



Extent and Characteristics of a Newly Discovered Unique Bryozoan Biogenic Reef Complex

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Using multibeam echosounding (R2Sonic, Austin, Texas) combined with underwater observation and imagery surveys, this study sought to estimate the size and describe the typology of newly discovered biogenic bryozoan reefs in Western Port (VIC, Australia). The bryozoan species forming the biogenic structure are *Triphyllozoon moniliferum*, *T. munitum* (both fenestrate forms and dominant on the reef) and *Celleporaria* (non-fenestrate form and sub-dominant). The bryozoan biogenic reefs occupy a total area of 1.74 km². Their distribution appears to be geomorphically and depth-controlled, which might limit their maximum possible extent. The Western Port bryozoan reefs are unique and globally significant because they: (1) occur in shallow water (typically 5–10 m), (2) are dominated by delicate fenestrate colonies of *T. moniliferum* and *T. munitum*, (3) have two distinct morphologies (linear and patch-like), (4) form continuous rows of reef interspersed with fine sediment, and (5) the size and vertical relief (up to approximately 1.5 m) are among the largest recorded in the world. In Western Port, these bryozoan reefs are habitat-forming and may represent sites of enhanced biodiversity, which is currently being investigated in another separate study and are likely important habitat utilised by commercially and recreationally fished species. The bryozoan biogenic reefs are a previously unrecognised feature of conservation significance in this Ramsar site. The reefs are potentially vulnerable to physical damage from recreational fishing practices, sediment smothering and scouring, algal encroachment and hydrological alteration. Recommendations for monitoring and potential management of this unique biotope are made.

Keywords: Western Port, biogenic habitat, bryozoan reefs, multibeam bathymetry, *Triphyllozoon* and *Celleporaria*

INTRODUCTION

Bryozoans are a diverse phylum of colonial, non-photosynthetic filter-feeding aquatic invertebrates, with ~5700 extant (Horowitz and Pachut, 1994) and ~15,000 fossil species (Amini et al., 2004). They are commonly referred to as “lace corals”, despite being quite unrelated to corals (phylum: Cnidaria). Marine bryozoan colonies (Class Stenolaemata) produce complex rigid calcareous skeletons (termed “hard” bryozoans), which creates structural microhabitat for a myriad of fauna (Wood and Probert, 2013), such as crustaceans, molluscs, fishes and worms and this ecological

function and service is enhanced when colonies combine to form bryozoan biogenic reefs “BBR” (Bradstock and Gordon, 1983). In contrast, “soft” marine bryozoan colonies (i.e., those that do not form a calcium carbonate matrix), can also occur locally in high abundance in the lower strata of kelp-dominated rocky reefs, however, these apparently do not form biogenic reefs. Consequently, complex habitats such as BBR are often biodiversity “hotspots” (Morgado and Tanaka, 2001; Wood et al., 2012; Wilson, 2021) compared to the surrounding habitats and provide food, attachment substrate for sessile organisms, shelter from strong currents as well as concealment from predators for both adult and larval stage organisms. Currently, there are only 54 high-density bryozoan sites known globally (Wood et al., 2012) with only three occurring in Australian waters prior to this study. Virtually nothing is known about the ecological role, biology, extent, and characteristics of the BBR of Western Port (a temperate Australian embayment).

It appears that modern BBR are confined to temperate zones, although bryozoan biogenic habitat has been noted through a latitudinal band between 13°N and 23°S (Wood et al., 2012). Bryozoan biogenic habitat tends to be confined to the lower limits of the photic zone or in shallow turbid waters where they are not out-competed by light-dependent algae, potentially allowing for larger colony sizes to be attained and diverse epifaunal communities to form (Bock et al., 2018). This light-based exclusion/competition scenario is analogous to tropical scleractinian corals and algae (McCook et al., 2001). Favourable growth conditions for bryozoans are thought to require adequate nutrients and an abundant supply of unarmoured phytoplankton that are their primary food source (James et al., 2008).

Several reports of bryozoan “reefs” actually relate to *quasi* bryozoan-encrusted rock formations [e.g., Abrolhos Shelf, Brazil (Bastos et al., 2018) and bryozoan-rich fossil bioherms (Cuffey et al., 1977; Ernst and Königshof, 2008) and sediments (e.g., Tasmanian continental shelf, James et al., 2008)]. So-called bryozoan “sands” are a feature of the southern Australian shelf sediments that are cited as Holocene in origin (James et al., 2008). Fenestrate bryozoan (i.e., net-like or reticulated skeletal structure) forms typically dominate these Holocene sands although fine branching forms can be locally abundant in deeper water (generally <80 m) where wave energy is substantially dissipated.

The bryozoan reef “thickets” on the Otago Shelf, New Zealand, are true extant bryozoan reefs, forming emergent mounds of various species (see **Table 1** for this and other notable extant sites). Wood et al. (2012) describes 50 mm as the size at which heavily calcified bryozoans become habitat-forming and the authors reported growth rates of 0.8 to 5 mm per year, indicating that the habitat-forming BBR likely establish on the scale of decades. As reported herein, epibenthic colonies in the Western Port (WP) bryozoan reefs attain sizes of some 2,000 mm in radius and 500 to 1,200 mm in vertical relief, suggesting century-scale age.

Bryozoa occur throughout most of WP as individual colonies or “clumps” (i.e., non-reef forming), often associated with circalittoral reef (i.e., below the zone of light penetration that supports algal growth) and among seagrasses, *Caulerpa* spp. beds

and growing on jetty pylons. The first indication of the potential presence of BBR comes from the historical records of the mud oyster (*Ostrea angasi*) dredge fishery that was prevalent between 1820s through to the 1920s. Historical records indicate that fishers interacted with abundant bryozoan reefs:

“Lumps of live coral were often brought up, so big that they blocked the dredge mouth. These lumps were broken up and thrown back. The coral was live and full of sealife, worms, octopus etc.” (Hannan and Bennett, 2010).

Indeed, the interaction between oysters and bryozoans is intriguing given the observations made in the present study of oysters occurring on BBR and associated biodiversity studies indicating the presence of a dense shell layer ~10–20 cm under the sediment surface in neighbouring habitats.

The general Western Port bryozoan reef area became known to commercial and recreational fishers as “The Corals”, probably named for the fragments of bryozoan skeleton that resemble coral, which is sometimes present upon the retrieval of anchors and fishing equipment. Today there is no commercial fishing in WP, but it is known as a productive recreational fishing area. Furthermore, this area has typically poor visibility, is far from shore and has strong currents making it unfavourable for diving and snorkelling activities, meaning the BBR have avoided attention until now.

The first scientific report of the presence of bryozoan structures in the area came from Blake et al. (2013) who employed towed camera transects and noted “*isolated colonies*”, comprising “*patches of low and high profile broken, and solid reef colonised by dense bryozoans and sparse sponges*”. The significance of that discovery was not appreciated until the same towed video footage was reviewed in 2016 during a biotope classification project by the Victorian Department of Environment, Land, Water and Planning “DELWP” (Department of Environment, Land, Water and Planning [DELWP], 2017). Biotope mapping in 2016 made use of multibeam bathymetry collected in 2009 which showed, at a coarse resolution, characteristic seabed textures that instigated the current study. This study sought to characterise the typology and quantify the extent of BBR, establish conservation values and provide critical information for management strategies. This was achieved by utilising high resolution multibeam echosounding (MBES) in conjunction with various ground-truthing surveys.

MATERIALS AND METHODS

Study Area and Ground-Truthing

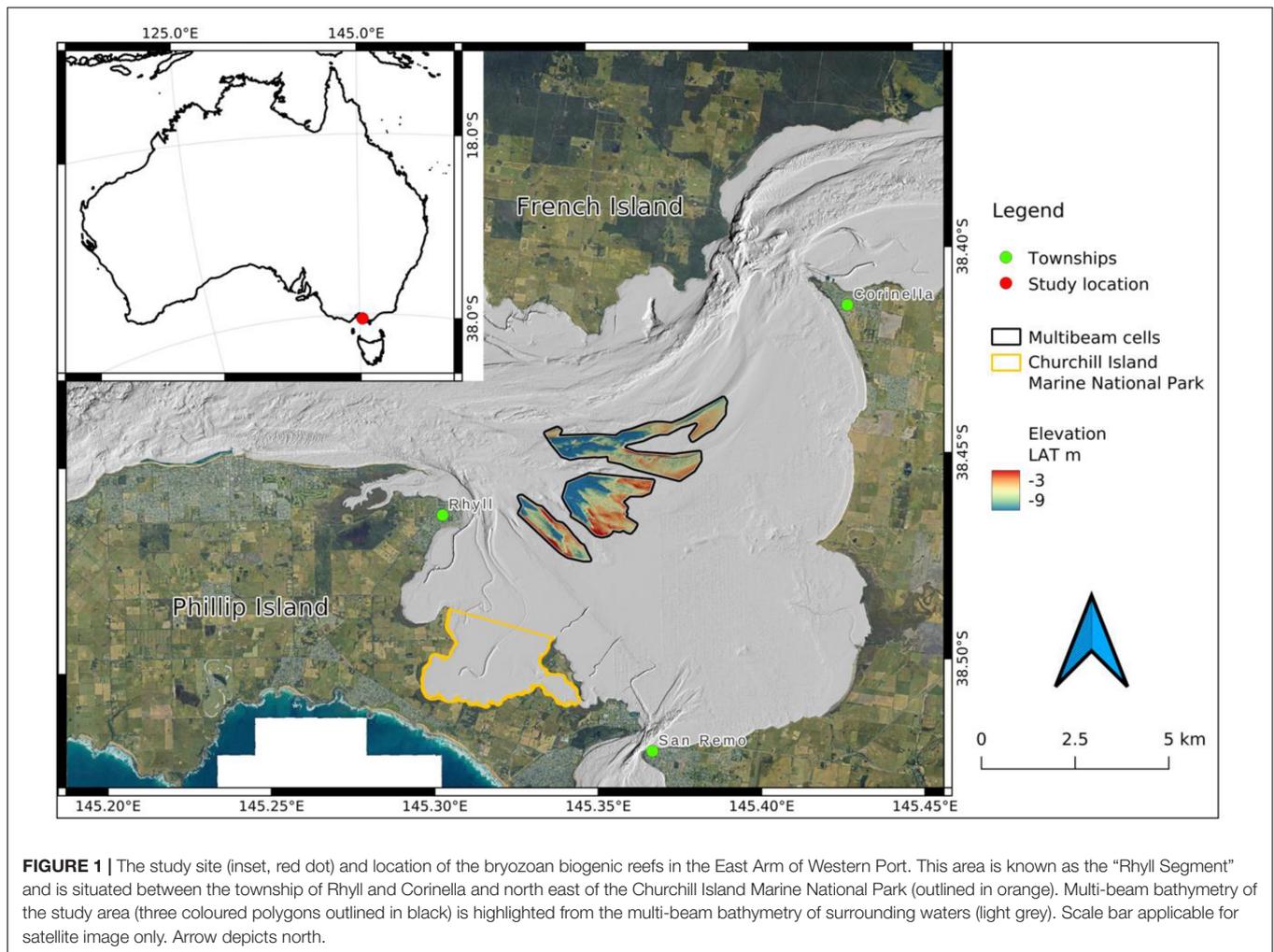
The bryozoan biogenic reef (BBR) area is situated in the Eastern Arm of Western Port (WP), VIC, Australia, in an area termed the “Rhyll Segment” (Harris et al., 1979; Jenkins et al., 2013) between French Island, Corinella and Rhyll (**Figure 1**). Potential reef areas were initially identified based on the seabed textures in a 2009 bay-wide multibeam bathymetry product (2.5 m horizontal resolution) generated by the Port of Hastings Development Authority. Image-based groundtruthing was achieved by inspection of linear seabed texture features detected in the multibeam bathymetry using a BlueROV2 Remote

TABLE 1 | Notable extant bryozoan habitats from around the world.

Location	Depth	Reef morphology	Dominant species	References
Otago Shelf, New Zealand	45–130 m	Habitat-forming bryozoans on gravels and sands, forming biogenic “thickets”	<i>Cinctipora elegans</i> , <i>Hornera robusta</i> , <i>Hippomenella vellicata</i> , <i>Adeonellopsis</i> spp., <i>Celloporina grandis</i>	Wood and Probert, 2013; Batson, 2000; Batson and Probert, 2000; Jones, 2006; Probert et al., 1979
Foveaux Strait, New Zealand	17–38 m	Biogenic reefs with ~ 20 cm relief	“ <i>Schizoporella</i> ” <i>spectabilis</i> and <i>Micropora</i> sp., <i>Penetrantia parva</i> , <i>Opaeophora lipida</i> , <i>Microporella agonistes</i> , <i>Beania elongata</i> , <i>Penetrantia irregularis</i>	Cranfield et al., 2003, 2004
Tasman Bay, New Zealand	10–35 m	<i>C. agglutinans</i> clumps up to 0.5 m in vertical height and <i>H. vellicata</i> foliaceous honeycomb structures to 0.15 m height	<i>Celleporaria agglutinans</i> , <i>Hippomenella vellicata</i>	Bradstock and Gordon, 1983
Abrolhos Shelf, South Atlantic Off Brazil	2–35 m	Erect mushroom-like reef formations.	Skeletal remains of multiple species dominating concretions	Bastos et al., 2018
Ligurian Sea, northwestern Mediterranean	11–22 m	Mounded, seabed covering reef formations	<i>Pentapora fascialis</i> (non-fenestrate species)	Cocito et al., 1998
Cap de Creus, northwestern Mediterranean	61–225 m	Bryozoans on coarse sands and gravels and canyon head	113 species of erect, encrusting frontal budding, shell epibionts, and dead skeleton	Madurell et al., 2013
Tasman Bay, New Zealand	10–35 m	Mounded, seabed covering reef formations of	<i>Celleporaria agglutinans</i> and <i>Hippomenella vellicata</i> (non-fenestrate species)	Bradstock and Gordon, 1983
Foveaux Strait, New Zealand	20–60 m	Reef “sward” aligned perpendicular to current flow	<i>Cinctipora elegans</i>	Cranfield et al., 1999
Bathurst Channel, Tasmania	5–20 m	Not reported but presumed individual colonies	Species and morphotypes not reported	Barrett et al., 2007
Lacepede Shelf, southern Australia	~30–250 m	Seafloor encrustation, branching colonies, erect colonies	<i>Adeona</i> spp., <i>Membranipora</i> spp., <i>Crisia</i> sp., <i>Cellaria pilosa</i> . >200 species present on shelf	Bone and James, 1993; Hageman et al., 1995
Coorong Lagoon, southern Australia	–	–	Intergrown bryozoa and serpulids	Bone and Wass, 1990
Great Australian Bight, southern Australia	100–240 m	Bryozoan-rich biogenic mounds. Pliocene–Pleistocene period	<i>Celleporaria</i> spp., nodular-arborescent, fenestrate, flat robust branching, encrusting, delicate branching	James et al., 2004
Bass Strait and Tasmanian shelf, southern Australia	~30–400 m	Bryozoan-rich shelf sediments.	Erect-flexible articulated zooidal and articulated forms, erect-rigid flat-robust branching and fenestrate forms, encrusting forms.	James et al., 2008
Joulters Cay, Bahamas	1–4 m	Historic bryozoan-rich reef formations now over-grown by corals. Reefs are circular, elongate, triangular or notched. plan view, to a max. of 10 m across.	<i>Cleidochasma</i> , <i>Holoporella</i> , <i>Parasmittina</i> , <i>Rhynchozoon</i> , <i>Schizoporella</i> , <i>Smittipora</i> , <i>Steginoporella</i> , and <i>Stylopoma</i>	Cuffey et al., 1977

Operated Vehicle (ROV) recording oblique high-definition video (Figure 2). Approximately 1.5 h of ROV footage was acquired in the northern cell. In addition, a drop-camera platform was used to collect imagery while station-keeping or slow drifting over seafloor features of interest and surrounding habitats. The drop camera platform comprised a Splashcam® high-definition video camera recording in an oblique orientation and a GoPro Hero4 with separate lighting collecting still images in planar orientation (see Figure 2). Further, *in situ* observations supported by underwater photography were made over 10 SCUBA dives in the BBR area (see Figure 2) during various sample collection events between December 2017 and February 2020. Collectively, the results of these groundtruthing efforts were

used in conjunction with the low-resolution bay-wide multibeam bathymetry to define the survey extents for high-resolution multibeam echosounding (MBES). BBRs were found to be comprised of three species *on species* and representative samples were collected for taxonomic confirmation by a bryozoan taxonomy expert at the *National Museum of Victoria* (Australia). Still images and video footage were also shared with bryozoan taxonomy experts at CSIRO (Australia) and National Institute of Water and Atmospheric Research (New Zealand) to confirm species identifications. For this study, the imagery was used to make qualitative assessments of the presence of BBR, its composition and characteristics and there was no quantitative analysis. In a partner study, epifauna species co-inhabiting the



reefs were scored from the imagery and these data will be reported in future publication.

Multibeam Echosounding Surveys

The vessel-based MBES surveys were completed in the northern, central and southern cells, for a total area of 6.4 km² at a resolution of 0.5 × 0.5 m (see **Figure 2**). MBES surveys were completed over two periods: 9–10 October 2019 and 11–14 February 2020. Transect direction was dictated by the shape and position of the survey area extents but was principally in east-west direction. The MBES was completed using an R2Sonic® Dual Head 2020 multibeam system (340, 400 kHz frequency). Positioning was achieved with an Applanix® POS MV Wavemaster system with the real-time kinematic (RTK) satellite position validated against the Victorian Permanent Survey Mark “Bittern PM 259”. The MBES were corrected to Lowest Astronomical Tide (LAT). Validation of the vertical data was achieved by deploying a cage to the seafloor at the survey site which had a RTK Global Navigation Satellite System (GNSS) receiver target installed at a known distance above the seabed. The MBES system captured data over the subsea target, resulting

in a Total Horizontal Uncertainty (THU) of ± 0.1 m and Total Vertical Uncertainty (TVU) of ± 0.05 m. Data processing used the Hypack® 2020 system and digital elevation models for the survey areas were produced at a horizontal resolution of 0.5 m × 0.5 m.

Reef Area Calculation

Bryozoan reefs were visible in the digital elevation models (DEMs) in each of the three cells. However, fully automated terrain analysis and habitat modelling routines failed to reliably distinguish BBR due to interference patterns introduced by seams in the DEM and textures associated with *Zostera nigricaulis* and seasonal *Caulerpa cactoides* beds. Therefore, a supervised approach that relied on the circularity of reef structures was applied as follows:

1. Apply slope calculation on bathymetry raster.
2. Bryozoan biogenic reefs were detected as circular features with slope ≥ 7 degrees in the northern cell and 5 degrees in the central and southern cells. Extract rasters of those slope values and convert to a binary raster that was a proxy for “reef” and “non-reef” pixels (**Figure 3**). As will

be discussed below, in the central and southern cell, there was a depth control of bryozoans and an additional rule of > 4 m water depth was applied to the slope-based detection.

3. Grow the resultant slope raster to omit fragmented cells using the Geographic Resources Analysis Support System (GRASS) tools:
 - a. r. grow function –1 m cell size.
 - b. r. neighbours' function – neighbourhood window of five cells.
4. Convert raster to polygon.
5. Use “delete holes” tool to fill annulus features using a minimum cell area of 1 m².
6. Use Fix Geometries tool to correct polygon spatial inconsistencies.
7. Visual check and manual polygon clean-up of obvious errors by overlaying on vertically exaggerated hill-shaded bathymetry.

Detections were checked by overlapping resultant polygons on a vertically exaggerated DEM and confirming against actual ground-truthing locations.

RESULTS

Biogenic bryozoan reefs (BBR) cover a total area of 1.74 km² (Figure 4A). The total area of BBR is comprised of 0.78 km², 0.91 km², and 0.05 km² for the northern, central and southern cells, respectively (Figures 2, 4A). Taxonomic collections confirmed that the linear form of BBR is composed exclusively of three bryozoan species: *Triphyllozoon moniliferum*, *Triphyllozoon munitum* and *Celleporaria foliata* (Figures 5, 6). The *Triphyllozoon* species are fenestrate, mounded colonies with intricate laminal folding, while *C. foliata* is a non-fenestrate species forming plate-like colonies that can be mounded, spreading or encrusting. Qualitative *in situ* observations suggest that the two *Triphyllozoon* species are present in approximately equal proportions and that these two species are dominant on the BBRs, with the cover of *C. foliata* being relatively low.

The BBR fall into two distinct structural categories; “linear” bryozoan reef (LBR) and “patch” bryozoan reef (PBR). The LBR are characterised by continuous bed coverage of large bryozoan colonies orientated in a north-south pattern whereas the PBR are characterised conglomerate colonies that can individually large, but not continuous bed coverage and with no regular orientation.

Linear Bryozoan Reefs

The LBRs morphology was observed only in the northern cell (see 3D reconstruction, Figure 3) and consisted of continuous reef mounds that are linearly aligned in approximately north-south orientated rows (see Figure 7). The reef rows ranged from ~ 3 m to some 30 m in width and the vertical relief of LBR rows ranges from ~ 50–120 cm above the surrounding sediment. Water depths in the LBR zones ranges from approximately 4.8–6 m. The sediment shoal separating the linear reef zones has a depth of approximately 4 m.

In the northern cell there were three discrete areas of LBR, interrupted by a sediment shoal (see Figure 2). This sediment shoal was present in the 2009 bathymetry in addition to this new bathymetry and so does not appear to be a recent feature or to have changed significantly since 2009. The northern edge of the contiguous reef rows terminates with a distinctive gully feature and a steep slope leading to a raised bank (see Figure 3).

In the northern cell, linear reef bands were separated by tracts of very fine silty muds with depauperate infauna. We hypothesise that the baffled colony structure and morphology of the reefs running perpendicular to the prevailing current flow has the effect of slowing currents causing the settlement and trapping of fine particles that are otherwise transported in tidal flows. The multibeam bathymetry showed signs of sediment processes, with potential sediment accretion on the eastern sides of reef rows. Due to the high density of habitat-forming bryozoans in the linear reefs, these are speculated to harbour the highest matrix-associated biodiversity of co-occurring fauna.

Patch Bryozoan Reef

The PBR were observed in the northern, central and southern cells. Patch reefs occurred in 4.5–8.5 m water depth (Figure 4B). In some areas, the PBR were locally dense but these patches were of a different form to the LBRs. Ground-truthing of PBR has not been as thorough as that for the strikingly unusual LBR and field visual observations have been hampered by poor visibility. However, field observations suggest that individual reef patches represent sites of enhanced localised biodiversity of matrix-associated fauna compared to the surrounding sediment beds.

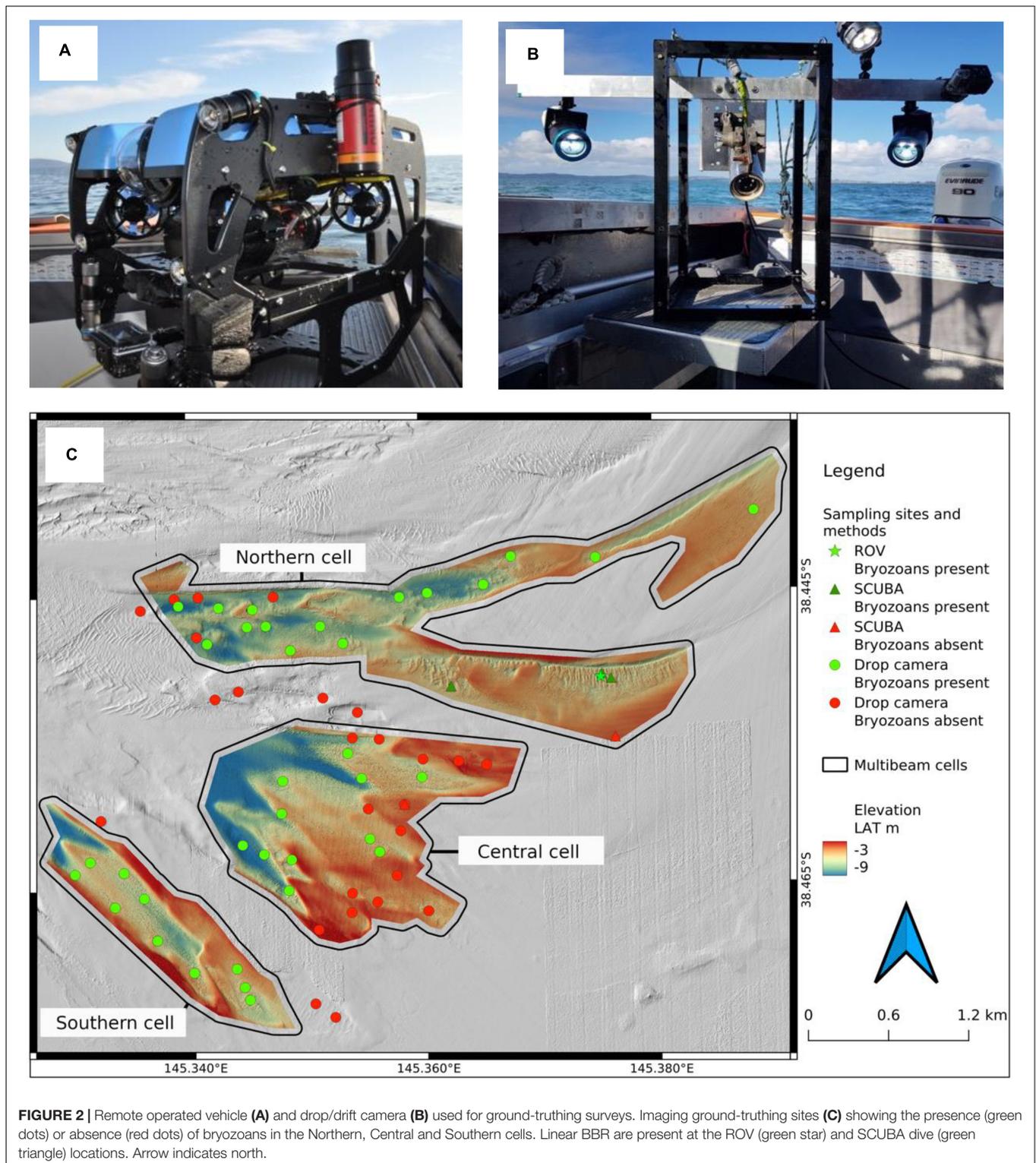
The discovery of PBR in the southern cell significantly expanded the initial expected distribution of bryozoan reefs. Furthermore, discovery of PBR in the far north and east portions of the northern cell represent new findings. Some PBR in the northern cell are associated with apparent remnant, or newly forming, LBR features, which requires future investigation.

DISCUSSION

Ecophysical Controls

The Western Port BBRs occur in seafloor depressions at the terminus of sub-channels. Harris et al. (1979) noted that due to the proximity of the Rhyll Segment to the Eastern Entrance of the bay, currents in this area do not display the directional symmetry that is typical of other sites but rather are generally counter-clockwise circulation pattern. During our surveys, underwater and surface observations were that tidal currents in the BBR areas are lower than those in the neighbouring main channels and there is significant eddying. This reduced current flow is likely to be required for the establishment and colonies with fine fenestrate lamellae that may be prohibited by the stronger currents in the main axis of nearby channels.

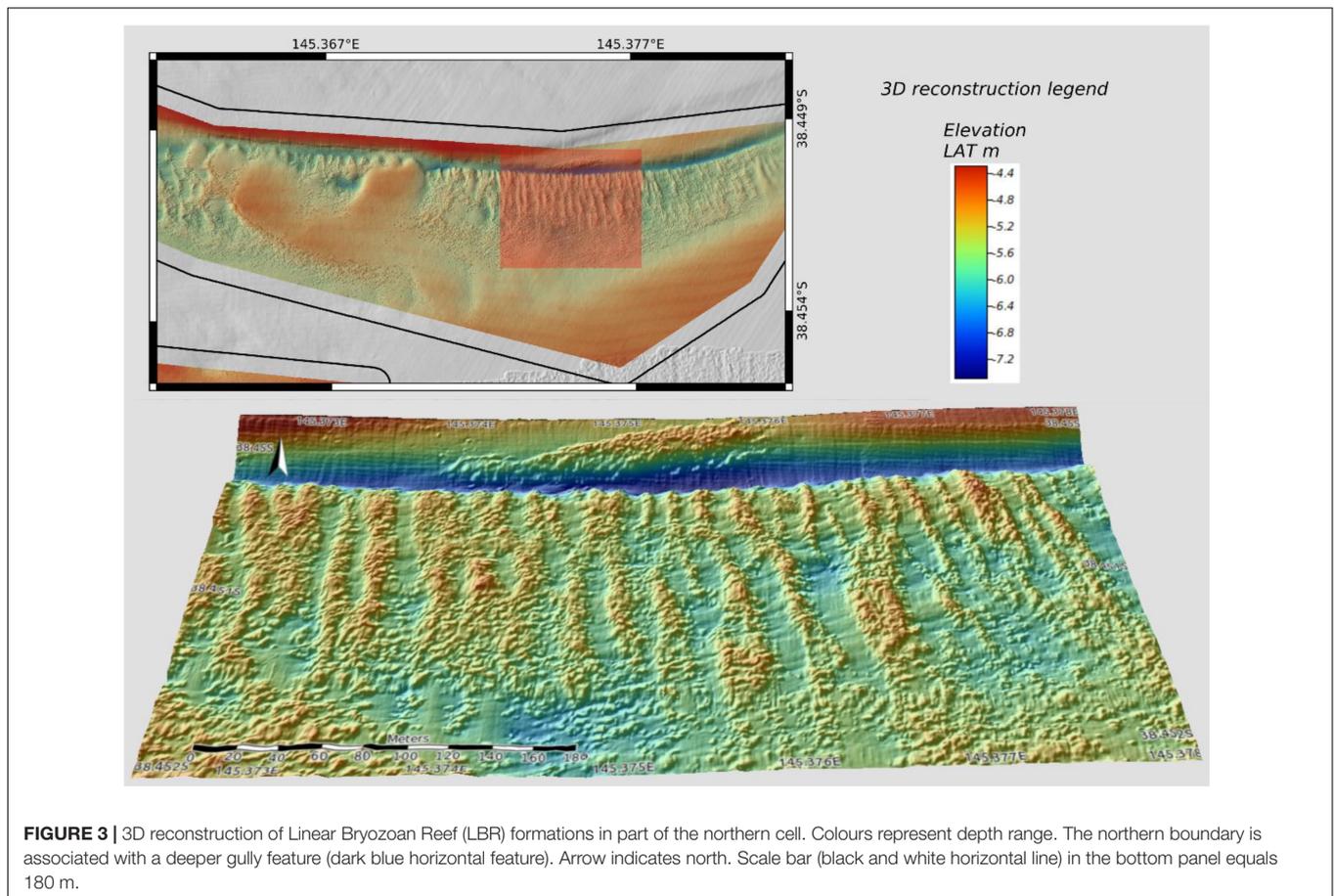
Linear bryozoan “swards” in Foveaux Strait, New Zealand are orientated parallel to the peak tidal current direction (Cranfield et al., 2003). The approximately north-south alignment of the LBR may be associated with the pattern of northerly movement of flood tide currents from eastern entrance, along the eastern



coast toward Corinella reported by Harris et al. (1979). The Western Port LBRs may further baffle current flow and eddying owing to their perpendicular orientation to the current flow in the area, potentially creating conditions that are conducive to

additional density-dependent growth and recruitment (Cranfield et al., 2003; James et al., 2008).

Subtle depth controls on bryozoan distribution were observed and these are exemplified in the central cell. Bryozoan reefs in the



central cell were generally constrained to the depth range ~ 4.5 – 8.5 m. Bryozoan reefs generally did not encroach onto the banks to the east and south of the central cell shallower than 4.5 m where *Zostera nigricaulis* seagrass and *Caulerpa cactoides* algae beds occur. Few bryozoans were detected deeper than 8.5 m in the central cell. Interestingly, this ecophysical model of bryozoan reefs bordering the offshore extent seagrass beds is reflected in the historical accounts from the oyster dredging fishery, in which fishers reported highest oyster catches at the offshore extent of seagrass (Hannan and Bennett, 2010).

On the Lacepede Shelf, South Australia, Hageman et al. (1995) identified factors of nutrients, temperature, water chemistry, environmental stability, substrate, competition, and secondary factors of water depth and currents as the potential abiotic drivers of bryozoan distributions. Additionally, on the Tasmanian continental shelf, Amini et al. (2004) noted bryozoan morphotype associations with water depth, energy regimes, sedimentation rate and substrate type, with fenestrate and foliose forms occupying deeper water where energy and sedimentation rates were low and where hard substrate is more abundant. The presence of shallow, dense growth of fenestrate bryozoan forms in the WP reefs is therefore unique and may be related to the relatively sheltered conditions, low wave action and current energy at this location. Across the northern, central, and southern cells in our study area, nutrient supply, water chemistry, temperature

and water circulation patterns are presumed to be constant. Therefore, the depth-related patterns observed may be related to competition with seagrass and *Caulerpa cactoides* at the shallow end of their distribution. Alternatively, historic differential parent substrate types at the time of reef formation may explain today's distribution. Limitations to distribution at the deep end of the distribution may be related to increased current exposure and substrate differences. Furthermore, given the generally turbid nature of this embayment, UV penetration would be typically reduced and hence closely mimic deeper habitats in which bryozoan reefs are more commonly found. Furthermore, the turbid conditions in the BBR area may be limiting competition with algae.

Link Between Bivalves and Bryozoans

There appears to be an association between the mud oysters (*Ostrea angasi*) and BBRs. Field surveys identified mud oyster clumps intermingled with bryozoan colonies, particularly *Celleporaria foliata*. This association is supported by the fact that the BBR area was an important mud oyster fishery between 1850–1900 (Hannan and Bennett, 2010). Our field surveys which involved diver hand-cores, also identified the presence of a dense subsurface bivalve shell layer consisting of *O. angasi* and *Anadara trapezia* shell. It is speculated that these observations are indicative of previously bivalve beds that may have served

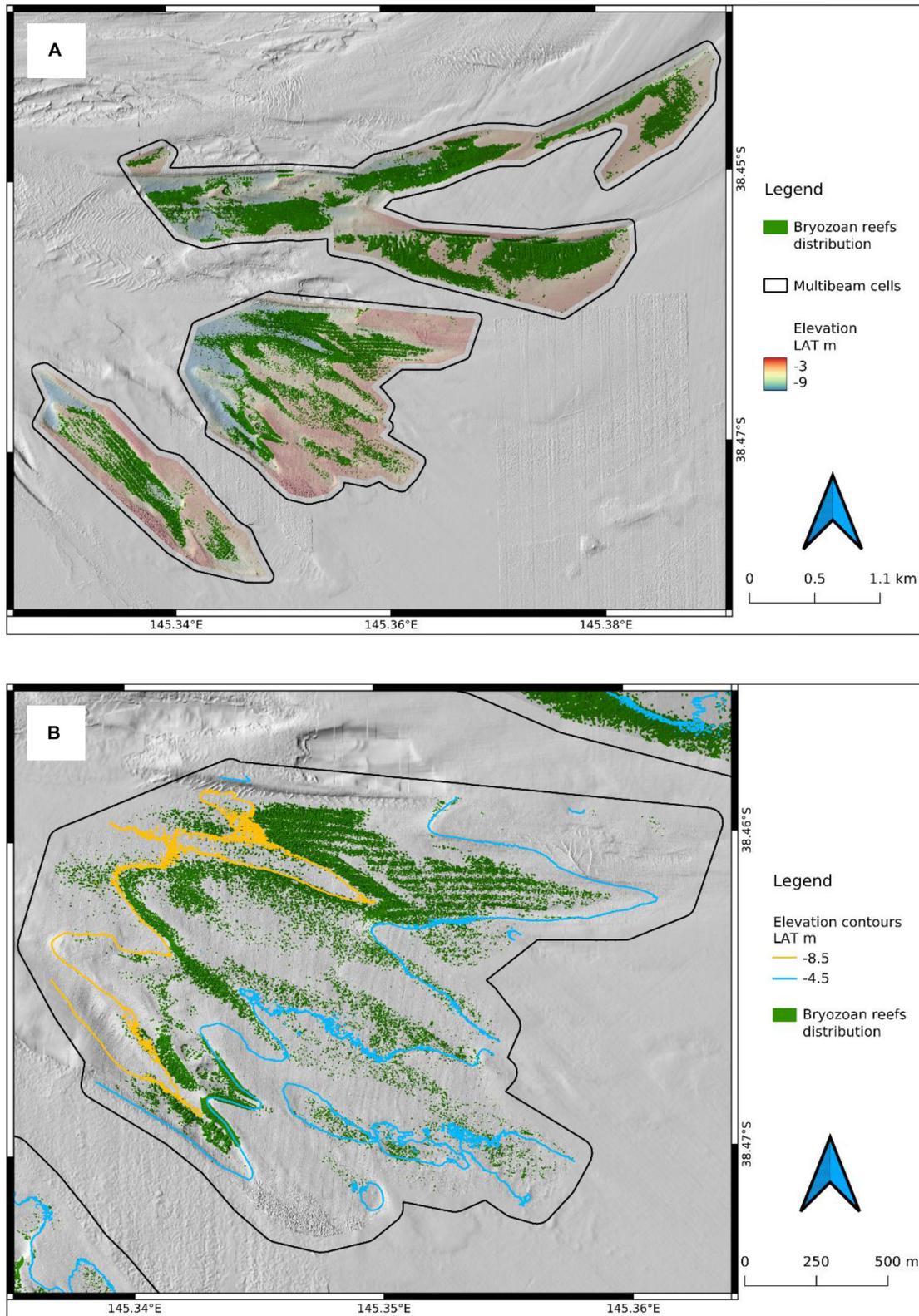


FIGURE 4 | Total extent of detected bryozoan reef **(A)** indicated in green in and the surveyed area (black polygons) of bryozoan reefs in the central cell (green, **B**) in relation to depth contours (orange = 8.5 m, blue = 4.5 m). Blue arrows denote north.

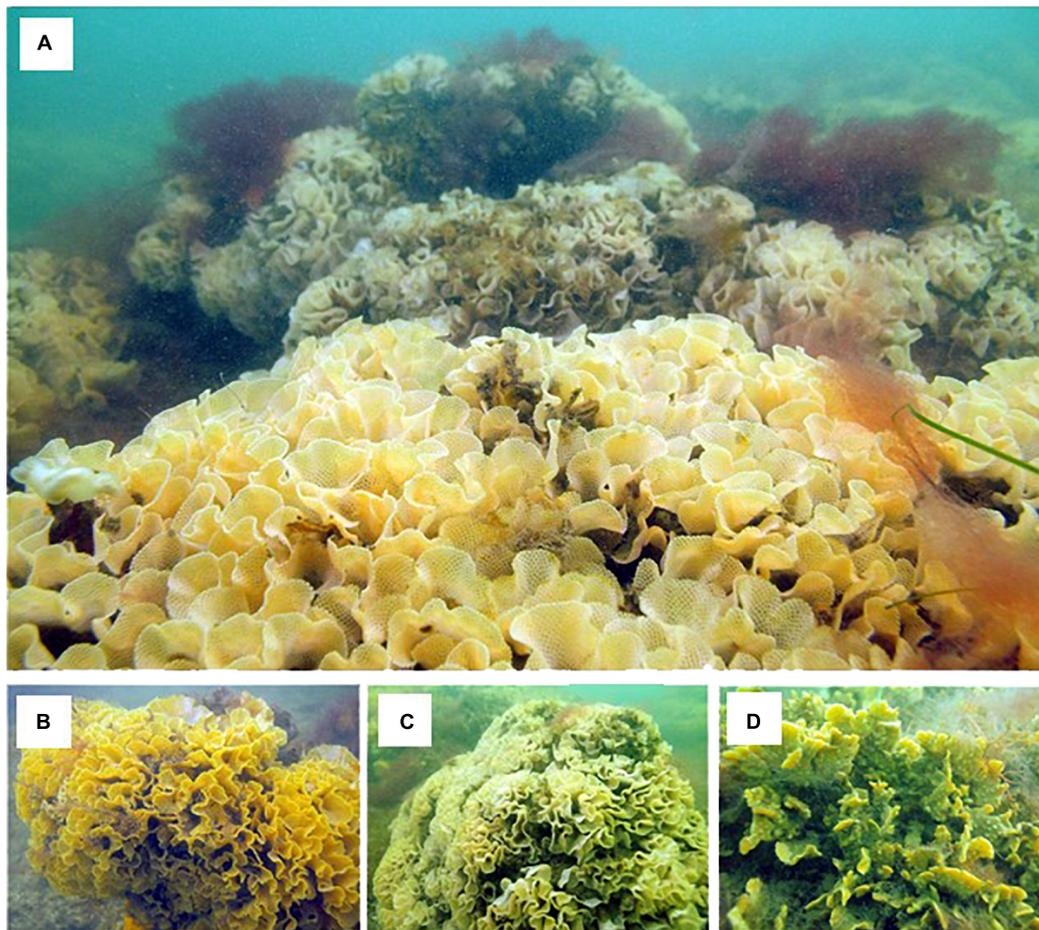


FIGURE 5 | Bryozoan biogenic reef (“BBR”, **A**) and the three dominant species of which they are comprised. *Triphyllozoon munitum* (**B**), *Triphyllozoon monilliferum* (**C**) and *Celleporaria foliata* (**D**).

as a settlement substrate for bryozoans and that is markedly different to the surface sediments today. Biogenic bivalve reefs have been severely impacted in Victoria (Ford and Hamer, 2016) and are now the subject of significant restoration investment. The apparent BBR-mud oyster association further adds to the uniqueness and ecological significance of the Western Port BBRs.

Potential Threats

The Victorian Government is actively seeking to increase recreational fishing activity (Target 1 Million initiative), which may increase the pressure associated with physical damage from anchors and fishing equipment (potentially visible in **Figure 6D**). GPS coordinates for the site, known to fishers as “The Corals”, are widely available. Recreational fishing activity should be monitored in terms of spatial intensity (has commenced as part of the *Western Port Bryozoan Reefs Project* and as per recommendation #4 in Ford and Gilmour, 2013) and linked to a program to monitor colony damage. Responsible anchoring practices and the use of new technological advancements in trawling motors, for example, could be investigated to minimise physical impacts. Precedence for this exists for the Port Phillip

Bay shellfish reef restoration project (Gillies et al., 2018), where fishers were educated about techniques to use current and wind conditions to maximise fishing on-reef structures while placing anchors in appropriate off-reef positions.

In WP, bryozoan reefs are potentially threatened by sedimentation and scouring. Sediment transport models have shown that the BBR site is exposed to southerly transport of sediment particles *via* ebb tidal water circulation from the north-eastern sector of the bay (north of French Island), through the tidal divide and into the Rhyll segment of East Arm. Once deposited, these fine sediment particles can become resuspended by wind-driven wave energy. No data are available on sedimentation specific to the BBRs, however, previous studies identified the Rhyll segment of East Arm as an area of very fine sediments (Marsden et al., 1979) and depositional processes (Hancock et al., 2001). Our field studies identified the presence of thick deposits of fine silt between the rows of LBRs that is differentiated from the muddy sands further away from the LBRs, suggesting localised sorting and trapping of fine particles. Initiatives currently underway to decrease sediment inputs include rehabilitation of mangrove and saltmarsh habitat

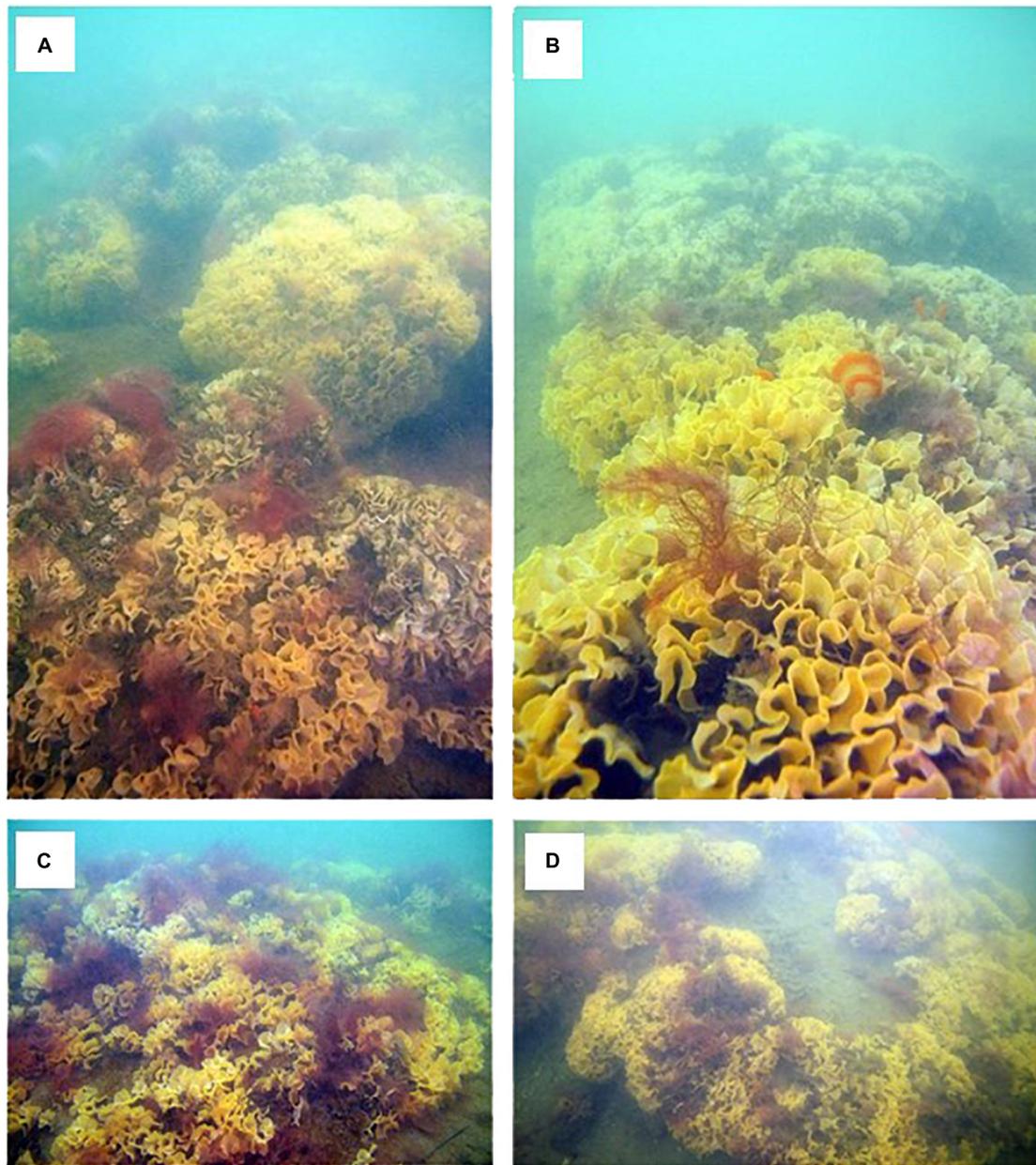


FIGURE 6 | Images from the linear reef – northern cell, **(A,B)**. Note presence of tufted fine red algae (potentially seasonal) that may be competition with bryozoans **(C)**. The presence of cavities and gaps in some large colonies may reflect damage **(D)**.

and replanting of intertidal seagrasses and riparian vegetation (Department of Environment, Land, Water and Planning [DELWP], 2017; Melbourne Water, 2018). While these initiatives are predominantly focussed on protecting seagrass meadows, they will also benefit BBRs.

To date, very limited marine pest incursions or outbreaks have been documented in the WP region. The algae wakame, *Undaria pinnatifida*, now prevalent throughout neighbouring Port Phillip Bay, presents a potential risk to BBRs as they present a large area of hard substrate that may be colonised by the algae.

The Northern Pacific seastar, *Asterias amurensis*, also prevalent throughout Port Phillip Bay, presents a risk to the abundant bivalve community associated with the BBR and a potentially large risk to bryozoans directly. Other *Asteroidea*, such as *Patiriella regularis* and *Coscinasterias calamaria* have are known to consume bryozoan zooids (Gordon, 1972), thus is it reasonable to speculate that Northern Pacific seastar could do likewise.

Increasing industrialisation and growing needs for larger and more frequent commercial and passenger shipping operations within WP, in addition to increasing recreational vessel activity,

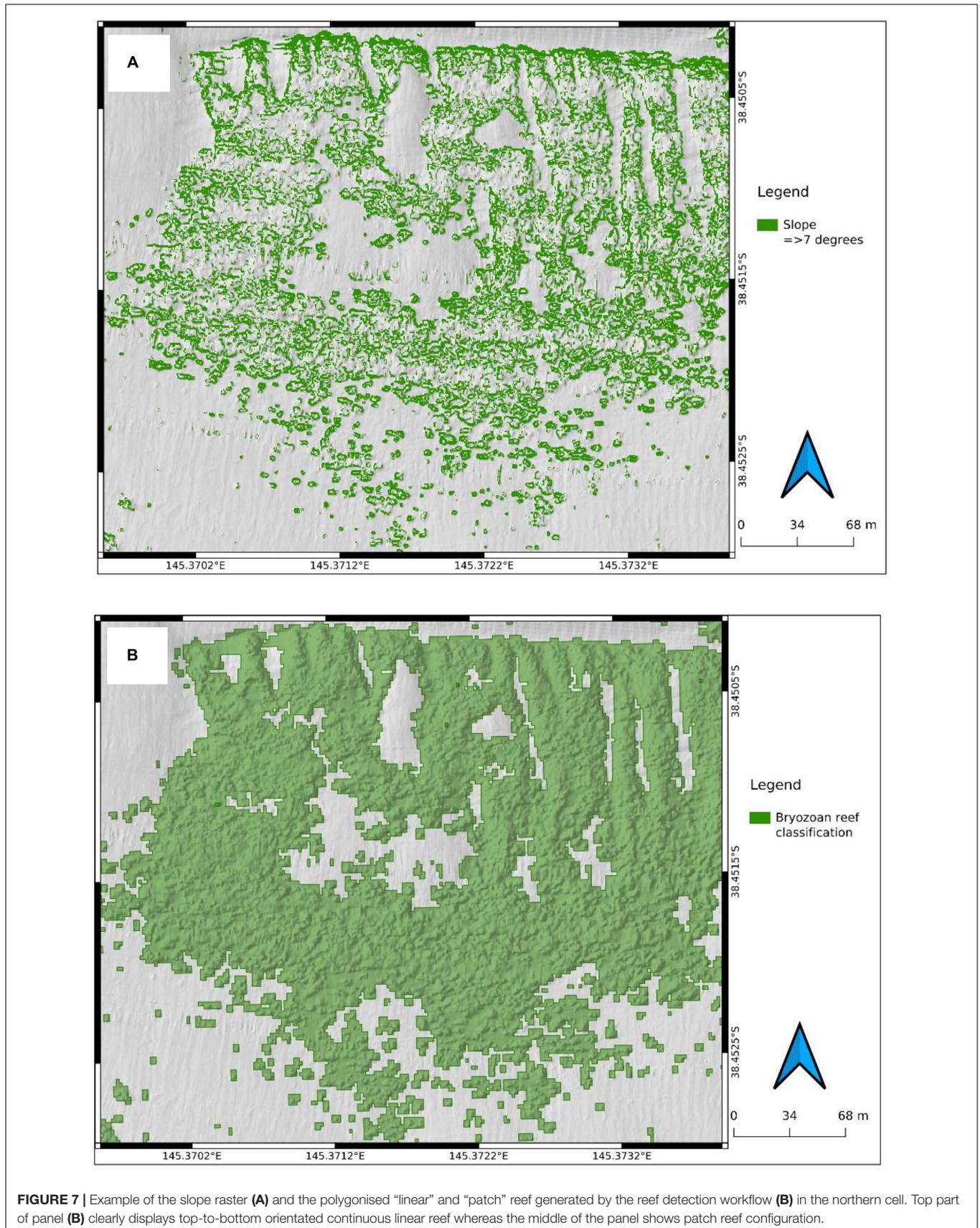


TABLE 2 | Eleven descriptors of Good Environmental Status (GES). These 11 descriptors are relevant for the BBRs of WP and based from European Commission [EC] (2008).

Descriptor	Objective
1	Biodiversity is maintained
2	Non-indigenous species do not adversely alter the ecosystem
3	The population of commercial fish species is healthy
4	Elements of food webs ensure long-term abundance and reproduction
5	Eutrophication is minimised
6	The sea floor integrity ensures functioning of the ecosystem
7	Permanent alteration of hydrographical conditions does not adversely affect the ecosystem
8	Concentrations of contaminants give no effects
9	Contaminants in seafood are below safe levels
10	Marine litter does not cause harm
11	Introduced energy (including underwater noise) does not adversely affect the ecosystem

TABLE 3 | Recommended GES indicators for monitoring the WP bryozoan reef complex. These relevant recommendations are based on European Commission [EC] (2008) and tailored to the BBRs of WP.

Indicator	Recommended indicators
(1) Habitat distribution and extent	(a) Distributional range of LBR and PBR (b) Extent of BBR (c) Relief of BBR
(2) Habitat composition, juxtaposition and fragmentation	(a) Juxtaposition of BBR habitat mosaic in the area (b) Multiscale patch dynamics and fragmentation
(3) Habitat condition	(a) Structural indices of BBR (b) Sedimentation and turbidity (c) Water quality (d) Algal cover (e) Colony health indices (f) Diversity and abundance of matrix epifauna (non-destructive)
(4) Marine pests	(a) Marine pest absence

increases the risk of potential marine pest introductions. Monitoring efforts should expand in step with this increasing risk, both at port locations and at highly sensitive locations such as the BBR complex.

Fortunately, the BBR of WP have, until now, have largely escaped anthropogenic destruction and so conservation efforts can focus on preservation rather than later restoration. In stark contrast, there are several examples in New Zealand whereby, despite many years of protection, various bryozoan populations have failed to sufficiently recover from past destructive fishing

practices (Cranfield et al., 2004; Rowden et al., 2004; Wood et al., 2012, 2013; Wood and Probert, 2013; Mello et al., 2021).

Further Research and Recommendations for Management

The Western Port Bryozoan Reef Project is engaged in concurrent studies of; reef macrofauna biodiversity, bryozoan colony age and growth, bryozoan reef fish biomass and citizen science acoustic monitoring methods. These studies are occurring at multiple spatial scales from the sub-meter scale to investigate biological interactions, to area-based management scales. The Victorian Government is embarking on a process of identifying indicators of Good Environmental Status (GES), adopting the general framework of the European Union's Marine Strategy Framework Directive (European Commission [EC], 2008). Good Environmental Status is a quantitative or qualitative description of the condition/character of the environment that managers wish to maintain or drive toward improvement (see **Table 2**). The GES is linked to ecosystem models and indicators for monitoring and establishing monitoring objectives. GES indicators that could be applied to BBR are listed in **Table 3**. Non-destructive techniques are required and data to inform these indicators can be derived from acoustics, imagery, physicochemical studies. Poor underwater visibility at the BBR location put a focus on acoustic methods of monitoring which has strongly informed related studies of the project.

These unique BBR complexes fall outside of any marine protected areas and are not afforded any legislated protection. The *Western Port Bryozoan Reefs Project* has nominated the bryozoan reefs for listing under the Victorian Flora and Fauna Guarantee Act 1998 as a community that is prone to a variety of future potential threats that may lead to extinction. Other formal legislative conservation options available include listing the community on the Environment Protection and Biodiversity Conservation Act 1999 (EPBC Act) List of Threatened Ecological Communities.

CONCLUSION

Whilst there are other kinds of extant BBR known around the world and extinct BBR from the fossil record, there are striking differences that make the WP biotope globally unique. To our knowledge the Western Port BBRs are unique because they are dominated by delicate fenestrate colonies of *Triphyllozoon* species, occur in shallow water, and can form continuous bed structures of a size and vertical relief (up to 1.5 m) that are among the largest recorded in the world. Other studies of the *Western Port Bryozoan Reef Project* have shown that the BBR support diverse and abundant invertebrate fauna. Studies into growth rates of the bryozoan species, fish biomass and cost-effective monitoring methods are underway. Increases in biodiversity and abundance of various co-occurring matrix fauna relative to neighbouring habitats have been reported (Wilson, 2021). The Western Port BBRs further contributes to the ecological character of the WP Ramsar site. Much more fundamental research is required to fully understand the ecology of BBRs and the findings

of this study can act as a baseline to track future changes in the BBR complex.

DATA AVAILABILITY STATEMENT

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

AUTHOR CONTRIBUTIONS

AFI and TD were the project organizers, conceived and designed the study, and prepared the manuscript. DD, TW, and AFe collected data, analysed data, and prepared figures. All authors contributed to the article and approved the submitted version.

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Conflict of Interest: AFL, AFe, and DD are employed by Fathom Pacific Pty Ltd. TW was employed by Total Hydrographic Pty Ltd.

The remaining author declares that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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