



# Insights on the marine microbial nitrogen cycle from isotopic approaches to nitrification

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The microbial nitrogen (N) cycle involves a variety of redox processes that control the availability and speciation of N in the environment and that are involved with the production of nitrous oxide (N<sub>2</sub>O), a climatically important greenhouse gas. Isotopic measurements of ammonium (NH<sub>4</sub><sup>+</sup>), nitrite (NO<sub>2</sub><sup>-</sup>), nitrate (NO<sub>3</sub><sup>-</sup>), and N<sub>2</sub>O can now be used to track the cycling of these compounds and to infer their sources and sinks, which has led to new and exciting discoveries. For example, dual isotope measurements of NO<sub>3</sub><sup>-</sup> and NO<sub>2</sub><sup>-</sup> have shown that there is NO<sub>3</sub><sup>-</sup> regeneration in the ocean's euphotic zone, as well as in and around oxygen deficient zones (ODZs), indicating that nitrification may play more roles in the ocean's N cycle than generally thought. Likewise, the inverse isotope effect associated with NO<sub>2</sub><sup>-</sup> oxidation yields unique information about the role of this process in NO<sub>2</sub><sup>-</sup> cycling in the primary and secondary NO<sub>2</sub><sup>-</sup> maxima. Finally, isotopic measurements of N<sub>2</sub>O in the ocean are indicative of an important role for nitrification in its production. These interpretations rely on knowledge of the isotope effects for the underlying microbial processes, in particular ammonia oxidation and nitrite oxidation. Here we review the isotope effects involved with the nitrification process and the insights provided by this information, then provide a prospectus for future work in this area.

**Keywords:** nitrification, isotopic fractionation, oxygen, nitrogen, nitrate, nitrous oxide

## NITRIFICATION IN THE OCEAN—ROLES IN NO<sub>3</sub><sup>-</sup> SUPPLY AND N<sub>2</sub>O PRODUCTION

Nitrification comprises a key link in the marine nitrogen (N) cycle converting the most reduced form of N (ammonia, NH<sub>3</sub>) to the most oxidized (nitrate, NO<sub>3</sub><sup>-</sup>). Although sunlight appears to partly inhibit nitrification (Olson, 1981a; Guerrero and Jones, 1996; Merbt et al., 2012), there are many indications that nitrification occurs in the euphotic zone (Ward, 1985, 2005; Wankel et al., 2007; Yool et al., 2007; Clark et al., 2008). Therefore, when reduced organic N is released into solution through cell lysis, grazing and digestion, it can be either reassimilated or oxidized back to NO<sub>3</sub><sup>-</sup> in the sunlit surface waters. Also, when particulate organic matter (in the form of detritus, fecal pellets, or marine snow) sinks out of the euphotic zone, it is gradually broken down into its component parts and remineralized into its inorganic forms: CO<sub>2</sub>, NH<sub>4</sub><sup>+</sup>, and PO<sub>4</sub><sup>3-</sup>. In oxic water columns, the NH<sub>4</sub><sup>+</sup> released from organic matter remineralization below the euphotic zone is rapidly oxidized to NO<sub>3</sub><sup>-</sup>. The distribution of nitrification rates in the ocean is therefore expected to follow the distribution of NH<sub>4</sub><sup>+</sup> supply from organic matter remineralization, which decreases exponentially with depth (Ward and Zafriou, 1988).

Nitrification is carried out through the combination of two microbial processes: ammonia oxidation to NO<sub>2</sub><sup>-</sup> and nitrite oxidation to NO<sub>3</sub><sup>-</sup>. Ammonia oxidation is a chemoautotrophic process carried out by ammonia-oxidizing bacteria (AOB) and ammonia-oxidizing archaea (AOA). These organisms use NH<sub>3</sub>

as their source of reducing power for CO<sub>2</sub> fixation and energy production. Nitrite oxidation is also a chemoautotrophic process and is carried out by nitrite-oxidizing bacteria (NOB). These bacteria use nitrite (NO<sub>2</sub><sup>-</sup>) as their source of reducing power for CO<sub>2</sub> fixation and energy production (Watson, 1965; Bock et al., 1989). Most ammonia and nitrite oxidizers are obligate chemoautotrophs (Watson and Waterbury, 1971), although a few are able to grow mixotrophically (Watson et al., 1986).

Although NO<sub>2</sub><sup>-</sup> is an intermediate in the nitrification process, it rarely accumulates in the ocean. NO<sub>2</sub><sup>-</sup> can be found at the base of the euphotic zone in a feature termed the primary nitrite maximum (PNM; Wada and Hattori, 1971). The processes contributing to NO<sub>2</sub><sup>-</sup> accumulation in the PNM are still debated, but most likely include a combination of ammonia oxidation and nitrite oxidation, as well as assimilatory nitrate and nitrite reduction by phytoplankton (Ward et al., 1982, 1989; Dore and Karl, 1996; Lomas and Lipschultz, 2006; Mackey et al., 2011). The relative contributions of these processes to NO<sub>2</sub><sup>-</sup> cycling have different implications for N biogeochemistry and the links between C and N cycling. Net production of NO<sub>2</sub><sup>-</sup> through nitrification (decoupling of ammonia and nitrite oxidation) can also have implications for the production of nitrous oxide (N<sub>2</sub>O), a climatically important greenhouse gas. It is therefore important to know how the processes contributing to the production and maintenance of the PNM vary in space and time.

NO<sub>2</sub><sup>-</sup> also accumulates in oxygen deficient regions of the water column in a feature termed the secondary nitrite maximum

(SNM; Brandhorst, 1959). The SNM is generally assumed to reflect active denitrification in oxygen deficient zones (ODZs), as SNM features are only found in the absence of dissolved oxygen (Brandhorst, 1959; Cline and Richards, 1972; Codispoti and Christensen, 1985). However, recent studies have shown that the presence of a SNM feature may not coincide with the most intense  $\text{NO}_2^-$  cycling, as active  $\text{NO}_2^-$  reduction occurs in the Omani upwelling region in the absence of  $\text{NO}_2^-$  accumulation (Jensen et al., 2011; Lam et al., 2011).  $\text{NO}_2^-$  consumption in the SNM may occur through many processes, including denitrification (reduction of  $\text{NO}_2^-$  to  $\text{N}_2$ ), anaerobic ammonia oxidation (reduction of  $\text{NO}_2^-$  to  $\text{N}_2$  and oxidation to  $\text{NO}_3^-$ ), and nitrite oxidation (oxidation of  $\text{NO}_2^-$  to  $\text{NO}_3^-$ ). Recent studies using natural abundance isotopes (Casciotti, 2009), profile modeling (Lam et al., 2011), isotope tracers (Lipschultz et al., 1990; Füssel et al., 2012), and gene markers (Füssel et al., 2012) suggest that a significant fraction of  $\text{NO}_2^-$  produced within the SNM may be consumed through oxidation to  $\text{NO}_3^-$ .

Several questions remain about the roles of AOB and AOA in marine nitrification, the controls on their distribution and activity, and the rates of these processes. These questions relate to the cycling of  $\text{NO}_3^-$ ,  $\text{NO}_2^-$ , and  $\text{NH}_4^+$  in the water column, and the production of  $\text{N}_2\text{O}$  linked to nitrification. These questions can be addressed with a variety of complementary approaches, including molecular community analysis and quantification, instantaneous rate measurements, natural abundance stable isotope measurements, and geochemical modeling.

Examples of applications involving the use of natural abundance stable isotopes to study nitrification include: (1) the role of euphotic zone nitrification in supplying  $\text{NO}_3^-$  for photosynthetic growth (Wankel et al., 2007; DiFiore et al., 2009), (2) the contributions of nitrification and nitrate reduction to  $\text{NO}_2^-$  accumulation in the PNM (Buchwald and Casciotti, unpublished), (3) the role of nitrification in near-surface  $\text{N}_2\text{O}$  production (Dore et al., 1998; Santoro et al., 2010, 2011), and (4) the role of nitrite oxidation in recycling  $\text{NO}_3^-$  in and around ODZs (Sigman et al., 2005; Casciotti and McIlvin, 2007; Casciotti, 2009). Understanding the isotopic systematics for nitrification is also important for tracking the balance of high-latitude and low-latitude productivity and N budget processes (N fixation and denitrification) through  $\text{NO}_3^-$  isotope distributions in the deep ocean (Sigman et al., 2009). In order to understand these applications we first review the N and O isotopic systematics of the nitrification process, including both ammonia and nitrite oxidation.

## ISOTOPE SYSTEMATICS FOR AMMONIA OXIDATION

The  $\delta^{18}\text{O}$  value of  $\text{NO}_2^-$  produced during ammonia oxidation ( $\delta^{18}\text{O}_{\text{NO}_2, \text{nit}} = ({}^{18}\text{O}/{}^{16}\text{O}_{\text{NO}_2} \div {}^{18}\text{O}/{}^{16}\text{O}_{\text{VSMOW}} - 1) \times 1000$ ) is dependent on the  $\delta^{18}\text{O}$  values of the oxygen atom sources ( $\text{O}_2$  and  $\text{H}_2\text{O}$ ), isotopic fractionation during their incorporation ( ${}^{18}\epsilon_{\text{k}, \text{O}_2}$  and  ${}^{18}\epsilon_{\text{k}, \text{H}_2\text{O}, 1}$ , respectively), as well as any exchange of oxygen atoms between nitrite and water ( $x_{\text{AO}}$ ) and the corresponding equilibrium isotope effect ( ${}^{18}\epsilon_{\text{eq}}$ ) (Equation 1; Casciotti et al., 2011). Throughout this review, kinetic isotope fractionation factors are defined by  $\alpha_{\text{k}} = k^l/k^h$  where  $k^l$  is the first order rate constant for reaction of the light isotope and  $k^h$  is that for reaction of the heavy isotope. Equilibrium fractionation factors are

defined as  $\alpha_{\text{eq}} = R_1/R_2$  where  $R_1$  and  $R_2$  are the isotope ratios of two species in equilibrium. Kinetic and equilibrium isotope effects are defined by  $\epsilon = (\alpha - 1) \times 1000$ .

$$\delta^{18}\text{O}_{\text{NO}_2, \text{nit}} = \left[ \frac{1}{2}(\delta^{18}\text{O}_{\text{O}_2} - {}^{18}\epsilon_{\text{O}_2}) + \frac{1}{2}(\delta^{18}\text{O}_{\text{H}_2\text{O}} - {}^{18}\epsilon_{\text{k}, \text{H}_2\text{O}, 1}) \right] \times (1 - x_{\text{AO}}) + (\delta^{18}\text{O}_{\text{H}_2\text{O}} + {}^{18}\epsilon_{\text{eq}})(x_{\text{AO}}) \quad (1)$$

Even though oxygen is incorporated enzymatically from  $\text{O}_2$  to  $\text{H}_2\text{O}$  in a 1:1 ratio during ammonia oxidation (Andersson and Hooper, 1983), early studies of AOB found that a large amount of oxygen atom exchange with water could be associated with ammonia oxidation (Dua et al., 1979; Andersson et al., 1982; Andersson and Hooper, 1983). The conditions favoring oxygen atom exchange included high cell densities and high  $\text{NO}_2^-$  concentrations. These findings, as well as the low variation of deep ocean  $\delta^{18}\text{O}_{\text{NO}_3}$  (Casciotti et al., 2002; Sigman et al., 2009) led researchers to assume that the O atoms in oceanic  $\text{NO}_3^-$  derive primarily from  $\text{H}_2\text{O}$  with little residual signal from dissolved  $\text{O}_2$ . In more recent studies, however, the amount of biologically-catalyzed exchange has been determined under lower cell densities and substrate concentrations and found to be much lower for marine AOB (Casciotti et al., 2010; Buchwald et al., 2012) and AOA (Santoro et al., 2011). Exchange levels were particularly low (5%) when  $\text{NO}_2^-$  concentrations were held near  $1 \mu\text{M}$  by co-cultivation with NOB (Buchwald et al., 2012). These results suggested that oxygen isotope exchange during nitrification may be quite low where ammonia and nitrite oxidation are tightly coupled, but may play a role when ammonia and nitrite oxidation become decoupled, such as in the PNM.

Given low amounts of biologically-catalyzed oxygen atom exchange with  $\text{H}_2\text{O}$ , the low  $\delta^{18}\text{O}$  values of  $\text{NO}_3^-$  in seawater may be surprising given the high  $\delta^{18}\text{O}$  values of dissolved  $\text{O}_2$  (Kroopnick and Craig, 1976). However, oxygen atom incorporation from  $\text{O}_2$  and/or  $\text{H}_2\text{O}$  during ammonia oxidation is associated with isotopic fractionation, such that the  ${}^{18}\text{O}:{}^{16}\text{O}$  of oxygen atoms incorporated into  $\text{NO}_2^-$  is significantly lower than the ambient pools of  $\text{O}_2$  and  $\text{H}_2\text{O}$  (Casciotti et al., 2010; Santoro et al., 2011). This leads to production of  $\text{NO}_2^-$  from ammonia oxidation with  $\delta^{18}\text{O}$  values between  $-3\text{‰}$  and  $5\text{‰}$  rather than near  $12\text{‰}$ , which would be expected from average  $\delta^{18}\text{O}_{\text{H}_2\text{O}}$  and  $\delta^{18}\text{O}_{\text{O}_2}$  values (Casciotti et al., 2010). Furthermore, since oxygen atom exchange occurs with an equilibrium isotope effect ( ${}^{18}\epsilon_{\text{eq}}$ ) of  $11\text{--}14\text{‰}$  (Casciotti et al., 2007; Buchwald and Casciotti, unpublished), this equilibration would tend to raise the  $\delta^{18}\text{O}$  value of  $\text{NO}_2^-$  relative to the initial  $\delta^{18}\text{O}_{\text{NO}_2}$  produced by ammonia oxidation.

Nitrogen isotopic fractionation during ammonia oxidation ranges from  $14\text{‰}$  to  $38\text{‰}$  for AOB (Mariotti et al., 1981; Yoshida, 1988; Casciotti et al., 2003) and  $20\text{--}22\text{‰}$  for AOA (Santoro and Casciotti, 2011). These values represent the isotope effect expressed under non-limiting concentrations of  $\text{NH}_4^+$ . In the ocean  $\text{NH}_4^+$  consumption generally goes to completion, so the isotope effect for ammonia oxidation may not be expressed. It may, however, be expressed at the branch point between ammonia assimilation and oxidation in the euphotic zone (Wankel

et al., 2007; DiFiore et al., 2009) or in the production of  $N_2O$  by ammonia oxidizers (Yoshida, 1988; Frame and Casciotti, 2010).

## ISOTOPE SYSTEMATICS FOR $N_2O$ PRODUCTION

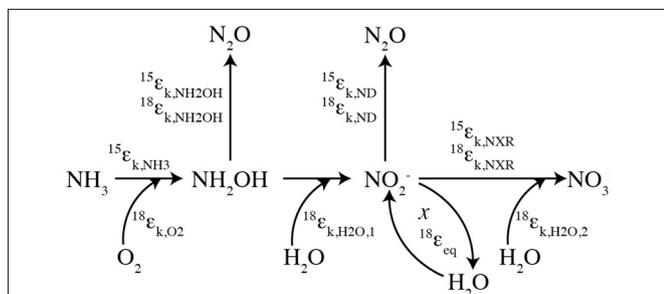
Production of  $N_2O$  by AOB occurs through two separate pathways: hydroxylamine decomposition and nitrite reduction, so-called “nitrifier denitrification” (Figure 1; Poth and Focht, 1985; Hooper et al., 1990). The isotopic compositions ( $\delta^{15}N^{bulk}$ ,  $\delta^{18}O$ ,  $\delta^{15}N^{\alpha}$ ,  $\delta^{15}N^{\beta}$ , and site preference (SP) =  $\delta^{15}N^{\alpha} - \delta^{15}N^{\beta}$ ) of the  $N_2O$  produced through these pathways may provide insight into the mechanisms of  $N_2O$  production under different growth conditions (Frame and Casciotti, 2010; Sutka et al., 2003, 2004). For example,  $N_2O$  production through nitrifier denitrification (enhanced by high cell densities, high  $NO_2^-$  concentrations, and low  $O_2$  concentrations; Frame and Casciotti, 2010) has low  $\delta^{15}N^{bulk}$  and low SPs relative to that produced by hydroxylamine decomposition (Figure 2). This is most likely due to the additional steps involved with the production of  $N_2O$  from  $NO_2^-$  and accumulation of the main product,  $NO_2^-$ , which enables fractionation associated with  $NO_2^-$  reduction to be expressed.

Oxygen isotopes have been underutilized in determining  $N_2O$  sources, primarily because the isotopic systematics are less well understood, but knowledge of the O isotope systematics is increasing (Frame and Casciotti, 2010; Snider et al., 2012). The  $N_2O$  produced via nitrifier denitrification has a slightly lower  $\delta^{18}O$  value than that produced from hydroxylamine decomposition (Figure 2; Frame and Casciotti, 2010). This is most likely because  $H_2O$  is incorporated into  $NO_2^-$ , leading to lower  $\delta^{18}O$  values in  $NO_2^-$  relative to  $NH_2OH$ . However, going from either  $NH_2OH$  or  $NO_2^-$  to  $N_2O$  involves the loss of O atoms, which can occur with fractionation. This fractionation leads to preferential loss of  $^{16}O$  and retention of  $^{18}O$  in the residual N oxides transferred to  $N_2O$ . The net

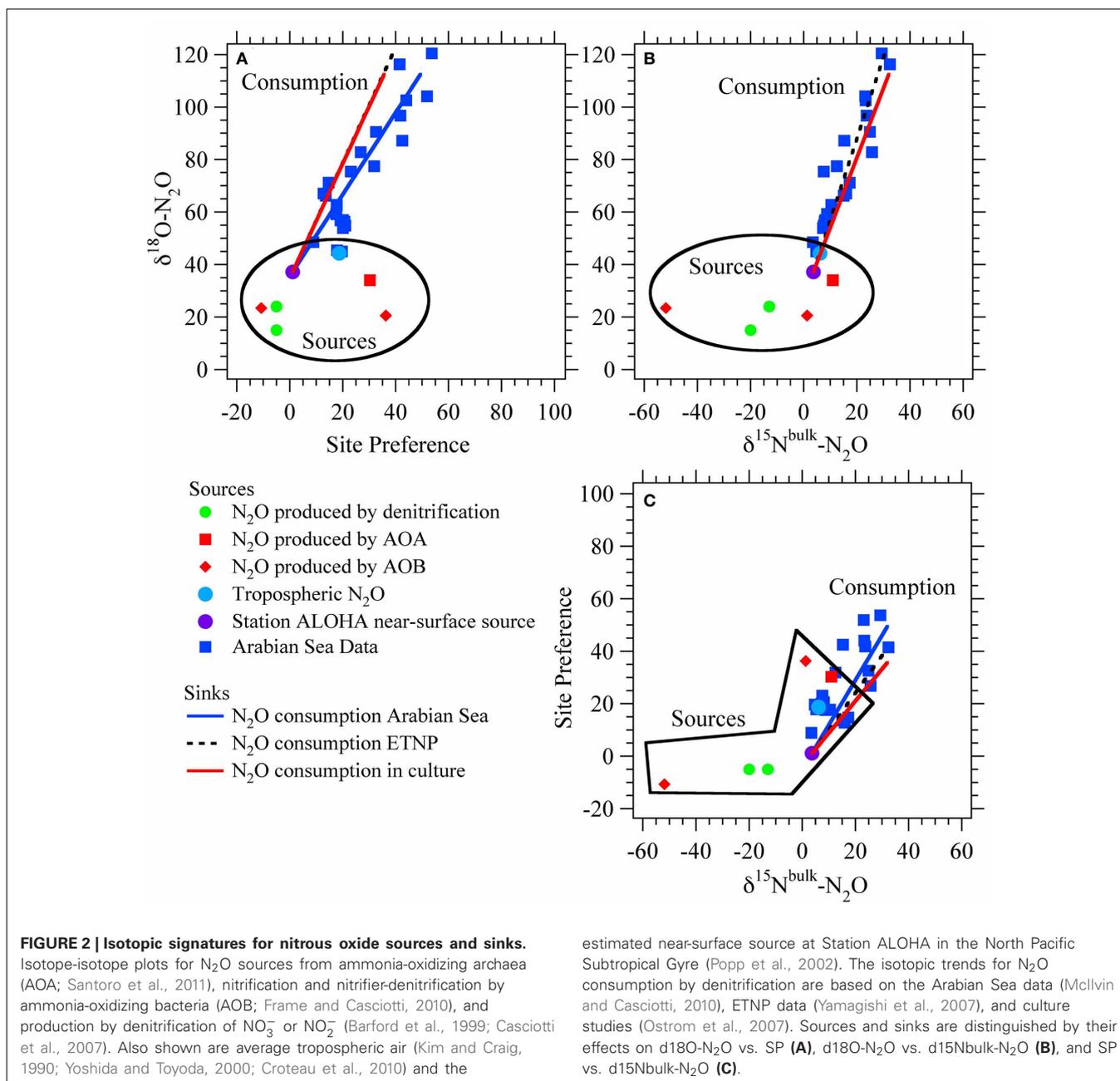
isotopic fractionation for oxygen isotopes in the hydroxylamine decomposition pathway ( $^{18}\epsilon_{k,NH_2OH}$ ), including both incorporation of  $O_2$  into  $NH_2OH$  and production of  $N_2O$  from  $NH_2OH$ , was  $2.9 \pm 0.8\%$  indicating that  $N_2O$  produced from this pathway had a lower  $^{18}O:^{16}O$  than the ambient  $O_2$  (Frame and Casciotti, 2010). The net isotope effect for  $N_2O$  production from  $NO_2^-$  via nitrifier denitrification ( $^{18}\epsilon_{k,ND}$ ) was  $-8.4 \pm 1.4\%$  (Frame and Casciotti, 2010). The negative value indicates that the  $N_2O$  produced from  $NO_2^-$  is enriched in  $^{18}O$  relative to  $NO_2^-$ , consistent with branching of O atoms and preferential loss of  $^{16}O$  during this reaction (Casciotti et al., 2007).

The  $N_2O$  site preference (SP) is determined mainly by the enzymatic mechanism, rather than the substrate  $\delta^{15}N$  value (Toyoda and Yoshida, 1999; Yoshida and Toyoda, 2000; Schmidt et al., 2004). The SP of  $N_2O$  produced during nitrification is  $+30\%$  to  $+38\%$  (Figure 2; Sutka et al., 2003, 2004; Frame and Casciotti, 2010), while  $N_2O$  produced from denitrification and nitrifier denitrification has a SP of  $-10\%$  to  $+5\%$  (Sutka et al., 2003, 2004; Toyoda et al., 2005; Frame and Casciotti, 2010). The large difference between the SP values of these two primary mechanisms for  $N_2O$  production provides a large signal with which to distinguish their contributions. The interpretation of SP values is therefore somewhat simplified relative to bulk  $\delta^{15}N$  and  $\delta^{18}O$  values that reflect both mechanism and substrate isotope ratios, which change over time. This seemingly simple distinction is complicated, however, by the fact that  $N_2O$  consumption during denitrification increases SP (Ostrom et al., 2007; Yamagishi et al., 2007; Koba et al., 2009). Therefore, a high SP value may arise through production of  $N_2O$  via nitrification or net  $N_2O$  consumption during denitrification. However, the  $\delta^{18}O$  signature of these two scenarios is quite different and can enable the scenarios to be distinguished (Figure 2).

Recently, the isotopic compositions of  $N_2O$  produced by AOA were found to be distinct from AOB (Santoro et al., 2011). In particular,  $N_2O$  produced by AOA is enriched in  $^{15}N$  and  $^{18}O$  relative to that produced by AOB, which may explain some of the elevated  $\delta^{15}N$  and  $\delta^{18}O$  values observed in oceanic  $N_2O$  (Santoro et al., 2011). The reasons for the isotopic distinction between AOA and AOB is not known, but may involve a different mechanism of  $N_2O$  production involving a unique intermediate or enzymatic pathway. However, the SP of  $N_2O$  produced by AOA is similar to that of  $N_2O$  produced by hydroxylamine decomposition by AOB (Santoro et al., 2011; Loescher et al., 2012). While it is not yet clear whether  $N_2O$  production (or nitrification in general) by AOA involves hydroxylamine, isotopic evidence to date shows that the  $N_2O$  produced aerobically by AOA does not have a SP consistent with denitrification or nitrifier-denitrification.  $\delta^{18}O$  data also show that the  $N_2O$  produced by AOA incorporates O primarily from  $O_2$ , rather than from  $H_2O$ , which supports production by decomposition of an intermediate, rather than from  $NO_2^-$  under the conditions tested (Santoro et al., 2011). It is still unknown whether AOA are able to produce  $N_2O$  through a second pathway similar to nitrifier denitrification and thus produce  $N_2O$  with a lower SP. Genetic analyses currently suggest that nitrification in AOA may proceed via a NO or HNO intermediate (Walker et al., 2010), which could potentially be converted to  $N_2O$ . Further work



**FIGURE 1 | Isotopic systematics for nitrification.** A schematic of the isotopic systematics for ammonia oxidation and nitrite oxidation during nitrification. The kinetic isotope effects for ammonia oxidation ( $^{15}\epsilon_{k,NH_3}$ ) and nitrite oxidation ( $^{15}\epsilon_{k,NXR}$  and  $^{18}\epsilon_{k,NXR}$ ) characterize the isotopic fractionation for the main N transformation processes while isotopic fractionation during oxygen atom incorporation ( $^{18}\epsilon_{k,O_2}$ ,  $^{18}\epsilon_{k,H_2O,1}$ , and  $^{18}\epsilon_{k,H_2O,2}$ ) controls the oxygen isotopes incorporated by the central pathway. Oxygen isotope exchange during ammonia oxidation and/or post-production abiotic exchange (x) may also play a role with the equilibrium fractionation ( $^{18}\epsilon_{eq}$ ) associated with it.  $N_2O$  production occurs with N and O fractionation through decomposition of hydroxylamine ( $NH_2OH$ ) and nitrifier-denitrification (ND).

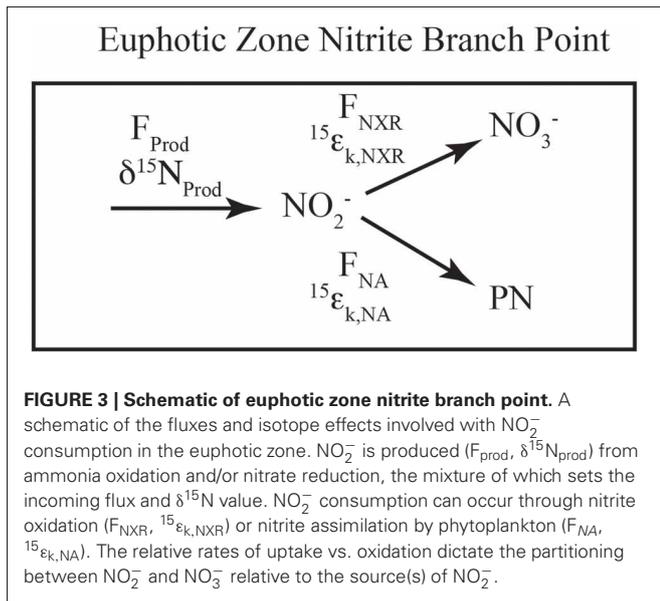


is required to determine the pathway and intermediates of nitrification and N<sub>2</sub>O production by AOA, and to further study its isotope systematics under a variety of growth conditions.

### ISOTOPE SYSTEMATICS FOR NITRITE OXIDATION

The isotopic systematics for nitrite oxidation to nitrate have also been studied recently, and were found to occur with extremely unique inverse kinetic isotope effects for N (Casciotti, 2009) and O isotopes (Buchwald and Casciotti, 2010). Because of these inverse isotope effects, when nitrite oxidation is active, the δ<sup>15</sup>N<sub>NO<sub>2</sub></sub> and δ<sup>18</sup>O<sub>NO<sub>2</sub></sub> values are expected to be lower than the NO<sub>2</sub><sup>-</sup> initially produced by ammonia oxidation or nitrate

reduction. As discussed below, this appears to occur in both primary and secondary nitrite maxima (Casciotti, 2009; Buchwald and Casciotti, unpublished). In most parts of the ocean, however, NO<sub>2</sub><sup>-</sup> does not accumulate and the isotope effects associated with nitrite oxidation can only be expressed through a branch point (Figure 3). Isotopic separation can occur at a branch point because there is more than one fate for NO<sub>2</sub><sup>-</sup> (e.g., NO<sub>2</sub><sup>-</sup> is either oxidized to NO<sub>3</sub><sup>-</sup> or assimilated into particulate N, PN) and the heavy isotope can be preferentially shunted in one direction vs. the other. This is analogous to the branch point that has been described during the oxidation or assimilation of ammonium (Sigman et al., 2005; Wankel et al., 2007; DiFiore et al., 2009). The



equations that describe the steady state N isotopic partitioning between  $\text{NO}_2^-$  and  $\text{NO}_3^-$  when nitrite oxidation and assimilation occur concurrently are:

$$\delta^{15}\text{N}_{\text{NO}_2} = \delta^{15}\text{N}_{\text{NO}_2,\text{produced}} + f_{\text{NA}} \times ^{15}\epsilon_{k,\text{NA}} + f_{\text{NXR}} \times ^{15}\epsilon_{k,\text{NXR}} \quad (2)$$

$$\delta^{15}\text{N}_{\text{NO}_3,\text{produced}} = \delta^{15}\text{N}_{\text{NO}_2} - ^{15}\epsilon_{k,\text{NXR}} \quad (3)$$

where  $f_{\text{NA}}$  and  $f_{\text{NXR}}$  are the fractions of  $\text{NO}_2^-$  consumed by assimilation and oxidation, respectively, and  $^{15}\epsilon_{k,\text{NA}}$  and  $^{15}\epsilon_{k,\text{NXR}}$  are the respective isotope effects. In general, nitrite oxidation will transfer  $\text{NO}_2^-$  with an elevated  $^{15}\text{N}:^{14}\text{N}$  ratio to the  $\text{NO}_3^-$  pool, while nitrite assimilation transfers the residual  $\text{NO}_2^-$  with a lower  $^{15}\text{N}:^{14}\text{N}$  ratio into the PN pool. If  $^{15}\epsilon_{k,\text{NA}}$  is 1‰ (Waser et al., 1998),  $^{15}\epsilon_{k,\text{NXR}}$  is -15‰ (Buchwald and Casciotti, 2010),  $\delta^{15}\text{N}_{\text{NO}_2}$  at steady state will be lower than the source of  $\text{NO}_2^-$ , unless nitrite assimilation is >95% of the  $\text{NO}_2^-$  sink. This has the opposite sense of the ammonia oxidation/assimilation branching where ammonia oxidation transfers low  $^{15}\text{N}:^{14}\text{N}$  material into the  $\text{NO}_2^-$  and  $\text{NO}_3^-$  pools and higher  $^{15}\text{N}:^{14}\text{N}$  material into the PN pool.

When nitrite oxidation is tightly coupled to ammonia oxidation and  $\text{NO}_2^-$  does not accumulate, the  $\delta^{18}\text{O}$  value of the  $\text{NO}_3^-$  produced primarily reflects the  $\delta^{18}\text{O}$  values of the O atom sources ( $\text{H}_2\text{O}$  and  $\text{O}_2$ ; Kumar et al., 1983) and the incorporation isotope effects for ammonia and nitrite oxidation (Buchwald et al., 2012). The oxygen isotope systematics of nitrite oxidation can be described by Equation 4, while the full oxygen isotope systematics of nitrification starting from  $\text{NH}_4^+$ , assuming no biologically-catalyzed oxygen atom exchange during nitrite oxidation ( $x_{\text{NO}} = 0$ ; DiSpirito and Hooper, 1986; Friedman et al.,

1986; Buchwald and Casciotti, 2010), is described by Equation 5.

$$\delta^{18}\text{O}_{\text{NO}_3,\text{final}} = \frac{2}{3} [(1-x_{\text{NO}})\delta^{18}\text{O}_{\text{NO}_2} + x_{\text{NO}}(\delta^{18}\text{O}_{\text{H}_2\text{O}} + ^{18}\epsilon_{\text{eq}})] + \frac{1}{3} (\delta^{18}\text{O}_{\text{H}_2\text{O}} - ^{18}\epsilon_{k,\text{H}_2\text{O},2}) \quad (4)$$

$$\delta^{18}\text{O}_{\text{NO}_3,\text{final}} = \left[ \frac{2}{3} + \frac{1}{3}x_{\text{AO}} \right] \delta^{18}\text{O}_{\text{H}_2\text{O}} + \frac{1}{3} [(\delta^{18}\text{O}_{\text{O}_2} - ^{18}\epsilon_{k,\text{O}_2} - ^{18}\epsilon_{k,\text{H}_2\text{O},1})(1-x_{\text{AO}}) - ^{18}\epsilon_{k,\text{H}_2\text{O},2}] + \frac{2}{3} ^{18}\epsilon_{\text{eq}}(x_{\text{AO}}) \quad (5)$$

Equation 5 indicates that the  $\delta^{18}\text{O}_{\text{NO}_3}$  produced by tightly-coupled ammonia and nitrite oxidation should reflect variations in both  $\delta^{18}\text{O}_{\text{O}_2}$  and  $\delta^{18}\text{O}_{\text{H}_2\text{O}}$  in a ratio of 1 to 2, with slight modification of this stoichiometry by biologically-catalyzed oxygen atom exchange during ammonia oxidation (Casciotti et al., 2010; Buchwald et al., 2012). As discussed below, when ammonia and nitrite oxidation are not tightly coupled, abiotic equilibration can affect  $\delta^{18}\text{O}_{\text{NO}_2}$  and the final  $\delta^{18}\text{O}_{\text{NO}_3}$  produced. Regardless of whether  $\text{NO}_2^-$  accumulates, isotopic fractionation during oxygen atom incorporation should lead to an isotopic offset between the substrates ( $\text{O}_2$  and  $\text{H}_2\text{O}$ ) and the produced  $\text{NO}_3^-$ . The expected  $\delta^{18}\text{O}_{\text{NO}_3}$  value produced in oxygenated seawater with little exchange is -1‰ to +1‰ (similar to  $\delta^{18}\text{O}_{\text{H}_2\text{O}}$ ), resulting from a complex series of fractionation factors rather than the unfractionated incorporation of and exchange with  $\text{H}_2\text{O}$  (Buchwald et al., 2012).

### ABIOTIC EQUILIBRATION OF OXYGEN ATOMS IN NITRITE

As introduced above, abiotic equilibration of oxygen atoms between  $\text{NO}_2^-$  and  $\text{H}_2\text{O}$  is likely to play a role in setting  $\delta^{18}\text{O}_{\text{NO}_2}$  and  $\delta^{18}\text{O}_{\text{NO}_3}$  values observed in the ocean. This process does not change the concentration of  $\text{NO}_2^-$  nor its  $\delta^{15}\text{N}$  value, only its  $\delta^{18}\text{O}$  value. Oxygen atom equilibration shifts a  $\delta^{18}\text{O}_{\text{NO}_2}$  value from its biological starting point or “end member,” set by the isotopic systematics for biological production and consumption, toward the equilibrated  $\delta^{18}\text{O}_{\text{NO}_2}$  value, dictated by ambient  $\delta^{18}\text{O}_{\text{H}_2\text{O}}$  and the equilibrium isotope effect for the exchange ( $^{18}\epsilon_{\text{eq}}$ ), which is dependent on temperature (McIlvin and Casciotti, 2006; Buchwald and Casciotti, unpublished). The relevance of abiotic exchange depends on the rates of biological turnover of nitrite relative to the rate of oxygen atom exchange with water. Where nitrite turns over quickly and does not accumulate, there is little opportunity for abiotic exchange to occur. Where nitrite turns over more slowly (several weeks-months), abiotic exchange can play an important role in  $\delta^{18}\text{O}_{\text{NO}_2}$  and  $\delta^{18}\text{O}_{\text{NO}_3}$  (Buchwald et al., 2012).

The tendency of  $\text{NO}_2^-$  to exchange oxygen atoms abiotically with  $\text{H}_2\text{O}$  at typical seawater pH and temperature conditions suggests a utility of  $\text{NO}_2^-$  oxygen isotopes as a tracer for determining the rate of biological turnover of  $\text{NO}_2^-$  (Buchwald and Casciotti, unpublished). This provides a unique approach to determining rates of biological processes based on static isotope measurements, without bottle incubation and associated perturbations

of the system. Applications such as this move us from laboratory studies of isotope effects to a deeper understanding of the cycling of N in the environment. There are many additional examples of how knowledge of the isotope effects for nitrification has enabled advances in our understanding of the marine N cycle, and we highlight a few below.

## IMPLICATIONS FOR UNDERSTANDING N CYCLING IN OXYGEN DEFICIENT ZONES

As mentioned above, processes that occur in ODZs are important for the marine N budget. Both denitrification and anammox can occur in these regions, producing N<sub>2</sub> gas from dissolved inorganic nitrogen (DIN) compounds thereby removing them from the nutrient inventory. The magnitudes of these fluxes have been estimated in many different ways: through isotope tracer experiments (Kuypers et al., 2005; Thamdrup et al., 2006; Hamersley et al., 2007; Lam et al., 2009; Ward et al., 2009; Bulow et al., 2010; Jensen et al., 2011), as well as geochemical techniques based on NO<sub>3</sub><sup>-</sup> deficit calculations (Cline and Richards, 1972; Naqvi et al., 1982; Codispoti and Christensen, 1985; Naqvi and Sen Gupta, 1985; Gruber and Sarmiento, 1997; Deutsch et al., 2001) and biogenic N<sub>2</sub> production (Devol et al., 2006; Chang et al., 2010). The <sup>15</sup>N experiments in particular showcase a complex series of interacting processes cycling N in and around ODZs that can vary sporadically in space and time. What controls the overall rate of N<sub>2</sub> production is not known with certainty, although it is most likely tied directly or indirectly to organic carbon supply (Ward et al., 2008). Natural abundance stable isotopes provide an integrative longer-term view of the average rates of the major fluxes of N that can be used to complement short-term incubation studies. For example, natural abundance <sup>15</sup>N<sub>NO<sub>3</sub> and <sup>18</sup>O<sub>NO<sub>3</sub> measurements have been used to estimate the relative rates of N cycle processes such as N fixation and denitrification (Brandes et al., 1998; Sigman et al., 2005).</sub></sub>

Another aspect of N cycling in ODZs that is of great interest is the fate of NO<sub>2</sub><sup>-</sup> that is produced in ODZs. Once produced, NO<sub>2</sub><sup>-</sup> can be consumed through oxidation, regenerating NO<sub>3</sub><sup>-</sup>, or reduction to N<sub>2</sub> and loss from the nutrient inventory. Since nitrite oxidation is believed to be an oxygen requiring process, the fate of NO<sub>2</sub><sup>-</sup> in the oxygen deficient zone has generally been assumed to be through nitrite reduction. However, it has been shown through a variety of approaches that NO<sub>2</sub><sup>-</sup> can also be oxidized to NO<sub>3</sub><sup>-</sup> in and around ODZs. For example, early 1-D modeling studies suggested that a large fraction of NO<sub>2</sub><sup>-</sup> produced by nitrate reduction is reoxidized to NO<sub>3</sub><sup>-</sup>, likely on the fringes of the oxygen deficient zone (Anderson et al., 1982). More recent nutrient profile modeling suggests that NO<sub>2</sub><sup>-</sup> could be oxidized to NO<sub>3</sub><sup>-</sup> within the oxygen deficient zone itself (Lam et al., 2011). Furthermore, direct evidence for NO<sub>2</sub><sup>-</sup> oxidation to NO<sub>3</sub><sup>-</sup> within the ODZ comes from short-term <sup>15</sup>N incubation experiments (Lipschultz et al., 1990; Füssel et al., 2012).

The importance of nitrite oxidation as a sink of NO<sub>2</sub><sup>-</sup> in and around ODZs is supported by natural abundance isotope measurements of NO<sub>3</sub><sup>-</sup> and NO<sub>2</sub><sup>-</sup>, which integrate over longer periods. Sigman et al. (2005) and Casciotti and McIlvin (2007) found that nitrite oxidation could be an important sink for NO<sub>2</sub><sup>-</sup> at the top of the SNM based on <sup>15</sup>N<sub>NO<sub>3</sub>,</sub>

and <sup>18</sup>O<sub>NO<sub>3</sub> measurements. Casciotti (2009) also showed the need for nitrite oxidation to explain the large <sup>15</sup>N differences between NO<sub>3</sub><sup>-</sup> and NO<sub>2</sub><sup>-</sup> ( $\Delta\delta^{15}\text{N} = \delta^{15}\text{N}_{\text{NO}_3} - \delta^{15}\text{N}_{\text{NO}_2}$ ) observed within ODZs (Casciotti and McIlvin, 2007). Although the isotope effect for NO<sub>3</sub><sup>-</sup> reduction to NO<sub>2</sub><sup>-</sup> is approximately 25‰ (Brandes et al., 1998; Voss et al., 2001),  $\Delta\delta^{15}\text{N}$  values within the SNM ranged from 25‰ to 40‰ (Casciotti and McIlvin, 2007). At steady state,  $\Delta\delta^{15}\text{N}$  is given by equation 6:</sub>

$$\Delta\delta^{15}\text{N} = \delta^{15}\text{N}_{\text{NO}_3} - \delta^{15}\text{N}_{\text{NO}_2} = {}^{15}\epsilon_{k,\text{NAR}} - F_{\text{NXR}}/F_{\text{NAR}} \times {}^{15}\epsilon_{k,\text{NXR}} - F_{\text{NIR}}/F_{\text{NAR}} \times {}^{15}\epsilon_{k,\text{NIR}} \quad (6)$$

where F<sub>NAR</sub>, F<sub>NXR</sub>, and F<sub>NIR</sub> are the fluxes from nitrate reduction, nitrite oxidation, and nitrite reduction, respectively, and <sup>15</sup>ε<sub>k,NAR</sub>, <sup>15</sup>ε<sub>k,NXR</sub>, and <sup>15</sup>ε<sub>k,NIR</sub> are the respective N isotope effects. At steady state, the large  $\Delta\delta^{15}\text{N}$  values cannot be explained by reductive processes alone since nitrite reduction would be expected to increase <sup>15</sup>N<sub>NO<sub>2</sub>, thereby decreasing  $\Delta\delta^{15}\text{N}$  below 25‰. The only known mechanism for increasing  $\Delta\delta^{15}\text{N}$  above 25‰ is through NO<sub>2</sub><sup>-</sup> consumption with an inverse kinetic isotope effect, such as observed in nitrite oxidation (Casciotti, 2009; Buchwald and Casciotti, 2010). If all NO<sub>2</sub><sup>-</sup> consumption occurs through oxidation (F<sub>NXR</sub>/F<sub>NAR</sub> = 1) with a kinetic isotope effect of -15‰, then  $\Delta\delta^{15}\text{N}$  at steady state should approach 40‰. If all NO<sub>2</sub><sup>-</sup> consumption occurs through nitrite reduction (F<sub>NXR</sub>/F<sub>NAR</sub> = 0) with a kinetic isotope effect of +15‰, then  $\Delta\delta^{15}\text{N}$  would be expected to approach 10‰ at steady state. The <sup>15</sup>N difference between NO<sub>3</sub><sup>-</sup> and NO<sub>2</sub><sup>-</sup> may therefore be diagnostic of NO<sub>2</sub><sup>-</sup> sinks in ODZs (Casciotti, 2009).</sub>

While nitrite oxidation is generally considered to be an oxygen requiring process, O<sub>2</sub> is not required as an enzymatic substrate for nitrite oxidation. Rather, O<sub>2</sub> is used as an electron acceptor to support the oxidation of NO<sub>2</sub><sup>-</sup> to NO<sub>3</sub><sup>-</sup>. Therefore, if an alternative electron acceptor could be substituted, nitrite oxidation may proceed in the absence of O<sub>2</sub>. The alternate electron acceptors that can be used by NOB for nitrite oxidation remain to be determined, but oxidation of NO<sub>2</sub><sup>-</sup> by species such as iodate (IO<sub>3</sub><sup>-</sup>), Fe(III), and Mn(IV) would be thermodynamically feasible. Moreover, as mentioned above, there is independent evidence based on <sup>15</sup>N incubations for nitrite oxidation occurring within the ODZs in the ETSP (Lipschultz et al., 1990) and Namibian upwelling (Füssel et al., 2012). The presence of nitrite oxidizing bacteria from the genera *Nitrospina* and *Nitrococcus* comprising up to 9% of the microbial community in the Namibian upwelling (Füssel et al., 2012) also gives strong support to their success even in low oxygen environments.

Of course, even if nitrite oxidation is occurring in ODZs, more than one process may contribute, as both bacterial nitrite oxidizers and anammox bacteria can oxidize NO<sub>2</sub><sup>-</sup> to NO<sub>3</sub><sup>-</sup>. The contribution of anammox to nitrite oxidation can be estimated by comparison of F<sub>NXR</sub>/F<sub>NIR</sub> required to explain the isotopic data with that observed during anammox (0.26:1.06; Strous et al., 2006). This ratio places an upper limit on the amount of nitrite oxidation that could be catalyzed by anammox. If the ratio of nitrite oxidation to nitrite reduction necessary to explain

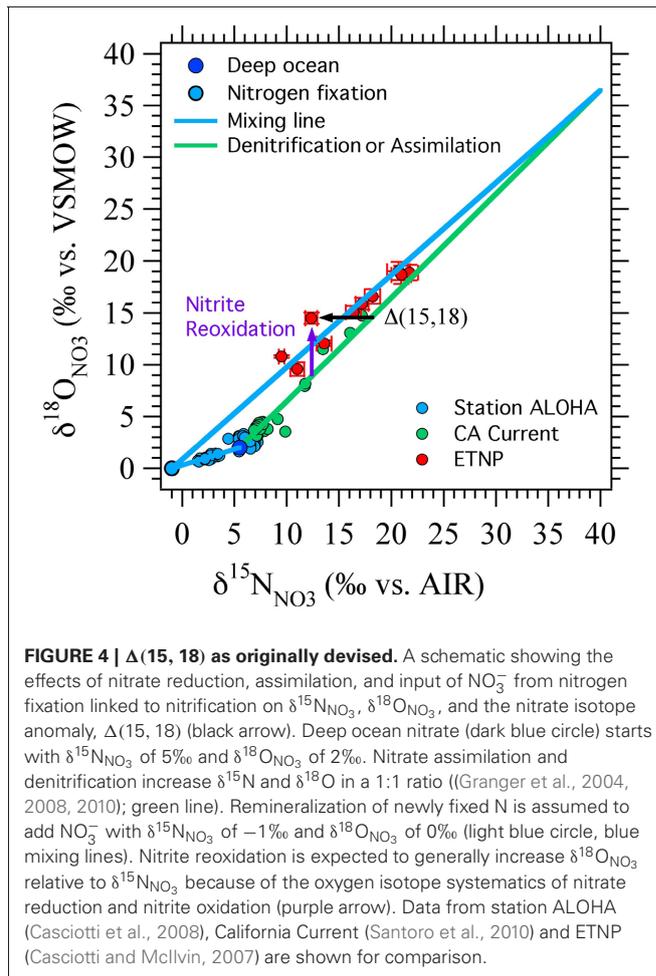
observed  $\Delta\delta^{15}\text{N}$  values is greater than this, then contributions from bacterial nitrite oxidation would be inferred (Casciotti, 2009). If the ratio of nitrite oxidation to nitrite reduction required to explain the isotopic data is less than this, then nitrite oxidation could potentially all be catalyzed by anammox, although denitrification may be required to explain the additional nitrite reduction. This analysis thus provides a new constraint on the relative rates of anammox and denitrification, integrated over long time periods. However, it assumes that the isotope effects for anammox are similar to denitrification for nitrite reduction and similar to nitrite oxidation for that step. Thus, the approach can be refined with additional information about the isotopic systematics of anammox.

### IMPLICATIONS FOR UNDERSTANDING $\text{NO}_3^-$ CYCLING AND BUDGETS: $\Delta(15, 18)$ REVISITED

Knowing the isotopic systematics of nitrification is critical for interpreting  $\delta^{18}\text{O}_{\text{NO}_3}$ ,  $\delta^{18}\text{O}_{\text{NO}_2}$ , and  $\delta^{18}\text{O}_{\text{N}_2\text{O}}$  measurements from the ocean. The culture studies described above have advanced our understanding of the oxygen isotope systematics of nitrification; however, there are also constraints from field data (Casciotti et al., 2002; Sigman et al., 2009). Casciotti et al. (2002) used the nitrate  $\delta^{18}\text{O}$  data to put the first constraints on the  $\delta^{18}\text{O}$  value of  $\text{NO}_3^-$  produced in the ocean. These estimates showed that  $\text{NO}_3^-$  is most likely produced with  $\delta^{18}\text{O}$  values close to those of seawater (0‰) and were used by Sigman et al. (2005) to constrain the rates of  $\text{N}_2$  fixation and nitrite reoxidation from  $\delta^{15}\text{N}_{\text{NO}_3}$  to  $\delta^{18}\text{O}_{\text{NO}_3}$  data. In order to do this, Sigman et al. (2005) introduced a  $\text{NO}_3^-$  isotope anomaly based on expected enrichments of  $\delta^{15}\text{N}_{\text{NO}_3}$  and  $\delta^{18}\text{O}_{\text{NO}_3}$  due to nitrate assimilation or nitrate reduction during denitrification:

$$\Delta(15, 18) = (\delta^{15}\text{N}_{\text{NO}_3} - \delta^{15}\text{N}_{\text{NO}_3,\text{deep}}) - {}^{18}\varepsilon_{\text{k,NAR}} / {}^{15}\varepsilon_{\text{k,NAR}} \times (\delta^{18}\text{O}_{\text{NO}_3} - \delta^{18}\text{O}_{\text{NO}_3,\text{deep}}) \quad (7)$$

where  $\delta^{15}\text{N}_{\text{NO}_3}$  and  $\delta^{18}\text{O}_{\text{NO}_3}$  are the measured isotopic values of the sample,  $\delta^{15}\text{N}_{\text{NO}_3,\text{deep}}$  and  $\delta^{18}\text{O}_{\text{NO}_3,\text{deep}}$  are the isotopic values of unaltered deep seawater, which define the starting point for fractionation.  ${}^{18}\varepsilon_{\text{k,NAR}}$  and  ${}^{15}\varepsilon_{\text{k,NAR}}$  are the isotope effects for O and N isotopes, respectively, during nitrate reduction. While there is a wide range in the absolute values of  ${}^{18}\varepsilon_{\text{k,NAR}}$  and  ${}^{15}\varepsilon_{\text{k,NAR}}$ , their ratio is very close to 1 (Granger et al., 2004, 2008, 2010). Therefore,  $\text{NO}_3^-$  consuming processes generally lead to  $\delta^{15}\text{N}_{\text{NO}_3}$  and  $\delta^{18}\text{O}_{\text{NO}_3}$  values that fall along a 1:1 line and produce samples with  $\Delta(15, 18) = 0\text{‰}$  (Figure 4). Non-zero  $\Delta(15, 18)$  values correspond to an enrichment of  $\delta^{18}\text{O}_{\text{NO}_3}$  relative to  $\delta^{15}\text{N}_{\text{NO}_3}$ , or a depletion in  $\delta^{15}\text{N}_{\text{NO}_3}$  relative to  $\delta^{18}\text{O}_{\text{NO}_3}$ , generally arising from production of  $\text{NO}_3^-$  with anomalous isotopic signatures. The most likely cause for depletion in  $\delta^{15}\text{N}$ , especially in the nitracline of oligotrophic ocean provinces, is through remineralization of newly fixed N with a  $\delta^{15}\text{N}$  value near  $-1\text{‰}$  (Capone et al., 1997; Karl et al., 1997; Meador et al., 2007). The particulate organic N produced by N fixation is remineralized to  $\text{NO}_3^-$  in the subsurface, gaining O atoms from nitrification, the same process that sets the oxygen isotopic signature of  $\text{NO}_3^-$  produced from other N sources. In scenario, the magnitude of

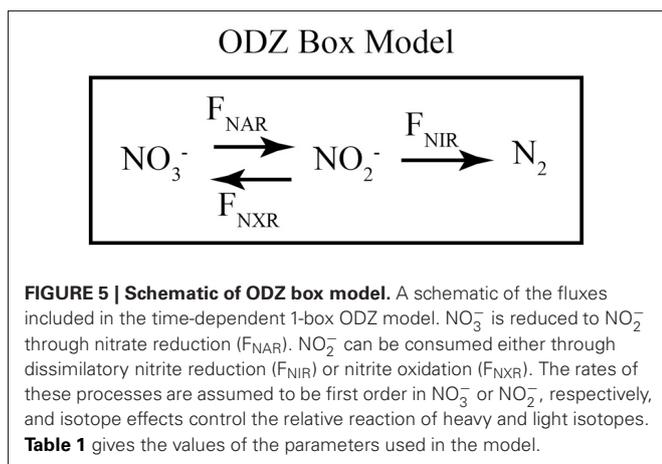


**FIGURE 4 |  $\Delta(15, 18)$  as originally devised.** A schematic showing the effects of nitrate reduction, assimilation, and input of  $\text{NO}_3^-$  from nitrogen fixation linked to nitrification on  $\delta^{15}\text{N}_{\text{NO}_3}$ ,  $\delta^{18}\text{O}_{\text{NO}_3}$ , and the nitrate isotope anomaly,  $\Delta(15, 18)$  (black arrow). Deep ocean nitrate (dark blue circle) starts with  $\delta^{15}\text{N}_{\text{NO}_3}$  of 5‰ and  $\delta^{18}\text{O}_{\text{NO}_3}$  of 2‰. Nitrate assimilation and denitrification increase  $\delta^{15}\text{N}$  and  $\delta^{18}\text{O}$  in a 1:1 ratio (Granger et al., 2004, 2008, 2010); green line). Remineralization of newly fixed N is assumed to add  $\text{NO}_3^-$  with  $\delta^{15}\text{N}_{\text{NO}_3}$  of  $-1\text{‰}$  and  $\delta^{18}\text{O}_{\text{NO}_3}$  of 0‰ (light blue circle, blue mixing lines). Nitrite reoxidation is expected to generally increase  $\delta^{18}\text{O}_{\text{NO}_3}$  relative to  $\delta^{15}\text{N}_{\text{NO}_3}$  because of the oxygen isotope systematics of nitrate reduction and nitrite oxidation (purple arrow). Data from station ALOHA (Casciotti et al., 2008), California Current (Santoro et al., 2010) and ETNP (Casciotti and McIlvin, 2007) are shown for comparison.

$\Delta(15, 18)$  would be proportional to the N fixation flux (Sigman et al., 2005).

A relative enrichment in  $^{18}\text{O}$ , especially in the vicinity of oceanic ODZs, could represent the cycling of  $\text{NO}_3^-$  through the reduction/reoxidation cycle, where the  $\text{NO}_3^-$  consumed by denitrification has a similar  $\delta^{15}\text{N}_{\text{NO}_3}$  but a lower  $\delta^{18}\text{O}_{\text{NO}_3}$  value than that returned to the  $\text{NO}_3^-$  pool from nitrite oxidation (Sigman et al., 2005). This formulation was successful at simulating data from regions of the ETNP where  $\text{NO}_2^-$  did not accumulate (Sigman et al., 2005) and where  $\text{NO}_2^-$  goes to zero at the top of the SNM (Casciotti and McIlvin, 2007). However, where  $\text{NO}_2^-$  accumulates, its isotopic composition can vary dramatically within the oxygen deficient zone itself (Casciotti and McIlvin, 2007), and an interpretation including  $\text{NO}_2^-$  isotope constraints is needed. The relationship between  $^{18}\text{O}$  enrichment in  $\text{NO}_3^-$  and the magnitude of the nitrite reoxidation flux depends critically on the N and O isotope systematics of nitrite oxidation, which we reviewed above. Here we revisit the implications of this new knowledge for interpretations of  $\Delta(15, 18)$  in euphotic zone and oxygen deficient zones.

Using a simple time-dependent 1-box model of the ODZ N cycle, we have reevaluated the impact of nitrite reoxidation on  $\delta^{15}\text{N}_{\text{NO}_3}$  and  $\delta^{18}\text{O}_{\text{NO}_3}$  in a hypothetical ODZ (Figure 5) and



show that nitrite oxidation can either raise or lower  $\Delta(15, 18)$ , depending on the relative  $\delta^{15}\text{N}$  and  $\delta^{18}\text{O}$  values of  $\text{NO}_2^-$  and  $\text{NO}_3^-$ . Our model focuses on determining the relative rates of  $\text{NO}_2^-$  reoxidation to  $\text{NO}_3^-$  ( $F_{\text{NXR}}$ ) and reduction (to  $\text{NO}$  or  $\text{NH}_4^+$ ;  $F_{\text{NIR}}$ ) from  $\text{NO}_3^-$  and  $\text{NO}_2^-$  isotopic data. The oxidative flux is assumed to have the N and O isotopic systematics of bacterial nitrite oxidation (Buchwald and Casciotti, 2010; **Table 1**), regardless of whether it is carried out by bacterial nitrite oxidizers or anammox bacteria, or some mixture of the two. The reductive processes are assumed to have  $^{15}\epsilon = ^{18}\epsilon = 15\text{‰}$  (**Table 1**) regardless of whether  $\text{NO}_2^-$  is reduced to  $\text{N}_2$  (via anammox or denitrification) or  $\text{NH}_4^+$  [via denitrification to ammonium (DNRA)]. Unfortunately, very little information is currently available on the N isotope effects for nitrite reduction by these processes (Bryan et al., 1983) and no information is available for the O isotope effects. In the absence of

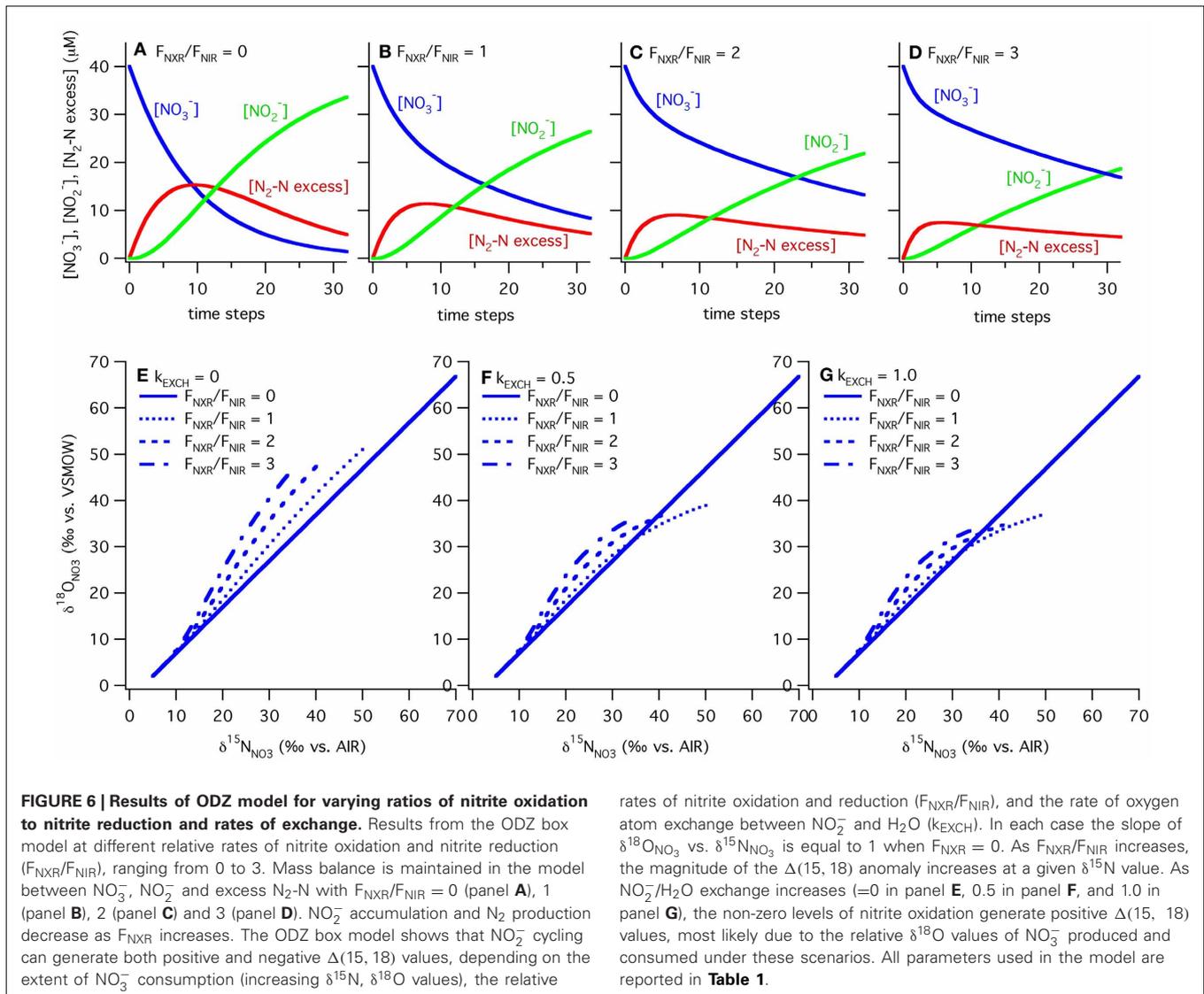
more specific information, we make the simplifying assumption that the different nitrite reductase enzymes have similar N and O isotope effects. Clearly, this is an important area of future research.

In our model, the processes are all represented as first order, and the rate constants ( $k$ 's) are given in units of  $\text{day}^{-1}$  to match measured rates of nitrate reduction, nitrite reduction, and nitrite oxidation in ODZs (**Table 1**). The isotope effects taken from the literature are also given in **Table 1**. We vary the relative rates of nitrite oxidation and nitrite reduction ( $F_{\text{NXR}}/F_{\text{NIR}}$ ) between 0 and 3 ( $F_{\text{NXR}}$  representing 0–75% of  $\text{NO}_2^-$  consumption) and the rate constant for exchange ( $k_{\text{EXCH}}$ ) between 0 and 1  $\text{day}^{-1}$  to evaluate the effects of changes in these parameters on simulated  $\delta^{15}\text{N}_{\text{NO}_3}$  and  $\delta^{18}\text{O}_{\text{NO}_3}$  (**Figure 6**). Maximum rate constants of exchange between  $\text{NO}_2^-$  and  $\text{H}_2\text{O}$  of 1  $\text{day}^{-1}$  appear reasonable based on recent laboratory studies (Casciotti et al., 2007; Buchwald and Casciotti, unpublished). As  $F_{\text{NXR}}/F_{\text{NIR}}$  increases from 0 to 3, the amount of  $\text{NO}_3^-$  retained in the system increases despite an unchanging rate constant for nitrate reduction. In fact, because the reaction is taken as first order, the higher concentrations of  $\text{NO}_3^-$  brought about by higher levels of  $F_{\text{NXR}}$  lead to higher overall rates of nitrate reduction. However, it is clear from the mass balances in the different scenarios that nitrite reoxidation helps buffer against excessive loss of  $\text{NO}_3^-$ , accumulation of  $\text{NO}_2^-$ , and production of  $\text{N}_2$  (**Figures 6A–D**), and may help explain why  $\text{NO}_3^-$  is never fully removed in oceanic ODZs.

The magnitude of nitrite oxidation also affects the  $\delta^{15}\text{N}_{\text{NO}_3}$  and  $\delta^{18}\text{O}_{\text{NO}_3}$  patterns. When  $F_{\text{NXR}}/F_{\text{NIR}} = 0$ , the  $\delta^{15}\text{N}_{\text{NO}_3}$  and  $\delta^{18}\text{O}_{\text{NO}_3}$  data fall along the 1:1 line prescribed by the isotope effects for nitrate reduction (**Figures 6E–G**). As  $F_{\text{NXR}}/F_{\text{NIR}}$  increases, increasingly negative  $\Delta(15, 18)$  values are produced. The strength of this effect is also dependent on the rate of

**Table 1 | Parameters used in oxygen deficient zone box model.**

Parameter	Description	Value	Reference
$\delta^{15}\text{N}_{\text{NO}_3, \text{initial}}$	Initial nitrate $\delta^{15}\text{N}$	5‰	Sigman et al., 2000
$\delta^{18}\text{O}_{\text{NO}_3, \text{initial}}$	Initial nitrate $\delta^{18}\text{O}$	2‰	Casciotti et al., 2002
$\delta^{18}\text{O}_{\text{H}_2\text{O}}$	Water $\delta^{18}\text{O}$ value	0‰	Craig and Gordon, 1965
$k_{\text{NAR}}$	First order rate constant for nitrate reduction	0.001 $\text{day}^{-1}$	Estimated to achieve a rate of 20 nM $\text{day}^{-1}$ ; Lam et al., 2011
$k_{\text{NXR}}$	First order rate constant for nitrite oxidation	0–0.003 $\text{day}^{-1}$	Estimated to achieve range of observed nitrite oxidation rates; Füssel et al., 2012; Lipschultz et al., 1990
$k_{\text{NIR}}$	First order rate constant for nitrite reduction	0.001 $\text{day}^{-1}$	Estimated to achieve a rate of 5 nM $\text{day}^{-1}$ ; Devol et al., 2006
$k_{\text{EXCH}}$	First order rate constant for nitrite/water exchange	0.01 $\text{day}^{-1}$	Buchwald and Casciotti
$^{15}\alpha_{k, \text{NAR}}$	N isotope effect for nitrate reduction	1.019	Deutsch et al., 2004; Granger et al., 2008
$^{15}\alpha_{k, \text{NXR}}$	N isotope effect for nitrite oxidation	0.985	Casciotti, 2009; Buchwald and Casciotti, 2010
$^{15}\alpha_{k, \text{NIR}}$	N isotope effect for nitrite reduction	1.015	Bryan et al., 1983
$^{18}\alpha_{\text{NAR}}$	O isotope effect for nitrate reduction	1.019	Granger et al., 2008
$^{18}\alpha_{k, \text{NXR}}$	O isotope effect for nitrite oxidation	0.997	Buchwald and Casciotti, 2010
$^{18}\alpha_{k, \text{NIR}}$	O isotope effect for nitrite reduction	1.015	Sigman et al., 2005
$^{18}\alpha_{k, \text{H}_2\text{O}, 2}$	O isotope effect for $\text{H}_2\text{O}$ incorporation	1.010	Buchwald and Casciotti, 2010
$^{18}\alpha_{\text{B}}$	Branching O isotope effect during nitrate reduction	0.975	Casciotti et al., 2007
$^{18}\alpha_{\text{eq}}$	Equilibrium isotope effect for nitrite/water O exchange	1.014	Casciotti et al., 2007; (Buchwald and Casciotti, unpublished)

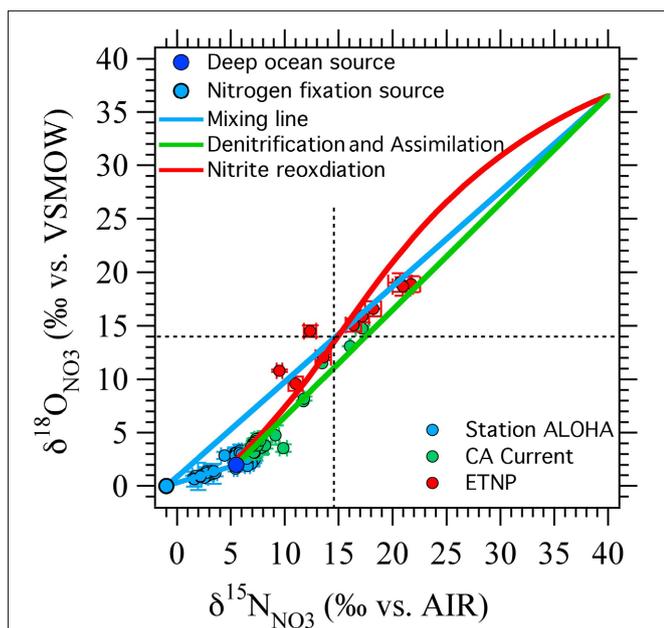


abiotic  $\text{NO}_2^-/\text{H}_2\text{O}$  exchange, with higher exchange rates partly diluting this effect and actually leading to positive  $\Delta(15, 18)$  values at high extents of  $\text{NO}_3^-$  consumption (the highest  $\delta^{15}\text{N}_{\text{NO}_3}$  values; **Figure 6**). This interesting phenomenon is most likely due to reversal of the impact of nitrite reoxidation on  $\delta^{18}\text{O}_{\text{NO}_3}$  at high  $\delta^{18}\text{O}_{\text{NO}_3}$  values, with nitrite oxidation returning  $\text{NO}_3^-$  with a lower  $\delta^{18}\text{O}_{\text{NO}_3}$  value than that removed by nitrite reduction. This would be exacerbated at high rates of exchange, which helps to maintain  $\delta^{18}\text{O}_{\text{NO}_2}$  values at a constant level regardless of  $\delta^{18}\text{O}_{\text{NO}_3}$ . Tuning the model to match observed  $\delta^{18}\text{O}_{\text{NO}_2}$  data requires a high rate of exchange relative to biological fluxes, and therefore most closely follows the  $k_{\text{EXCH}} = 1$  scenario.

Larger ratios of  $F_{\text{NXR}}/F_{\text{NIR}}$  could be imagined, but the model results from such simulations produce unrealistic  $\Delta(15, 18)$  anomalies at a given  $\delta^{15}\text{N}_{\text{NO}_3}$  value. Furthermore, because excess  $\text{N}_2$  does accumulate in ODZs, we know that some  $\text{NO}_2^-$  is ultimately reduced to  $\text{N}_2$ . Indeed, we could potentially use the

stoichiometry of  $\text{N}_2$  production in ODZs to interrogate the importance of nitrite oxidation. If nitrite oxidation is not important, the standard stoichiometry (Richards, 1965; Devol et al., 2006) of 106  $\text{CO}_2$ : 55.2  $\text{N}_2$  would be expected, whereas higher amounts of  $\text{CO}_2$  would be expected if a significant fraction of the produced  $\text{NO}_2^-$  is reoxidized to  $\text{NO}_3^-$ . This may seem counterintuitive because autotrophic nitrite oxidation should fix  $\text{CO}_2$  back into organic matter, but the excess  $\text{NO}_3^-$  reduction required to supply the  $\text{NO}_2^-$  in the first place should far outweigh the  $\text{CO}_2$  fixed by nitrite oxidation.

It is interesting to note that the two scenarios for producing negative  $\Delta(15, 18)$  values ( $\text{N}_2$  fixation and nitrite reoxidation) are each more effective at different points in  $\text{NO}_3^-$  isotope space (**Figure 7**).  $\text{N}_2$  fixation is most effective at generating negative  $\Delta(15, 18)$  signals at  $\delta^{15}\text{N}_{\text{NO}_3}$  and  $\delta^{18}\text{O}_{\text{NO}_3}$  values less than 10‰, near the base of the euphotic zone. In contrast, nitrite reoxidation is most effective at generating negative  $\Delta(15, 18)$  signals at intermediate  $\delta^{15}\text{N}_{\text{NO}_3}$  and  $\delta^{18}\text{O}_{\text{NO}_3}$  values and extents of  $\text{NO}_3^-$



**FIGURE 7 | Results from ODZ model compared against published data.**

Nitrogen fixation (blue circle and blue lines) is efficient at generating negative  $\Delta(15, 18)$  values at the lower range of  $\delta^{15}\text{N}_{\text{NO}_3}$  and  $\delta^{18}\text{O}_{\text{NO}_3}$  values, while nitrite reoxidation (red line) has a stronger effect at higher  $\delta^{15}\text{N}_{\text{NO}_3}$  and  $\delta^{18}\text{O}_{\text{NO}_3}$  values. The nitrite reoxidation curve shown here was generated from the box model with  $k_{\text{EXCH}} = 0.5$  and  $F_{\text{NKR}}/F_{\text{NIR}} = 2$  (Figure 6F, dashed line). Data from station ALOHA (Casciotti et al., 2008) in the north Pacific subtropical gyre is well explained by an input from  $\text{N}_2$  fixation. Data from the California Current (Santoro et al., 2010) falls close to the 1:1 line suggesting little influence of nitrogen fixation or nitrite oxidation in the euphotic zone. Most of the data from the ETNP (Casciotti and McIlvin, 2007) could be explained by either nitrite reoxidation or nitrogen fixation, but two points (which fall in the shallow oxycline at the top of the SNM) require inputs from both nitrite reoxidation and nitrogen fixation.

consumption by denitrification, where  $\text{N}_2$  fixation has relatively little effect on the  $\Delta(15, 18)$ . Therefore, we may be able to distinguish between the processes responsible for  $\Delta(15, 18)$  generation by where the anomaly lies in  $\delta^{15}\text{N}_{\text{NO}_3}$  vs.  $\delta^{18}\text{O}_{\text{NO}_3}$  space, as well as from other water column indicators. For example, using a steady state model, Casciotti and McIlvin (2007) showed that the  $\text{NO}_3^-$  isotope anomaly at the top of the SNM could not be generated by  $\text{N}_2$  fixation alone and was consistent with oxidation of  $\text{NO}_2^-$  leaking out of the top of the SNM. However, they suggested that a combination of  $\text{N}_2$  fixation and nitrite reoxidation may best fit the observations. This conclusion is echoed here where it is difficult to generate large  $\Delta(15, 18)$  signals at these  $\delta^{15}\text{N}_{\text{NO}_3}$  and  $\delta^{18}\text{O}_{\text{NO}_3}$  values through either  $\text{N}_2$  fixation or nitrite reoxidation alone (Figure 7).

In addition to oxygen deficient zone and near-surface processes,  $\text{NO}_3^-$  isotopes have also been used to examine the global ocean cycle and budget of  $\text{NO}_3^-$  in the ocean interior (Sigman et al., 2009). This was done using an 18-box model of the global ocean where the implications of different assumptions about the oxygen isotopic systematics of nitrification could be tested. Their model was also used to constrain the relative rates of the internal N cycle ( $\text{NO}_3^-$  uptake, export, and nitrification) and N budget

processes ( $\text{N}_2$  fixation and denitrification) and the ratio of low latitude productivity, where nutrient consumption goes to completion, to high latitude productivity, where nutrient uptake is incomplete. By comparing model results to  $\delta^{15}\text{N}_{\text{NO}_3}$  and  $\delta^{18}\text{O}_{\text{NO}_3}$  data from a variety of oceanographic profiles representing the major ocean basins, the impacts of partial  $\text{NO}_3^-$  assimilation in polar regions on the N and O isotopes of  $\text{NO}_3^-$  in the ocean interior, and of low latitude productivity on the  $^{18}\text{O}$  enrichment in preformed  $\text{NO}_3^-$  was diagnosed. N budget processes ( $\text{N}_2$  fixation and denitrification) led to variations in subsurface  $\delta^{15}\text{N}_{\text{NO}_3}$  and  $\delta^{18}\text{O}_{\text{NO}_3}$ , but in their absence, the large scale steady state  $\delta^{18}\text{O}$  value of subsurface  $\text{NO}_3^-$  was set by nitrate assimilation in polar regions. Nitrate uptake in the southern ocean leads to heavy isotope enrichment in preformed  $\text{NO}_3^-$ , while nitrate assimilation in low latitudes removes the  $\delta^{18}\text{O}$  signal of the preformed  $\text{NO}_3^-$  and replaces it with the nitrification signal (Sigman et al., 2009). Overall, when only internal processes were active in the model, the mean ocean  $\delta^{18}\text{O}_{\text{NO}_3}$  value was 1.1‰ higher than the nitrification source. When the N budget was added to the model, the mean ocean  $\delta^{18}\text{O}_{\text{NO}_3}$  value was 2.4‰ higher than the nitrification source value. This analysis provides additional constraints on the  $\delta^{18}\text{O}$  value of newly produced  $\text{NO}_3^-$  in the ocean to fall between  $-1\%$  and  $+1\%$  (Sigman et al., 2009), which is consistent with culture studies that illustrate how these values are controlled biochemically (Buchwald et al., 2012).

## NITROGEN CYCLING IN THE EUPHOTIC ZONE

Several studies have now used N and O isotope ratio measurements to study the relative rates of N cycling in the euphotic zone. In particular, knowledge of the isotopic systematics of nitrate uptake (Granger et al., 2004, 2010) and nitrification (Buchwald and Casciotti, 2010; Casciotti et al., 2010, 2011; Buchwald et al., 2012) enables the assessment of the relative rates of nitrification and nitrate uptake from euphotic zone  $\text{NO}_3^-$  isotope data.

Wankel et al. (2007) used a steady-state box model to interpret the amount of nitrification contributing to nitrate uptake by phytoplankton in Monterey Bay, CA using  $\delta^{15}\text{N}_{\text{NO}_3}$  and  $\delta^{18}\text{O}_{\text{NO}_3}$  variations. Assuming that nitrate assimilation leads to equivalent fractionation of N and O isotopes (Granger et al., 2004), and that  $\delta^{18}\text{O}_{\text{Ntr}} = 2.9\%$ , they estimated that nitrification could supply up to 30% of  $\text{NO}_3^-$  assimilated by phytoplankton in Monterey Bay, consistent with intensive isotope tracer incubation studies (Ward, 2005). Because  $\delta^{18}\text{O}_{\text{Ntr}}$  was uncertain at that time, they performed sensitivity studies to address the impact of different  $\delta^{18}\text{O}_{\text{Ntr}}$  values on their interpretation. We now believe that  $\delta^{18}\text{O}_{\text{Ntr}}$  is between  $-1\%$  and  $+1\%$  (Buchwald et al., 2012), and applying this to the model from Wankel et al. (2007), leads to a smaller increase in  $\delta^{18}\text{O}_{\text{NO}_3}$  for the same amount of nitrification. Thus, to achieve the same  $\delta^{18}\text{O}_{\text{NO}_3}$  enrichment in their model requires more nitrification than originally estimated.

DiFiore and colleagues (2009) estimated the amount of nitrification contributing to nitrate uptake in the euphotic zone of the Polar Antarctic Zone using a time-dependent 1-box model. Like Wankel et al. (2007), they assumed that  $^{18}\text{ENR} = ^{15}\text{ENR}$  for nitrate

uptake and allowed branching of  $\text{NH}_4^+$  (and  $\text{NO}_2^-$ ) between nitrification and assimilation to partition isotopes between the  $\text{NO}_3^-$  and particulate N pools. One important difference from the Wankel et al. (2007) model is that they assumed  $\delta^{18}\text{O}_{\text{ntr}} = +1.1\text{‰}$  based on more recent constraints on this value (Sigman et al., 2009). They inferred that  $\delta^{15}\text{N}_{\text{NO}_3}$  should be lowered slightly due to nitrification (offsetting the isotopic fractionation during uptake) and  $\delta^{18}\text{O}_{\text{NO}_3}$  should be raised (because the  $\delta^{18}\text{O}$  of newly produced  $\text{NO}_3^-$  was higher than that removed). Both of these factors should lead to negative  $\Delta(15, 18)$  values, as discussed above, but they found that nitrification had a relatively small impact on  $\delta^{15}\text{N}_{\text{NO}_3}$  and  $\delta^{18}\text{O}_{\text{NO}_3}$  values in the Polar Antarctic Zone. They concluded that in the Polar Antarctic Zone less than 1% of  $\text{NO}_3^-$  assimilated by phytoplankton is likely to have been produced by nitrification in the euphotic zone (DiFiore et al., 2009). This is consistent with other estimates from the southern ocean (Olson, 1981b; Bianchi et al., 1997; Law and Ling, 2001) and quite a bit lower than other regions (Yool et al., 2007; Wankel et al., 2007; Clark et al., 2008). This elegant study provides an excellent example of how  $\text{NO}_3^-$  isotopes can be used to constrain N cycle processes in an appropriate model framework.

$\text{NO}_3^-$  and  $\text{NO}_2^-$  isotopes have also been used to understand the sources and cycling of  $\text{NO}_2^-$  in the PNM at the base of the euphotic zone. Mackey et al. (2011) used natural abundance  $\text{NO}_3^- + \text{NO}_2^-$  isotope data and isotope tracer experiments to determine the sources of  $\text{NO}_2^-$  to the PNM in the Gulf of Aqaba. They found active nutrient regeneration and nitrification throughout the water column. In the transition from well mixed to stratified conditions,  $\text{NO}_2^-$  was generated by incomplete  $\text{NO}_3^-$  reduction by light-limited phytoplankton creating a broad band of  $\text{NO}_2^-$ . After stratification was established,  $\text{NO}_2^-$  generation by ammonia oxidation contributed to maintenance of the PNM. In both cases,  $\text{NO}_2^-$  was consumed by nitrite oxidation below the PNM. Once again, nitrification was interpreted to play an important role in  $\text{NO}_3^-$  isotope dynamics in the upper water column where increases in  $\delta^{18}\text{O}_{\text{NO}_3}$  were much higher than increases in  $\delta^{15}\text{N}_{\text{NO}_3}$ .

In another recent study of PNM dynamics, natural abundance  $\delta^{18}\text{O}_{\text{NO}_2}$  and  $\delta^{15}\text{N}_{\text{NO}_2}$  values were used to infer the sources and average age of  $\text{NO}_2^-$  in the PNM of the Arabian Sea (Buchwald and Casciotti, unpublished). Because the  $\delta^{15}\text{N}_{\text{NO}_2}$  and  $\delta^{18}\text{O}_{\text{NO}_2}$  values produced from ammonia oxidation and nitrate reduction are distinct, the sources can be readily distinguished. Based on natural abundance  $\delta^{15}\text{N}_{\text{NO}_2}$  and  $\delta^{18}\text{O}_{\text{NO}_2}$  data, ammonia oxidation was inferred to be the main source of  $\text{NO}_2^-$  to the PNM in the Arabian Sea.

### IMPLICATIONS FOR INTERPRETING $\text{N}_2\text{O}$ SOURCES

Uncertainty in the isotopic composition of  $\text{N}_2\text{O}$  produced during ammonia oxidation has hampered the interpretation of near-surface  $\text{N}_2\text{O}$  production rates and fluxes using two-component end member models (Dore et al., 1998; Popp et al., 2002; Santoro et al., 2010). Better understanding of the oxygen isotopic systematics of nitrification can provide further insight into outstanding questions in  $\text{N}_2\text{O}$  oxygen isotope variations, such as why  $\delta^{18}\text{O}_{\text{N}_2\text{O}}$  in seawater is so high (Ostrom et al., 2000; Popp et al., 2002), what

mechanisms of  $\text{N}_2\text{O}$  production operate in oxyclines surrounding oceanic ODZs (Codispoti and Christensen, 1985), and what the mechanisms and controls on  $\text{N}_2\text{O}$  production are in the near-surface ocean (Dore et al., 1998; Popp et al., 2002; Santoro et al., 2011).

For example,  $\text{N}_2\text{O}$  production in the near-surface ocean is largely believed to be the result of nitrification. However, the isotopic composition of  $\text{N}_2\text{O}$  in the near surface and the inferred near surface source (Dore et al., 1998) have higher  $\delta^{15}\text{N}$  and  $\delta^{18}\text{O}$  values than are characterized by bacterial ammonia oxidation (Yoshida, 1988; Frame and Casciotti, 2010). Recent evidence suggests that AOA are important for nitrification in such environments (Wuchter et al., 2006; Beman et al., 2008; Mincer et al., 2007; Church et al., 2010; Santoro et al., 2010) and that they produce  $\text{N}_2\text{O}$  with bulk  $\delta^{15}\text{N}$  and  $\delta^{18}\text{O}$  values similar to the near-surface source (Santoro et al., 2011). These data support a role for them in near-surface  $\text{N}_2\text{O}$  production. As discussed above, the mechanisms of  $\text{N}_2\text{O}$  production by AOA are currently unknown, and more work is needed to characterize the  $\text{N}_2\text{O}$  production and isotopic composition of marine AOA under a variety of growth conditions. For example, the SP of  $\text{N}_2\text{O}$  produced by AOB varies widely with dissolved oxygen levels (Frame and Casciotti, 2010) but so far the isotopic composition of  $\text{N}_2\text{O}$  produced by AOA has only been examined under aerobic growth conditions (Santoro et al., 2011; Loescher et al., 2012). Therefore, we do not know whether they are capable of producing  $\text{N}_2\text{O}$  with a SP similar to near surface  $\text{N}_2\text{O}$  (Popp et al., 2002).

### CONCLUDING REMARKS

Understanding the nitrogen and oxygen isotopic systematics of nitrification can contribute greatly to our understanding of nitrogen cycling in the ocean, as nitrification is involved with transformations between the major pools of DIN ( $\text{NH}_4^+$ ,  $\text{NO}_2^-$ ,  $\text{NO}_3^-$ , and  $\text{N}_2\text{O}$ ). Both ammonia and nitrite oxidation are involved with large and distinctive isotope effects, leading to predictable patterns in the isotope ratios of compounds that they transform. The discovery of AOA and their importance in ocean biogeochemistry necessitates renewed study of the isotopic systematics of nitrification. In preliminary studies, the isotopic systematics of AOA appear similar to AOB for N isotope fractionation and O atom incorporation into  $\text{NO}_2^-$  (Santoro and Casciotti, 2011; Santoro et al., 2011). However, the production of  $\text{N}_2\text{O}$  and the isotopic systematics of this process need to be further investigated.

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