



OPEN ACCESS

EDITED BY

Andrea D. Phillott,
Flame University, India

REVIEWED BY

Adriana A. Cortés-Gómez,
Systématique et Évolution (ESE), France
Celine Alexia Godard-Codding,
Texas Tech University, United States

*CORRESPONDENCE

Arthur D. Barraza

✉ Arthur.Barraza@griffithuni.edu.au

RECEIVED 12 June 2023

ACCEPTED 17 October 2023

PUBLISHED 13 November 2023

CITATION

Barraza AD, Young L, Smith CE, Booth DT,
Hof CAM, Finlayson KA, Leusch FDL and
van de Merwe JP (2023) Exploring
contaminants as a disruptor of
temperature-dependent sex
determination in sea turtle hatchlings.
Front. Mar. Sci. 10:1238837.
doi: 10.3389/fmars.2023.1238837

COPYRIGHT

© 2023 Barraza, Young, Smith, Booth, Hof,
Finlayson, Leusch and van de Merwe. This is
an open-access article distributed under the
terms of the [Creative Commons Attribution
License \(CC BY\)](https://creativecommons.org/licenses/by/4.0/). The use, distribution or
reproduction in other forums is permitted,
provided the original author(s) and the
copyright owner(s) are credited and that
the original publication in this journal is
cited, in accordance with accepted
academic practice. No use, distribution or
reproduction is permitted which does not
comply with these terms.

Exploring contaminants as a disruptor of temperature-dependent sex determination in sea turtle hatchlings

Arthur D. Barraza^{1*}, Larissa Young², Caitlin E. Smith^{3,4},
David T. Booth², Christine A. Madden Hof³,
Kimberly A. Finlayson¹, Frederic D. L. Leusch¹ and
Jason P. van de Merwe¹

¹Australian Rivers Institute, School of Environment and Science, Griffith University, Gold Coast, QLD, Australia, ²School of Biological Science, The University of Queensland, St. Lucia, QLD, Australia, ³UniSC Fraser Coast, University of Sunshine Coast, Hervey Bay, QLD, Australia, ⁴World Wildlife Fund for Nature-Australia, Brisbane, QLD, Australia

Sea turtle nesting beaches are experiencing increased sand temperatures as climate change progresses. In one major green turtle (*Chelonia mydas*) nesting beach in the northern Great Barrier Reef, over 99 percent of hatchlings are female. The effects of contaminants on sea turtle hatchling sex determination are not often explored. Liver samples were collected from green turtle hatchlings that were sacrificed for histological sex determination in a parallel study on the effects of sand cooling on sex ratios, and analysed for trace elements via acid digestion and organic contaminants via *in vitro* cytotoxicity bioassays. Chromium, antimony, barium, and cadmium have previously been demonstrated to be estrogenic, and concentrations of these elements were used to calculate three estrogenic indexes for each clutch: predicted relative estrogenic potency (PEEQA), the sum of percent trace elements above the median of all samples (TEOM), and the sum of percent estrogenic elements above the median of all samples (EstroEOM). Excluding an outlier clutch, cadmium, antimony, and EstroEOM had significant positive relationships with sex ratio deviation. Mean clutch cobalt, lead, antimony and barium, also had a significant positive relationship with clutch sex ratio. There was no relationship between *in vitro* cytotoxicity of liver extracts and sex ratio, however, 9% of hatchlings had organic contaminants high enough to suggest potential cellular damage. Contaminant effects on sex determination are likely to be caused by a mixture of contaminant interactions as well as temperature. Many trace elements detected in this study have also been linked to negative health effects on hatchlings in previous studies. Considering the risks of feminization due to climate change and potential contaminant effects on hatchling health and sex determination, future studies exploring contaminant effects on sea turtle hatchling sex determination are recommended.

KEYWORDS

organic contaminant, sea turtle, sex ratio, trace elements, sex determination

1 Introduction

The world's ocean and surface temperatures are increasing due to climate change (Loarie et al., 2009; Abraham et al., 2013), adding challenges to the recovery of threatened species (Hawkes et al., 2009; Hamann et al., 2010). Sea turtles travel, sometimes thousands of kilometres, from their foraging grounds to sandy beaches in the regions where they hatched to lay their eggs (Meylan et al., 1990; Bowen et al., 1992; Limpus et al., 1992). Sea turtles have temperature-dependent sex determination (TSD), in which nest temperature causes a series of epigenetic and hormonal changes that determine hatchling sex development (Valenzuela and Lance, 2004; Warner, 2011). As a result, there are many studies investigating whether climate change could negatively affect sea turtle population recovery. For example, increased sand temperature on nesting beaches on the northern Great Barrier Reef (GBR) is already causing 99% female bias in hatchling production of green turtle (*Chelonia mydas*) (Jensen et al., 2018).

As a reproductive strategy, TSD can offer resilience to increased temperatures by producing more females, and thereby, increasing reproductive capacity (Santidrián Tomillo et al., 2015; Santidrián Tomillo and Spotila, 2020). However, more recent studies have found that moderate- to high-temperature scenarios still lead to declining northern GBR green turtle populations (Blechschmidt et al., 2020), and that female-skewed sex ratios take time to manifest within populations (Hays et al., 2023). Additional concerns that can compound nest feminization include sea turtle nests reaching upper thermal thresholds, reducing hatchling health and survival (Fuentes et al., 2011; Fleming et al., 2020; Page-Karjian et al., 2022). For example, loggerhead sea turtles nesting within Cabo Verde islands are predicted to reach over 99% female hatchling production, with 90% of nests reaching upper thermal limits under current climate change predictions (Tanner et al., 2019). To combat these issues, recent studies have explored using artificial means, such as cool water irrigation, to decrease sand and nest temperatures, and thus increase the proportion of male hatchlings produced on warmer nesting beaches (Lolavar and Wyneken, 2021; Porter et al., 2021; Smith et al., 2021). Both freshwater and sea water irrigation have been found to be effective at reducing nest temperatures and producing males, without decreasing hatching success *in situ* (Lolavar and Wyneken, 2021; Porter et al., 2021; Smith et al., 2021).

Due to the invasive nature of assessing the sex of sea turtle hatchlings (removal and histological investigation of the gonads), the sex ratio of incubating clutches is often predicted or modelled. Depending on the methods used, models estimating sex ratio from sand temperature can vary significantly, and may not be as accurate as previously believed (Georges et al., 1994; Wyneken and Lolavar, 2015). There are many other factors, such as nest depth, shading, and metabolic heating that can affect nest temperatures, and hence sex ratios (van de Merwe et al., 2006). The current study proposes that anthropogenic contaminants may be an additional factor that can influence hatchling sex determination.

Anthropogenic contaminants have been found to disrupt a variety of endocrine and reproductive systems in reptiles (Gardner and Oberdorster, 2005; Boggs et al., 2011; Barraza et al.,

2021). In sea turtles, only dichlorodiphenyldichloroethylene (DDE) has been investigated directly, and was found not to influence sea turtle hatchling sex determination at the concentrations tested (Podreka et al., 1998). However, when examining the literature on other reptile species, many studies have found that contaminants can skew sex ratios of TSD species such as alligators, freshwater turtles, and caimans (Bergeron et al., 1994; Crain et al., 1997; Beldomenico et al., 2007; Canesini et al., 2018). For example, pond slider (*Trachemys scripta*) eggs dosed with a mixture of polychlorinated biphenyls (PCBs) produced more female-biased clutches than dosing with single PCB congener (Bergeron et al., 1994).

It is known that contaminants accumulate in green turtles (Keller, 2003; Finlayson et al., 2016; Cortes-Gomez et al., 2017), mainly from foraging near contaminated areas (manuscript in review), and that these contaminants can be maternally offloaded to their eggs, resulting in *in ovo* hatchling exposure (van de Merwe et al., 2009; Perrault et al., 2011; Perrault et al., 2016). We hypothesised that this maternal transfer of contaminants may influence sex determination in green turtle hatchlings. Considering the logistical and ethical difficulties in conducting a direct dosing experiment on green turtle embryos, the influence of contaminants in altering sex determination was investigated in conjunction with a concurrent study that assessed the impact of fresh and sea water irrigation on hatchling sex ratios *in situ*. Hatchlings that emerged from these manipulated nests were analysed for inorganic and organic contaminants, and concentrations were interpreted in the context of expected and actual sex ratios to assess possible evidence of contaminant-induced sex ratio skewing.

2 Methods

2.1 Animal ethics

In the concurrent nest cooling study, research procedures and methods were approved by animal ethics committee at the University of Queensland (SBS/237/20), and egg and hatchling collection was completed under a Queensland Parks and Wildlife Services (QPWS) scientific purposes permit (PTU19-002377-1). All efforts were made to minimise suffering and maximise the amount of data produced from each animal.

2.2 Egg collection and incubation

The concurrent nest cooling experiment was a part of WWF-Australia's Turtle Cooling Project, which aimed to reduce the risk of excessive feminization of green turtles across Asia-Pacific (Young et al. in press). Green turtle clutches were collected from Heron Island (latitude -23.8416, longitude 151.2498), a small coral sand cay on the southern GBR, roughly 80 km off the coast of Gladstone, Queensland, Australia. Heron Island is a medium density green turtle nesting site (200 – 1800 nesting females per season) and was

selected because it is a long-term monitoring site for the southern GBR nesting population of green turtles.

Clutches ($n = 17$) were relocated within two hours of oviposition with HOBO MX2303 wireless temperature probes placed amongst eggs, with the nest and temperature logged every hour throughout incubation. The relocated clutches were arranged in a Latin square design pattern to avoid column or row effects between clutches. Five different irrigation regimens were used: a control with no nest irrigation (Control), freshwater irrigation of the nest (Freshwater), sea water irrigation of the nest (Seawater), chilled seawater irrigation of the nest (Cold Seawater), and twice the amount of sea water irrigation (2x Seawater; Table 1). Data on the relationships between incubation temperature, incubation period, and hatchling sex-ratio were gathered from published literature (Bustard and Greenman, 1967; Miller and Limpus, 1981; Booth and Astill, 2001; Booth et al., 2004; Burgess et al., 2006). Using these data, a predictive non-linear algorithm for the relationship between incubation temperature and embryonic development rate and hatchling sex-ratio was formulated (Smith et al., 2021). Nest temperatures were downloaded periodically from HOBO temperature probes throughout incubation, without disturbing the sand, and used to calculate when the developing embryos were midway through development so that the irrigation treatments could be applied at the time when a decrease in temperature is effective in altering hatchling sex ratios (Porter et al., 2021).

A constant temperature equivalent (CTE; Georges et al., 1994) was then calculated using hourly nest temperature data during the sex determining period (SDP) for each nest, and used to calculate the predicted hatchling sex ratio for each nest from the relationship between incubation temperature and the proportion of female hatchlings produced at different temperatures. Briefly, a quadratic model was used to predict the hourly embryonic development based on temperature readings (Booth et al., 2022). The product of the incubation temperature and the amount of development at the temperature is summed during the sex determining period to calculate a constant temperature equivalent. The actual sex ratio of each nest was determined from histological examination of the gonads of 10 hatchlings from each nest (Young et al., 2023).

2.3 Gonad and liver sample collection

Once hatchlings had emerged from their nests, they were euthanised by placing them inside a sealed container for 1 h with 2-4 mL of isoflurane (gaseous anesthetic). Once deceased, each hatchling was dissected, and both left and right kidney-gonad complexes were removed. Gonad histology was carried out as described in Young et al. (in press) for sex identification. Hatching success was also calculated for each clutch post-emergence.

Hatchling liver samples (0.5 - 1 g wet weight) were dissected from 16 of the clutches (no hatchlings from clutch 7 were collected) that

TABLE 1 Summary data for each green turtle clutch including mean sex determining period (SDP) temperature, the predicted % females (based on constant temperature equivalents), the measured % females (based on histological examination of gonads), and the percent sex ratio deviation of the clutch from predicted % females, with positive percentages indicating more females than expected.

Clutch	Treatment	Mean SDP temp (°C)	Predicted % females	Measured % females	Sex ratio deviation
1	Control	28.18	81%	100%	19%
2	Freshwater	27.30	0%	90%	90%
3	Cold Seawater	28.14	73%	60%	-13%
4	Seawater	27.43	0%	0%	0%
5	2x Seawater	27.18	0%	20%	20%
6	Control	28.24	89%	80%	-9%
8	Cold Seawater	28.14	73%	40%	-33%
9	Seawater	27.67	1%	36%	35%
10	2x Seawater	27.37	0%	30%	30%
11	Control	27.93	22%	70%	48%
12	Freshwater	27.47	0%	60%	60%
13	Cold Seawater	27.45	0%	90%	90%
14	Seawater	27.51	0%	20%	20%
15	2x Seawater	27.50	0%	60%	60%
16	Control	28.26	92%	100%	8%
17*	Cold Seawater	30.84	100%	100%	0%

*As noted, clutch 17 was excluded from further analysis as it was identified as an outlier.

were sacrificed for histological examination of gonads. Liver samples were then partitioned into two halves, one stored in an Eppendorf tube for trace element analysis, and one wrapped in aluminium foil and stored in an Eppendorf tube for organic contaminant analysis. Liver samples were freeze dried (Christ Alpha 1-4 LSCbasic, Germany) for 12 h, and kept frozen at -20°C until analysis.

2.4 Trace element analysis by ICP-MS

Inductively coupled plasma mass spectrometry (ICP-MS) was used to analyse 22 trace elements (see [Supplementary Material; Table 6.6](#)) in 78 hatchling liver samples, or roughly five samples per clutch. Briefly, liver samples (~0.25 g dry weight, accurately weighed) were microwave-digested (CEM Mars 6; model 910905; EPA method 3052) with 4.5 mL nitric acid, 1 mL hydrochloric acid, and 1 mL hydrogen peroxide within pressurised Teflon vessels. Sample batches included laboratory blanks (MilliQ water) and dogfish liver standard reference material DOLT-5 (National Research Council Canada). Digests were diluted 1:50 in MilliQ water, and 100 µL was injected into the ICP-MS (Agilent Technologies; 7900). Metal concentrations were analysed using Agilent Chemstation software (ICP-MS MassHunter version 4.3). Trace elements concentrations were blank adjusted within batches with the mean laboratory blank concentrations. All trace elements were within 85 – 115% of DOLT-5 certified values with less than 5% standard error.

2.5 Organic chemical analysis by *in vitro* bioassay

Organic contaminants were extracted via a QuEChERS (quick, easy, cheap, effective, rugged, and safe) method used in previous research with sea turtles ([Dogruer et al., 2018](#); [Finlayson et al., 2020](#)). The QuEChERS method extracts a wide range of organic contaminants, such as polycyclic aromatic hydrocarbons (PAHs), polychlorinated biphenyls (PCBs), and polybrominated diphenyl ethers (PBDEs) ([Anastassiades et al., 2003](#); [Escher et al., 2021](#)). This method has previously been used to assess organic contaminant loads in foraging green turtles along the eastern Australia coast ([Finlayson et al., 2022](#)).

Briefly, freeze dried liver samples (n = 78) were thawed and homogenised in 50 mL Falcon tubes with 10 mL of MilliQ water, using a laboratory hand blender. The hand blender was solvent-cleaned with acetone and water between each sample. Two ceramic homogenisers were added to centrifuge tubes and vortexed for 1 min, followed by addition of 15 mL of acetonitrile and further vortexing for 1 min. Five grams of magnesium sulfate and 1 g of sodium chloride were added to the Falcon tube and shaken for 1 min. Mixtures were then centrifuged at 1990×g for 8 min at 10°C. The supernatant was aliquoted into glass vials and evaporated under a gentle stream of nitrogen gas. Dried extracts were reconstituted in 1 mL methanol and transferred into glass amber vials with Teflon lids. Extracts were stored at -20°C until use in the cytotoxicity bioassay.

The toxicity of the liver extracts (containing a mixture of organic chemicals) to green turtle primary skin fibroblasts was measured using a resazurin cell viability assay, following previously established methods ([Finlayson et al., 2019a](#); [Finlayson et al., 2021](#)). Briefly, green turtle primary skin fibroblasts (GT10s-p), established in [Finlayson et al. \(2019b\)](#) were grown in RPMI-1640 medium with 10% foetal bovine serum (FBS) and incubated at 30°C with 5% CO₂. A subsample of each methanol extract was evaporated using nitrogen gas and reconstituted in RPMI-1640 medium with 5% FBS, resulting in a relative enrichment factor of 2, corresponding to twice the contaminant concentration in liver samples ([Finlayson et al., 2020](#)).

A flat clear bottom 96-well microtiter plate was seeded with 3×10⁴ cells per well and incubated for 24 h at 30°C with 5% CO₂. After 24 h, a separate dosing plate was used to prepare extracts in a serial dilution (2-fold, 4 point) along with: Triton X-100 (0.1%) as a positive control, chromium in a 6-point, 2-fold dilution as a reference compound, and a negative control of RPMI medium (supplemented with 5% FBS). Medium from the seeded plate was then replaced with medium from the dosing plate (containing samples and controls) for a 24-h exposure. All controls and extracts were added to the seeded plate in duplicate. After 24 h, 20 µL of 0.15 mg/mL resazurin in phosphate buffered saline and 80 µL of RPMI medium (5% FBS) were added to each well for a further 24-h incubation. Following resazurin incubation, each well's fluorescence was measured at λ_{ex} = 544 nm and λ_{em} = 590 nm using a Spark plate reader (Tecan, Switzerland). Each plate analysed 10 extracts/controls, and extracts were tested in at least two independent assay runs on separate days.

The percent viability of cells for each sample/reference dilution with an assay was calculated by Eq (1). F signifies mean fluorescence of sample/reference duplicates, NC signifies mean fluorescence of the negative control duplicates, and PC signifies mean fluorescence of the positive control duplicates.

$$Viability (\%) = \left(1 - \left(\frac{F - NC}{PC - NC}\right)\right) * 100 \quad (1)$$

Mean percent viability (from multiple assay runs) of the sample/reference dilutions were plotted against the log concentration of sample/reference (expressed as relative enrichment factor) in GraphPad Prism 5 (GraphPad Software Inc), and the Hill Slope equation was used to calculate the relative enrichment factor (REF) required to produce 20% decrease in cell viability (IC₂₀) of each sample, and the reference compound. The results were then expressed as a Toxic Unit (TU), the reciprocal of the IC₂₀. A high TU therefore indicates greater toxicity, and TU = 1 indicates that contaminant concentrations present in the sample decreased cell viability by exactly 20% (at a relative enrichment factor of 1).

2.6 Statistical analysis

An α of 0.05 was used for statistical analyses in R (v. 4.2.0) with RStudio (v. 2022.06.0) and GraphPad Prism 5 (GraphPad Software Inc). Trace elements below limits of detection were replaced by

detection frequency times the limit of detection, which was calculated as three times the standard deviation of laboratory blanks. Kaplan-Meier adjustments were not possible with these data as most elements measured were either all above detection limits or had >70% data below LOD (Helsel, 2012). Elements with over 70% below LOD include caesium and mercury. Sex ratio deviation (%) was calculated as: measured % females – predicted % females, with a positive sex ratio deviation indicating more females than predicted.

To assess the potential impact of pollutants on skewing hatchling clutch sex ratio, several parameters were calculated to represent estrogenic potential or overall contamination levels: 1) predicted 17 β -estradiol equivalent activity (PEEQA), 2) percent trace elements over median (TEOM), and 3) percent estrogenic elements over median (EstroEOM). Choe et al. (2003) tested 28 metals in a reporter gene assay for estrogenic activity (MCF7-ERE) and reported the relative estrogenic potencies for four elements: antimony (1.62×10^{-3}), chromium (7.68×10^{-4}), cadmium (2.45×10^{-4}), and barium (3.57×10^{-5}). Estrogenic potencies were determined as concentrations needed to achieve the same effect as 17 β -estradiol to MCF 7-ERE human breast cancer cells (Choe et al., 2003). These relative potencies were multiplied by the concentration of the elements measured in hatchling livers and summed to calculate a PEEQA (in ng 17 β -estradiol per g) for each hatchling. For the second measure (TEOM), the number of trace elements per hatchling over the median of all samples was summed for each clutch then divided by the number of elements and multiplied by 100%. This value was calculated to highlight unusually high contaminant loads in a clutch. For example, if all five hatchlings in a clutch had 10 out of 20 elements over the median of those elements across the corresponding clutch samples, that would correspond to a TEOM value of $(5 \text{ [hatchlings]} \times 10 \text{ [out of 20 elements]}) / 100 \text{ [representing a maximum of five hatchlings having all 20 elements]} * 100\% = 50\%$. A similar method was used to calculate the last measure (EstroEOM), except that only four estrogenic elements (chromium, barium, antimony, and cadmium) over the median of all hatchlings was calculated for each clutch. For example, if all five hatchlings in a clutch had three of the four estrogenic elements over the median of those elements across all samples, that would result in a EstroEOM value of $(5 \text{ [hatchlings]} \times 3 \text{ [out of four elements]}) / 20 \text{ [representing a maximum of five hatchlings with all 4 elements]} * 100\% = 75\%$.

To compare the relationship between non-parametric means of individual trace elements, a generalised additive model was used to calculate the relationship between these calculated parameters and sex ratio or sex ratio deviation. To offset the influence of temperature, clutch 17, identified as an outlier by Grubbs test, was excluded from generalised additive models between trace elements and sex ratio. Spearman's correlations were used to assess the relationship between mean clutch contaminants and clutch hatching success.

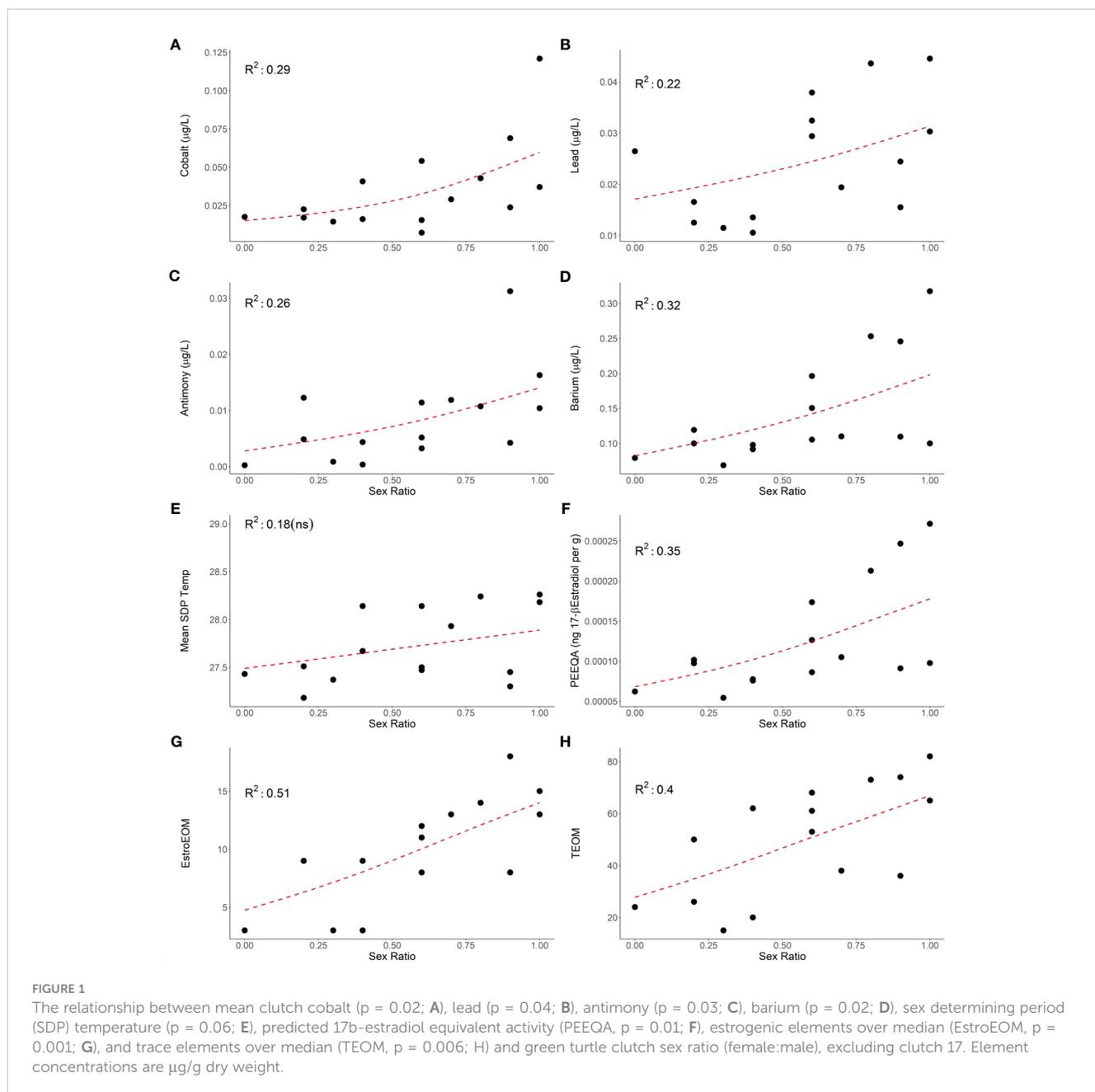
3 Results and discussion

This study found that contaminant loads in hatchling livers had significant effects on sex ratio and sex ratio deviation. Of the 16

clutches sampled, eleven clutches had more females than predicted (8 – 90% more females), three had more males (9 – 33% more males), and two matched the predicted sex ratio (Table 1). As cobalt, lead, antimony, barium, PEEQA, and EstroEOM increased, so did the female ratio of the clutch (Figure 1). Lead, chromium, antimony, barium, and cadmium have been previously studied for their endocrine disrupting effects on animal reproductive health, especially in mammals and fish (Adeel et al., 2017; Lecomte et al., 2017). A previous study found these elements to be estrogenic in a human estrogen reporter gene assay, particularly barium, antimony, chromium, and cadmium (Choe et al., 2003). Antimony is commonly found in waterways due to plastic bottle degradation (Mihucz and Zárny, 2016), and has been found in higher concentration in blood of turtles foraging near agriculture centers in Eastern Australia (Villa et al., 2016). Barium exposure in zebrafish caused increased production of the female-specific hormone estradiol in male fish (Kwon et al., 2016). Cadmium and chromium have been extensively studied for their negative reproductive effects on fish (Chakraborty, 2021), including reducing sperm quality in pejerrey (*Odontesthes bonariensis*) fish (Gárriz and Miranda, 2020). Cadmium exposure in pond sliders (*Trachemys scripta*) has been shown to alter the transcription of *dmrt1* and *aromatase* (Mizoguchi et al., 2022), both of which are important for sex determination. Other trace elements, such as lead, have been directly linked to endocrine disruption in sea turtles, with a previous study finding that increasing concentrations of lead inhibited estrogen and testosterone binding in plasma of nesting green turtles (Ikonomopoulou et al., 2009).

The relationship between sex ratio deviation and contaminants was less evident in the current study. Cadmium had a significant positive relationship with sex ratio deviation, while PEEQA and EstroEOM trended with sex ratio deviation but did not show a significant relationship (Figure 2). The current study used previous research to calculate relative estrogen potency, but the same study demonstrated that different metal species can have different relative estrogenic potencies (Choe et al., 2003). The methods used in the current study cannot account for these differences and the potential for mixture effects with organic compounds. In addition, previous research has found that the ability for organic compounds such as PCBs to inhibit estrogen receptor affinity can vary across mammalian and reptilian models (Matthews and Zacharewski, 2000). However, cadmium, a known endocrine disrupting compound, did have a significant relationship with sex ratio deviation, highlighting the need for future research investigating how contaminants could cause disruptions to sex determination in sea turtle hatchlings, as they could have population-level effects.

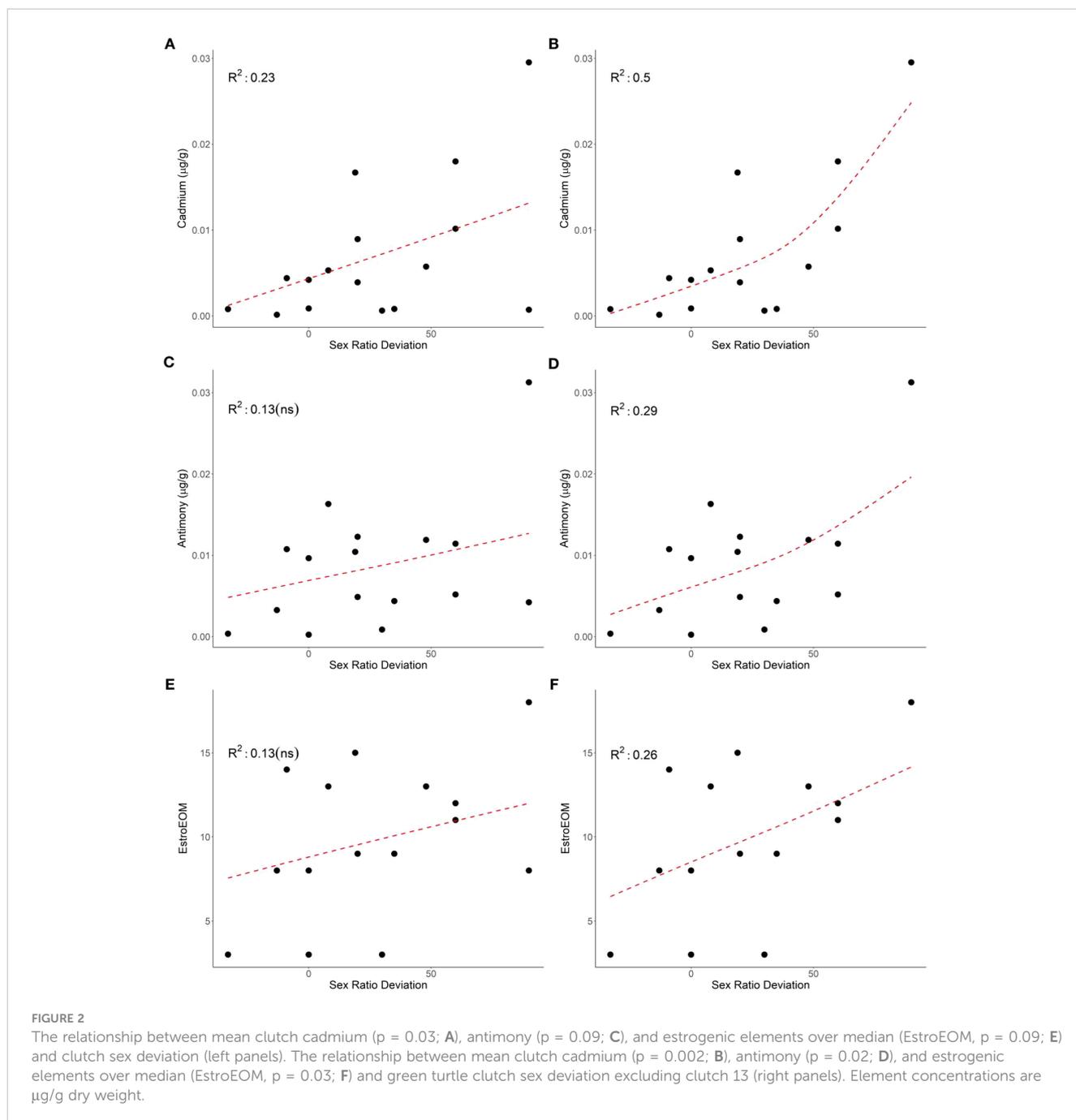
Multiple reptilian studies have shown that contaminants can override temperature and potentially impact reproductive output of TSD animals (Barraza et al., 2021). In previous research, freshwater turtle and crocodilian studies found contaminant-driven sex determination disruption, but direct evidence of contaminant effects on the sex determination of sea turtles has been elusive (Barraza et al., 2021), with a single study showing no effect on sex ratios when dosing with DDE (Podreka et al., 1998). However, synergistic effects have been found in pond sliders as eggs dosed with PCB mixtures produced more female-bias clutches than with



single PCB congener dosing (Bergeron et al., 1994). PCB dosing in another pond slider study found that some PCB congeners had no effect on TSD, while other congeners acted synergistically to disrupt sex determination (Matsumoto et al., 2014). In a model using snapping turtle (*Chelydra serpentina*) data, endocrine disruption via PCBs caused small variations in predicted pivotal temperature and sex determination towards feminization (Salice et al., 2014). Additional modeling from the same study suggested that even small increases in feminization could lead to population declines in long-lived species (Salice et al., 2014). Conversely, if contaminants masculinise hatchlings, they may provide some resistance to feminization as climate change increases temperature, but at the potential cost of individual health. There has been some research demonstrating that exogenous estrogens and higher temperatures produce more females pond sliders than estrogen dosing alone

(Wibbels et al., 1991). Therefore, as nesting beach temperatures increase across the globe, maternally transferred estrogenic contaminants could further exacerbate the feminizing effect of increased temperature. Exogenous estrogen dosing has also been found to have no effect on sex determination during extreme temperature events, but a significant effect at milder temperatures in lizards (Warner et al., 2017). For sea turtles, these studies suggest that nesting females with high contaminant loads nesting on beaches with mild temperatures may have the highest risk of altered sex ratios. This becomes especially important as beaches with mild temperatures are expected to offset the further increasing female bias already seen in northern GBR nesting beaches (Jensen et al., 2018).

In this study, cell-based methods were used to measure the overall effect of the complex mixture of organic contaminants on green turtle cell viability, instead of attempting to analyse individual



organic contaminants. We did not find a significant relationship between the organic contaminant load (as measured by cell viability) and sex ratio ($p = 0.5$) or sex ratio deviation ($p = 0.4$) (data not shown). However, organic contaminant load differences were found between males and females from the same clutch in some comparisons. Including all clutches, significantly higher toxicity (TU_{IC20}) was measured in extracts from male compared to female hatchlings ($p = 0.02$). Of the eight hatchlings with a $TU > 1$, seven (88%) were male. These data indicate that among the clutches examined, male hatchlings had higher organic contaminant concentrations in the liver than female hatchlings, and in some cases, high enough to cause cellular damage (as measured by our cytotoxicity assay).

It is important to note that not all contaminant comparisons with hatchling sex ratio or sex ratio deviation matched expectations. In addition, many of the clutches had individuals that were outliers in one or two elements compared to other individuals within their clutch. Typically, contaminant loads tend to be similar among hatchlings from the same clutch (Guirlet et al., 2010; Páez-Osuna et al., 2010; van de Merwe et al., 2010; Perrault et al., 2011; Perrault et al., 2016). There were no significant relationships between any contaminant and size (mass or carapace length), temperature, or nest irrigation treatment group. Outlier concentrations were not consistently from the same clutch and only confined to one or two elements, reducing the chance that these data are due to sampling or laboratory errors. These data may demonstrate that hatchlings

contaminant loads differ within a clutch, potentially due to differences in egg contamination. One particularly unusual clutch, clutch 13, had high sex deviation but low estrogenic and overall trace element concentrations (Table 1; Figure 2). Removing clutch 13 from sex ratio deviation comparisons resulted in cadmium ($p = 0.002$), antimony ($p = 0.02$), and EstroEOM ($p = 0.03$) having a significant positive relationship with sex ratio deviation (Figure 2). Many organic contaminants or endocrine disruptors can be far more estrogenic than trace elements (Choe et al., 2003; Leusch et al., 2009), and estrogen receptor binding of contaminants can vary by species (Matthews and Zacharewski, 2000; Matthews et al., 2000). However, our measure of organic contaminants did not correlate to sex ratio deviation. One possibility for this pattern is that the estrogenic effects of organic contaminants is not necessarily associated with cytotoxicity, but rather driven by a few specific and highly potent estrogenic mimics. Estrogenic chemicals induce an estrogenic response at concentrations well below cytotoxic concentrations, and it is therefore not possible to estimate the estrogenic activity associated with organic contaminants in hatchling livers from the cytotoxicity results. As make-up of the organic contaminants in hatchling samples can vary based on maternal contaminant loads, assessing the estrogenic potential from organic contaminants could be beneficial for future research. Regardless, cadmium, antimony, cobalt, lead, barium, TEOM, EstroEOM, and PEEQA had significant relationships with clutch sex ratio, while mean sex determining period temperature did not (Figure 1E). There have been studies that have found synergistic relationships between temperature and contaminant accumulation, such as in amphipods (Jacobson et al., 2007). It is possible that sea turtles may have inherent differences in male and female trace element accumulation due to incubation temperature. However, our data suggest that, at least in a temperature range where a mixed sex ratio can develop, some contaminants have a stronger correlation to sex ratio than temperature. Studies find contaminants in eggshell and albumen. There may be preferential maternal transfer in certain tissues such as liver, which is often used to assess organic contaminants due to its high lipid content. Considering the relationships between other contaminants and sex ratio deviation, these data indicate that there may be a potential for contaminants to influence hatchling sex determination.

Finally, it is important to place these hatchling liver contaminant concentrations in the context of sea turtle health, more generally. Many of the trace elements and toxicity measures have been studied for their negative health effects in sea turtles. Few studies have examined hatchling livers for trace elements with two studies examining green turtle hatchlings in Oman (Al-Rawahy et al., 2007) and Saudi Arabia (Tanabe et al., 2022), and one study examining loggerhead hatchlings in Turkiye (Kaska and Furness, 2001). Most trace elements, except selenium, were similar or less than the concentrations found in these other studies (Supplementary Material; Table 6.7 and 6.8). In green turtles, barium was correlated with reduced hatching success (Souza et al., 2018), and vanadium has been negatively correlated with hatching success in leatherback turtles (Dennis et al., 2020). In this study, only iron negatively correlated ($p = 0.03$, $\rho = -0.54$) with hatching success. To date, there have not been any studies finding a

negative correlation between iron and hatchling success in green turtles. However, while iron is an essential element, there are studies that have found excess iron to cause reduced hatching success in freshwater insects and lampreys (Myllynen et al., 1997; Rousch et al., 1997). Organic contaminants such as persistent organic pollutants have been correlated with reduced hatchling size in green turtles (van de Merwe et al., 2010). Liver extracts from male hatchlings exhibited significantly higher cytotoxicity than female, often above the threshold of 1 TU (suggestive of cellular toxicity at the concentrations detected), indicating that either they might be more vulnerable to contaminants or that male livers contain more cytotoxic endogenous compounds. The current study is the first to assess hatchling organic contaminants with an effects-based method, using primary turtle cell bioassays. Out of 78 hatchling liver samples tested, only one sample did not induce a cytotoxic response in green turtle primary skin fibroblasts. Compared to previous research in foraging immature green turtles throughout Queensland (Finlayson et al., 2021), hatchling liver samples had toxic unit values comparable to blood extracts from turtles foraging within foraging grounds throughout Queensland. Previous research has shown that maternal organic contaminants are transferred to offspring (van de Merwe et al., 2010), with less lipophilic PCB congeners being transferred more readily. Potentially, less lipophilic contaminants are more cytotoxic within green turtle fibroblasts than more lipophilic contaminants. However, the *in vitro* cytotoxicity of two similarly lipophilic compounds, 4,4'-dichlorodiphenyldichloroethylene (4,4'-DDE) and perfluorononanoic acid (PFNA), were compared within green turtle, loggerhead turtle, dugong, Burrunan dolphin, and common bottlenose dolphin cell lines (Finlayson and van de Merwe, 2021). There were species differences in cytotoxicity in PFNA, but not 4,4'-DDE, and PFNA was less cytotoxic in green turtle cells than 4,4'-DDE. These data suggest that there is likely wide variation in *in vitro* cytotoxicity between compounds, and that variation is likely to increase when accounting for mixture effects and differences in compounds that are maternally transferred.

4 Conclusions

While the current study cannot point to a specific contaminant that causes feminizing in sea turtle hatchlings, this study demonstrates that chemical mixtures can possibly affect sex ratio and sex ratio deviation, trending towards feminization. Considering the relatively low sample size of the current study (16 nests), finding significant positive relationships to sex ratio and sex ratio deviation with contaminant mixtures within hatchling sea turtles suggests that contaminants may affect sex determination. Cadmium was the only element measured with a moderate relationship with sex ratio deviation, and cadmium has been shown to alter the transcription of two genes important for sex determination in freshwater turtles (Mizoguchi et al., 2022). However, it is likely that other (unmeasured) contaminants may also contribute to the effects reported here. As contaminant mixtures and temperature can have interactive estrogenic effects, the risk for hatchling feminization from contaminants can be a combination of maternal contaminant

exposure from foraging and nesting site temperatures. *In vitro* bioassays were a useful tool for assessing organic contaminant exposure in hatchling organ tissue. Many of the contaminants found in the current study have demonstrated negative health effects in previous research among multiple sea turtle species. Considering the risks of feminization due to climate change and the potential negative health effects of anthropogenic contaminants, the possibility that hatchling development is disrupted by anthropogenic contaminants should be investigated further.

Data availability statement

The original contributions presented in the study are included in the article/Supplementary Material. Further inquiries can be directed to the corresponding author.

Ethics statement

The animal study was approved by Griffith University Animal Ethics Committee. The study was conducted in accordance with the local legislation and institutional requirements.

Author contributions

AB, KF, FL, and JM contributed to the writing, and conception of this submission. AB contributed to the data analysis. LY, CS, CH, and DB contributed to the data collection of this submission. All authors contributed to the article and approved the submitted version.

Funding

AB was supported by the Griffith University Postgraduate Research Scholarship, the Griffith University International

Postgraduate Research Scholarship, and the Griffith University Completion Assistance Postgraduate Research Scholarship. We thank the World Wildlife Fund for Nature – Australia (WWF-AU) and its donor, [Koala.com](https://www.koala.com), for funding this study as part of the WWF-led “Turtle Cooling Project”.

Acknowledgments

The authors would like to thank Melissa Staines for their assistance in the field as well as Natali Marinovski, and the various Griffith University volunteers for their assistance in the laboratory.

Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

Publisher's note

All claims expressed in this article are solely those of the authors and do not necessarily represent those of their affiliated organizations, or those of the publisher, the editors and the reviewers. Any product that may be evaluated in this article, or claim that may be made by its manufacturer, is not guaranteed or endorsed by the publisher.

Supplementary material

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fmars.2023.1238837/full#supplementary-material>

References

- Abraham, J. P., Baringer, M., Bindoff, N. L., Boyer, T., Cheng, L. J., Church, J. A., et al. (2013). A review of global ocean temperature observations: Implications for ocean heat content estimates and climate change. *Rev. Geophys.* 51, 450–483. doi: 10.1002/rog.20022
- Adeel, M., Song, X., Wang, Y., Francis, D., and Yang, Y. (2017). Environmental impact of estrogens on human, animal and plant life: A critical review. *Environ. Int.* 99, 107–119. doi: 10.1016/j.envint.2016.12.010
- Al-Rawahy, S. H., AlKindi, A. Y., Elshafie, A., Ibrahim, M., Al Bahry, S. N., Al Siyabi, S. S., et al. (2007). Accumulation of metals in the egg yolk and liver of hatchling of green turtles *Chelonia mydas* at Ras Al Hadd, Sultanate of Oman. *J. Biol. Sci.* 7, 925–930. doi: 10.3923/jbs.2007.925.930
- Anastasiades, M., Lehotay, S. J., Štajnbaher, D., and Schenck, F. J. (2003). Fast and easy multiresidue method employing acetonitrile extraction/partitioning and “dispersive solid-phase extraction” for the determination of pesticide residues in produce. *J. AOAC Int.* 86, 412–431. doi: 10.1093/jaoac/86.2.412
- Barraza, A. D., Finlayson, K. A., Leusch, F. D. L., and van de Merwe, J. P. (2021). Systematic review of reptile reproductive toxicology to inform future research directions on endangered or threatened species, such as sea turtles. *Environ. Pollut.* 286, 117470. doi: 10.1016/j.envpol.2021.117470
- Beldomenico, P. M., Rey, F., Prado, W. S., Villarreal, J. C., Muñoz-de-Toro, M., and Luque, E. H. (2007). In ovum exposure to pesticides increases the egg weight loss and decreases hatchlings weight of *Caiman latirostris* (Crocodylia: Alligatoridae). *Ecotoxicol. Environ. Saf.* 68, 246–251. doi: 10.1016/j.ecoenv.2006.12.018
- Bergeron, J. M., Crews, D., and McLachlan, J. A. (1994). PCBs as environmental estrogens: turtle sex determination as a biomarker of environmental contamination. *Environ. Health Perspect.* 102, 780–781. doi: 10.1289/ehp.94102780
- Blechs Schmidt, J., Wittmann, M. J., and Blüml, C. (2020). Climate change and green sea turtle sex ratio—preventing possible extinction. *Genes (Basel)*. 11. doi: 10.3390/genes11050588
- Boggs, A. S. P., Botteri, N. L., Hamlin, H. J., and Guilette, L. J. (2011). “Endocrine disruption of reproduction in reptiles,” in *Hormones and Reproduction of Vertebrates*. 3, 373–396. doi: 10.1016/B978-0-12-374930-7.10014-7
- Booth, D. T., and Astill, K. (2001). Incubation temperature, energy expenditure and hatchling size in the green turtle (*Chelonia mydas*), a species with temperature-sensitive sex determination. *Aust. J. Zool.* 49, 389–396. doi: 10.1071/ZO01006
- Booth, D. T., Burgess, E., McCosker, J., and Lanyon, J. M. (2004). The influence of incubation temperature on post-hatching fitness characteristics of turtles. *Int. Congr. Ser.* 1275, 226–233. doi: 10.1016/j.ics.2004.08.057

- Booth, D. T., Turner, A. G., Laloë, J., and Limpus, C. J. (2022). How well do embryo development rate models derived from laboratory data predict embryo development in sea turtle nests? *J. Exp. Zool. Part A. Ecol. Integr. Physiol.* 337, 516–526. doi: 10.1002/jez.2585
- Bowen, B. W., Meylan, A. B., Ross, J. P., Limpus, C. J., Balazs, G. H., and Avise, J. C. (1992). Global population structure and natural history of the green turtle (*Chelonia mydas*) in terms of matriarchal phylogeny. *Evol. (N. Y.)* 46, 865–881. doi: 10.2307/2409742
- Burgess, E. A., Booth, D. T., and Lanyon, J. M. (2006). Swimming performance of hatchling green turtles is affected by incubation temperature. *Coral Reefs* 25, 341–349. doi: 10.1007/s00338-006-0116-7
- Bustard, R., and Greenman, P. (1967). Physical and chemical factors affecting hatching in the green sea turtle, *Chelonia mydas* (L.). *Ecol. Soc. Am.* 49, 269–276. doi: 10.2307/1934455
- Canesini, G., Stoker, C., Galoppo, G. H., Durando, M. L., Tschopp, M. V., Luque, E. H., et al. (2018). Temperature- vs. estrogen-induced sex determination in *Caiman latirostris* embryos: Both females, but with different expression patterns of key molecules involved in ovarian development. *Gen. Comp. Endocrinol.* 259, 176–188. doi: 10.1016/j.ygcen.2017.11.024
- Chakraborty, S. B. (2021). Non-essential heavy metals as endocrine disruptors: evaluating impact on reproduction in teleosts. *Proc. Zool. Soc* 74, 417–431. doi: 10.1007/s12595-021-00399-x
- Choe, S. Y., Kim, S. J., Kim, H. G., Lee, J. H., Choi, Y., Lee, H., et al. (2003). Evaluation of estrogenicity of major heavy metals. *Sci. Total. Environ.* 312, 15–21. doi: 10.1016/S0048-9697(03)00190-6
- Cortes-Gomez, A. A., Romero, D., and Girondot, M. (2017). The current situation of inorganic elements in marine turtles: A general review and meta-analysis. *Environ. Pollut.* 229, 567–585. doi: 10.1016/j.envpol.2017.06.077
- Crain, D. A., Guillette, L. J., Rooney, A. A., and Pickford, D. B. (1997). Alterations in steroidogenesis in alligators (*Alligator mississippiensis*) exposed naturally and experimentally to environmental contaminants. *Environ. Health Perspect.* 105, 528–533. doi: 10.1289/ehp.97105528
- Dennis, M. M., Poppenga, R., Conan, A., Hill, K., Hargrave, S., Maroun, V., et al. (2020). Leatherback sea turtle (*Dermochelys coriacea*) hatch success and essential and nonessential metals in eggs and embryos from nests in St. Kitts, (2015). *Mar. Pollut. Bull.* 161, 111726. doi: 10.1016/j.marpolbul.2020.111726
- Dogruer, G., Weijs, L., Tang, J. Y., Hollert, H., Kock, M., Bell, I., et al. (2018). Effect-based approach for screening of chemical mixtures in whole blood of green turtles from the Great Barrier Reef. *Sci. Total. Environ.* 612, 321–329. doi: 10.1016/j.scitotenv.2017.08.124
- Escher, B., Neale, P. A., and Leusch, F. D. L. (2021). *Bioanalytical Tools in Water Quality Assessment, Water Intelligence Online*. (New York and London: IWA publishing). doi: 10.2166/9781843393689
- Finlayson, K. A., Leusch, F. D. L., and van de Merwe, J. P. (2016). The current state and future directions of marine turtle toxicology research. *Environ. Int.* 94, 113–123. doi: 10.1016/j.envint.2016.05.013
- Finlayson, K. A., Leusch, F. D. L., Limpus, C. J., and van de Merwe, J. P. (2019a). Towards the development of standardised sea turtle primary cell cultures for toxicity testing. *Ecotoxicol. Environ. Saf.* 173, 63–70. doi: 10.1016/j.ecoenv.2019.01.117
- Finlayson, K. A., Leusch, F. D. L., and van de Merwe, J. P. (2019b). Cytotoxicity of organic and inorganic compounds to primary cell cultures established from internal tissues of *Chelonia mydas*. *Sci. Total. Environ.* 664, 958–967. doi: 10.1016/j.scitotenv.2019.02.052
- Finlayson, K. A., Leusch, F. D. L., Villa, C. A., Limpus, C. J., and van de Merwe, J. P. (2021). Combining analytical and *in vitro* techniques for comprehensive assessments of chemical exposure and effect in green sea turtles (*Chelonia mydas*). *Chemosphere* 274, 129752. doi: 10.1016/j.chemosphere.2021.129752
- Finlayson, K. A., Limpus, C. J., and van de Merwe, J. P. (2022). Temporal changes in chemical contamination of green turtles (*Chelonia mydas*) foraging in a heavily industrialised seaport. *Sci. Total. Environ.* 817, 152848. doi: 10.1016/j.scitotenv.2021.152848
- Finlayson, K. A., Madden Hof, C. A., and van de Merwe, J. P. (2020). Development and application of species-specific cell-based bioassays to assess toxicity in green sea turtles. *Sci. Total. Environ.* 747, 142095. doi: 10.1016/j.scitotenv.2020.142095
- Finlayson, K. A., and van de Merwe, J. P. (2021). Differences in marine megafauna *in vitro* sensitivity highlights the need for species-specific chemical risk assessments. *Aquat. Toxicol.* 239, 105939. doi: 10.1016/j.aquatox.2021.105939
- Fleming, K. A., Perrault, J. R., Stacy, N. I., Coppenrath, C. M., and Gainsbury, A. M. (2020). Heat, health and hatchlings: Associations of *in situ* nest temperatures with morphological and physiological characteristics of loggerhead sea turtle hatchlings from Florida. *Conserv. Physiol.* 8, 1–17. doi: 10.1093/conphys/coaa046
- Fuentes, M. M. P. B., Limpus, C. J., and Hamann, M. (2011). Vulnerability of sea turtle nesting grounds to climate change. *Glob. Change Biol.* 17, 140–153. doi: 10.1111/j.1365-2486.2010.02192.x
- Gardner, S. C. M., and Oberdorster, E. (2005). *Toxicology of Reptiles* (Boca Raton: CRC Press).
- Gárriz, Á., and Miranda, L. A. (2020). Effects of metals on sperm quality, fertilization and hatching rates, and embryo and larval survival of pejerrey fish (*Odontesthes bonariensis*). *Ecotoxicology* 29, 1072–1082. doi: 10.1007/s10646-020-02245-w
- Georges, A., Limpus, C., and Stoutjesdijk, R. (1994). Hatchling sex in the marine turtle *Caretta caretta* is determined by proportion of development at a temperature, not daily duration of exposure. *J. Exp. Zool.* 270, 432–444. doi: 10.1002/jez.1402700504
- Guirlet, E., Das, K., Thomé, J. P., and Girondot, M. (2010). Maternal transfer of chlorinated contaminants in the leatherback turtles, *Dermochelys coriacea*, nesting in French Guiana. *Chemosphere* 79, 720–726. doi: 10.1016/j.chemosphere.2010.02.047
- Hamann, M., Godfrey, M. H., Seminoff, J. A., Arthur, K., Barata, P. C. R., Bjørndal, K. A., et al. (2010). Global research priorities for sea turtles: Informing management and conservation in the 21st century. *Endanger. Species. Res.* 11, 245–269. doi: 10.3354/esr00279
- Hawkes, L. A., Broderick, A. C., Godfrey, M. H., and Godley, B. J. (2009). Climate change and marine turtles. *Endanger. Species. Res.* 7, 137–154. doi: 10.3354/esr00198
- Hays, G. C., Laloë, J. O., Lee, P. L. M., and Schofield, G. (2023). Evidence of adult male scarcity associated with female-skewed offspring sex ratios in sea turtles. *Curr. Biol.* 33, R14–R15. doi: 10.1016/j.cub.2022.11.035
- Helsel, D. R. (2012). *Statistics for Censored Environmental Data*. (Hoboken New Jersey: John Wiley & Sons, Inc.).
- Ikonomopoulou, M. P., Olszowy, H., Hodge, M., and Bradley, A. J. (2009). The effect of organochlorines and heavy metals on sex steroid-binding proteins *in vitro* in the plasma of nesting green turtles, *Chelonia mydas*. *J. Comp. Physiol. B.* 179, 653–662. doi: 10.1007/s00360-009-0347-3
- Jacobson, T., Prevodnik, A., and Sundelin, B. (2007). Combined effects of temperature and a pesticide on the baltic amphipod *Monoporeia affinis*. *Aquat. Biol.* 1, 269–276. doi: 10.3354/ab00028
- Jensen, M. P., Allen, C. D., Eguchi, T., Bell, I. P., LaCasella, E. L., Hilton, W. A., et al. (2018). Environmental warming and feminization of one of the largest sea turtle populations in the world. *Curr. Biol.* 28, 154–159 e4. doi: 10.1016/j.cub.2017.11.057
- Kaska, Y., and Furness, R. W. (2001). Heavy metals in marine turtle eggs and hatchlings in the mediterranean. *Zool. Middle. East.* 24, 127–132. doi: 10.1080/09397140.2001.10637891
- Keller, J. M. (2003). *Occurrence and effects of organochlorine contaminants in sea turtles* (University Microfilms).
- Kwon, B., Ha, N., Jung, J., Kim, P. G., Kho, Y., Choi, K., et al. (2016). Effects of barium chloride exposure on hormones and genes of the hypothalamic-pituitary-gonad axis, and reproduction of zebrafish (*Danio rerio*). *Bull. Environ. Contam. Toxicol.* 96, 341–346. doi: 10.1007/s00128-016-1731-9
- Lecomte, S., Habauzit, D., Charlier, T. D., and Pakdel, F. (2017). Emerging estrogenic pollutants in the aquatic environment and breast cancer. *Genes (Basel)* 8. doi: 10.3390/genes8090229
- Leusch, F. D. L., Moore, M. R., and Chapman, H. F. (2009). Balancing the budget of environmental estrogen exposure: the contribution of recycled water. *Water Sci. Technol.* 60, 1003–1012. doi: 10.2166/wst.2009.398
- Limpus, C. J., Milier, J. D., Parmenter, C. J., Reimer, D., McLachlan, N., and Webb, R. (1992). Migration of green (*Chelonia mydas*) and loggerhead (*Caretta caretta*) turtles to and from eastern Australian rookeries. *Wildl. Res.* 19, 347–357. doi: 10.1071/WR9920347
- Loarie, S. R., Duffy, P. B., Hamilton, H., Asner, G. P., Field, C. B., and Ackerly, D. D. (2009). The velocity of climate change. *Nature* 462, 1052–1055. doi: 10.1038/nature08649
- Lolavar, A., and Wyneken, J. (2021). Effects of supplemental watering on loggerhead (*Caretta caretta*) nests and hatchlings. *J. Exp. Mar. Bio. Ecol.* 534, 151476. doi: 10.1016/j.jembe.2020.151476
- Matsumoto, Y., Hannigan, B., and Crews, D. (2014). Embryonic PCB exposure alters phenotypic, genetic, and epigenetic profiles in turtle sex determination. A biomarker of environmental contamination. *Endocrinology* 155, 4168–4177. doi: 10.1210/en.2014-1404
- Mathews, J., Celiuș, T., Halgren, R., and Zacharewski, T. (2000). Differential estrogen receptor binding of estrogenic substances: a species comparison. *J. Steroid Biochem. Mol. Biol.* 74, 223–234. doi: 10.1016/S0960-0760(00)00126-6
- Mathews, J., and Zacharewski, T. (2000). Differential binding affinities of PCBs, HO-PCBs, and aroclors with recombinant human, rainbow trout (*Onchorhynchus mykiss*), and green anole (*Anolis carolinensis*) estrogen receptors, using a semi-high throughput competitive binding assay. *Toxicol. Sci.* 53, 326–339. doi: 10.1093/toxsci/53.2.326
- Meylan, A. B., Bowsen, B. W., and Avise, J. C. (1990). A genetic test of the natal homing versus social facilitation models for green turtle migration. *Mol. Ecol. Evol. Org. Side. Sel. Writings. Avis. Lab.* 248, 71–74. doi: 10.1142/9789814317764_0001
- Mihucz, V. G., and Záray, G. (2016). Occurrence of antimony and phthalate esters in polyethylene terephthalate bottled drinking water. *Appl. Spectrosc. Rev.* 51, 163–189. doi: 10.1080/05704928.2015.1105243
- Miller, J. D., and Limpus, C. J. (1981). *Incubation period and sexual differentiation in the green turtle Chelonia mydas (L.)*, in: *Proceedings of the Melbourne Herpetological Symposium* (Australia: The Zoological Board of Victoria Melbourne), 66–73.
- Mizoguchi, B., Topping, N. E., Lavin, A. M., and Valenzuela, N. (2022). Cadmium cotoxic effects on embryonic Dmrt1 and aromatase expression in chrysemys picta turtles may implicate changes in DNA methylation. *Genes (Basel)* 13. doi: 10.3390/genes13081318

- Myllynen, K., Ojutkangas, E., and Nikinmaa, M. (1997). River water with high iron concentration and low pH causes mortality of lamprey roe and newly hatched larvae. *Ecotoxicol. Environ. Saf.* 36, 43–48. doi: 10.1006/eesa.1996.1484
- Páez-Osuna, F., Calderón-Campuzano, M. F., Soto-Jiménez, M. F., and Ruelas-Inzunza, J. R. (2010). Lead in blood and eggs of the sea turtle, *Lepidochelys olivacea*, from the Eastern Pacific: Concentration, isotopic composition and maternal transfer. *Mar. Pollut. Bull.* 60, 433–439. doi: 10.1016/j.marpolbul.2009.10.004
- Page-Karjian, A., Stacy, N. I., Morgan, A. N., Coppenrath, C. M., Manire, C. A., Herbst, L. H., et al. (2022). Morphologic and physiologic characteristics of green sea turtle (*Chelonia mydas*) hatchlings in southeastern Florida, USA. *J. Comp. Physiol. B. Biochem. Syst. Environ. Physiol.* 192, 751–764. doi: 10.1007/s00360-022-01450-9
- Perrault, J. R., Bauman, K. D., Greenan, T. M., Blum, P. C., Henry, M. S., and Walsh, C. J. (2016). Maternal transfer and sublethal immune system effects of brevetoxin exposure in nesting loggerhead sea turtles (*Caretta caretta*) from western Florida. *Aquat. Toxicol.* 180, 131–140. doi: 10.1016/j.aquatox.2016.09.020
- Perrault, J., Wyneken, J., Thompson, L. J., Johnson, C., and Miller, D. L. (2011). Why are hatching and emergence success low? Mercury and selenium concentrations in nesting leatherback sea turtles (*Dermochelys coriacea*) and their young in Florida. *Mar. Pollut. Bull.* 62, 1671–1682. doi: 10.1016/j.marpolbul.2011.06.009
- Podreka, S., Georges, A., Maher, B., and Limpus, C. J. (1998). The environmental contaminant DDE fails to influence the outcome of sexual differentiation in the marine turtle *Chelonia mydas*. *Environ. Health Perspect.* 106, 185–188. doi: 10.1289/ehp.98106185
- Porter, E., Booth, D. T., Limpus, C. J., Staines, M. N., and Smith, C. E. (2021). Influence of short-term temperature drops on sex-determination in sea turtles. *J. Exp. Zool. Part A. Ecol. Integr. Physiol.* 335, 649–658. doi: 10.1002/jez.2509
- Rousch, J. M., Simmons, T. W., Kerans, B. L., and Smith, B. P. (1997). Relative acute effects of low pH and high iron on the hatching and survival of the water mite (*Arrenurus manubriator*) and the aquatic insect (*Chironomus riparius*). *Environ. Toxicol. Chem.* 16, 2144–2150. doi: 10.1897/1551-5028(1997)016<2144:RAEOLP>2.3.CO;2
- Salice, C. J., Rowe, C. L., and Eisenreich, K. M. (2014). Integrative demographic modeling reveals population level impacts of PCB toxicity to juvenile snapping turtles. *Environ. Pollut.* 184, 154–160. doi: 10.1016/j.envpol.2013.08.031
- Santidrián Tomillo, P., Genovart, M., Paladino, F. V., Spotila, J. R., and Oro, D. (2015). Climate change overruns resilience conferred by temperature-dependent sex determination in sea turtles and threatens their survival. *Glob. Change Biol.* 21, 2980–2988. doi: 10.1111/gcb.12918
- Santidrián Tomillo, P., and Spotila, J. R. (2020). Temperature-dependent sex determination in sea turtles in the context of climate change: uncovering the adaptive significance. *BioEssays* 42, 1–6. doi: 10.1002/bies.202000146
- Smith, C., Booth, D., Crosby, A., Miller, J., Staines, M., Versace, H., et al. (2021). Trialling seawater irrigation to combat the high nest temperature feminisation of green turtle *Chelonia mydas* hatchlings. *Mar. Ecol. Prog. Ser.* 667, 177–190. doi: 10.3354/meps13721
- Souza, N. L. N., Carneiro, M. T. W. D., Pimentel, E. F., Frossard, A., Freire, J. B., Endringer, D. C., et al. (2018). Trace elements influence the hatching success and emergence of *Caretta caretta* and *Chelonia mydas*. *J. Trace Elem. Med. Biol.* 50, 117–122. doi: 10.1016/j.jtemb.2018.06.007
- Tanabe, L. K., Scott, K., Dasari, V., and Berumen, M. L. (2022). An assessment of heavy metals in green sea turtle (*Chelonia mydas*) hatchlings from Saudi Arabia's largest rookery, Ras Baridi. *PeerJ* 10. doi: 10.7717/peerj.13928
- Tanner, C. E., Marco, A., Martins, S., Abella-Perez, E., and Hawkes, L. A. (2019). Highly feminised sex-ratio estimations for the world's third-largest nesting aggregation of loggerhead sea turtles. *Mar. Ecol. Prog. Ser.* 621, 209–219. doi: 10.3354/meps12963
- Valenzuela, N., and Lance, V. (2004). "Temperature-Dependent sex determination," in *Reptilian Incubation: Environment, Evolution, and Behavior*. (Smithsonian Books), 211–227.
- van de Merwe, J. P., Hodge, M., Olszowy, H. A., Whittier, J. M., Ibrahim, K., and Lee, S. Y. (2009). Chemical contamination of green turtle (*Chelonia mydas*) eggs in Peninsular Malaysia: Implications for conservation and public health. *Environ. Health Perspect.* 117, 1397–1401. doi: 10.1289/ehp.0900813
- van de Merwe, J. P., Hodge, M., Whittier, J. M., Ibrahim, K., and Lee, S. Y. (2010). Persistent organic pollutants in the green sea turtle *Chelonia mydas*: Nesting population variation, maternal transfer, and effects on development. *Mar. Ecol. Prog. Ser.* 403, 269–278. doi: 10.3354/meps08462
- van de Merwe, J., Ibrahim, K., and Whittier, J. (2006). Effects of nest depth, shading, and metabolic heating on nest temperatures in sea turtle hatcheries. *Chelonian. Conserv. Biol.* 5, 210–215. doi: 10.2744/1071-8443(2006)5[210:EONDSA]2.0.CO;2
- Villa, C. A., Flint, M., Bell, I., Hof, C., Limpus, C. J., and Gaus, C. (2016). Trace element reference intervals in the blood of healthy green sea turtles to evaluate exposure of coastal populations. *Environ. Pollut.* 220, 1465–1476. doi: 10.1016/j.envpol.2016.10.085
- Warner, D. A. (2011). *Sex determination in reptiles, in: Hormones and Reproduction of Vertebrates*. Vol. 3 (Elsevier), 1–38. doi: 10.1016/B978-0-12-374930-7.10001-9
- Warner, D. A., Mitchell, T. S., Bodensteiner, B. L., and Janzen, F. J. (2017). The effect of hormone manipulations on sex ratios varies with environmental conditions in a turtle with temperature-dependent sex determination. *J. Exp. Zool. Part A. Ecol. Integr. Physiol.* 327, 172–181. doi: 10.1002/jez.2085
- Wibbels, T., Bull, J. J., and Crews, D. (1991). Synergism between temperature and estradiol: A common pathway in turtle sex determination? *J. Exp. Zool.* 260, 130–134. doi: 10.1002/jez.1402600117
- Wyneken, J., and Lolavar, A. (2015). Loggerhead sea turtle environmental sex determination: Implications of moisture and temperature for climate change based predictions for species survival. *J. Exp. Zool. Part B. Mol. Dev. Evol.* 324, 295–314. doi: 10.1002/jez.b.22620
- Young, L. R., Booth, D. T., Smith, C. E., and Hof, C. A. M. (2023). Seawater irrigation on nests can increase male marine turtle production. *Mar. Ecol. Prog. Ser.* 718, 85–97. doi: 10.3354/meps14392