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Characteristics and ecological effects of dissolved organic matter released by the main causative macroalgae of green and golden tides in the yellow sea

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Green and golden tides caused by *Ulva prolifera* and *Sargassum horneri* erupt annually along the coasts of China. During the late stages of algal blooms, massive macroalgal die-offs result in the release of dissolved organic matter (DOM), which can profoundly affect local marine environments and elemental cycles. However, studies addressing these impacts remain insufficient and urgently needed. This study investigates the two primary macroalgal contributors, *Ulva prolifera* and *Sargassum horneri*, by simulating their decay and DOM release in seawater using Bacteria-active (BA) and Bacteria-inhibited (BI) groups. The release characteristics of carbon, nitrogen, and phosphorus, as well as their effects on microbial communities, were analyzed. Results show that *U. prolifera* and *S. horneri* rapidly release DOM and nutrients during their decay. Significant differences in the quantity and rate of DOM and nutrient release were observed due to structural differences between the two macroalgae. Carbohydrates constitute the major component of the released organic matter and are rapidly utilized by microbes for growth and reproduction. Furthermore, the DOM and nutrients released by macroalgae reshape the composition of microbial communities in the marine microenvironment. This study provides new insights into the potential regional marine ecological impacts of green and golden tide outbreaks from a biogeochemical perspective.

KEYWORDS

macroalgae, dissolved organic matter, nutrients, microbial communities, algal blooms

Introduction

The phenomenon of large-scale green tide outbreaks was first recorded in the early 1970s in the Brittany region (Schreyers et al., 2021). Since then, the distribution of green tides has progressively expanded to coastal regions across multiple continents, including Europe, Asia, America, and Australia (Yuan et al., 2022; Song et al., 2022). Moreover, golden tide events caused by *Sargassum* macroalgae have erupted on a large scale in regions such as West Africa, the Gulf of Mexico, the Brazilian coast, the Caribbean Sea, and the East China Sea, causing severe ecological damage, the main manifestations are nutrient release and increased eutrophication, biotoxicity and disruption of community structure (García-Robledo et al., 2012; Byeon et al., 2019; Gordillo Sierra et al., 2022). Since 2007, green tides dominated by *Ulva prolifera* have occurred consecutively for 17 years in the South Yellow Sea of China, characterized by their vast scale and severe impacts, significantly damaging the marine ecosystem (Zhou et al., 2015). In some localized areas of the South Yellow Sea, concurrent outbreaks of green and golden tides have been observed, intensifying the severity of these algal blooms. As a result, the outbreaks of green and golden tides have become a prominent marine environmental issue and a global ecological disaster (Xiao et al., 2021).

In recent years, the frequency, geographic extent, and adverse impacts of green and golden tides have shown an increasing trend, driven by rapid industrialization and population growth (Xiao et al., 2021). The outbreak of green tides sharply reduces light penetration into the marine environment, disrupting photosynthesis and leading to the mortality of benthic macroalgae (Van Alstyne et al., 2013; Feng et al., 2024). The outbreak of green tides sharply reduces light penetration into the marine environment, disrupting photosynthesis and leading to the mortality of benthic macroalgae (Van Alstyne et al., 2013; Feng et al., 2024). Additionally, dense algal biomass hinders the movement of plankton, suppressing the growth and reproduction of marine organisms (Du et al., 2023). Furthermore, the visual pollution and unpleasant odors associated with algal blooms adversely affect local tourism. At the later stages of blooms, as macroalgal masses decay, large quantities of organic matter and nutrients are rapidly released from the substantial biomass. This can exert detrimental effects on the regional ecological environment and disrupt carbon and nutrient cycles (Zhang and Wang, 2017; Xiu et al., 2019; Zhang et al., 2021). In addition, microbial utilization and remineralization of DOM is a central part of the marine biogeochemical cycle. Microorganisms remineralize organic carbon, nitrogen, and phosphorus into inorganic forms by degrading algal-derived DOM, affecting regional carbon sink capacity and eutrophication processes (Li et al., 2017; Zhang and Wang, 2017; Wang et al., 2024).

Current research on green and golden tides primarily focuses on the life cycles, reproductive strategies, growth conditions of the causative algae, and the factors driving these macroalgal outbreaks, including environmental variables such as water temperature, light availability, and nutrient concentrations (Gross et al., 2015; Wu et al., 2019, 2024). However, studies on the DOM released by the main macroalgae responsible for these tides (*Ulva prolifera* and *Sargassum horneri*) during the late decay stages of blooms, as well as

their effects on regional environmental factors and microbial communities, remain relatively limited (Zhao et al., 2018). This study simulates the dynamics of DOM release during the decay of two macroalgae species (*Ulva prolifera* and *Sargassum horneri*) and investigates the changes in microbial utilization of DOM, as well as its effects on the composition and abundance of microbial communities. The aim is to explore the potential impacts of green and golden tide outbreaks on regional marine ecosystems and elemental cycling.

Materials and methods

Sample collection and cultivation

The sampling sites' coordinates are shown in Figure 1. Macroalgae samples were collected during the routine marine environmental monitoring cruise in the South Yellow Sea, off Yancheng, Jiangsu Province, China, in June 2023. The sampling locations were designated as S1 (121°32'52"N; 34°17'53"E), S2 (122°0'55"N; 34°13'52"E), and S3 (122°25'37"N; 34°15'27"E). At each site, no less than 300 g of each of the two algae species was collected, along with 100 L of natural seawater. The collected algae were placed in sealed bags, stored in a cooler (4 °C), and transported back to the laboratory. Upon arrival, the macroalgae were thoroughly rinsed with Milli-Q water to remove attached sediments and other impurities. Excess surface water was removed using absorbent paper (80 × 25, Changde Bkmm Biotechnology Co., Ltd.), with the process repeated three times. For the natural seawater, impurities were removed by filtering the collected seawater through a 0.45 µm mixed cellulose fiber filter membrane (Tianjin Jinteng Experiment Equipment Co., Ltd.) to eliminate autotrophic microorganisms such as algae. Filtration may remove a small amount of bacteria; however, the process of removing particulate matter is still necessary because particulates in seawater could influence subsequent measurements and experimental results. After pretreatment, 20 g of *Ulva prolifera* and *Sargassum horneri* were individually placed in 4 L of natural seawater for cultivation as the bacteria-active group (BA). Simultaneously, a bacteria-inhibited group (BI) was established by adding microbial inhibitors (5 mL of saturated mercuric chloride solution), alongside a blank seawater control group without algae. All seawater used in this experiment was *in-situ* collected during routine monitoring cruises in the South Yellow Sea, and remained unmodified to preserve its natural physicochemical properties. Each treatment (BA: algae-seawater co-culture, BI: algae-seawater with microbial inhibitor, and blank control: seawater only) was established with two independent biological replicates per algal species (*U. prolifera* and *S. horneri*), yielding a total of 12 sample replicates (2 species × 3 treatments × 2 replicates) (Zhang and Wang, 2017; Zhang et al., 2020). The macroalgae and pretreated natural seawater were placed in 5 L glass Erlenmeyer flasks, covered with gas-permeable oxygen-sealing polypropylene membranes to prevent contamination while ensuring adequate air exchange. The flasks were incubated in a culture chamber (model LRH-70, Rex (Shanghai) Technology Co.,

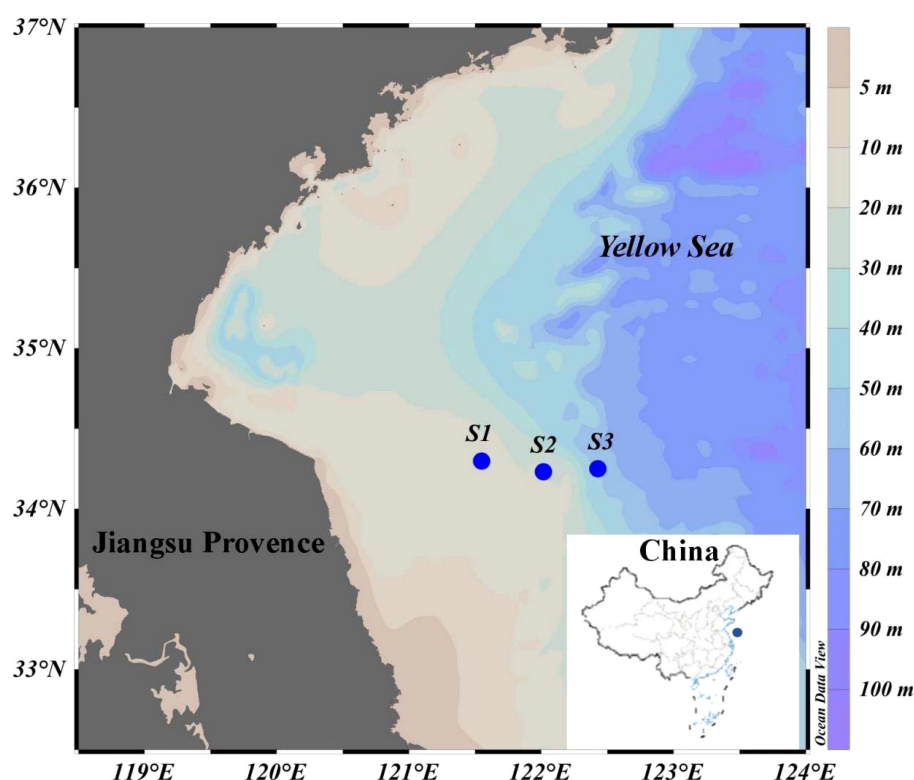


FIGURE 1
Geographical position of the sampling sites.

Ltd.) at a constant temperature of 25°C under dark conditions for 150 days. All glassware used for sample handling and cultivation was pre-cleaned by acid washing, rinsed with Milli-Q water, and combusted at 550°C for 5 hours to eliminate organic impurities.

Chemical analysis

At specific intervals during the cultivation period (0, 1, 2, 3, 5, 7, 9, 11, 13, 15, 18, 21, 24, 27, 30, 35, 40, 45, 50, 60, 90, 120, and 150 days), 40 mL of culture solution was sampled for analysis. After each sampling, 40 mL of filtered natural seawater was promptly added to maintain the initial volume. A polyethersulfone (PES) syringe filter (Changde Bkmm Biotechnology Co., Ltd.) with a pore size of 0.45 µm was used for filtration. The filter was pre-rinsed with Milli-Q water, and the filtered seawater was subsequently used for chemical analysis. The concentrations of dissolved organic carbon (DOC) and total nitrogen (TN) were determined using a multi-element carbon-nitrogen analyzer (Multi N/C 2100, Analytik Jena AG, Germany). Total sugar concentration was measured according to GB 500.9-2016, nitrite concentration following GB 7493-87, ammonia-nitrogen concentration according to HG 536-2009, nitrate concentration according to GB 7480-87, and

phosphate concentration following HJ 670-2013. Ammonium nitrogen ($\text{NH}_4^+\text{-N}$), nitrite nitrogen ($\text{NO}_2^-\text{-N}$), and nitrate nitrogen ($\text{NO}_3^-\text{-N}$) were the dominant components of the TN released by *Ulva prolifera* and *Sargassum horneri* (Lv et al., 2018). During each sampling, the pH and dissolved oxygen levels of the culture solution were monitored in real-time using a pH meter (PHS-25, DELIXI Group Limited Company) and a dissolved oxygen meter (ARB804, INASE Scientific Instrument Co., Ltd., Shanghai). The relevant calculation formula refers to previous study (Weigel and Pfister, 2021).

Microbial community abundance and composition analysis

Samples from both the control and experimental groups were collected on the 150th day of cultivation to analyze microbial diversity in the water after the decomposition and release of *Ulva prolifera* and *Sargassum horneri*. The culture solution was first filtered through a 3 µm mixed fiber membrane, followed by filtration through a 0.22 µm mixed fiber membrane to enrich microbial communities. The enriched samples were subjected to high-throughput 16S rRNA sequencing, following protocols outlined in previous studies (Liao et al., 2018).

Data analysis

Statistical analysis and graphical visualization were performed using Origin 2024 and SPSS 27.0. Prior to parametric tests, the normality of all datasets was confirmed by the Shapiro-Wilk test ($P > 0.05$), and homogeneity of variances was validated via Levene's test ($P > 0.05$). One-way analysis of variance (ANOVA) was applied to evaluate differences in microbial community richness among the *Ulva prolifera*, *Sargassum horneri*, and blank control groups. The confidence interval for all tests was set at 95%.

Results

DOC release and microbial utilization by decomposing macroalgae

The kinetics of dissolved organic carbon (DOC) release by *Ulva prolifera* and *Sargassum horneri* varied between the Bacteria-inhibited (BI) and Bacteria-active (BA) groups (Figure 2a). In the BI group, DOC content steadily accumulated, whereas in the BA group, DOC levels increased rapidly at first, followed by a rapid decline and stabilization. Carbohydrates constituted the largest proportion of the algal biomass in both *Ulva prolifera* and *Sargassum horneri*, and their total sugar release was measured (Figure 2b). In the BI group, the total sugar concentration from *Ulva prolifera* increased from 0.32 mg/L on day 1 to 42.38 mg/L on day 35, reaching 54.88 mg/L by the end of the experiment, with a release rate of 5.57 mg/L/day. Similarly, for *Sargassum horneri*, the total sugar concentration rose to 24.77 mg/L on day 27 and eventually reached 30.82 mg/L, with a release rate of 3.41 mg/L/day. In the BA group, the total sugar concentration for *Ulva prolifera* and *Sargassum horneri* peaked on day 11 at 32.71 mg/L and 17.98 mg/L, respectively, before rapidly declining. By day 55, the concentrations stabilized at 5.50 ± 0.5 mg/L and 3.50 ± 0.5 mg/L, respectively.

The proportion of total sugar within DOC (Figure 2c) exhibited fluctuations between 38.50% and 50.53% during the first eight days in the BA group. After day 18, due to microbial uptake, the total sugar proportion decreased to 18.58% and 14.73% for *Ulva prolifera* and *Sargassum horneri*, respectively (Figure 3). In the later stages of the experiment, the total sugar proportion stabilized, with final proportions of 30.51% and 20.74% for *Ulva prolifera* and *Sargassum horneri* in the BA group. Microbial utilization rates of DOC released by *Ulva prolifera* and *Sargassum horneri* were 89% and 84%, respectively, while the differences in total sugar concentrations were 91% and 90%.

Nutrient release and microbial utilization from decomposing macroalgae

The release kinetics of total nitrogen (TN) by *Ulva prolifera* and *Sargassum horneri* differed between the BI and BA groups (Figure 4a). In the BI group, TN content steadily accumulated,

whereas in the BA group, TN levels rapidly increased initially, followed by a sharp decline and subsequent stabilization. By day 30, TN concentrations in the BI group increased from 0.14 mg/L to 6.72 mg/L for *Ulva prolifera* and eventually reached 8.01 mg/L, with an overall release rate of 0.88 mg/L/day. The TN concentration for *Sargassum horneri* in the BI group reached 4.89 mg/L by day 30 and eventually stabilized at 6.13 mg/L, with a release rate of 0.64 mg/L/day. In the BA group, TN concentrations for *Ulva prolifera* and *Sargassum horneri* initially increased and then decreased. For *Ulva prolifera*, the TN concentration peaked at 5.06 mg/L on day 18, while for *Sargassum horneri*, it peaked at 3.82 mg/L on day 21. By the end of the experiment, the TN concentrations in the BA group were 1.37 mg/L for *Ulva prolifera* and 1.02 mg/L for *Sargassum horneri*, with release rates of 0.50 mg/L/day and 0.37 mg/L/day, respectively. The microbial utilization of TN was calculated as 26.56 mg and 20.44 mg for *Ulva prolifera* and *Sargassum horneri*, accounting for 82.9% and 83.4% of the total TN released, respectively.

Therefore, their concentrations were measured in this study (Figure 4b-d). In the BI group, $\text{NH}_4^+\text{-N}$, $\text{NO}_2^-\text{-N}$ and $\text{NO}_3^-\text{-N}$ concentrations steadily accumulated over time. For *Ulva prolifera*, the $\text{NH}_4^+\text{-N}$ concentration increased from 0.01 mg/L on day 1 to 1.03 mg/L on day 3, with a release rate of 0.08 mg/L/day. For *Sargassum horneri*, the $\text{NH}_4^+\text{-N}$ concentration reached 0.49 mg/L on day 5, with a release rate of 0.09 mg/L/day. Similarly, the $\text{NO}_2^-\text{-N}$ concentration for *Ulva prolifera* increased from 0.0012 mg/L to 0.0080 mg/L by day 5, while that for *Sargassum horneri* increased from 0.0013 mg/L to 0.0062 mg/L, with release rates of 0.0014 mg/L/day and 0.0011 mg/L/day, respectively. The $\text{NO}_3^-\text{-N}$ concentrations released by *Ulva prolifera* and *Sargassum horneri* increased from 0.012 mg/L to 0.058 mg/L and from 0.013 mg/L to 0.057 mg/L, respectively, by day 5. The overall release rates were 0.0102 mg/L/day and 0.0089 mg/L/day, respectively. In the BA group, the $\text{NH}_4^+\text{-N}$ concentrations released by *U. prolifera* and *S. horneri* increased rapidly, reaching their peaks of 0.89 mg/L and 0.8 mg/L on days 18 and 7, respectively, before quickly declining and stabilizing at 1.12 ± 0.02 mg/L and 0.11 ± 0.02 mg/L by day 80. The $\text{NO}_2^-\text{-N}$ concentrations of *U. prolifera* and *S. horneri* also rose rapidly, peaking at 0.018 mg/L and 0.013 mg/L on days 9 and 5, respectively, and subsequently decreasing to 0.021 mg/L and 0.019 mg/L by days 35 and 50. For $\text{NO}_3^-\text{-N}$, the concentrations of *U. prolifera* and *S. horneri* rose sharply, peaking at 0.27 mg/L and 0.20 mg/L on day 13, before stabilizing at 0.196 ± 0.02 mg/L and 0.086 ± 0.02 mg/L, respectively.

The release kinetics of phosphate ($\text{PO}_4^{3-}\text{-P}$) by *U. prolifera* and *S. horneri* differed significantly between the BI and BA groups (Figure 4e). In the BI group, $\text{PO}_4^{3-}\text{-P}$ content accumulated gradually, whereas in the BA group, $\text{PO}_4^{3-}\text{-P}$ concentrations initially increased rapidly, followed by a sharp depletion and subsequent stabilization. By day 2 in the BI group, the $\text{PO}_4^{3-}\text{-P}$ concentration of *U. prolifera* increased from 0.0037 mg/L to 0.0725 mg/L, ultimately reaching 0.1268 mg/L with an overall release rate of 0.0127 mg/L/day. The $\text{PO}_4^{3-}\text{-P}$ concentration released by *Sargassum horneri* reached 0.0849 mg/L on day 2, eventually stabilizing at 0.1590 mg/L with a release rate of 0.0191 mg/L/day. In the BA group, the $\text{PO}_4^{3-}\text{-P}$ concentrations of *Ulva prolifera* and *S. horneri* initially increased

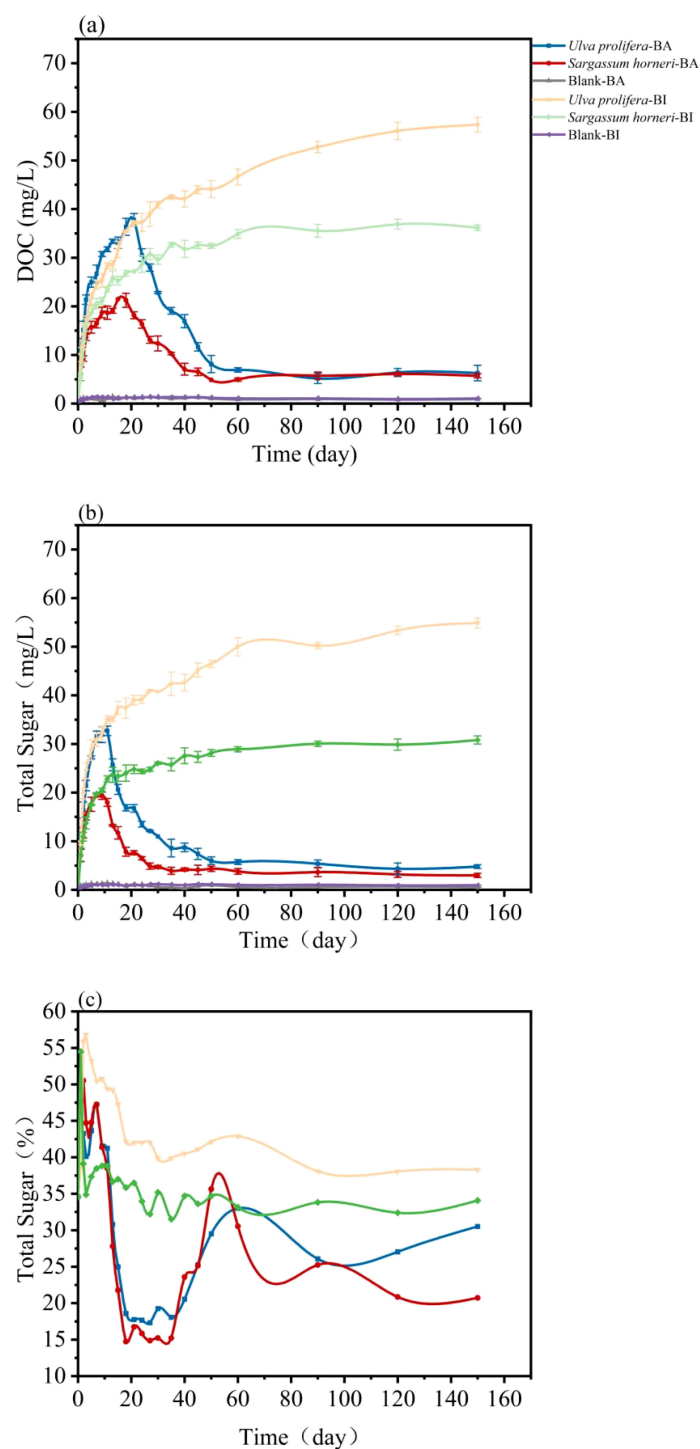


FIGURE 2

Temporal variation of DOC concentration (a), total sugar concentration (b), and the proportion of total sugar in DOC (c) released by *Ulva prolifera* and *Sargassum horneri* in the Bacteria-active and Bacteria-inhibited groups.

and then decreased, with *U. prolifera* peaking at 0.1649 mg/L on day 9 and *S. horneri* peaking at 0.2114 mg/L on day 11. By the end of the experiment, the PO_4^{3-}P concentrations in the BA group were 0.0285 mg/L and 0.0440 mg/L for *U. prolifera* and *S. horneri*,

respectively, with release rates of 0.0127 mg/L/day and 0.0164 mg/L/day. The microbial utilization of PO_4^{3-}P accounted for 0.39 mg and 0.46 mg for *U. prolifera* and *S. horneri*, respectively, representing 77.5% and 72.3% of their total release.

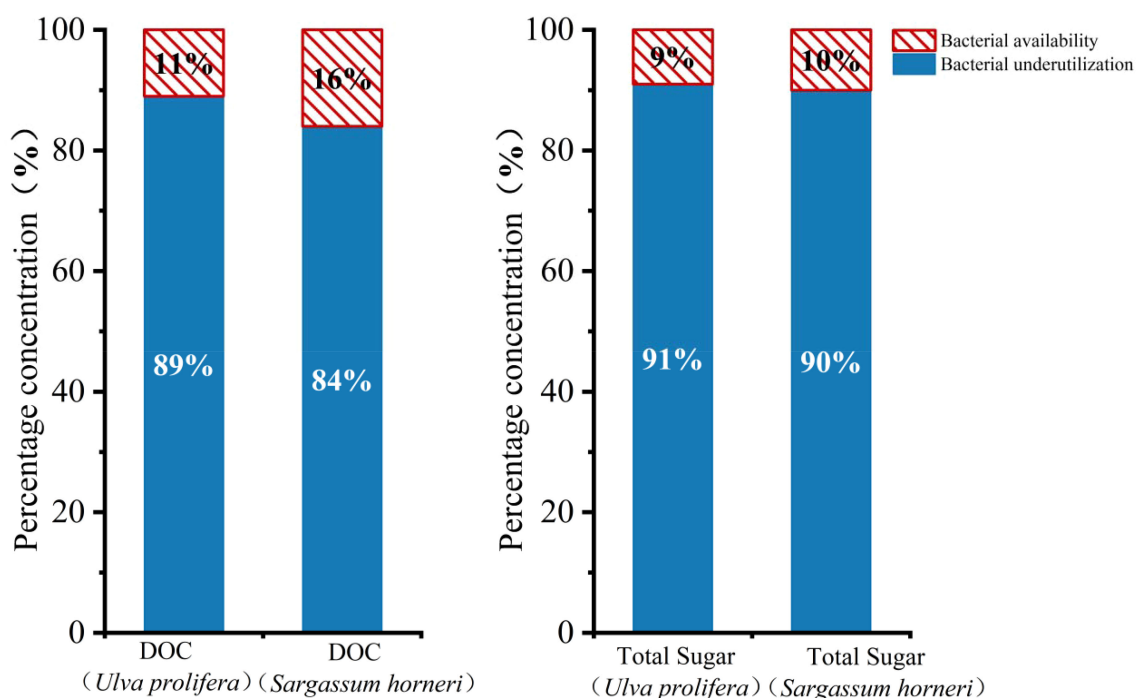


FIGURE 3

Proportional contributions of DOC and total sugar concentrations released by *Ulva prolifera* and *Sargassum horneri* in the Bacteria-active and Bacteria-inhibited groups.

Effects of macroalgae decomposition on environmental pH and DO

During the decomposition process of *U. prolifera* and *S. horneri*, significant temporal changes in pH were observed in both the BI and BA groups (Figure 5a). In both groups, pH initially dropped rapidly, followed by a sharp recovery. By day 2, the pH values in the BI group decreased from 8.11 to 7.84 for *Ulva prolifera* and from 8.11 to 7.91 for *Sargassum horneri*, eventually stabilizing at 8.06 and 8.09, respectively. In the BA group, the lowest pH values were recorded on day 9 (6.68) for *U. prolifera* and on day 7 (7.05) for *S. horneri*. By the end of the experiment, the pH in the BA group stabilized at 7.73 for *U. prolifera* and 7.83 for *S. horneri*.

The dissolved oxygen (DO) release kinetics of *U. prolifera* and *S. horneri* also exhibited distinct patterns between the BI and BA groups (Figure 5b). In the BI group, DO levels dropped rapidly, followed by fluctuations and eventual stabilization. In contrast, in the BA group, DO levels first declined sharply, then quickly recovered and stabilized. By day 9, the DO concentration of *U. prolifera* in the BI group had decreased from 7.65 mg/L to 7.01 mg/L, ultimately stabilizing at 7.05 mg/L. Similarly, *S. horneri*'s DO concentration in the BI group dropped from 7.65 mg/L to 7.12 mg/L, stabilizing at 7.13 mg/L, with release rates of 1.103 mg/L/day and 1.114 mg/L/day, respectively. In the BA group, the DO concentrations of *U. prolifera* and *S. horneri* initially decreased and then rebounded. The minimum DO levels were observed on day 13 (1.66 mg/L) for *U. prolifera* and on day 11 (1.96 mg/L) for *S. horneri*. By the end of the experiment, the DO concentrations stabilized at 4.67 mg/L for *U. prolifera* and 5.44 mg/L for *S. horneri*,

with release rates of 0.57 mg/L/day and 0.62 mg/L/day, respectively. The microbial utilization of DO accounted for 9.52 mg (33.8%) of the total release in *U. prolifera* and 6.76 mg (23.7%) in *S. horneri*.

Effects of macroalgae decomposition on environmental microbial community structure

High-throughput sequencing of 16S rRNA was conducted on the microbial communities in the *Ulva prolifera*, *Sargassum horneri*, and blank control groups. The results revealed significant differences in microbial community composition at various taxonomic levels across the three experimental groups. Overall, the groups exhibited 23–25 phyla, 35–49 classes, 80–120 orders, 117–187 families, and 150–257 genera. The number of Amplicon Sequence Variants (ASVs)—representing distinct microbial species identified through 16S rRNA sequencing—was 482, 389, and 330 in the *U. prolifera*, *S. horneri*, and blank control groups, respectively. At the genus level, 257, 189, and 150 genera were identified in these groups, respectively. Microbial richness followed the order of *U. prolifera* > *S. horneri* > blank, and the evenness of microbial communities in the *U. prolifera* and *S. horneri* groups was higher than that in the blank control group (Figure 6). At the ASV level, a total of 482, 389, and 330 ASVs were detected in the *U. prolifera*, *S. horneri*, and blank groups, respectively. Of these, 42 ASVs were shared among all three groups. There were 88 ASVs shared between *U. prolifera* and *S. horneri*, 72 ASVs shared between *U. prolifera* and the blank group, and 33 ASVs shared between *S. horneri* and the

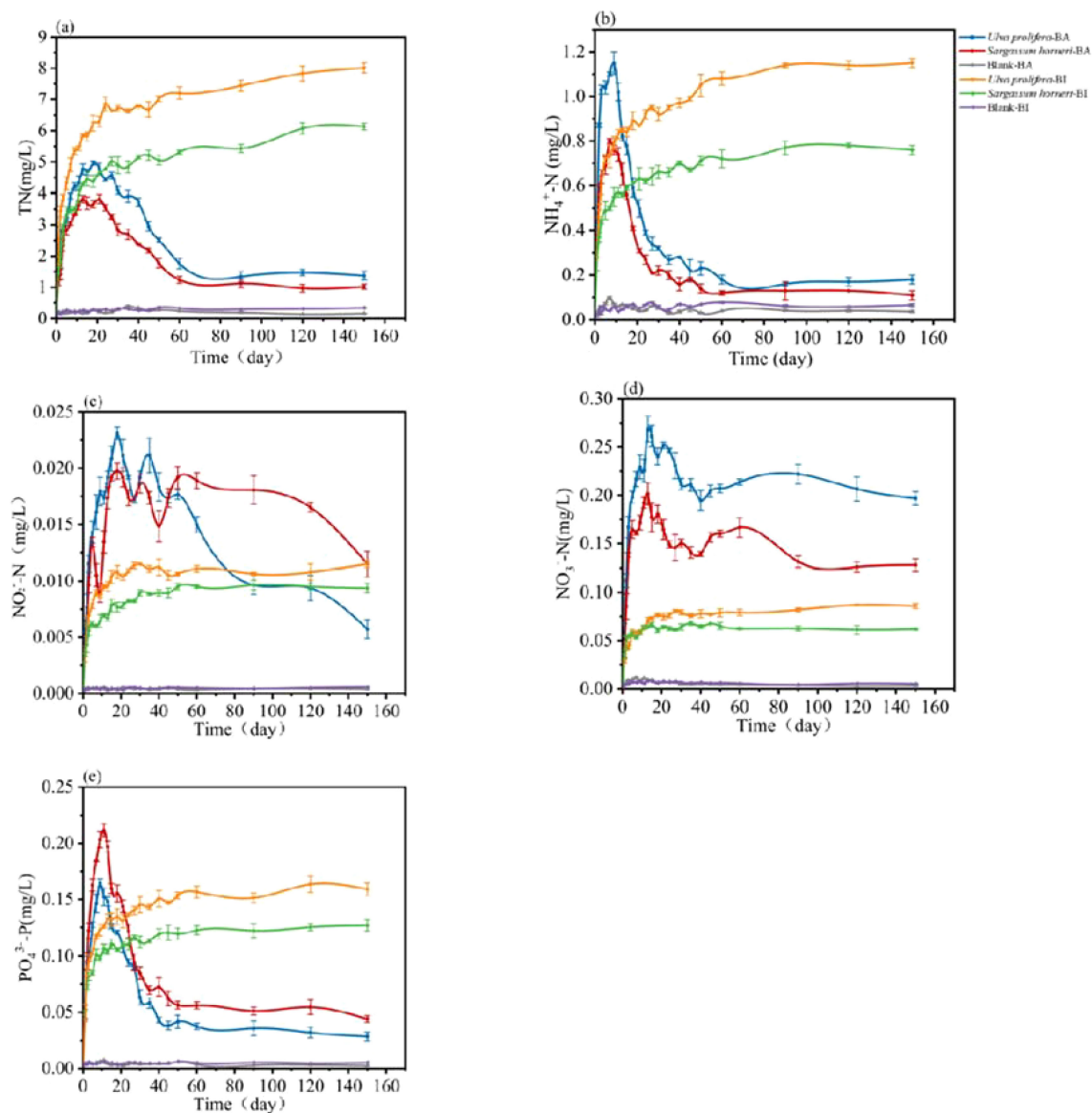


FIGURE 4

Temporal changes in the concentrations of total nitrogen (TN; a), ammonium nitrogen ($\text{NH}_4^+\text{-N}$; b), nitrite nitrogen ($\text{NO}_2^-\text{-N}$; c), nitrate nitrogen ($\text{NO}_3^-\text{-N}$; d), and phosphate ($\text{PO}_4^{3-}\text{-P}$; e) released by *Ulva prolifera*, *Sargassum horneri*, and blank controls in the Bacteria-active and Bacteria-inhibited groups.

blank group. The *Ulva prolifera* group exhibited the highest number of unique ASVs, with 288 species, while the blank control group had the fewest, at 183 species. During the decomposition process, the dominant microbial communities associated with *U. prolifera* and *Sargassum horneri* displayed distinct compositions. For *U. prolifera*, the dominant phyla included Proteobacteria (54.42%), Bacteroidota (13.46%), and Bdellovibrionota (10.73%). In contrast, the dominant phyla for *S. horneri* were Proteobacteria (74.04%), Bacteroidota (8.89%), and Firmicutes (5.22%). At the genus level, the dominant species in the *U. prolifera* group were Magnetospiraceae and Pseudohongiella, while Magnetospiraceae dominated the *S. horneri* group.

Discussions

In this study, both macroalgae exhibited rapid release of DOC during the initial stage, typically within a few hours, releasing substantial amounts of DOC. This phase was potentially controlled by physical and chemical processes, such as hydrolysis and organic matter dissolution effects (Zhang and Wang, 2017; Zhang et al., 2020). Following the initial stage, DOC release controlled by chemical processes gradually slowed down, while microbial growth and reproduction played a dominant role in the degradation of algal biomass, resulting in a gradual decrease in DOC concentration in the culture medium. Total sugars were

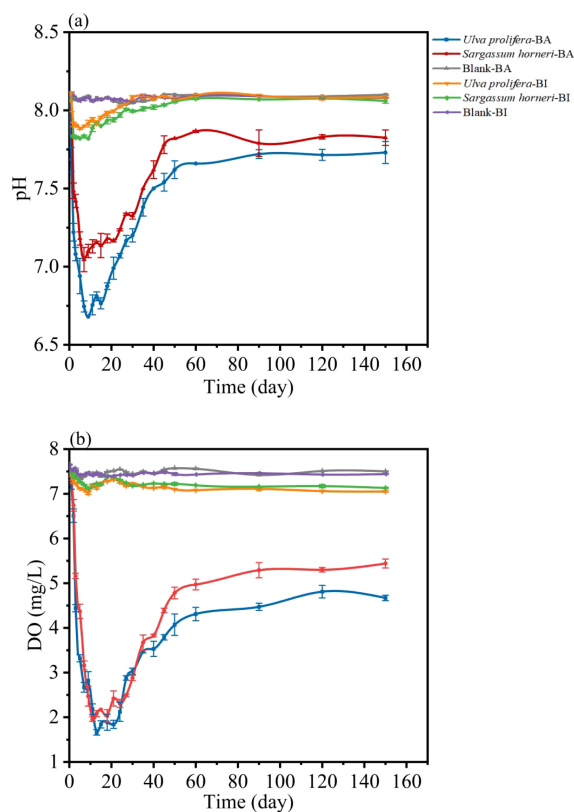


FIGURE 5
Changes in pH (a) and DO (b) concentrations of *Ulva prolifera*, *Sargassum horneri* and blank over culture time in Bacteria-active and Bacteria-inhibited groups.

identified as the main components of the organic matter released by both macroalgae species, which has also been described in other studies (Dave et al., 2021), and differences were observed in the rates and amounts of DOC and total sugar release between the two species. These differences may be attributed to the higher pectin and cellulose content in the cellular structure of *Sargassum horneri* compared to *Ulva prolifera* (Dave et al., 2021). These high-molecular-weight substances are more resistant to microbial degradation and utilization (Mei et al., 2019; Sun et al., 2024). In contrast, the hollow tubular structure of *U. prolifera* provides a larger specific surface area, enhancing physical and chemical processes as well as microbial contact, thereby accelerating the degradation of algal biomass (Pan et al., 2022). Moreover, these structural differences further influence the growth and reproduction of environmental microbial communities associated with the two macroalgae (Pan et al., 2022; Chen et al., 2023; Xie et al., 2024). On the other hand, the total sugar concentration exhibited a strong positive correlation with the organic carbon concentration for *Ulva prolifera* ($R^2 = 0.97, p < 0.01$) and *Sargassum horneri* ($R^2 = 0.98, p < 0.01$). In contrast, the data for the BA groups of *U. prolifera* ($R^2 = 0.55, p < 0.01$) and *S. horneri* ($R^2 = 0.47, p < 0.01$) were more dispersed and showed weaker correlations (Figure 7), indicating that bacteria utilized sugar components more rapidly. Previous studies have found that sugars serve not only as critical sources of

energy and structural components in microbial growth but also play key roles in metabolic regulation and microbial community dynamics (Pan et al., 2022; Chen et al., 2023).

It has been demonstrated that during the decomposition and senescence of plants, nutrients such as N and P from their tissues are released into the surrounding water (Zhang et al., 2021). The processes of DOC release during the senescence of *U. prolifera* and *S. horneri* are accompanied by the dissolution of large amounts of nutrients, complicating the changes in nutrient concentrations and their effects in the aquatic environment (Chen et al., 2020). The relationship between DOC released by *U. prolifera* and *S. horneri* and TN during their cultivation period is shown in Figure 7. The BA group of *Ulva prolifera* ($R^2 = 0.87, p < 0.01$) exhibited a significant correlation with its BI group ($R^2 = 0.92, p < 0.01$). Similarly, the BA group of *Sargassum horneri* ($R^2 = 0.85, p < 0.01$) showed a strong linear correlation with its BI group ($R^2 = 0.98, p < 0.01$). These findings suggest that carbon and nitrogen are released in an inherent proportion during the senescence of *U. prolifera* and *S. horneri*, indicating that TN release is predominantly in the form of organic nitrogen (Zhang and Wang, 2017). Microbial nitrification and denitrification play critical roles in altering nitrogen dynamics during plant senescence (Zhao et al., 2023). Additionally, the deposition of dead algal biomass and other organic matter through physical settling may influence the concentration and distribution of nitrogen in the water column (Li et al., 2017). Through ammonification and nitrification processes, microorganisms can convert organic nitrogen into inorganic nitrogen ions such as $\text{NH}_4^+\text{-N}$, NO_2^-N and NO_3^-N , integrating these into the aquatic environment (Raut et al., 2018). As microbial growth and reproduction progress, microorganisms not only consume DOC and TN as nutritional sources but also alter their concentrations and compositions through metabolic processes (Yang et al., 2021; Zhang et al., 2020). Consequently, this microbially mediated transformation may increase variability in the relationship between DOC and TN release, reducing their correlation (Li et al., 2017). A similar phenomenon is observed in the release of P, as shown in Figure 7.

In bacterial-active environments, the DOC released by the BA group of *Ulva prolifera* ($R^2 = 0.86, p < 0.01$) and *Sargassum horneri* ($R^2 = 0.89, p < 0.01$) exhibited significant negative correlations with environmental pH (Figure 7). Studies have shown that higher rates of plant detritus addition lead to lower pH levels, likely due to the release of organic matter from plants into the water, which is subsequently decomposed by microorganisms into acidic substances, such as carbon dioxide and humic acids, thereby reducing the pH of the aquatic environment (Xu et al., 2022; Zhang et al., 2024). Similarly, significant negative correlations were observed between DO and DOC in the BA group for both *U. prolifera* ($R^2 = 0.67, p < 0.01$) and *S. horneri* ($R^2 = 0.67, p < 0.01$) (Figure 8). This phenomenon may result from oxygen consumption by microbial respiration in the water, leading to a decrease in DO concentration. The rapid proliferation of bacteria accelerates oxygen consumption beyond the reoxygenation rate, causing the aquatic environment to enter an anaerobic state within a short period (Wang et al., 2014). As the release rate of organic carbon

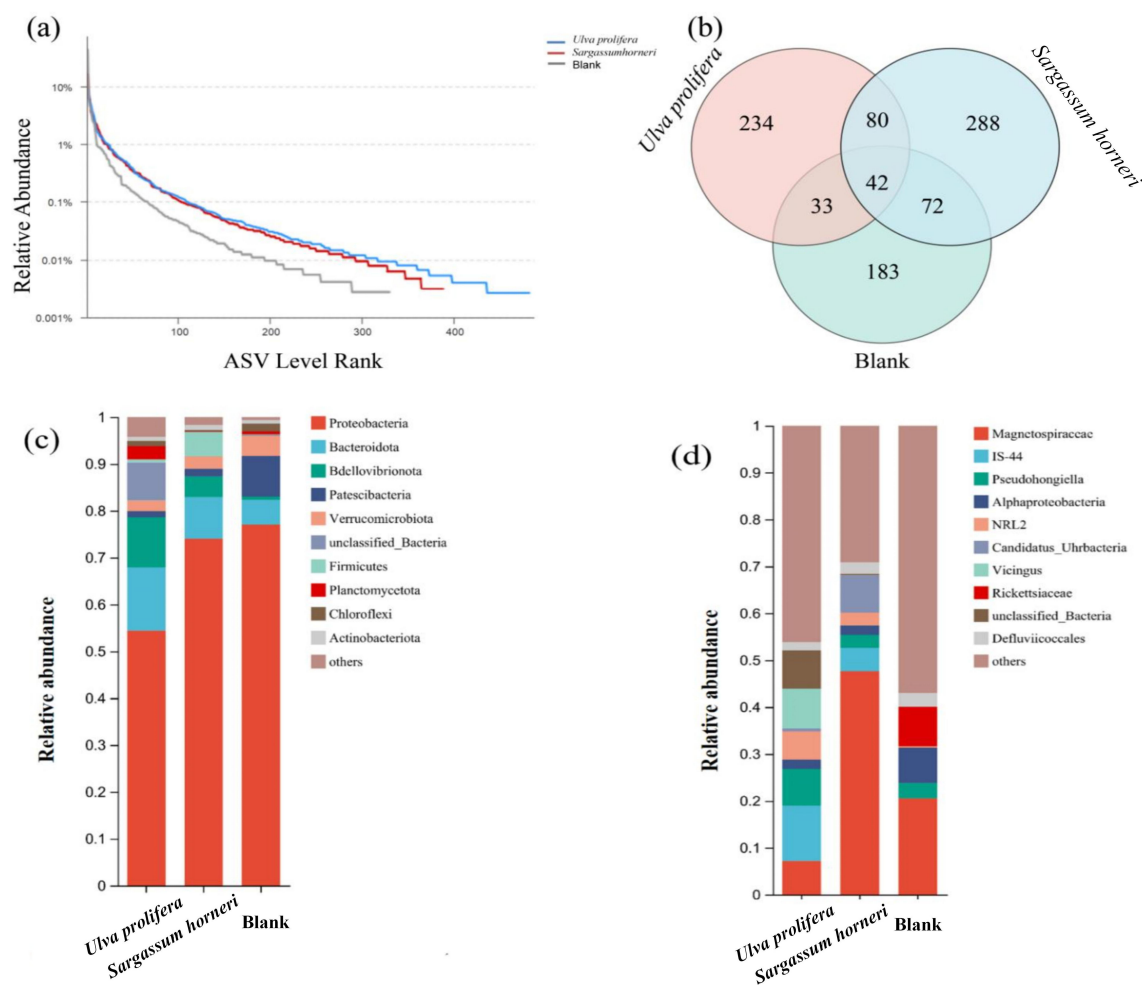


FIGURE 6

(a) Rank-abundance curves of microbial diversity in seawater during the decomposition of *Ulva prolifera*, *Sargassum horneri*, and the blank control. (b) Venn diagram showing ASV overlap between *U. prolifera*, *S. horneri*, and blank seawater. (c) Composition of microbial communities at the phylum level during the decomposition of *U. prolifera* and *S. horneri*. (d) Composition of microbial communities at the genus level during the decomposition of *U. prolifera* and *S. horneri*.

from the plants decreased and microorganisms continuously consumed the existing organic matter, the concentration of DOC gradually declined. This reduction in DOC concentration slowed the microbial oxidation rate, eventually allowing atmospheric reoxygenation to exceed oxygen consumption. Consequently, during the later stages of the experiment, the DO concentration in the water began to recover gradually (Wang et al., 2024). These findings indicate that microorganisms rapidly impact the environment by utilizing organic matter released from macroalgae, which may pose potential short-term adverse effects on nearby marine organisms.

The Alpha diversity of microbial communities was evaluated using the Sobs index, Chao index, Shannon index, and Simpson index (Figure 9). The Coverage index for all samples was 100%,

indicating the reliability of the sequencing results. The Sobs and Chao indices of *U. prolifera* were higher than those of *S. horneri*, indicating that the substantial organic matter released by *U. prolifera* more effectively promoted microbial proliferation. The Shannon and Simpson indices revealed that microbial communities in the senescent environments of both *U. prolifera* and *S. horneri* exhibited higher richness and evenness. Therefore, the apoptosis process of *Ulva prolifera* and *Sargassum horneri* may significantly affect the structure and function of microbial communities, further emphasizing the potential impacts of macroalgal apoptosis on aquatic ecosystems (García-Robledo et al., 2012; Zhu et al., 2022). Based on the analysis of the relative abundance of microbial phyla, the Proteobacteria phylum accounted for the highest proportion in the aquatic environment of this study, followed by the Bacteroidota

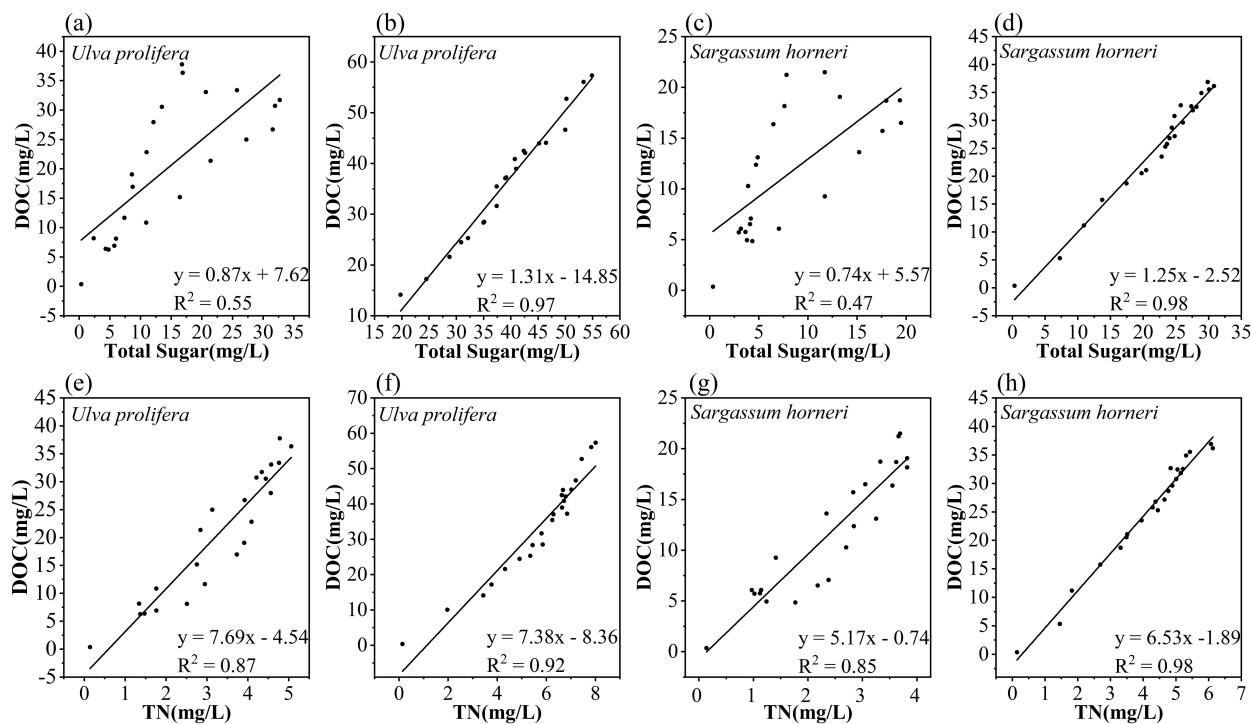


FIGURE 7

Correlation between DOC and total sugar concentrations released by *Ulva prolifera* and *Sargassum horneri* in the Bacteria-active (a, c) and Bacteria-inhibited (b, d) groups; Correlation between DOC and TN released by *Ulva prolifera* and *Sargassum horneri* in the Bacteria-active (e, g) and Bacteria-inhibited (f, h) groups.

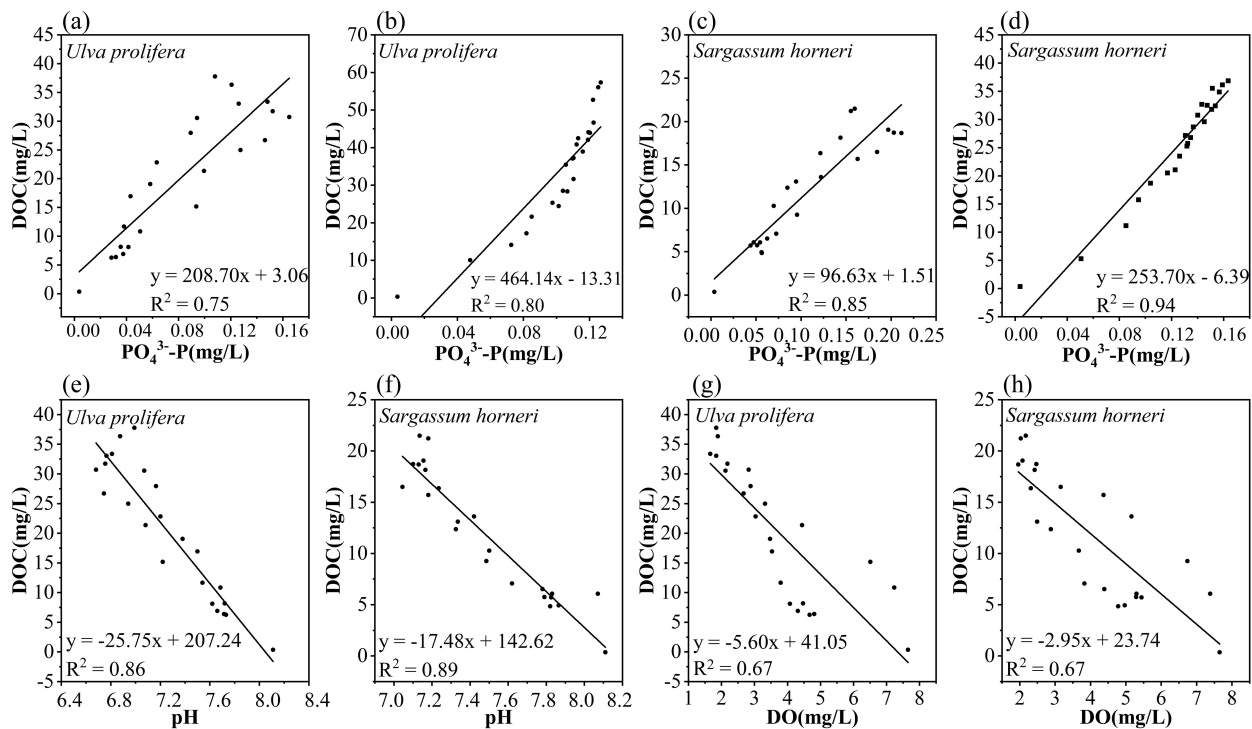


FIGURE 8

Correlation between DOC and PO_4^{3-} -P released by *Ulva prolifera* and *Sargassum horneri* in the Bacteria-active (a, c) and Bacteria-inhibited (b, d) groups; Correlation between DOC and pH (e, f) and DO concentration (g, h) in the Bacteria-active group of *Ulva prolifera* (left) and *Sargassum horneri* (right).

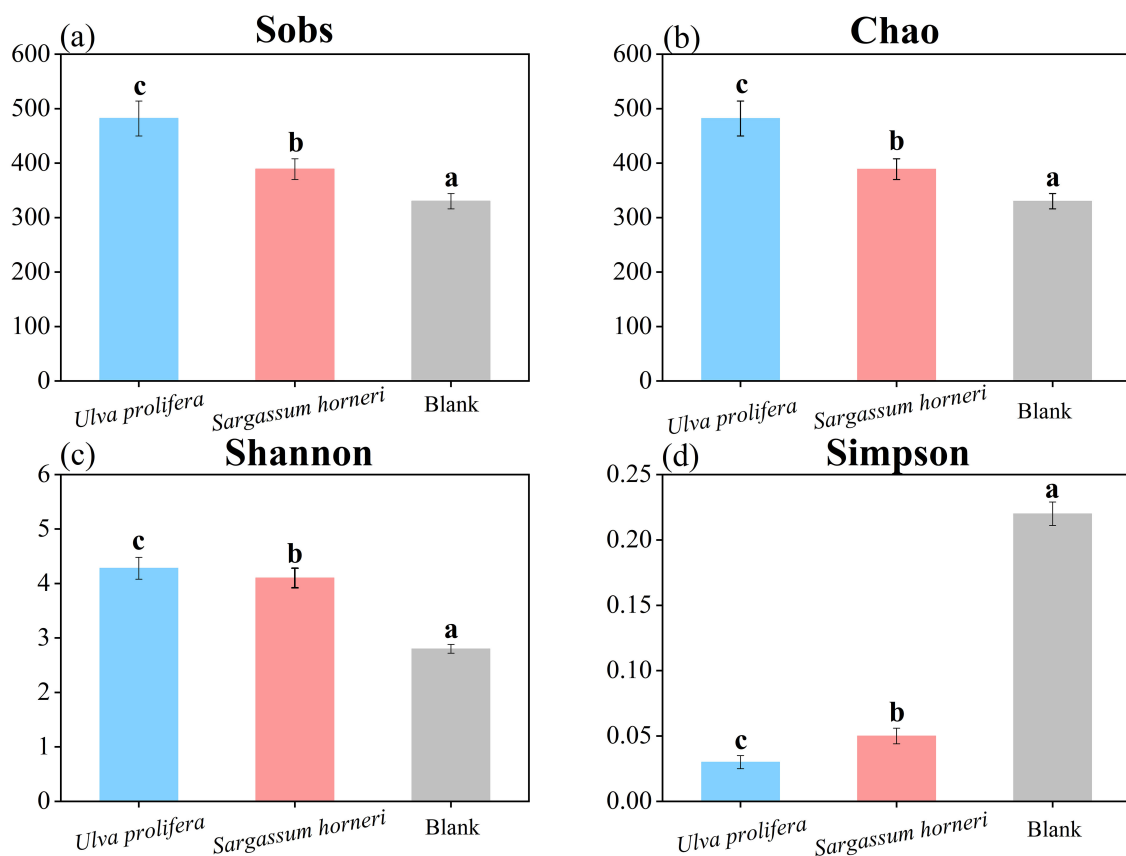


FIGURE 9

Alpha diversity indices (Sobs, Chao, Shannon, and Simpson) of microbial communities in the *Ulva prolifera*, *Sargassum horneri*, and blank control groups.

phylum (Figure 6). Research indicates that the Proteobacteria and Bacteroidota phyla play crucial roles in organic matter degradation and nitrogen cycling. Specifically, Alphaproteobacteria can utilize DOM and decompose carbon and nitrogen compounds such as ammonium and methane, thereby facilitating carbon and nitrogen cycling. Betaproteobacteria primarily function as heterotrophs, participating in the decomposition of organic matter and the cycling of nutrients in water bodies, thus contributing to the maintenance of ecological balance (Bi et al., 2019). Additionally, Gammaproteobacteria include key ammonia-oxidizing bacteria, which are considered major contributors to the nitrogen cycle (Supty et al., 2024). Meanwhile, Deltaproteobacteria have been shown to promote the cycling of nitrogen, phosphorus, and sulfur in wetland ecosystems (Mehdizadeh Allaf and Peerhossaini, 2022). In this study, bacteria from the Proteobacteria and Bacteroidota phyla facilitated the cycling of elements such as carbon and nitrogen through the decomposition of organic matter, thereby influencing the material cycling and energy flow within the entire ecosystem (Ballot et al., 2005; Zhu et al., 2022; Zhao et al., 2022).

The findings revealed that each gram of *Ulva prolifera* released 11.466 mg of organic carbon, 1.602 mg of nitrogen, and 0.025 mg of phosphorus into the coastal waters, while each gram of *Sargassum*

horneri released 7.234 mg of organic carbon, 1.226 mg of nitrogen, and 0.032 mg of phosphorus. Based on estimations, the outbreak biomass of *U. prolifera* and *S. horneri* in 2023 was approximately 5.70×10^5 tons and 5.25×10^5 tons, respectively (Yu et al., 2023; Zhang et al., 2023). Through calculations, the total release of organic carbon by *U. prolifera* and *S. horneri* was estimated to be 6535.62 tons and 3797.85 tons, respectively, resulting in a combined release of 10,333.47 tons. Similarly, *U. prolifera* and *S. horneri* released 913.14 tons and 643.65 tons of nitrogen, respectively, with a combined total of 1556.79 tons. For phosphorus, the respective releases were 16.80 tons and 14.25 tons, summing to 31.05 tons. The short-term influx of these organic substances and nutrient salt into coastal waters may significantly alter the aquatic environment, leading to changes in the composition and abundance of microbial communities, as well as impacting carbon and nutrient cycling. Such disruptions of balance are likely to have profound implications for the coastal marine ecosystem (Zhang and Wang, 2017). These ecological disruptions highlight the importance of enhancing monitoring and management strategies to mitigate the negative impacts on marine biodiversity and ecosystem health. Therefore, this study recommends implementing necessary removal measures during the early stages of algal blooms to reduce subsequent adverse ecological effects.

Conclusion

In this study, the two main causative macroalgae of green and golden tides, *Ulva prolifera* and *Sargassum horneri*, by simulating their decay and DOM release in seawater using Bacteria-active (BA) and Bacteria-inhibited (BI) groups were investigated. The results show that *U. prolifera* and *S. horneri* rapidly release DOM and nutrients during their decay. Significant differences in the quantity and rate of DOM and nutrient release were observed due to structural differences between the two macroalgae. Carbohydrates constitute the major component of the released organic matter and are rapidly utilized by microbes for growth and reproduction. Furthermore, the DOM and nutrients released by macroalgae reshape the composition of microbial communities in the marine microenvironment.

Data availability statement

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

Author contributions

TZ: Conceptualization, Data curation, Formal Analysis, Funding acquisition, Investigation, Methodology, Project administration, Resources, Software, Supervision, Validation, Visualization, Writing – original draft, Writing – review & editing. JS: Conceptualization, Investigation, Methodology, Project administration, Software, Supervision, Writing – original draft, Writing – review & editing. XH: Methodology, Resources, Supervision, Validation, Visualization, Writing – original draft. BW: Formal Analysis, Investigation, Methodology, Project administration, Validation, Writing – review & editing. JR: Investigation, Methodology, Validation, Writing – original draft. SH: Conceptualization, Data curation, Formal Analysis, Funding acquisition, Investigation, Project administration, Resources, Validation, Visualization, Writing – original draft, Writing – review & editing. XC: Conceptualization, Data curation, Funding acquisition, Investigation, Methodology, Project administration,

Resources, Validation, Visualization, Writing – original draft, Writing – review & editing.

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Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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