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# Guarding net effects on landings and discards in Mediterranean trammel net fishery: Case analysis of Egadi Islands Marine Protected Area (Central Mediterranean Sea, Italy)

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Discards remain among the main negative impacts of fishing activities, and their reductions are strengthened by the European Common Fisheries Policy (European Regulation 1380/2013). Trammel net fisheries appear more sustainable compared with other fishing techniques, especially from an ecological viewpoint. Despite this, reports show that trammel net fisheries deliver discard quantities between 10% and 43% of the total catch biomass. To supplement existing information, this current work attempts to address the discard reduction using guarding net in the small-scale fisheries of Egadi Islands MPA (Western Sicily, Central Mediterranean Sea). To assess the reduction of unwanted catches, 48 experimental fishing trials were conducted within a 6-month period. The experimental fishing trial employed a trammel net made up of 20 panels alternated with two different net configurations. The control panels (CN) held a large outer (180 mm) and small inner (31.25 mm) meshes. The test panels (GN) with guarding net constituted a three-mesh-high (50-mm mesh size) net placed between trammel net panels and a lead line. A total of 3,310 individuals belonging to 106 taxa and nine phyla were caught. Crustaceans were the most abundant unwanted catches in the control panels, whereas bioconstructions occurred in the guarding net panels. The discard ratios of CN and GN panels were statistically different ( $t$ -value =  $-2.55$ ;  $p < 0.05$ ). The analysis of catch per unit effort showed higher catches of CN panels for both commercial and discard fractions ( $p < 0.05$ ). Moreover, the guarding net panels caught the main discarded species at 20% lower compared with the control. The overall value of the catch at the CN panels (€ 3,366.90) was higher than the total income (€ 2,043.70) generated using the GN panels, which suggests a significant commercial loss of 40% ( $p < 0.05$ ).

## KEYWORDS

multivariate analysis, benthic assemblages, sustainability, unwanted catches, discard, by-catch reduction device, conservation

## 1 Introduction

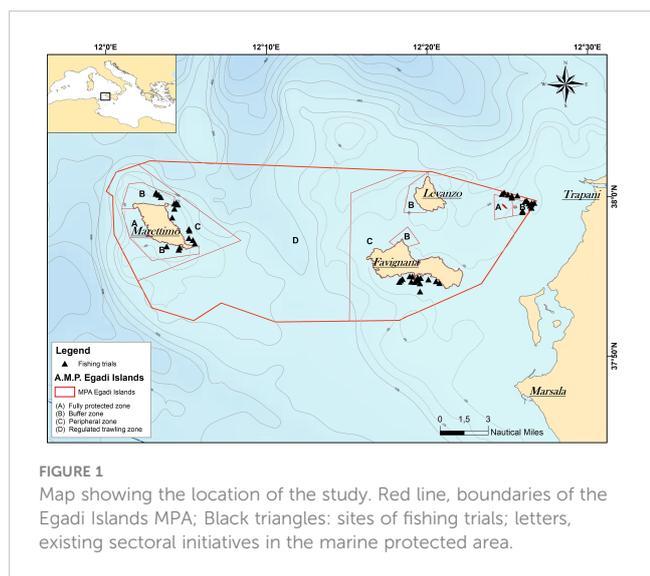
Small-scale fishery (SSF) remains a tradition that has long been performed across countries bordering the Mediterranean Sea (Lucchetti et al., 2020) and is an important socio-economic sector at the local level for tourism, cultural implications, and fisher employment (Colloca et al., 2004; Vitale et al., 2011; Cohen and Foale, 2013; Falsone et al., 2020). The main fishing gears employed in SSF include set nets (trammel nets and gillnets) or longlines, which catch a variety of demersal resources (Tzanatos et al., 2005; Stergiou et al., 2006). Specifically, trammel nets with different characteristics are crucial in southern European countries such as Spain, Italy, and Greece (Salas and Gaertner, 2004; Tzanatos et al., 2005; Erzini et al., 2006; Grati et al., 2022). The trammel nets are more selective compared with the towed gears (Vitale et al., 2011; Petetta et al., 2021) and, in the Mediterranean Sea, are characterized by moderate discarding rates (Veiga et al., 2016). Moreover, small-scale fleets account for 83% (71,400 vessels) of the EU Mediterranean fleet and contribute to 15% of the EU catches (Maynou et al., 2013; FAO, 2020). The estimate of SSF discards in European waters depends on such factors as the fishing practices and market influence as well as *métier* and varies between 10% and 43% of the total catch biomass (European Commission (EC), 2002; Vassilopoulou, 2011; Tsagarakis et al., 2014). Typically, SSFs have moderate levels of discards per vessel, but the large small-scale fleets (e.g., Falsone et al., 2020) in the Mediterranean can produce a considerable amount of discards (Bellido et al., 2011). Discards are an integral part of most fishing operations since all gears catch species that are thrown back into the sea (Bellido et al., 2011; Roda et al., 2019). However, the discards involving marine organisms are not negligible and depict a waste of natural resources that impact negatively the marine ecosystem and become a worldwide problem for the sustainable management of marine fisheries (Kelleher 2005; Vitale et al., 2018; Sardo et al., 2020; Geraci et al., 2021a). Over the past few decades, several efforts by the European Commission (EC) have aimed to tackle fishery discard-related challenges/issues—for instance, the European Commission introduced the landing obligation in Art. 15 of the EU Common Fisheries Policy (EC Regulation 1380/2013) (European parliament and Council (EU), 2013) and adopted measures such as spatio-temporal fishery closures (i.e., fisheries restricted areas), landing quotas (e.g., tuna and swordfish) as well as minimum mesh sizes (EU STECF, 2008; Damalas and Vassilopoulou, 2013; De Vos et al., 2016; European parliament and Council (EU), 2019; Di Maio et al., 2022). Importantly, the EU encourages the adoption of more selective fishing gears and better control about the record of catches (Catanese et al., 2018). As trammel nets are one of the most commonly used gears in coastal waters worldwide (Gonçalves et al., 2008; Gökçe et al., 2016), there have been some studies on discards (Purbayanto et al., 2001; Coelho et al., 2005; Martínez-Baños and Maynou, 2018), but as with discard studies for most gears, the emphasis was on vertebrates and commercially valuable or protected species (Catanese et al., 2018; Brownell et al., 2019; Geraci et al., 2019; Swimmer et al., 2020; Buscaino et al., 2021). Nevertheless, ecologically important species such as habitat-builder species and invertebrates, species at risk, or small-sized individuals

are also affected by discards, and such concerns regarding the ecological impact of trammel net fisheries on benthic communities within the coastal zone (Gonçalves et al., 2008; Metin et al., 2009; Aydın et al., 2013; Gökçe et al., 2016) and the unwanted by-catch (according to the ICES, 2020 classification) of benthic species with no commercial interest—for instance, crustaceans (crabs and hermit crabs), echinoderms (starfish, sea urchins and sea cucumbers), and gastropods (Gökçe et al., 2016). Previous works have focused on trammel net modifications (hanging ratio, different materials used to construct trammel nets, and mesh size) and devices such as acoustic deterrents and artificial lights to improve the selectivity of the gear and to reduce the interaction with unwanted by-catch species (Aydın et al., 2011; Maccarrone et al., 2014; Martínez-Baños and Maynou, 2018; Bruno et al., 2021). Other investigations have highlighted that adding a mono-cloth net with a larger mesh from 10 to 30 cm in height, called “guarding net”, at the bottom of the trammel net could represent an efficient solution to reduce the amount of unwanted species in trammel net fisheries (Sartor et al., 2007; Metin et al., 2009; Aydın et al., 2013; Gökçe et al., 2016). To supplement the existing information, this current work attempts to address the discard reduction using a guarding net in the small-scale fisheries of Egadi Island MPA (Western Sicily, Central Mediterranean Sea). The current study was designed within the framework of Project “GRECA”, a pilot project based on a measure of 3.5 of the EFF (European Fishery Fund) 2007/2013 which targets the importance of adopting low-impact fishing gears in the marine protected areas (MPAs).

## 2 Materials and methods

### 2.1 Study area

The study took place at the Egadi Islands Marine Protected Area (Central Mediterranean, Italy; hereinafter referred to as “E-MPA”), located at about 7–9 km from the western coast of Sicily, across the Strait of Sicily and the Tyrrhenian Sea (Figure 1). This archipelago comprises three main islands (Favignana, Marettimo, and Levanzo) plus a few small rocky outcrops (Galeotta, Galera, Preveto, Formica, and Maraone) (Mannino et al., 2017). The 53,992-ha E-MPA, one of the largest in the Mediterranean Sea, was created in 1991 and subdivided into four management zones (Guidetti et al., 2008; D’Anna et al., 2016): (A) fully protected zone, where only scientific research is allowed, (B) buffer zone, where small-scale fishing gears are allowed, (C) peripheral zone, in which stipulated selective fishing gears and recreational activities are permitted, and (D) regulated trawling zone, where all legal fishery activities are permitted, including trawl fishing. Specifically, both SSF and trawl fishery activities require a prior authorization of the E-MPA managing body. In particular, local management rules allow the SSF activities within zones B and C using a minimum mesh size (knot to knot) and a total length of trammel nets of 50 m and 2,000 m, respectively (DM 715/2010). In the E-MPA, SSF is practiced throughout the year in the shallow waters below 100 m in depth surrounding the Egadi Islands and characterized by soft



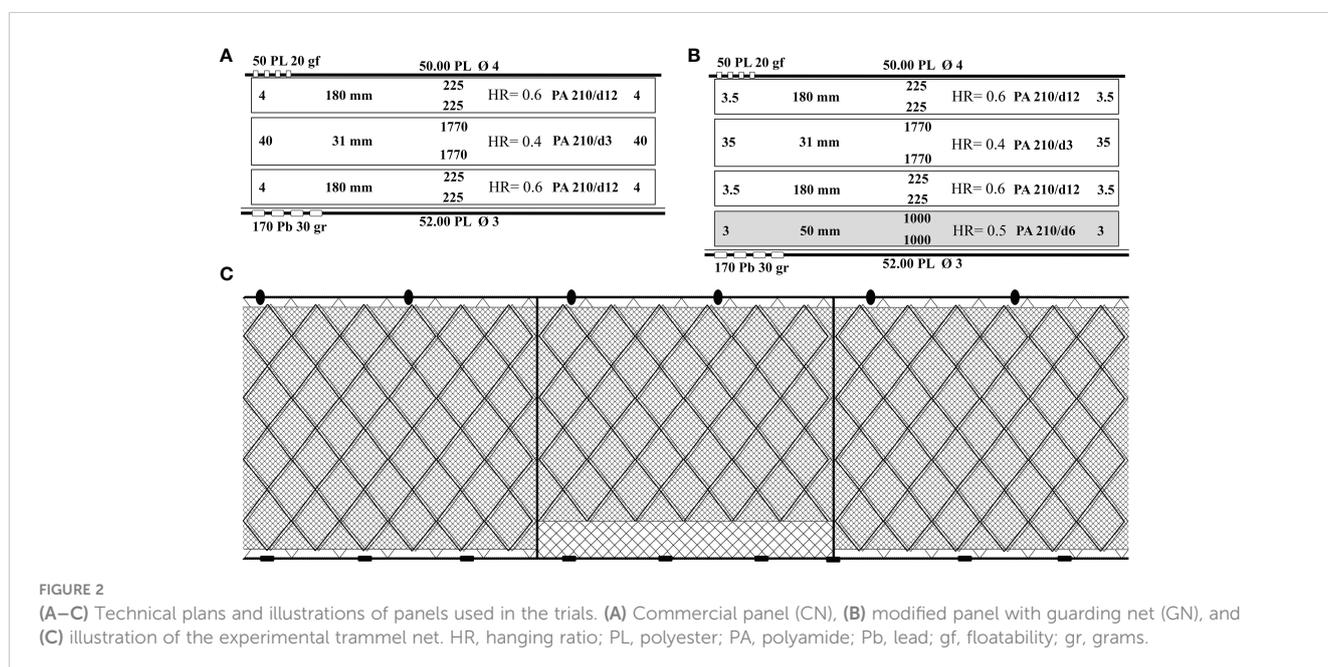
muds with fluid surface film (Garofalo et al., 2004), playing an important role in terms of employment and harvest of fishery products (Maccarrone et al., 2014). Trammel nets are normally set on the seafloor in inshore areas during the afternoon and retrieved in the morning of the next day. In particular, *Sepia officinalis* (Linnaeus, 1758), *Palinurus elephas* (Fabricius, 1787) as well as *Mullus surmuletus* Linnaeus, 1758 are generally targeted.

## 2.2 Trammel net setup and experimental program

The experimental trammel net was designed so as to alternate standard commercial panels, designated as control (CN), and modified panels with guarding net (GN), designated as test. Thus,

10 control panels were constructed with two large mesh outer panels (225 meshes in depth, 180-mm mesh size, twine thickness 210/d12 PA, four meshes in height, and hanging ratio 0.6) and one smaller mesh inner panel (1,770 meshes in depth, 31-mm mesh size, twine thickness 210/d3 PA, 40 meshes in height, and hanging ratio 0.4). Moreover, 10 modified panels that differed with a monofilament guarding net made of three-mesh-high polyamide mesh (50 mm mesh size, 1,000 meshes in depth, and twine thickness 210/d6 PA) between the trammel net and the lead line and a different height of the mesh of the outer (3.5 meshes in height) and inner (35 meshes in height) panels were constructed (Figures 2A, B). In particular, the mesh size was measured from knot to knot. The floatline and the lead line of each panel were 50- and 52-m long, with 4- and 3-mm-diameter polyester, respectively. The donut-shaped polyester floats were 50 mm in diameter. The experimental trammel net that comprised 20 panels, each of 50 m in length and 1.8 m in height (Figure 2C).

A schematic outlay of the experimental design is shown in Figure 3. Overall, there were six fishing vessels, with a length overall (LOA) comprised between 5.5 and 6.5 m, harbored in Favignana, Marettimo, and Trapani. Three different areas—namely, Favignana, Marettimo, and Formica—were randomly chosen within the “B” and “C” zones of the E-MPA (Figure 3). In particular, the vessels deployed in each area were the same during the entire period of experimental fishing trials and adhered to normal fishing practices. Eight fishing trials, with the setting of the gear in the afternoon (17:00 p.m.) and the hauling after sunrise (6:00 a.m.) of the next day (soak time of about 12 hours), were performed per vessel at a depth between 15 and 30 m. A total of 48 experimental fishing trials were carried out from January to June 2015, during the fishing season of *Sepia officinalis*. In total, 10 replicates were carried out per panel during each fishing trial, which amounted to a total of 960 replicates. Catches of each replicate were classified by the fishers



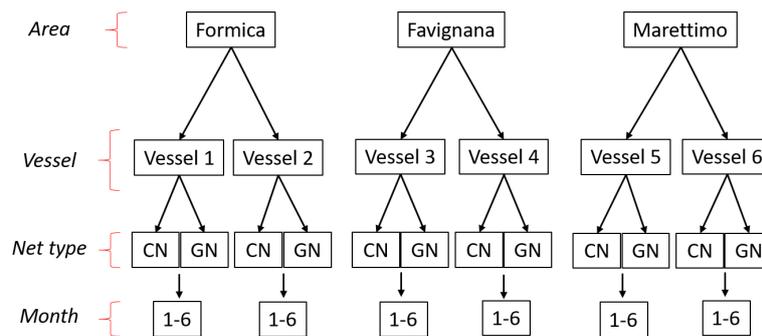


FIGURE 3  
Schematic outlay of the experimental design applied in the presently reported study.

as commercial or discard and labeled according to panel type and their order in the trammel net. Thereafter, the samples were stored in ice and transported to the laboratory of CNR–IRBIM of Mazara del Vallo where the identification of species at the lowest taxonomic level possible and biometric data were implemented. All organisms were numbered and weighed (0.1-g accuracy) both individually and as total by species.

## 2.3 Data analysis

The emergent data were standardized as the number of individuals per square meter ( $N/m^2$ ) and grams per square meter ( $g/m^2$ ) and divided into two groups, namely: (i) commercial catch and (ii) discards. Furthermore, the discard fraction comprised non-marketable undersized specimens, species with low commercial value, non-commercial invertebrate discard such as crustaceans, molluscs, echinoderms, bioconstructions, as well as the seagrass *Posidonia oceanica* (Linnaeus) Delile, 1813 and algae. Specifically, bioconstructions are characterized by non-identified species that belonged to Alcyonacea and Bryozoa taxa. Hence, three matrices (total catch, commercial catch, and discards) of species density ( $N/m^2$ ) per CN and GN panel combination and area were constructed to analyze the composition of catch and discard. Data were transformed using the square root transformation to minimize the dominant effect of abundant species (Field et al., 1982). Variables were checked for collinearity using a Pearson's correlation matrix and a scatterplot plot of each pair of variables. Furthermore, the homoscedasticity assumption was assessed with a scatter plot of the residuals (Zuur et al., 2009). Matrices were computed into triangular matrices of similarities using the Bray–Curtis similarity index (Bray and Curtis, 1957). Subsequently, a four-way permutational multiple analysis of variance (PERMANOVA, Anderson, 2001) was performed, which considered four factors, namely: net type, fixed with two levels (CN and GN); area, fixed and orthogonal with three levels (Favignana, Marettimo, and Formica); month, random and orthogonal, with five levels (number of the months); and vessel, fixed and nested in area, with two levels (number of vessels per area). Furthermore, the similarity of percentages (SIMPER) analysis (Clarke, 1983) was performed in order to identify species that

mainly contribute to the similarity either within CN and GN panels and across the groups of total catch, commercial catch, and discards. In addition, non-metric multidimensional scaling (NMDS) of the three matrices was performed to graphically highlight differences in assemblage of catches across control and experimental panels. In particular, due to the high number of replicates, only the centroids from the interaction between factors month, area, and net were considered. Stress coefficient values  $<0.2$  (Field et al., 1982) were considered as a good representation (Clarke, 1993). The probability level of  $p < 0.05$  was considered as statistically significant. Furthermore, catch per unit effort (CPUE; expressed as  $g/km$  of commercial and discard fractions) as well as the most abundant commercial species were subjected to Shapiro–Wilk normality, which resulted in a non-parametric ( $p < 0.05$ ) outcome. Thus, a Kruskal–Wallis H-test was performed to determine the overall differences in the CPUEs of commercial and discard fractions between CN and GN panels. In addition, a Dunn *post-hoc* test was performed to establish differences in the CPUEs of taxa between commercial and discard fractions. Probability level  $p < 0.05$  was considered statistically significant. Furthermore, fishers involved in the fishing trials set the price per species according to the local fish markets (Favignana, Marettimo, and Trapani). Thus, total income per vessel and panel, respectively, were recorded, and one-way analysis of variance (ANOVA) was performed to compare the economic performance of the two types of panels, which were subsequently converted to economic mean value (€).

### 2.3.1 Discard modeling

Discard ratio was defined as the discard fraction of the total catch for each type of panel (CN and GN) and calculated as:

$$D_{\text{ratio}} = \frac{\text{Discard biomass}}{\text{Discard biomass} + \text{Commercial catch biomass}} \quad (1)$$

with the biomass expressed in weight and the  $D_{\text{ratio}}$  ranging between 0 and 1.

The GAMLSS model was fitted by specifying a beta error family distribution, which allowed for responded variables of a value between 0 and 1 (Stasinopoulos and Rigby, 2007). Through the model, it was possible to test whether the discard rate (discard ratio) was significantly different between the two levels of the explanatory variable “trammel type” (CN vs. GN). The model validation was

performed through checking the absence of a residual pattern and their normality. The statistical analysis employed Primer 6 and Permanova+ (Clarke & Gorley, 2006; Anderson et al., 2008) software for Windows (Plymouth Routines in Multivariate Ecology Research, Auckland, New Zealand) and R Statistical Environment (R Core Team, 2022) using the gamlss library (Wood, 2006).

### 3 Results

A total of 106 taxa, including 98 species and five genera, were caught by the experimental trammel nets (Supplementary Table S1). Specifically, the fishing trials resulted in a catch of 3,310 individuals belonging to seven zoobenthic phyla: Cnidaria (Anthozoa), Mollusca (Gastropoda, Bivalvia, and Cephalopoda), Arthropoda (Malacostraca), Echinodermata (Asteroidea, Echinoidea, and Holothuroidea), Chordata (Chondrichthyes and Osteichthyes), and two phytobenthic phyla: Tracheophyta and Rhodophyta. The biomass of taxa using control and experimental panels per area are shown in Table 1. The total catch weight was 498 kg, and 30.5% of it was represented by discards.

The NMDS ordination for the matrices of total catch, commercial catch, and discards showed patterns of catches in relation to CN and GN panels (Figures 4A–C). Furthermore, the PERMANOVA results showed a significant value for area/net type as well as the interaction between month and vessel for both total catch and non-commercial invertebrate discard (Supplementary Tables S2, S4), whereas net type/vessel and the interaction between month and vessel were significant for the commercial catch (Supplementary Table S3). The absence of a significant interaction between area and net type for both commercial catch and discards suggests that species assemblages for CN and GN panels in different areas of the E-MPA did not differ. The results of the SIMPER analysis identified species that contributed to the separation between CN and GN panels among commercial and discard catches. In particular, *Scorpaena porcus* Linnaeus, 1758 and

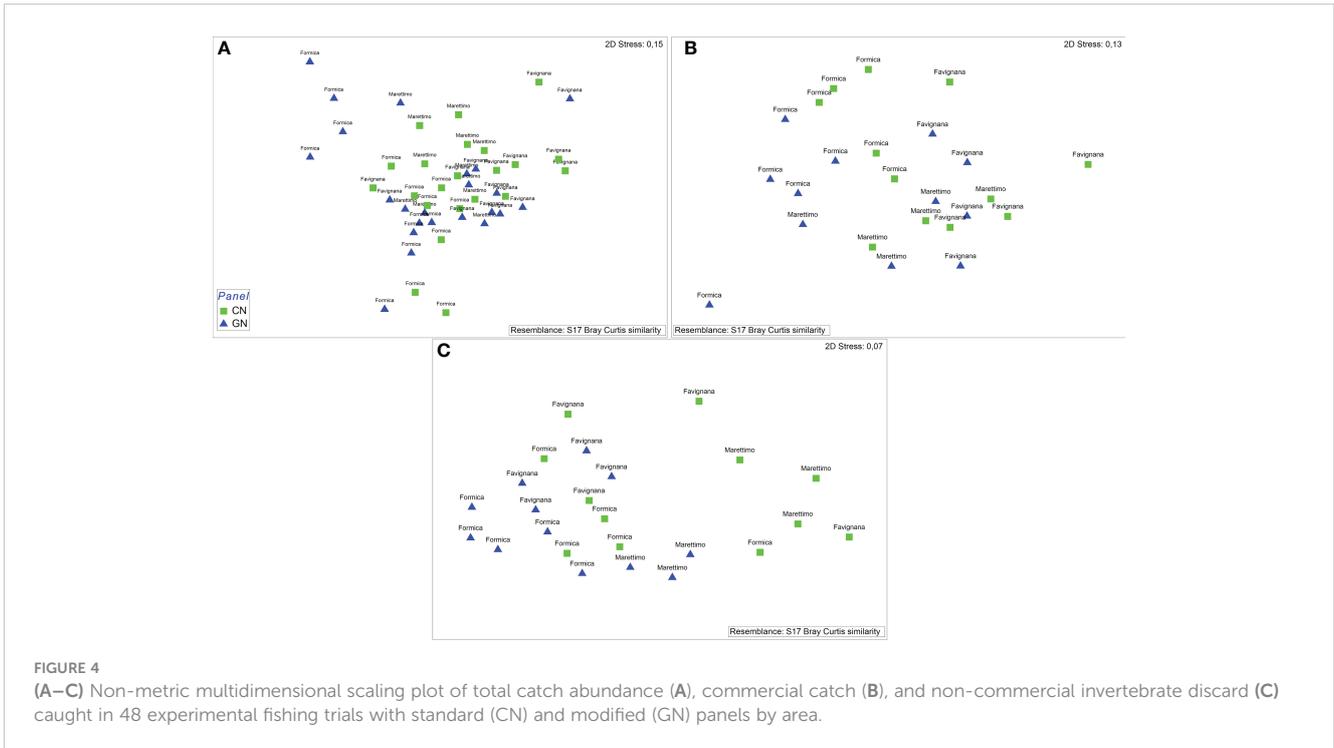
*Sepia officinalis* Linnaeus, 1758 were the most abundant species, with *S. porcus* having a contribution to the similarity for the total catch and commercial catch matrices of 18.92% in CN and of 24.16% in GN, while *S. officinalis* contributed 27.26% in CN and 28.33% in GN (Tables 2, 3). Specific to the discards, the SIMPER analysis showed the most representative groups using CN and GN panels to be Crustacea and bioconstructions, with 28.21% and 15.13% similarity, respectively (Table 4). Regarding CN panels, Crustacea was followed by bioconstructions, Osteichthyes, Gastropoda, and Asteroidea, whereas Crustacea, Osteichthyes, Gastropoda, and Chondrichthyes followed bioconstructions for GN panels (Table 4).

The analysis of CPUEs showed higher catches of CN panels for both commercial (CN, 6,442 ± 1,887; GN, 3,377 ± 1,294) and discard (CN, 1,437 ± 685; GN, 659 ± 521) fractions ( $p < 0.05$ ) (Figure 5). Moreover, Osteichthyes (CN, 5,393 ± 1,367; GN, 2,510 ± 681), Cephalopoda (CN, 38,333 ± 6,009; GN, 26,000 ± 4,932), and Crustacea (CN, 3,328 ± 1,032; GN, 1,395 ± 1,050) were the most caught taxa of CN panels ( $p < 0.05$ ). As for discard, Mollusca (CN, 33,333 ± 4,409; GN, 18,918 ± 4,441), Echinodermata (CN, 4,100 ± 1,209; GN, 972 ± 366), bioconstructions (CN, 11,000 ± 2,081; GN, ±) (Figure 6), and *P. oceanica* (CN, 1,198 ± 178; GN, 721 ± 86) (Figure 7) were the most caught taxa/groups ( $p < 0.05$ ) of CN panels, while Chondrichthyes (CN, 16,678 ± 4,602; GN, 29,238 ± 5,524) was the most caught taxon of GN panels ( $p < 0.05$ ) (Figure 6).

The discard composition was constituted by 50 species, mainly crustaceans—such as *Hexaplex trunculus* (Linnaeus, 1758), *Liocarcinus corrugatus* (Pennant, 1777), and *Dardanus calidus* (Risso, 1827)—and bioconstructions (Alcyonacea, Actiniaria, and Bryozoa) for control and experimental panels, respectively (Supplementary Table S1). The discard composition was significantly different upon comparing area and vessel ( $p < 0.05$ ). The discarded species with the highest weight at CN panels included *P. oceanica* (9.7 kg), *Torpedo marmorata* Risso, 1810 (8.4 kg), and *Aplysia* spp. (4.6 kg), while those at GN panels included *T. marmorata* (21.6 kg) and *P. oceanica* (6.1 kg). Overall, there was a significant reduction (20%) in the total

TABLE 1 Biomass of taxa for standard (CN) and modified (GN) panels per area.

Taxa	Total biomass (kg)					
	Formica		Favignana		Marettimo	
	CN	GN	CN	GN	CN	GN
<i>Osteichthyes</i>	31.61	10.60	52.27	25.48	94.70	43.14
<i>Chondrichthyes</i>	4.63	6.17	2.62	3.85	6.93	13.94
<i>Cephalopoda</i>	30.32	19.39	28.76	22.68	28.34	24.18
<i>Crustacea</i>	0.62	0.89	2.58	0.77	4.89	1.57
<i>Mollusca</i>	1.77	0.05	4.58	3.34	5.30	1.54
<i>Echinodermata</i>	0.46	0.06	0.80	0.08	1.00	0.10
Bioconstruction	0.52	0.13	0.15	0.01	1.00	0.18
<i>Posidonia oceanica</i>	1.78	1.49	5.10	3.35	2.83	1.21
Algae	0.00	0.01	0.00	0.00	0.00	0.00



**FIGURE 4** (A–C) Non-metric multidimensional scaling plot of total catch abundance (A), commercial catch (B), and non-commercial invertebrate discard (C) caught in 48 experimental fishing trials with standard (CN) and modified (GN) panels by area.

weight of discards of GN (35 kg) compared with CN (43 kg). The discard ratios of CN and GN panels were statistically different ( $t$  value =  $-2.55$ ;  $p < 0.05$ ). The GN panels showed a lower discard ratio (0.18) than the CN panels (0.23).

The income and loss of commercial catch for both CN and GN panels are shown in Table 5. The overall value of the catch at CN panels (€ 3,366.90) was higher than the total income (€ 2,043.70) using the GN panels, which suggests a significant commercial loss of 40% ( $p < 0.05$ ). Moreover, a higher loss of income was recorded in Marettimo (41.17%). The mean CPUEs of commercial species contributing to the difference in revenues between the two types of panels are shown in Figure 8. Notably, catches of *Sepia officinalis* (CN,  $9,754 \pm 736$ ; GN,  $9,102 \pm 641$ ) seem unaffected by guarding net panels ( $p > 0.05$ ), whereas non-targeted commercial species such as *Symphodus tinca* (Linnaeus, 1758) (CN,  $4,092 \pm 292$ ; GN,  $2,525 \pm 118$ ), *Mullus surmuletus* (CN,  $5,470 \pm 808$ ; GN,  $2,641 \pm 227$ ), *Scorpaena scrofa* Linnaeus, 1758 (CN,  $6,888 \pm 847$ ; GN,  $4,999 \pm 542$ ), *Scorpaena porcus* Linnaeus, 1758 (CN,  $3,392 \pm 204$ ; GN,  $2,379 \pm 145$ ), *Diplodus annularis* (Linnaeus, 1758) (CN,  $3,052 \pm 606$ ;

GN,  $1,916 \pm 196$ ), and *Spondyliosoma cantharus* (Linnaeus, 1758) (CN,  $1,888 \pm 198$ ; GN,  $1,113 \pm 169$ ) appeared with somewhat noticeable reduced catches ( $p < 0.05$ ).

## 4 Discussion

This current study has demonstrated that the adoption of a guarding net fitted to the lead line of net reduced the catch of benthic invertebrates, and our findings appear to be in agreement with previous studies in the Mediterranean Sea (Sartor et al., 2007; Metin et al., 2009; Aydin et al., 2013; Gökçe et al., 2016; Martínez-Baños and Maynou, 2018; Szynaka et al., 2018). The composition of bycatch taxa in the control panels suggests that the discards, both in terms of abundance and biomass, comprised mainly of invertebrates such as crustaceans and gastropods. High fishing mortality of crustaceans would produce cascading effects on the various species that depend on this trophic resource, which suggests that their removal could reduce the diversity and complexity of the

**TABLE 2** Percentage contribution of the first five typifying species (over 1.5%) to within-group similarity for the standard (CN) and modified (GN) panels of total catch as assessed by similarity percentage analysis.

CN		GN	
Species	Contribution (%) to catch	Species	Contribution (%) to catch
<i>Scorpaena porcus</i>	18.92	<i>Sepia officinalis</i>	24.16
<i>Sepia officinalis</i>	16.89	<i>Scorpaena porcus</i>	14.39
<i>Posidonia oceanica</i>	15.09	<i>Posidonia oceanica</i>	14.30
<i>Symphodus tinca</i>	13.45	<i>Symphodus tinca</i>	7.44
<i>Diplodus annularis</i>	2.40	<i>Scorpaena scrofa</i>	1.61

TABLE 3 Percentage contribution of the first five typifying species (over 1.5%) to within-group similarity for the standard (CN) and modified (GN) panels of commercial catch as assessed by similarity percentage analysis.

CN		GN	
Species	Contribution (%) to catch	Species	Contribution (%) to catch
<i>Scorpaena porcus</i>	27.26	<i>Sepia officinalis</i>	28.33
<i>Sepia officinalis</i>	23.76	<i>Scorpaena porcus</i>	17.47
<i>Symphodus tinca</i>	18.86	<i>Symphodus tinca</i>	9.50
<i>Diplodus annularis</i>	3.61	<i>Scorpaena scrofa</i>	2.42
<i>Spondylisoma cantharus</i>	3.57	<i>Mullus surmuletus</i>	1.70

TABLE 4 Percentage contribution of the invertebrate discard (over 1.5%) to within-group similarity for the standard (CN) and modified (GN) panels as assessed by similarity percentage analysis.

CN		GN	
Species	Contribution (%) to catch	Species	Contribution (%) to catch
Crustacea	28.21	Bioconstruction	15.13
Bioconstruction	15.04	Crustacea	11.72
Osteichthyes	8.95	Osteichthyes	10.54
Gastropoda	8.65	Gastropoda	9.12
Asteroidea	6.72	Chondrichthyes	9.11

benthic community. Disentanglement of discarded invertebrates might increase the labor time on board and result to some damage to the net with the addition of weight and increasing contact with the seabed (Catanese et al., 2018; Szynaka et al., 2018). Among the benefits associated with the guarding net could be the catch reduction of predatory epifaunal invertebrates, which would be potentially feasible when a physical barrier is effected to the climbing scavengers that might damage the capture (Metin et al., 2009; Szynaka et al., 2018). Furthermore, the fragments of sessile organisms such as rhizomes and leaves of *P. oceanica* represented a non-negligible fraction of the discard and provided an insight about the interaction between the trammel net and the seagrass bed. In

this study, the control panels caught more *P. oceanica* than the test panels. There is evidence that seagrass meadows serve as source of food and shelter as well as nursery for numerous marine species (Vlachopoulou et al., 2013)—for instance, crustaceans and gastropods were among the best-represented groups in the vagile fauna of *P. oceanica* beds (Russo and Terlizzi, 1997). Given its important role in coastal ecosystems and the status of endangered species, *P. oceanica* is protected through the EC Habitats Directive (92/43/EEC, 1992), and this encourages the establishment of MPAs for priority habitats. In addition, reductions in the bycatch of *P. oceanica* especially in sensitive ecosystems should be considered an important issue for sustainable fisheries. In particular, the E-MPA

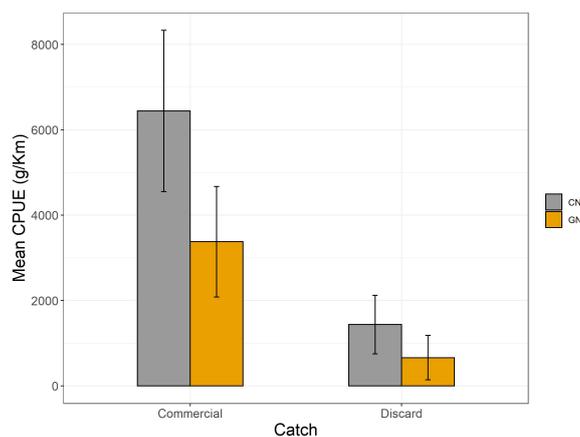
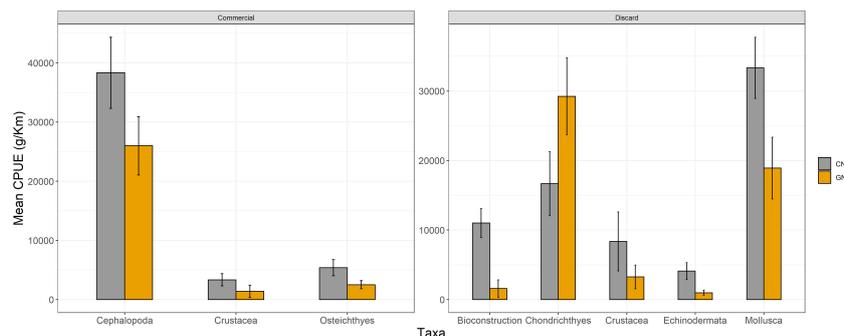


FIGURE 5

Mean catch per unit effort expressed as g/km for commercial and discarded catches per commercial (CN) and modified (GN) panel, respectively.



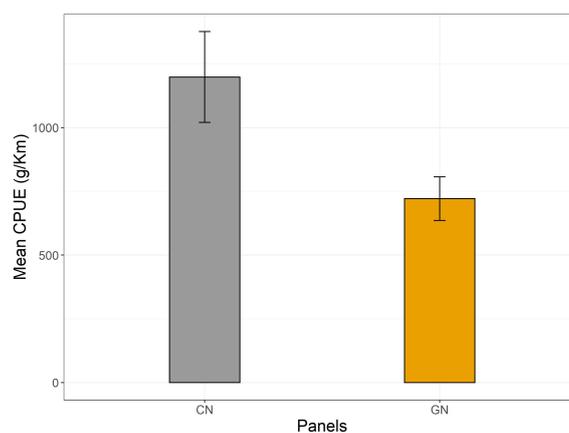
**FIGURE 6**  
Mean catch per unit effort expressed as g/km of most abundant taxa for commercial and discarded fractions per commercial (CN) and modified (GN) panel, respectively.

includes the best-preserved and the largest *P. oceanica* meadow of the Mediterranean Sea (about 7,700 ha) (Agius and Chaperon, 2021). Therefore, the adoption of the guarding net is advisable in order to reduce the impact of the trammel net on *P. oceanica*. Nevertheless, further investigations to identify technical devices aimed to mitigate interactions with *P. oceanica* should be considered.

In this study, some specimens of elasmobranchs such as *T. marmorata* showed great importance for the total discard biomass both in control and experimental panels despite being captured in low numbers. Plausibly, the higher catches of *T. marmorata* in the guarding net panels might be related to the size of the specimens. Small individuals could most likely pass through the guarding net as larger individuals are retained. The bycatch and discards of elasmobranchs would, therefore, be considered high (33.2%) for trawl fisheries (Coelho and Erzini, 2008; FAO, 2020; Geraci et al., 2021b; Falsone et al., 2022). Nevertheless, artisanal fishery such as trammel net (37.8%) and longline (7.7%) also provide a significant bycatch of elasmobranch species (Erzini et al., 2002; Baeta et al., 2010; Catanese et al., 2018; FAO, 2020). Notably, the yields of SSF depend on environmental and seasonal conditions as well as the

quantity of target species caught during the fishing trip (Battaglia et al., 2010). It is important to mention that fishers in the Egadi Islands SSF seem to retain bycatch species with potential commercial value, such as *Raja* spp., especially when catches are low, probably to sustain their economic activity. Either the retention or discarding of elasmobranchs could differ based on the local fisher's behavior—for instance, *T. marmorata* is retained for fishers' own consumption, and large-sized specimens are marketed along the southern coasts of Sicily (Tiralongo et al., 2018), Algarve (Gonçalves et al., 2007), and Portugal (Baeta et al., 2010). On the contrary, *T. marmorata* is among the discards in the Ionian coast of Sicily (Tiralongo et al., 2018), Gulf of Cadiz (Gonçalves et al., 2007), and E-MPA. According to Gil et al. (2018), *T. marmorata* showed a low survival rate, using trammel nets, when immediately released to the sea.

Concerning the catch composition of commercial species, the present study corroborate with the seasonal dynamics of Sicilian SSF (Grati et al., 2018; Falautano et al., 2018; Falsone et al., 2020). Trammel net fishery in southwestern Sicily might provide high species variability of landings during the year according to the seasonal patterns of target species. Specifically, *S. officinalis* is



**FIGURE 7**  
Mean catch per unit effort expressed as g/km of *Posidonia oceanica* per commercial (CN) and modified (GN) panel, respectively.

TABLE 5 Comparison of income and loss per area and vessel with standard (CN) and modified (GN) panels.

Area	Vessel	Income (€)		Total income (€)		Total Loss (€)	Loss (%)
		CN	GN	CN	GN		
Formica	1	425.93	232.97	1,042.31	652.57	389.74	37.39
	2	616.38	419.59				
Favignana	1	849.14	494.55	1,519.53	917.52	602.01	39.62
	2	670.39	422.97				
Marettimo	1	484.04	280.24	805.06	473.61	331.45	41.17
	2	321.02	193.37				

exploited mostly in winter–spring, when mature specimens aggregate inshore for spawning from late fall to spring (Gharbi and Ktari, 1981; Colloca et al., 2004). In this study, guarding net panels caught more *S. officinalis* than *S. porcus* when compared with the control panels. Thus, guarding net panels might improve the catchability of trammel nets' target cuttlefish within the E-MPA. The potential loss of revenue of about 40% seems likely in the commercial value of landings when a commercial trammel net is replaced with an alternative net aiming to reduce by-catch and discards. Such loss might be due to the lower catches of non-target marketable species, such as *Spondyliosoma cantharus*, *Scorpaena scrofa*, and *Mullus surmuletus*, that have been highly priced. Such occurrence might plausibly corroborate the mesh size of the guarding net, which could decrease the chance for the abovementioned fish species to be trapped. To the best of our knowledge, the contrasting results about the effects of guardian nets in SSF appears to be the situation that we found, which is consistent with those reported in scientific literature (Sartor et al., 2007; Metin et al., 2009; Martínez-Baños and Maynou, 2018). In particular, Sartor et al. (2007) showed that discards in trammel nets in the Tyrrhenian Sea were significantly reduced along with abundance of the target species carapace prawn *Penaeus kerathurus* (Forskål, 1775). On the other hand, Metin et al. (2009) in the Aegean Sea showed that the height of the guarding net is an important factor for discard reduction, whereas the decrease of commercial catch was considered not significant for the fishermen at Izmir Bay. Martínez-Baños and Maynou (2018) showed that guarding net reduced unwanted by-catch and increased the catches of commercial species in the Murcia Region, which would target cuttlefish and

some fishes, by 30%. Sartor et al. (2018) considered that the economic loss that could happen due to the reduced catch of commercial species is offset by decreased sorting time and labor costs in the immediate short term. Other research, nonetheless, believe that a difference of incomes between the standard and test panels might create resistance toward the adoption of the guarding net by fishers (Szynaka et al., 2018). However, overcoming the economic loss might require the adoption of an eco-label marker for fishery products harvested with more sustainable fishing gears mounting the guarding net, as supported by the E-MPA managing body. The use of guarding net on the lead line of the trammel nets at Egadi Islands significantly reduced the catch rate of discards, which benefits the benthic invertebrate species. Nevertheless, further investigation on catch composition, specific to commercial and discarded fraction within the Sicilian SSF, should consider the effectiveness and suitability of guarding net by *métier* (Szynaka et al., 2018), season and depth (Stergiou et al., 2006), soak time, fishing grounds as well as the use of different combinations of mesh sizes both for inner and outer panels (Gonçalves et al., 2008).

## Data availability statement

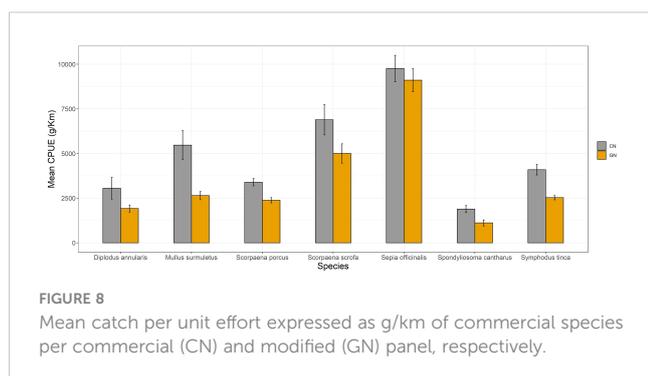
The raw data supporting the conclusions of this article will be made available by the authors without undue reservation.

## Author contributions

LV, GM, and SV: conceptualization. LV and GM: formal analysis. GS, MG, FF, DS, and DM: data curation. MG, FF, DS, PR, and DM: data collection and figures. SV, DS, FF, and GS: validation. GS, LV, and GM: writing—original draft. MG, DS, FF, DS, DM, and SV: writing—review and editing. All authors contributed to the article and approved the submitted version.

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## Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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## Supplementary material

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fmars.2023.1011630/full#supplementary-material>

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