixas, Fisterra-Bares and Cantabrian. These two locations (CA and AC) would be found in the Fisterra-Bares region, which is the region with a greater intensity of upwelling, although the period favorable for upwelling is longer in the region of Rias Baixas (sampled here i.e., Baiona). Galician juveniles showed clear genetical differences from those of Portugal and Asturias. The main explanation for this distinction is that the Biscay Bay Current, characterized by a wide gyre, can trap larvae, and thus should favor self-recruitment and perhaps local importations from the French and Cantabrian populations. This ocean circulation could also be responsible for the differentiation between juveniles from Asturian and Galicia. Previous studies on adult barnacles have found significant differences between the Asturian and Galician localities (Quinteiro et al., 2007). However, it should be noted that most of the Asturian sites sampled in this study are located to the West of Cape Peñas, while the site sampled by Quinteiro et al. (2007) was located to the East of the same cape, which has been described as a biogeographic barrier (Anadón and Niell, 1980). Rivera et al. (2013) showed that during a year of high upwelling activity (2009), the theoretical *P. pollicipes* recruitment success was 94%, with a recruitment peak predicted 56 km west of the emission point. Consistently, migration rates derived from genetic analyses showed that westward dispersal was much more likely along the Cantabrian coast, which matches the upwelling driven circulation typical of the stalked barnacle larval season in summer/autumn (Figure 1). Thus, the recurrence of upwelling may not only define the spatial scale and direction of the dispersal process but also the genetic structure of the barnacle metapopulation.

The Western Iberian upwelling system represents an important crossroad between Lusitanian and boreal temperate species (Jolly et al., 2006; Maggs et al., 2008). Upwelling/downwelling wind-driven circulation and tides are recurrent physical processes along the Atlantic Iberian coastlines and are among the most energetic phenomena that can affect near-shore circulation during the spring and summer periods when reproduction occurs and, during the summer and beginning of autumn in the case of recruitment (Queiroga et al., 2007). However, when studying a strong upwelling region in the northeastern Pacific coast, Morgan et al. (2009) observed that the larvae of most invertebrate species remain close to the shore even during strong upwelling, where high local retention and limited connectivity have been evidenced in populations of several species, such as *Petrolisthes cinctipes* (Hameed et al., 2016) or in the red rock lobster *Panulirus interruptus* (Iacchei et al., 2013). Despite this phenomenon, upwelling areas have been pointed out as probable climate change refuges for the distribution of *Fucus guiryi*, other barnacles such as *S. balanoides* and other sessile marine species (Gómez and Lunt, 2007 for a review; Hoarau et al., 2007; Provan and Bennett, 2008; Lourenço et al., 2016; Herrera, 2019). In addition, Campo et al. (2010) suggested the existence of a Pleistocene refuge area off the coast of North Africa and two additional northern glacial refuges for *P. pollicipes*, in the English Channel/Brittany region and in the northwestern Iberian Peninsula.

Previous studies have mentioned that the southern region of Portugal also represents a well-known upwelling area (Lourenço et al., 2016) with a high level of barnacle larval settlement (Queiroga et al., 2007) and recruitment (Aguión et al., in prep.). Remarkably, the number of private alleles was significantly higher in adults there when compared with those from Galicia and in Asturias. Portuguese juveniles were, however, significantly less genetically diversified and more related to each other than expected based on random mating. Moreover, Nolasco et al. (in prep.) show that connectivity matrices integrated over the period of the observations (July 2017 to July 2019) indicate high levels of larval retention. Such retention is probably caused by the recurrent eddies driven by upwelling circulation observed off southern-central Portugal in between Cape Roca and Cape San Vicente (Figure 1; Haynes et al., 1993; Batteen et al., 2000; Sánchez and Relvas, 2003; Peliz et al., 2004). These findings suggest that Portuguese populations are likely to export more migrants than they receive. As Queiroga et al. (2007) hypothesized, regular exchanges of larvae over the distance separating the southern and northern parts of Portugal are unlikely. Conversely, the Portugal Current, which shows a north- or southward direction, depending on the season, could be an important factor in promoting gene flow between our sampling locations in southern Portugal and other, unsampled, *P. pollicipes* southernmost areas such as the Canarian and North African coasts. Nevertheless, microsatellite markers have recently shown a genetic differentiation between European and African *P. pollicipes* populations (Fernandes et al., in prep.).

The correct management of marine ecosystems relies on understanding the scale and magnitude of connectivity among populations through the identification of adaptive genetic differences (Almany et al., 2009; Aceves-Bueno et al., 2017), because locally adapted populations should be considered poorly-connected, separate management units (Waples, 1998). Our results suggested that *P. pollicipes* populations in the Iberian Peninsula possibly exhibit a “CGP” structure, which extends from a few kilometers apart to as much as hundreds of kilometers apart. This phenomenon has clear consequences for the sustainable management of resources. Currently, an increasing number of small-scale fisheries have successfully implemented co-managed TURFs; a governance arrangement that enables the collaboration across diverse stakeholders, develops new knowledge and increases the capacity of the system to deal with new drivers (Rivera et al., 2014). However, the design of TURFs does not usually account for the spatial configuration of resources (Aceves-Bueno et al., 2017) due to the multi-species nature of fisheries. This mismatch between management and biological scales can compromise the sustainability of sessile stocks (Ouréns et al., 2015), like barnacles. However, a better understanding of the spatial structure and larval dynamics of the population, permits the redefinition of management units according to population boundaries. In addition to these management measures, it would be interesting to implement networks of protected areas at detailed scales to ensure that propagules are available when and where conditions are favorable for their survival (Larson and Julian, 1999; Ouréns et al., 2015).

CONCLUSION

New molecular markers have been developed in the highly valued species *P. pollicipes* and offer useful tools to provide a better fine-tuning assessment of its population dynamics along the Iberian Peninsula. *P. pollicipes* displays high genetic diversity, which is attributable to large effective population sizes representing a well-connected network of local populations. However, temporal and spatial genetic differentiation of populations over regional scales, on one hand, and a significant reduction in genetic diversity in juveniles, on the other hand, clearly indicate that patterns of exchanges together with seasonal wind-induced upwelling may induce genetic differences between settlers throughout generations. Such patterns of CGP are likely due to sweepstake reproductive success with possible collective dispersal or episodic self-recruitment events. Therefore, our *P. pollicipes* genetic dataset suggests that recruitment may be stochastic and highly dependent on climatic conditions with multiple sources of emissions. These phenomena may have strong implications in terms of management plans over the whole Iberian Peninsula with the need to protect a series of putative sources within each region. Future research should combine genetic information at broader spatiotemporal scales with larval dispersal models based on ecological and biological characteristics of *P. pollicipes*. This means, among others, mapping the complete species distribution and tracking the genetic structure of age groups over time and space. It also means applying new sequencing technologies to fully understand the dynamics of larval exchanges and the post-larval settlement of the stalked barnacle but also to better apprehend how environmental variations shape genomic variation in this species.

DATA AVAILABILITY STATEMENT

The datasets presented in this study can be found in online repositories. The names of the repository/repositories and accession number(s) can be found in the article/supplementary material.

AUTHOR CONTRIBUTIONS

MP: methodology, formal analysis, investigation, writing—original draft, and writing—review and editing. PM:

conceptualization, supervision, methodology, and writing—review and editing. MB and NW: methodology and writing—review and editing. JLA: conceptualization, supervision, project administration, funding acquisition, and writing—review and editing. AA, JA, JNF, LG-F, and KJG: writing—review and editing. JC: methodology. TC and GM: conceptualization and writing—review and editing. EG-V: conceptualization, funding acquisition, and writing—review and editing. ET and DJ: conceptualization, supervision, and writing—review and editing. YJB: conceptualization, supervision, funding acquisition, and writing—review and editing. All authors contributed to manuscript revision, read, and approved the submitted version.

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