Check for updates

#### **OPEN ACCESS**

EDITED BY Yul Roh, Chonnam National University, Republic of Korea

REVIEWED BY Yuanfeng Cai, Nanjing University, China Ji-Hoon Lee, Jeonbuk National University, Republic of Korea

\*CORRESPONDENCE Jixia Zhao ⊠ zhaoiixiacc@163.com

<sup>†</sup>These authors have contributed equally to this work and share first authorship

RECEIVED 13 February 2023 ACCEPTED 05 June 2023 PUBLISHED 29 June 2023

#### CITATION

Li C, Fan M, Wang X, Li X, Zhao G, Liu G and Zhao J (2023) Effects of nitrogen fertilizers on the bacterial community diversity and the weathering of purple mudstone in Southwest China. *Front. Microbiol.* 14:1164826.

doi: 10.3389/fmicb.2023.1164826

### COPYRIGHT

© 2023 Li, Fan, Wang, Li, Zhao, Liu and Zhao. This is an open-access article distributed under the terms of the Creative Commons Attribution License (CC BY). The use, distribution or reproduction in other forums is permitted, provided the original author(s) and the copyright owner(s) are credited and that the original publication in this journal is cited, in accordance with accepted academic practice. No use, distribution or reproduction is permitted which does not comply with these terms.

## Effects of nitrogen fertilizers on the bacterial community diversity and the weathering of purple mudstone in Southwest China

Chunpei Li<sup>1†</sup>, Maopan Fan<sup>1†</sup>, Xuan Wang<sup>2,3</sup>, Xue Li<sup>1</sup>, Guang Zhao<sup>4</sup>, Gangcai Liu<sup>2</sup> and Jixia Zhao<sup>1\*</sup>

<sup>1</sup>College of Resources and Environment, Yunnan Agricultural University, Kunming, China, <sup>2</sup>Key Laboratory of Mountain Surface Processes and Ecological Regulation, Chinese Academy of Sciences, Institute of Mountain Hazards and Environment, Chinese Academy of Sciences, and Ministry of Water Conservancy, Chengdu, China, <sup>3</sup>University of the Chinese Academy of Sciences, Chinese Academy of Sciences, Beijing, China, <sup>4</sup>Key Laboratory of Ecosystem Network Observation and Modeling, Institute of Geographic Sciences and Natural Resources Research, Chinese Academy of Sciences, Beijing, China

**Introduction:** Rock weathering is crucial in the development of soil. Yet the role of bacteria in the fine particle-forming process of purple mudstone is not fully understood, especially under nitrogen fertilization.

**Methods:** In this study, the particles (0.25mm to 1mm) of purple mudstone from Penglai Group ( $J_{3p}$ ) were selected as the test material. Two nitrogen fertilizers, i.e., urea (U) and ammonium bicarbonate (AB), and four application levels (0, 280, 560, and 840Nkg•ha<sup>-1</sup>) with 18 replications were designed in an incubation experiment. The weathering indices and bacterial community structure of the purple mudstone particles were investigated after 120 days of incubation.

**Results:** The results showed that the weathering indices of purple mudstone particles in the AB treatment were higher than that in the U treatment at the same fertilization levels and a reducing trend was observed with increasing nitrogen fertilizer levels under the same nitrogen fertilizer application types. The diversities of the bacterial community were extremely significantly altered by nitrogen fertilizer application (*p*<0.01). The effect of the nitrogen fertilizer application level on the beta diversity of the bacterial community (*R*<sup>2</sup>=0.34) was greater than that of the nitrogen fertilizer application type (*R*<sup>2</sup>=0.20). Through stepwise regression analysis, the positive effects of nitrification of Nitrobacter (*Nitrolancea*) (*R*<sup>2</sup>=0.35) on the weathering indices of J<sub>3</sub>p purple mudstone particles could be observed. Structural equation modelling indicated that nitrogen fertilizer application level affects the abundance of the dominant species at the genus level (*Nitrolancea* and *Massilia*), and key environmental factor (N-NO<sub>3</sub><sup>-</sup>), which in turn accelerated the weathering indices (59%).

**Discussion and Conclusion:** Our findings imply that the enhancements of nitrification of Nitrobacter (*Nitrolancea*) and of phosphorus solubilization of Phosphorous-dissolving bacteria (*Massilia*) by nitrogen fertilization are the key factors affecting the weathering indices of  $J_{3p}$  purple mudstone particles.

#### KEYWORDS

purple mudstone weathering, bacterial community structure, urea, ammonium bicarbonate, weathering indices

### 1. Introduction

Weathering of parent rock is an integral part of the Earth's surface processes of soil formation and plays a critical role in maintaining terrestrial ecosystems (Dixon et al., 2012; Zhu et al., 2013; Israeli et al., 2021), as well as controlling the long-term chemical composition of groundwaters, rivers, lakes, and oceans (Duan et al., 2002; Godderis et al., 2009). The weathering of rocks has long been a research focus in geosciences, as it also plays a critical role in the global carbon cycle (Li and Elderfield, 2013; Torres et al., 2016) and in regulating atmospheric CO<sub>2</sub> on geological time scales (Them et al., 2017). It has even been suggested that enhanced weathering would curb present-day anthropogenic carbon emissions (Xu and Liu, 2010; Torres et al., 2016; Strefler et al., 2018; Beerling et al., 2020). As the key process of soil formation, weathering of parent rocks is an aspect that cannot be ignored in influencing global climate change.

The weathering rate of rock is affected by both natural (e.g., geological settings and topographical conditions, the provenance of source rock, temperature, precipitation, and chemical composition) and artificial factors (Köhler et al., 2003; Kirschbaum et al., 2005; Caspari et al., 2006; Chetelat et al., 2008; Huang et al., 2013; Moore et al., 2013). Mass balance calculations indicated that the contribution of anthropogenic factors to parent rock weathering accounts for 16-40% (Liu et al., 2018). Analyzing the hydrochemistry in regional rivers indicated that nitric acid and sulfuric acid produced by anthropogenic activities promote the weathering of minerals (Yu et al., 2016) and increase the weathering rate of carbonate rocks by approximately 6% (Zhu et al., 2019). Some studies indicated that exogenous H<sup>+</sup> addition induces clay minerals dissolving, i.e., Montmorillonite, Kaolinite, Mica, and Illite (Tombácz and Szekeres, 2006; Zhang et al., 2009; Sokolova, 2013; Bibi et al., 2014), while more difficult-to-dissolve minerals are retained, i.e., Quartz.

Fertilization, an artificial disturbance factor in agricultural production activities, is becoming increasingly prominent, and will affect soil and other resources, including the impact on the weathering and soil formation of the parent rock. Nitrogen fertilizer is the most commonly used fertilizer, accounting for almost 50% of fertilizer consumption, and the amount of nitrogen application in China accounts for 32% of the world at present (Yu et al., 2019). According to the National Statistical Yearbook, the amount of nitrogen fertilizer applied in China increased from 93.4 kg•ha<sup>-1</sup> in 1980 to 153.0 kg•ha<sup>-1</sup> in 2018. The long-term large-scale or excessive use of nitrogen fertilizer application, especially ammonium nitrogen fertilizer, can cause soil acidification (Chao et al., 2011; Goulding, 2016; Yang et al., 2016), which is widely known to affect the weathering of parent rock (Zhao et al., 2021). Furthermore, the application of nitrogen fertilizers has a critical effect on soil microbial processes, key microbial community structures, and N availability (Zeng et al., 2015; Gai et al., 2016). Previous studies also implied nitrification of fertilizer nitrogen could prominently promote the weathering of minerals (Pacheco et al., 2013). Although microbial action is one of the most significant factors affecting the weathering of rocks and minerals (Krumbein, 1983), the alterations of bacterial community structure and pH on the weathering of parent rocks driven by nitrogen fertilizer addition is not fully understood yet.

Purple rock is a sedimentary rock, mainly distributed in the upper Yangtze River in China, especially in Sichuan and Yunnan provinces. The coverage of purple soil in these two provinces accounts for over 75% of the national purple soil area, of which Sichuan Province accounts for 51.5% and Yunnan Province accounts for 23.4% (Liu, 2008). Due to its fast physical weathering, purple rock is easily broken into rock particles or gravel by environmental factors as well as anthropogenic activities, and crops can be planted directly on weathered products. The 1 mm particle formation process was dominated by physical and chemical weathering. However, biological weathering was enhanced in the <1 mm particles formation process. Soil nutrients derived from purple rocks are largely from the weathering processes, especially P- and K-rich minerals. Purple rock is generally considered the nutrient source for purple soil (Gao et al., 1999; He, 2003). The purple rock area has practical significance in food and ecological security in China. N fertilization is a major practice in crop production in this area. Long-term application of nitrogen fertilizer causes acidification of purple soil (Huang, 2012). A number of previous studies have been carried out on physical and chemical weathering (Gao et al., 1999; Zhang et al., 2017; Zhao et al., 2021), but relatively few studies have been reported on the effects of biological weathering on purple rocks under N application.

In order to understand the bacterial effects driven by nitrogen fertilizer application on the fine particle formation of purple mudstone particles, an indoor incubation experiment was carried out in this study. Our aim was to provide information for purple soil formation under current agricultural production.

### 2. Materials and methods

### 2.1. Rock sample collection

The purple mudstone of the Penglai Group  $(J_3p)$  (31°16'N, 105°27'E), one of the most widely distributed alkaline purple parent rocks in the Sichuan Basin (He, 2003), was collected from a rock layer 60 cm below the earth surface without anthropogenic influences. This was then air-dried to a constant weight at room temperature and mechanically broken. The particles with a particle size of 0.25-1 mm were selected with the main components of clay minerals and main oxide, presented in Supplementary Table S1 and Supplementary Table S2, respectively.

### 2.2. Experimental design

Urea (U, 46%) and ammonium bicarbonate (AB, 17%), as the main N fertilizers applied in the study area, were selected. Four nitrogen fertilizer addition levels (0, 280, 560, and 840 N kg•ha<sup>-1</sup> described as CK, 100% CL, 200% CL, and 300% CL) were used, of which the 280 N kg•ha<sup>-1</sup> is the conventional application level. Specifically, seven treatments were used in the incubation experiment: (1) CK: no fertilizers, (2) U1: urea of 280 N kg•ha<sup>-1</sup>, (3) U2: urea of 560 N kg•ha<sup>-1</sup>, (4) U3: urea of 840 N kg•ha<sup>-1</sup>, (5) AB1: ammonium bicarbonate of 280 N kg•ha<sup>-1</sup>, (6) AB2: ammonium bicarbonate of 560 N kg•ha<sup>-1</sup>, and (7) AB3: ammonium bicarbonate of 840 N kg•ha<sup>-1</sup>. Each treatment was replicated 18 times.

The incubation experiment was conducted as follows. After weighing, 100 g  $J_3p$  of mudstone particles (0.25-1 mm) were placed into 240 mL culture bottles. These culture bottles were put in an incubator with a constant temperature of 25°C and soil moisture was

maintained at 40% of the field water capacity for 7 days as the preparation phase. Then, relevant rate and already dissolved nitrogen fertilizer was added into the culture bottles according to the treatments, while soil moisture content was adjusted to  $60\% \sim 70\%$  of field water capacity. This incubation condition was maintained during the whole experiment and the whole cultivation process lasted for 120 days.

### 2.3. Sampling of J<sub>3</sub>p mudstone particles

Sampling was conducted on the 1st, 7th, 15th, 30th, 60th, and 120th days from the start of the incubation. Each time, three culture bottles were randomly selected from 18 replicates of each treatment. After sampling, fresh samples were used to determine the dissolved organic carbon (DOC), total dissolved nitrogen (TDN), N-NH<sub>4</sub><sup>+</sup>, N-NO<sub>2</sub><sup>-</sup>, and N-NO<sub>3</sub><sup>-</sup>. Samples collected on day 120 were also used to analyze the bacterial community (samples kept at  $-4^{\circ}$ C) and the physical and chemical properties (samples air-dried under normal temperature).

### 2.4. Measurement of indicators

The dissolved organic carbon (DOC), total dissolved nitrogen (TDN), N-NH4<sup>+</sup>, and N-NO3<sup>-</sup> were measured using a continuous flow analytical system (AutoAnalyzer 3) (Wang et al., 2007; Yin et al., 2015). The content of dissolved organic carbon and nitrogen in purple mudstone particles after 120 days of culture with nitrogen fertilizer was measured by methods of agrochemical analysis of soil (Lu, 1999) and is shown in Table 1. The pH was measured in deionized water at a ratio of 1:2.5 (soil: water) using an air-dried sample by a pH meter (Sartorius PB-10). After the lithium metaborate melt, the content of SiO<sub>2</sub>, Al<sub>2</sub>O<sub>3</sub>, Fe<sub>2</sub>O<sub>3</sub>, MgO, Na<sub>2</sub>O, K<sub>2</sub>O, and CaO of minerals was determined by inductively coupled plasma atomic emission spectrometry (ICP-AES), and this test was carried out in Chengdu Baihui Biotechnology Co. The mineral composition and the clay mineral content were measured by an X-ray diffractometer (XD-3010301) (Zhang and Fan, 2003). The mechanical composition was determined by the screening method (Liu, 1996), and the field capacity of the purple rock weathering product was measured by the bubble water method (Xu, 2007). Moreover, the water-soluble Ca<sup>2+</sup> was measured in deionized water at a ratio of 1:10 (soil: water) by atomic absorption spectrophotometers.

# 2.5. DNA extraction, PCR amplification, and sequencing

According to the manufacturer's protocol, bacterial DNA samples were extracted from 0.5g fresh samples using the Fast DNA SPIN extraction kit (MP Biomedicals, Santa Ana, CA, United States). The V3-V4 region of the bacterial 16S rRNA gene was amplified (upstream primers 338F 5'-ACTCCTACGGGAGGCAGCA-3' and the downstream 806R 5'- GGACTACHVGGGTWTCTAAT -3') by PCR. Specifically, sample-specific 7-bp barcodes were incorporated into the primers for multiplex sequencing. The PCR components contained  $5 \mu L$  of Q5 reaction buffer (5×),  $5 \mu L$  of Q5 High-Fidelity GC buffer (5×), 0.25 µL of Q5 High-Fidelity DNA Polymerase (5 U/  $\mu l),\,2\,\mu L$  (2.5 mM) of dNTPs,  $1\,\mu L$  (10 uM) of each Forward and Reverse primer, 2 µL of DNA Template, and 8.75 µL of ddH<sub>2</sub>O. Thermal cycling consisted of initial denaturation at 98°C for 2 min, followed by 25 cycles consisting of denaturation at 98°C for 15 s, annealing at 55°C for 30 s, and extension at 72°C for 30 s, with a final extension of 5 min at 72°C. PCR amplicons were purified with Agencourt AMPure Beads (Beckman Coulter, Indianapolis, IN) and quantified using the PicoGreen dsDNA Assay Kit (Invitrogen, Carlsbad, CA, United States). After the quantification step, applicants of PCR were pooled in equal amounts, and the sequencing was performed using the Novaseq-PE250 pattern of the Illumina MiSeq platform at Shanghai Personal Biotechnology Co., Ltd. (Shanghai, China).

## 2.6. Evaluation index of the rocks weathering

The chemical index of alteration (*CIA*-Nesbitt and Young, 1982), chemical index of weathering (*CIW*- Harnois, 1988), chemical proxy of alteration (*CPA*- Buggle et al., 2011), and modified *CIA* (*CIX*-Garzanti et al., 2014) were widely applied to evaluate the weathering of the rocks. Ion composition ratios of the minerals were used to calculate *CIA*, *CIW*, *CPA*, and *CIX* by the following formula:

TABLE 1 The content of dissolved organic carbon and nitrogen in purple mudstone particles after 120 days	of culture.
--	-------------

Nitrogen treatment	рН <sub>н20</sub>	DOC g•kg⁻¹	TDN g∙kg <sup>-1</sup>	N-NH₄⁺ g∙kg⁻¹	N-NO <sub>3</sub> <sup>-</sup> g•kg <sup>-1</sup>
СК	$9.11 \pm 0.02b$	$0.45 \pm 0.03 d$	$0.03\pm0.00g$	$0.01 \pm 0.00d$	$0.02 \pm 0.00 d$
U1	8.03 ± 0.02c	$0.60 \pm 0.00c$	1.13 ± 0.01e	$0.04 \pm 0.00d$	$1.01 \pm 0.00a$
U2	9.38 ± 0.01a	0.27 ± 0.01e	$2.26\pm0.06b$	$1.55 \pm 0.08b$	$0.05 \pm 0.01 d$
U3	9.41 ± 0.01a	$1.95 \pm 0.03a$	$2.64\pm0.03a$	$2.37\pm0.04a$	$0.10 \pm 0.01c$
AB1	7.88 ± 0.05d	$1.20 \pm 0.02b$	$1.01 \pm 0.07 f$	$0.05 \pm 0.00d$	$0.92 \pm 0.03b$
AB2	$9.40 \pm 0.02a$	0.34 ± 0.01de	1.25 ± 0.02d	$1.20 \pm 0.02c$	$0.05 \pm 0.01 d$
AB3	9.43 ± 0.01a	$0.60 \pm 0.10c$	$1.57 \pm 0.01c$	$1.54 \pm 0.01 b$	0.03 ± 0.00d

The no fertilizers (CK), urea of 280 N kg•ha-1 (U1), urea of 560 N kg•ha-1 (U2), urea of 840 N kg•ha-1 (U3), ammonium bicarbonate of 280 N kg•ha-1 (AB1), ammonium bicarbonate of 560 N kg•ha-1 (AB2), and ammonium bicarbonate of 840 N kg•ha-1 (AB3) were observed. And the pH, dissolved organic carbon (DOC), total dissolved nitrogen (TDN), N-NH4+, and N-NO3– concentrations after 120 days of nitrogen fertilizer application were displayed in Table 1. Numbers depict means ± standard deviations (n = 3), the same below. ANOVA was conducted followed by the Duncan test for multiple comparisons, and different lowercase letters indicate that there were significant differences among all treatments (p < 0.05).

$$CIA = Al_2O_3 / (Al_2O_3 + Na_2O + K_2O + CaO^*) \times 100 \cdots$$
 Formula 1

$$CIW = Al_2O_3 / (Al_2O_3 + Na_2O + CaO^*) \times 100 \cdots$$
 Formula 2

 $CPA = Al_2O_3 / (Al_2O_3 + Na_2O +) \times 100 \cdots$  Formula 3

 $CIX = Al_2O_3 / (Al_2O_3 + Na_2O + K_2O) \times 100 \cdots$  Formula 4

where *CaO*<sup>\*</sup> in formula 1, 2, and 3 is the amount of CaO contained in the silicate fraction of the rocks.

### 2.7. Data statistics analysis

Microbiome bioinformatics was performed with QIIME 22019.4 (Bolyen et al., 2019) with slight modifications according to the official tutorials.1 Raw sequence data were demultiplexed using the demux plugin followed by primers cutting with the cutadapt plugin (Martin, 2011). Sequences were then quality filtered, denoised, merged, and chimera removed using the DADA2 plugin (Callahan et al., 2016). Non-singleton amplicon sequence variants (ASVs) were aligned with MAFFT (Katoh et al., 2002) and used to construct a phylogeny with fasttree2 (Price et al., 2009). Alphadiversities metrics (Chao1, Shannon, and Simpson) and betadiversity metrics were estimated using the diversity plugin. Taxonomy was assigned to ASVs using the classify-sklearn naïve Bayes taxonomy classifier in a feature-classifier plugin (Bokulich et al., 2018) against the Greengenes 13\_8 99% OTUs reference sequences (McDonald et al., 2012). Based on the result of the ASV taxonomic unit processed by the method of DATA2 (Callahan et al., 2016), the relative abundance of the bacterial composition at the genus level was counted to explore the dominant bacterial species under the different nitrogen fertilizer treatments, and the graph of the bacterial composition at the genus level was drawn by Origin 2021. The differences in diversities (alpha and beta diversities) of the bacterial community were analyzed under different nitrogen fertilizer treatments based on Bray Curtis, permutational multivariate analysis of the variance (PERMANOVA) (Mcardle and Anderson, 2001), and nonmetric multidimensional scaling (NMDS) of Bray Curtis. One-way ANOVA was applied to the data by the Duncan method via IBM SPSS Statistics 25 to compare the differences in bacterial alpha-diversities and physicochemical indices among the nitrogen fertilizer treatments.

In addition, the multiple linear stepwise regression analysis was applied to measure the effectiveness of the bacterial dominant species and the physicochemical indices on weathering indices. Based on the result of the multiple linear stepwise regression analysis, the partial least squares structural equation modelling (PLS-SEM) was used to construct structural equation models by SmartPLS 3 tools. The variance inflation factor (VIF), coefficient of determination ( $R^2$ ), and normed fit index (NFI) were applied to evaluate the model (Liu et al., 2022).

### 2.8. Accession numbers

The raw sequencing data from 16S rRNA genes were deposited in NCBI's sequence read archive under the accession number PRJNA881222.<sup>2</sup>

### 3. Results

# 3.1. The content of dissolved organic carbon and nitrogen in purple mudstone particles after 120 days of culture

After 120 days of incubation, the pH (except for the 100% CL treatment), DOC (except for the 200% CL treatment), TDN, N-NH<sub>4</sub><sup>+</sup>, and N-NO<sub>3</sub><sup>-</sup> were significantly more enhanced by nitrogen fertilizer addition than the CK treatment (p < 0.05). As seen in Table 1, for the same nitrogen fertilizer type, the pH, TDN, and N-NH<sub>4</sub><sup>+</sup> increased whilst the N-NO3<sup>-</sup> decreased with increasing nitrogen fertilizer application. The DOC at U treatments displayed an increasing trend but a decreasing trend was observed at AB treatments with increasing nitrogen fertilizer addition. Additionally, under the same nitrogen fertilizer application levels, compared with the U treatment, the pH and N-NO<sub>3</sub><sup>-</sup> of the AB treatment were increased by  $-1.87\% \sim 0.28\%$ and  $-73.33\% \sim 7.14\%$ , respectively, and the increase rate shows a trend of increasing first and then decreasing with the increasing of nitrogen fertilizer application. The N-NH4+ and DOC under the AB treatment were enhanced by 69.28% ~ 99.45% and - 35.25% ~ 33.33% compared with the U treatment, respectively, and the decrease rate displayed a decreasing trend with increasing of nitrogen fertilizer application. The TDN of the AB treatment was reduced by 10.36% ~ 44.69% compared with the U treatment, and the decrease rate displayed decreasing first and then an increasing trend with increasing nitrogen fertilizer application.

## 3.2. Nitrogen fertilizer addition changes bacterial community composition

After 120 days of incubation, the bacterial community composition (Figure 1) and the dominant species (Supplementary Figure S1) were measured under different nitrogen fertilizer applications at the genus level. Among all samples, the nine most abundant phylum of bacteria were *Nitrolancea* (7.14%), *Faecalibacterium* (4.94%), *Massilia* (4.94%), *Domibacillus* (4.38%), *Sericytochromatia* (3.35%), *Lysobacter* (2.53%), *Sphingomonas* (2.50%), *Nitrosospira* (2.39%), and *Luteimonas* (2.23%). Compared with CK treatment, the relative abundance of the *Nitrolancea*, *Nitrosospira*, and *Luteimonas* at U1 treatment were significantly

<sup>1</sup> https://docs.qiime2.org/2019.4/tutorials/

<sup>2</sup> https://www.ncbi.nlm.nih.gov/sra/PRJNA881222

100 Relative abundance of bacteria (%) Others 80 Luteimonas Nitrosospira Sphingomonas Lysobacter 60 Sericytochromatia Domibacillus Massilia 40 Faecalibacterium Nitrolancea 20 СК U1 U2 U3 AB1 AB2 AB3 Fertilizer treatment FIGURE 1 Relative abundances of the bacterial genera in J<sub>3</sub>p purple mudstone particles after 120 days of nitrogen fertilizer addition in the laboratory. Displayed is a bacterial genus that made up the top 10 of relative abundance; moreover, other relative abundance of bacterial were grouped as "Others." The no fertilizers (CK), urea of 280Nkg•ha<sup>-1</sup> (U1), urea of 560Nkg•ha<sup>-1</sup> (U2), urea of 840Nkg•ha<sup>-1</sup> (U3), ammonium bicarbonate of 280Nkg•ha<sup>-1</sup> (AB1),

ammonium bicarbonate of 560Nkg•ha<sup>-1</sup> (AB2), and ammonium bicarbonate of 840Nkg•ha<sup>-1</sup> (AB3) were observed.

enhanced (p < 0.05), but other main bacterial species had no significant difference (p > 0.05); the relative abundance of the Domibacillus, Lysobacter, and Sphingomonas at U3 treatment were significantly increased but the relative abundance of the *Faecalibacterium* was significantly decreased by 90.15% (p < 0.05), and other main bacterial species had no significant difference (p > 0.05); the relative abundance of the *Nitrolancea*, *Nitrosospira*, and Luteimonas at AB1 treatment were significantly increased but the relative abundance of the Faecalibacterium was significantly decreased by 90.32% (p < 0.05), and other main bacterial species had no significant difference (p > 0.05); the relative abundance of the Massilia, Sericytochromatia, and Lysobacter at AB2 treatment were significantly enhanced but the relative abundance of the *Faecalibacterium* was significantly decreased by 83.16% (p < 0.05), and other main bacterial species had no significant difference (p > 0.05); the relative abundance of the *Massilia* and Sericytochromatia at AB3 treatment were significantly enhanced but the relative abundance of the Faecalibacterium was significantly decreased by 91.15% (p < 0.05), and other main bacterial species had no significant difference (p > 0.05). Moreover, the relative abundance of bacteria was not significantly altered between U2 and CK treatment.

Under U fertilizer application treatment, the relative abundance of *Domibacillus* was observed to have an increasing trend but the relative abundance of *Nitrolancea*, *Nitrosospira*, and *Luteimonas* displayed a decreasing trend with the nitrogen fertilizer application levels increasing. Additionally, the relative abundance of *Faecalibacterium*, *Massilia*, *Sericytochromatia*, *Lysobacter*, and *Sphingomonas* displayed a non-linear characteristic. Under AB fertilizer application treatment, the relative abundance of *Sericytochromatia* displayed an increasing trend but the relative abundance of *Nitrolancea*, *Domibacillus*, *Nitrosospira*, and *Luteimonas*  displayed a decreasing trend with the nitrogen fertilizer application levels increasing. Additionally, the relative abundance of *Faecalibacterium*, *Massilia*, *Lysobacter*, and *Sphingomonas* displayed a non-linear characteristic.

# 3.3. The difference in the bacterial community diversity under nitrogen fertilizer application

After 120 days of incubation, nitrogen fertilizer application significantly altered the alpha and beta diversities of the bacterial community, and the effect of the nitrogen fertilizer addition levels on the alpha and beta diversities of the bacterial community was greater than the effect of the nitrogen fertilizer addition types (Figures 2, 3). The Chao1 of bacteria community under U1 and U3 treatment were enhanced by 72.15 and 101.65% compared to CK treatments, respectively, but other treatments (including U2, AB1, AB2, and AB3) decreased by 12.30 to 30.84% compared to CK treatments. The Shannon of the bacterial community under the 200% CL treatment was enhanced by 2.58 to 3.44%, but under the 100 CL % and 300 CL treatments was decreased by 2.38 to 8.76%. And the Simpson of bacteria community under nitrogen fertilizer addition was increased by 1.08 to 5.85%.

Under the same nitrogen fertilizer addition levels, the Chao1 of the bacteria community under 100% CL and 200% CL of AB treatments were enhanced by 19.84 and 22.31% compared to U treatments, but Chao1 of the bacteria community under 300% CL of AB treatment was reduced by 51.87% compared to U treatment. And the Shannon and Simpson of the bacteria community at 100% CL and 200% CL of AB treatments were reduced by 0.83 to 1.55% compared to U treatments, but the Shannon and Simpson of the bacteria



community at 300% CL of AB treatment was enhanced by 1.52 to 5.06% compared to U treatments.

Under the U treatment condition, a trend of the decreasing and then increasing of the Chao1 of the bacteria community was observed, which was opposite to the changing trend of Shannon and Simpson with increasing of nitrogen fertilizer addition levels. With increasing of AB addition levels, the Chao1 of the bacteria community displayed a decreasing trend, and the Shannon of the bacteria community showed an increasing and then decreasing trend, and an increasing trend of the Simpson of the bacteria community was observed.

Nonmetric multidimensional scaling (NMDS) indicated that the bacterial community beta diversity was significantly altered by AB fertilizer application but was not significantly changed by U treatment compared to CK treatment (Figure 3A). The bacterial community beta diversity was significantly altered by 100% CL (U1 and AB1) and 300% CL (U3 and AB3) treatment but was not significantly changed by 200% CL (U2 and AB2) compared to CK treatment (Figure 3B). Under the nitrogen fertilizer types, nitrogen fertilizer levels, and combination with the type and level of the nitrogen fertilizer application, the explanation of bacterial beta diversity was calculated as 20, 34, and 59%, respectively, *via* the method of permutational multivariate analysis of variance.

# 3.4. The effect of bacterial on the weathering of purple mudstone particles under nitrogen fertilizer application

After 120 days of the laboratory experiment, the mineral elements' content of purple mudstone particles was measured to calculate the chemical index of alteration (CIA), the chemical index of weathering (CIW), the chemical proxy of alteration (CPA), and the modified CIA (CIX) (Table 2). The CIA, CIW, CPA, and CIX of the purple mudstone particles after 120 days of incubation with fertilizer application were enhanced by 1.22%~9.85, 47.01%~70.23, 1.07%~6.47, and 0.15%~6.08%, respectively, compared to before the incubation treatments. Compared with the CK treatment, the CIA, CIW, and CPA of purple mudstone particles under nitrogen fertilizer application treatments were enhanced by 0.61 to 8.52%, 1.75 to 15.80%, and 1.19 to 5.35%, respectively, and the CIX of purple mudstone particles under U2 and U3 treatment were reduced by 0.18 and 0.27%, respectively, but other treatments were enhanced by 1.86 to 5.63%. Under the same nitrogen fertilizer level application, the weathering indices (CIA, CIW, CPA, and CIX) of the purple mudstone particles under AB treatment were higher than that under U treatments. Additionally, weathering indices (CIA, CIW, CPA, and CIX) of the purple mudstone particles under both the U and AB treatments showed a decreasing trend with increasing nitrogen fertilizer application levels.



#### FIGURE 3

The beta diversity was calculated based on nonmetric multidimensional scaling (NMDS) after 120 days of nitrogen fertilizer addition in the laboratory (C), and the bacterial beta diversity was restricted by the nitrogen fertilizer types (A) and levels (B) *via* permutational multivariate analysis of variance. The no fertilizers (CK), urea of 280Nkg•ha<sup>-1</sup> (U1), urea of 560Nkg•ha<sup>-1</sup> (U2), urea of 840Nkg•ha<sup>-1</sup> (U3), ammonium bicarbonate of 280Nkg•ha<sup>-1</sup> (AB1), ammonium bicarbonate of 560Nkg•ha<sup>-1</sup> (AB2), and ammonium bicarbonate of 840Nkg•ha<sup>-1</sup> (AB3) were observed.

TABLE 2 The chemical index of alteration response to nitrogen fertilization addition after 120 days in the laboratory.

Treatment	CIA	CIW	СРА	CIX
Original rock	61.8 ± 0.2	$34.9 \pm 0.4$	83.0 ± 0.3	$70.8 \pm 0.2$
СК	62.6 ± 0.8a	52.8 ± 0.9ab	83.9 ± 0.4a	71.1 ± 0.8ab
U1	65.6 ± 0.5bcd	56.3 ± 1.5b	86.2 ± 0.5bcd	73.3 ± 0.3abc
U2	63.1 ± 1.2ab	52.7 ± 1.3ab	85.1 ± 0.7abc	71.0 ± 1.1a
U3	62.9 ± 0.7ab	51.6 ± 1.0a	84.8 ± 0.4ab	70.9 ± 0.5a
AB1	67.9 ± 0.7d	56.9 ± 0.8b	87.6 ± 0.4d	75.1 ± 0.6c
AB2	$66.0 \pm 0.8  \text{cd}$	53.3 ± 2.2ab	$86.7\pm0.4cd$	73.4 ± 0.7bc
AB3	64.6 ± 1.0abc	52.9 ± 1.3ab	85.6 ± 0.8abc	72.4 ± 0.8ab

ANOVA was conducted followed by the Duncan test for multiple comparisons. Different lowercase letters indicate that there were significant differences among all treatments (*p* <0.05). The weathering indices of the purple mudstone before incubation (original rock) and after incubation (CK, U1, U2, U3, AB1, AB2, and AB3) were displayed in Table 2, and the indicators to characterize the weathering of purple mudstone included the Chemical index of alteration (CIA-Nesbitt and Young, 1982), Chemical index of weathering (CIW- Harnois, 1988), Chemical proxy of alteration (CPA- Buggle et al., 2011), and Modified CIA (CIX-Garzanti et al., 2014), which were calculated and showed in Table 2.

Stepwise regression analysis was used to analyze the effect of the dominant bacterial species at the genus level and environmental factors on the weathering indices (*CIA*, *CIW*, *CPA*, and *CIX*) of purple mudstone particles in this study (Table 3). The stepwise regression analysis indicated that Nitrobacter (*Nitrolancea*) and Phosphorusdissolving bacteria (*Massilia*) play the main roles in the weathering of the purple mudstone particles, and the N-NO<sub>3</sub><sup>-</sup> was measured as a key environmental factor during the purple mudstone particles' weathering process. The proportions of *Nitrolancea*, *Massilia*, and N-NO<sub>3</sub><sup>-</sup> that affected the weathering indices (*CIA*, *CIW*, *CPA*, and *CIX*) of J<sub>3</sub>p purple mudstone particles were calculated to be 36, 12, and 35% by the stepwise regression analysis, respectively (Table 3).

### 4. Discussion

# 4.1. The effect of nitrogen application on the characteristics of bacterial community diversity

Nitrogen fertilizer application altered the alpha and beta diversities of the bacterial community (Figures 2, 3), and the effect of the nitrogen fertilizer addition levels on the alpha and beta diversities of the bacterial community was more prominent. The bacterial community was controlled by the environmental factors and interaction relationships of bacteria (Zhang et al., 2017; Nie et al.,

Dependent variable	Enter variable	Remove variable	Parameter variable	Partial <i>R</i> <sup>2</sup>	Model R <sup>2</sup>	p
Weathering indices	Equation intercept		-0.68			
	Nitrolancea		5.89	0.36	0.36	0.00
		Faecalibacterium	0.04	0.00	0.36	0.83
	Massilia		5.33	0.12	0.48	0.03
		Domibacillus	-0.04	0.00	0.48	0.81
		Sericytochromatia	0.10	0.00	0.48	0.65
		Lysobacter	-0.05	0.00	0.48	0.78
		Sphingomonas	-0.05	0.00	0.48	0.80
		Nitrosospira	0.47	0.07	0.48	0.06
		Luteimonas	-0.21	0.00	0.48	0.59
		Nocardioides	-0.03	0.00	0.48	0.87
Weathering indices	Equation intercept		-0.44			
	N-NO <sub>3</sub> <sup>-</sup>		1.44	0.35	0.35	0.00
		рН	0.24	0.00	0.35	0.76
		DOC	-0.14	0.00	0.35	0.47
		TDN	-0.04	0.00	0.35	0.84
		Net nitrification rate	-7.43	0.00	0.35	0.98

TABLE 3 The weathering indices and the content-dissolved organic carbon and nitrogen of the  $J_{3p}$  purple mudstone fragaments' response to the bacteria of the top 10 at the genus level after 120 days of nitrogen fertilizer addition in the laboratory.

The dependent variable (weathering indices) was obtained by reducing dimensionality (Principal component analysis) of CIA, CIW, CPA, and CIX. And based on the multivariable linear regression model, the effect of dominated species of bacteria at genus and the environment factors (pH, dissolved organic carbon (DOC), total dissolved nitrogen (TDN), N-NH4+, and N-NO3–) on weathering indices were calculated, and key factor were separated from all factors. In addition, this model's results were applied to construct a structural equation model.

0.00

N-NH4+

2018; Sun et al., 2020). In terms of environmental factors, the bacterial community diversity was determined by the pH of the soil environment (Rousk et al., 2010; Zhang et al., 2017). Previous studies have demonstrated that the number of bacterial species showed a trend of increasing and then decreasing with increasing pH of the soil environment in alkaline soil (pH range 7.43 ~ 8.66), and the optimal environmental conditions for bacterial growth were observed to be pH 8.04 (Shi et al., 2021). Another study also implied that the Shannon index increased following a decreasing trend with increasing soil pH (pH range 4.31-8.31) in a 7-year long-term study, and the soil pH was significantly altered by fertilizer addition (Zhang et al., 2017). This may be due to the poorer tolerance of the bacteria in the alkaline environment (Bárcenas-Moreno et al., 2016) or the narrower pH for optimal growth of bacteria (Rousk et al., 2010). In this study, the effect of the nitrogen fertilizer application levels on the pH of the soil environment was higher than that of the nitrogen fertilizer application types (Table 1), and the pH (range 7.88–9.46) displayed an increasing trend with increasing nitrogen fertilizer addition levels, which resulted in the dominant species of bacteria at the genus level significantly altering with increasing pH (Supplementary Figure S2). The relative abundance of the Nitrolancea and Luteimonas displayed a decreasing trend but other dominant species in the genus observed an increasing trend with the pH increasing (Supplementary Figure S2). Based on the polynomial fitting model analysis, it was found that with an increase in pH, the Chao1 of the bacterial community exhibits a nonlinear trend ( $R^2 = 0.09$ , p = 0.39) of decreasing and then increasing (Supplementary Figure S3a), while the Simpson ( $R^2 = 0.24$ , p < 0.05) and Shannon ( $R^2 = 0.22$ , p < 0.05) of the bacterial community showed

nonlinear characteristics of first increasing and then decreasing (Supplementary Figures S3b,S3c), which was caused by changes in dominant bacterial species. In addition, the alpha diversities displayed nonlinear characteristics of increasing and then decreasing with increasing nitrogen fertilizer addition levels in previous studies (Zhang et al., 2008; Li et al., 2019; He et al., 2021). However, another environmental factor in the bacterial community was determined by the N-NH<sub>4</sub><sup>+</sup> content in the soil (Nie et al., 2018). The N-NH<sub>4</sub><sup>+</sup> content increased with increasing nitrogen fertilizer addition levels, but the N-NH4<sup>+</sup> content was not different under the different nitrogen fertilizer addition types at the same fertilizer levels (Table 1). The high N-NH<sub>4</sub><sup>+</sup> concentration in the soil was toxic to some bacterial groups, and the low N-NH4<sup>+</sup> concentration in the soil promoted bacterial growth (Zhou et al., 2017) in a previous study. According to the polynomial fitting model analysis, the alpha diversities (Chao1, Shannon, and Simpson) of the bacterial community showed nonlinear characteristics with increasing N-NH4+, and the Chao1  $(R^2 = 0.19, p = 0.06)$  of the bacterial community displayed a decreasing and following increasing trend (Supplementary Figure S3d), while the Simpson ( $R^2 = 0.23$ , p < 0.05) and Shannon ( $R^2 = 0.14$ , p = 0.09) of the bacterial community (Supplementary Figures S3e,S3f) showed increasing first and then decreasing. Additionally, a significantly negative relationship was calculated between pH ( $R^2 = 0.80$ , p < 0.05), N-NH<sub>4</sub><sup>+</sup> ( $R^2 = 0.21$ , p < 0.05), and beta diversity of the bacterial community (Supplementary Figures S3g,S3h). Based on the above analysis in terms of environmental factors, changes in the pH and N-NH<sub>4</sub><sup>+</sup> in the soil environment were induced by the application of different nitrogen fertilizer levels, which was one of the main reasons

0.00

0.35

0.41

inducing the succession of the dominant species in the soil environment, thereby inducing differences in the structural alpha and beta diversities of the bacterial community.

In terms of the interrelationship of the bacterial communities, antagonistic and synergistic effects were widely found between bacterial communities (Meidute et al., 2008; Cai et al., 2019). Another previous study showed that the antagonistic effect of inorganic fertilizer application treatment was observed to have a main role in the bacterial community compared to no fertilizer addition and organic fertilizer treatment (Sun et al., 2020). Previous studies reported that the toxic effector proteins secreted by the type VI secretion system (T6SS) were one of the main ways that the antagonistic effect was induced between bacterial communities (Trunk et al., 2018), and the toxic effector proteins were transported into the nucleus of neighboring fungi or bacteria to promote bacterial or fungal lethality (Tang et al., 2018; Jurenas and Journet, 2021). The dominant species may be altered by antagonistic effects of bacteria under different nitrogen fertilizer applications, thereby affecting the beta diversity of bacteria. In this study, compared to the dominant species structure in the CK treatment, smaller differences in the dominant species structure under different nitrogen fertilizer addition types were observed, and larger differences in the dominant species structure under different nitrogen fertilizer addition levels were observed, which resulted in a higher effect of nitrogen fertilizer addition levels on bacterial beta diversity than nitrogen fertilizer addition types. The difference in nitrogen fertilizer addition treatment may be due to the composition of the bacterial community being codetermined by the adaptation of the bacteria to the soil environment change (Bárcenas-Moreno et al., 2016) and the interrelationship between species (Meidute et al., 2008), thereby affecting the beta diversity of the bacterial community. In addition, a global meta-analysis implied that the beta diversity of the bacterial community increased with increasing nitrogen fertilizer addition levels (Zhou et al., 2022). The antagonistic effects of the bacteria at nitrogen fertilizer application may be one of the factors that induced the dominant species structure alteration in this study, thereby promoting the change in the beta diversity of bacteria.

# 4.2. The effect of nitrogen application on the weathering of purple mudstone particles

This result indicated that the weathering indices (*CIA*, *CIW*, *CPA*, and *CIX*) of the purple mudstone particles were promoted by nitrogen fertilizer addition compared with the CK treatment and were measured with a decreasing trend with increasing nitrogen fertilizer levels applied under the same nitrogen fertilizer types (Table 2). Our previous studies have confirmed that acid input will accelerate purple rock weathering (Zhao et al., 2021). More specifically, the H<sup>+</sup> was expanded by alkaline anions in the soil solution (Rice and Herman, 2012) and the exchangeable base ions in clay minerals (Warby et al., 2009). After the above reaction, the clay minerals (Warby et al., 2009). After the above reaction, the clay minerals were decomposed by residual H<sup>+</sup> to accelerate the base ion release from the clay minerals (Tabatabai and Olson, 1985), thereby promoting the weathering of rock. Additionally, the water-soluble Ca<sup>2+</sup> was measured and displayed a decreasing trend with the nitrogen fertilizer addition levels increasing in this study (Supplementary Table S3), and a significant

positive correlation was calculated between the content of watersoluble Ca<sup>2+</sup> and the weathering indices (Supplementary Figure S4a) but a negative correlation of the content of CaO in clay minerals and weathering indices was observed (Supplementary Figure S4b). The previous studies implied that the weathering carbonate was accelerated by nitrogen fertilizer addition, and the leaching amount of the base ion was enhanced due to the lack of a buffering reaction in carbonatefree soil, however, the soil acidification after nitrogen fertilizer application was offset by the buffering reaction that occurred in carbonated soils (Gandois et al., 2011). And another study indicated that an increasing trend of weathering rate of carbonate, CO<sub>2</sub> concentration, Ca<sup>2+</sup>, Mg<sup>2+</sup>, and NO<sub>3</sub><sup>-</sup> was measured after nitrogen fertilizer addition, which was attributed to the release of protons from the nitrification reaction (Song et al., 2017). This provides ample evidence that nitrogen fertilizer application affects the weathering indices of purple mudstone particles that are rich in Calcite (Supplementary Table S1). Moreover, the H<sup>+</sup> by nitrification of N-NH4<sup>+</sup> was one of the main sources of H<sup>+</sup> in the soil environment, yet the relative abundance of Nitrolancea, a nitrite-oxidizing bacteria (NOB) (Sorokin et al., 2014; Jeong and Bae, 2021), showed a decreasing trend with increasing nitrogen fertilizer addition levels (Figures 1, 4), which resulted in the nitrification of  $N-NH_4^+$  decreasing with increasing nitrogen fertilizer addition levels. The main reason for this phenomenon may be due to the high N-NH<sub>4</sub><sup>+</sup> concentration in the soil being toxic to some bacterial groups, and the low N-NH<sub>4</sub><sup>+</sup> concentration in the soil promoted bacterial growth (Zhou et al., 2017). The growth of bacteria with nitrification was inhibited by the concentration of N-NH4+ with increasing nitrogen fertilizer application levels. This results in a reduction in the H<sup>+</sup> produced by nitrification of N-NH4+ and therefore a reduction trend in the weathering of the purple mudstone particles with increasing nitrogen fertilizer application levels.

The weathering indices (CIA, CIW, CPA, and CIX) of the purple mudstone particles with AB application was higher than that with U application under the same nitrogen fertilizer levels but did not reach a significant level (p > 0.05), and the weathering indices (CIA, CIW, CPA, and CIX) of the purple mudstone particles were reduced with the nitrogen fertilizer addition level increasing (Table 2). The weathering indices (CIA, CIW, CPA, and CIX) of the purple mudstone particles were controlled by H<sup>+</sup>, from net nitrification of Nitrolancea. And net nitrification rate was significantly reduced (Supplementary Figure S5) with the nitrogen fertilizer addition level increasing. The average nitrification rate at 200% CL and 300% CL can be ignored because the average nitrification rate was observed by  $-0.03 \text{ mg} \cdot \text{kg}^{-1} \cdot \text{d}^{-1}$  to  $0.14 \text{ mg} \cdot \text{kg}^{-1} \cdot \text{d}^{-1}$  (Figure 5), therefore, the nitrification of Nitrolancea at 100% CL was one of the main factors to affect the weathering of the purple mudstone particles. Additionally, Massilia may be another key bacterial in the weathering of purple mudstone particles. In this study, under the 200% CL and 300% CL nitrogen fertilizer treatments, Massilia was observed in the AB treatment but not in the U treatment. According to previous research, Massilia has a dissolved phosphorus function (Zheng et al., 2017). The phosphorus dissolution process of bacteria may promote the release of calcium and magnesium, which may affect rock weathering. In addition, a previous study indicated that highly effective Al solubilizers from Bacillus secretion promote the release of Al from surface rocks (Wang et al., 2017). Another study also reported that acetic acid was found in the metabolites of mucilaginous Bacillus, and the



complexation of acetic acid was one of the main mechanisms that accelerated silicate clay mineral weathering (Mo and Lian, 2011). These studies provide some strong evidence that the secretions of bacteria promote rock weathering. The effects of nitrification of N-NH<sub>4</sub><sup>+</sup> or phosphorus dissolution of *Massilia* were shown to be stronger with the AB treatment than with the U treatment, thereby indicating that the weathering indices (*CIA*, *CIW*, *CPA*, and *CIX*) of the purple mudstone particles in the AB treatment were greater than that in the U treatment.

The PLS-SEM indicated that the explanation of the effect of bacteria on the weathering of purple mudstone particles was calculated as 59% after nitrogen fertilizer application, and the nitrogen fertilizer addition levels had a negative effect on Nitrobacter (p < 0.01) and N-NO<sub>3</sub><sup>-</sup> (p = 0.50) but had a positive effect on Phosphorousdissolving bacteria (p = 0.16). And Nitrobacter had an extremely significant positive effect on the N-NO<sub>3</sub><sup>-</sup> (p < 0.01), while the N-NO<sub>3</sub><sup>-</sup> (p < 0.05) (Figure 6). Noteworthy is the opposite positive effect of the N-NO<sub>3</sub><sup>-</sup> (p < 0.01), and Phosphorous-dissolving bacteria (p < 0.05) (Figure 6). Noteworthy is the opposite positive effect of the N-NO<sub>3</sub><sup>-</sup> (p < 0.01), and Phosphorous-dissolving bacteria (p < 0.01) on weathering indices (*CIA*, *CIW*, *CPA*, and *CIX*) (Figure 6). Based on the above analysis, the H<sup>+</sup> from nitrification of Nitrobacter (*Nitrolancea*) (Sorokin et al., 2014; Jeong and Bae, 2021) and the phosphorus dissolution of Phosphorous-dissolving bacteria (*Massilia*) (Zheng et al., 2017) may be the main processes affecting the weathering of purple mudstone particles, which was confirmed by the PLS-SEM calculation. The relative abundances of *Nitrolancea* and *Massilia* were significantly altered by different nitrogen fertilizer additions in this study (Figure 6). *Nitrolancea* was observed in the 100% CL treatment but was not detected in the 200% CL and 300% treatment, and *Massilia* was observed in the 200% CL and 300% treatment but was not detected in the 100% CL treatment. In summary, nitrogen fertilizer addition, especially AB application, can promote the weathering of purple mudstone particles in this study. Nitrification (*Nitrolancea*) was the main pathway for nitrogen fertilizer at conventional levels (100% CL), but phosphorus dissolution of *Massilia* may be the main pathway for nitrogen fertilizer affecting the weathering of purple mudstone particles at excessive application (200% CL and 300% CL).

## 5. Conclusion

At the same fertilization levels, the weathering indices (*CIA*, *CIW*, *CPA*, and *CIX*) of purple mudstone particles under the AB treatment were higher than that under the U treatment, and a reducing trend was observed with increasing nitrogen fertilizer levels under the same nitrogen fertilizer application types. Compared with the CK treatment,



### FIGURE 5

The net nitrification rate during the laboratory under nitrogen fertilizer addition. The no fertilizers (CK), urea of 280Nkg•ha<sup>-1</sup> (U1), urea of 560Nkg•ha<sup>-1</sup> (U2), urea of 840Nkg•ha<sup>-1</sup> (U3), ammonium bicarbonate of 280Nkg•ha<sup>-1</sup> (AB1), ammonium bicarbonate of 560Nkg•ha<sup>-1</sup> (AB2), and ammonium bicarbonate of 840Nkg•ha<sup>-1</sup> (AB3) were observed.



The model of the bacteria on weathering indices (*CIA*, *CIW*, *CPA*, and *CIX*) of  $J_3p$  purple mudstone particles under different nitrogen fertilizer additions. The names inside the boxes are measured variables, and the labels below the circles are latent variables. The numbers associated with the arrows are path coefficients. The percentages inside the boxes refer to the variance explained by the Model ( $R^2$ ). The significance correlations (p<0.05) and extremely significant correlations (p<0.01) of the physicochemical properties are expressed in "\*" and "\*\*," respectively.

the *CIX* of purple mudstone particles at U2 and U3 treatments were reduced by 0.18, and 0.27%, respectively, however, the weathering indices (*CIA*, *CIW*, *CPA*, and *CIX*) of purple mudstone particles at

other fertilizer treatments were enhanced by 0.61 to 15.80%. Nitrogen fertilizer application significantly altered the diversity of the bacterial community. Specifically, the alpha diversities of the bacterial

community appeared to increase first and then decline as the nitrogen fertilizer levels increased, and the effect of the nitrogen fertilizer application levels ( $R^2 = 0.34$ ) on beta diversity was higher than that of the nitrogen fertilizer type ( $R^2 = 0.20$ ). Compared with the CK treatment, the beta diversity of bacteria under the AB treatment was significantly changed, while it was not significantly altered under the U treatment. The beta diversity was significantly enhanced compared to the CK treatment in the 100% CL and 300% CL treatments, but it was not significantly different from the CK treatment in the 200% CL treatment. The impact proportions of Nitrolancea, Massilia, and N-NO<sub>3</sub><sup>-</sup> on the weathering indices (CIA, CIW, CPA, and CIX) of J<sub>3</sub>p purple mudstone particles were calculated to be 36, 12, and 35% by the stepwise regression analysis, respectively. Notably, the result of the calculations for the stepwise regression analysis and the structural equation model maintained a high degree of consistency. Additionally, the result of the structural equation model implied that the nitrification of Nitrobacter (Nitrolancea) and the phosphorus dissolution of Phosphorous-dissolving bacteria (Massilia) are the main drivers of the dominant weathering process of the J<sub>3</sub>p purple mudstone particles ( $R^2 = 0.59$ ).

### Data availability statement

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

### Author contributions

CL, MF, XW, XL, GZ, GL, and JZ contributed to the study's conception and design. Material preparation, data collection, and analysis were performed by CL, XW, and XL. The first draft of the manuscript was written by CL. GL, GZ, MF, and JZ commented on

### References

Bárcenas-Moreno, G., Bååth, E., and Rousk, J. (2016). Functional implications of the pH-trait distribution of the microbial community in a re-inoculation experiment across a pH gradient. *Soil Biol. Biochem.* 93, 69–78. doi: 10.1016/j.soilbio.2015.10.024

Beerling, D. J., Kantzas, E. P., Lomas, M. R., Wade, P., Eufrasio, R. F., Renforth, P., et al. (2020). Potential for large-scale CO2 removal via enhanced rock weathering with croplands. *Nature* 583, 242–248. doi: 10.1038/s41586-020-2448-9

Bibi, I., Singh, B., and Silvester, E. (2014). Dissolution kinetics of soil clays in sulfuric acid solutions: ionic strength and temperature effects. *Appl. Geochem.* 51, 170–183. doi: 10.1016/j.apgeochem.2014.10.004

Bokulich, N. A., Kaehler, B. D., Rideout, J. R., Dillon, M., Bolyen, E., Knight, R., et al. (2018). Optimizing taxonomic classification of marker-gene amplicon sequences with QIIME 2's q2-feature-classifier plugin. *Microbiome* 6:90. doi: 10.1186/s40168-018-0470-z

Bolyen, E., Rideout, J. R., Dillon, M. R., Bokulich, N. A., Abnet, C. C., al-Ghalith, G. A., et al. (2019). Author correction: reproducible, interactive, scalable and extensible microbiome data science using QIIME 2. Nat. Biotechnol. 37:1091. doi: 10.1038/ s41587-019-0209-9

Buggle, B., Glaser, B., Hambach, U., Gerasimenko, N., and Marković, S. (2011). An evaluation of geochemical weathering indices in loess–paleosol studies. *Quat. Int.* 240, 12–21. doi: 10.1016/j.quaint.2010.07.019

Cai, P., Sun, X. J., Wu, Y. C., Gao, C. H., Mortimer, M., Holden, P. A., et al. (2019). Soil biofilms: microbial interactions, challenges, and advanced techniques for ex-situ characterization. *Soil. Ecol. Lett.* 1, 85–93. doi: 10.1007/s42832-019-0017-7

Callahan, B. J., McMurdie, P. J., Rosen, M. J., Han, A. W., Johnson, A. J., and Holmes, S. P. (2016). DADA2: high-resolution sample inference from Illumina amplicon data. *Nat. Methods* 13, 581–583. doi: 10.1038/nmeth.3869

previous versions of the manuscript. All authors contributed to the article and approved the submitted version.

### Funding

This work was supported by the National Natural Science Foundation of China (No.42007002), the Yunnan Basic Research Project-general projects (No. 202101AT070220), and the Doctoral Research Foundation of Yunnan Agricultural University (No. KY2018-26).

### **Conflict of interest**

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

### Publisher's note

All claims expressed in this article are solely those of the authors and do not necessarily represent those of their affiliated organizations, or those of the publisher, the editors and the reviewers. Any product that may be evaluated in this article, or claim that may be made by its manufacturer, is not guaranteed or endorsed by the publisher.

### Supplementary material

The Supplementary material for this article can be found online at: https://www.frontiersin.org/articles/10.3389/fmicb.2023.1164826/ full#supplementary-material

Caspari, T., Bäumler, R., Norbu, C., Tshering, K., and Baillie, I. (2006). Geochemical investigation of soils developed in different lithologies in Bhutan, Eastern Himalayas. *Geoderma* 136, 436–458. doi: 10.1016/j.geoderma.2006.04.017

Chao, S., Liu, C. L., Wang, J. K., Zhang, Y., and Hou, H. B. (2011). Impact of the addition of a compound fertilizer on the dissolution of carbonate rock tablets: a column experiment. *Appl. Geochem.* 26, S170–S173. doi: 10.1016/j.apgeochem.2011.03.095

Chetelat, B., Liu, C. Q., Zhao, Z. Q., Wang, Q. L., Li, S. L., Li, J., et al. (2008). Geochemistry of the dissolved load of the Changjiang Basin rivers: anthropogenic impacts and chemical weathering. *Geochim. Cosmochim. Acta* 72, 4254–4277. doi: 10.1016/j.gca.2008.06.013

Dixon, J. L., Hartshorn, A. S., Heimsath, A. M., DiBiase, R. A., and Whipple, K. X. (2012). Chemical weathering response to tectonic forcing: a soils perspective from the San Gabriel Mountains, California. *Earth Planet. Sci. Lett.* 323-324, 40–49. doi: 10.1016/j. epsl.2012.01.010

Duan, L., Hao, J., Xie, S., Zhou, Z. P., and Ye, X. M. (2002). Determining weathering rates of soils in China. *Geoderma* 110, 205–225. doi: 10.1016/S0016-7061(02)00231-8

Gai, X. P., Liu, H. B., Zhai, L. M., Tan, G. C., Liu, J., and Ren, T. Z. (2016). Vegetable yields and soil biochemical properties as influenced by fertilization in southern China. *Appl. Soil Ecol.* 107, 170–181. doi: 10.1016/j.apsoil.2016.06.001

Gandois, L., Perrin, A. S., and Probst, A. (2011). Impact of nitrogenous fertilizerinduced proton release on cultivated soils with contrasting carbonate contents: a column experiment. *Geochim. Cosmochim. Acta* 75, 1185–1198. doi: 10.1016/j. gca.2010.11.025

Gao, M. R., Zhu, B., and Jiang, M. F. (1999). Distribution of zinc forms and its influence factors in calcareous purple soils. *Chin. J. Appl. Ecol.* 10, 415–418. (In Chinese)

Garzanti, E., Padoan, M., Setti, M., López-Galindo, A., and Villa, I. M. (2014). Provenance versus weathering control on the composition of tropical river mud (southern Africa). *Chem. Geol.* 366, 61–74. doi: 10.1016/j.chemgeo.2013.12.016

Godderis, Y., Roelandt, C., Schott, J., Pierret, M. C., and François, L. M. (2009). Towards an integrated model of weathering, climate, and biospheric processes. *Rev. Mineral. Geochem.* 70, 411–434. doi: 10.2138/rmg.2009.70.9

Goulding, K. W. T. (2016). Soil acidification and the importance of liming agricultural soils with particular reference to the United Kingdom. *Soil Use Manag.* 32, 390–399. doi: 10.1111/sum.12270

Harnois, L. (1988). The CIW index: a new chemical index of weathering. Sediment. Geol. 55, 319–322. doi: 10.1016/0037-0738(88)90137-6

He, Y. R. (2003). Purple soils in China. Chinese Science Press, Beijing (In Chinese).

He, W. Y., Zhang, M. M., Jin, G. Z., Sui, X., Zhang, T., and Song, F. Q. (2021). Effects of nitrogen deposition on nitrogen-mineralizing enzyme activity and soil microbial community structure in a Korean pine plantation. *Microb. Ecol.* 81, 410–424. doi: 10.1007/s00248-020-01595-6

Huang, L. L. (2012). Effect of nitrogen application on acidification and surface electrochemical properties of purple soil [master's thesis]. [China (Chongqi Province)]. Southwest University (In Chinese).

Huang, L. M., Zhang, G. L., and Yang, J. L. (2013). Weathering and soil formation rates based on geochemical mass balances in a small forested watershed under acid precipitation in subtropical China. *Catena* 105, 11–20. doi: 10.1016/j.catena.2013.01.002

Israeli, Y., Salhov, E., and Emmanuel, S. (2021). Impact of textural patterns on modeled rock weathering rates and size distribution of weathered grains. *Earth Surf. Process. Landf.* 46, 1177–1187. doi: 10.1002/esp.5093

Jeong, D., and Bae, H. (2021). Insight into functionally active bacteria in nitrification following Na<sup>+</sup> and Mg<sup>2+</sup> exposure based on 16S rDNA and 16S rRNA sequencing. *Sci. Total Environ.* 758:143592. doi: 10.1016/j.scitotenv.2020.143592

Jurenas, D., and Journet, L. (2021). Activity, delivery, and diversity of type VI secretion effectors. *Mol. Microbiol.* 115, 383–394. doi: 10.1111/mmi.14648

Katoh, K., Misawa, K., Kuma, K., and Miyata, T. (2002). MAFFT: a novel method for rapid multiple sequence alignment based on fast Fourier transform. *Nucleic Acids Res.* 30, 3059–3066. doi: 10.1093/nar/gkf436

Kirschbaum, A., Martínez, E., Pettinari, G., and Herrero, S. (2005). Weathering profiles in granites, sierra Norte (Córdoba, Argentina). *J. S. Am. Earth Sci.* 19, 479–493. doi: 10.1016/j.jsames.2005.06.001

Köhler, S. J., Dufaud, F., and Oelkers, E. H. (2003). An experimental study of illite dissolution kinetics as a function of pH from 1.4 to 12.4 and temperature from 5 to 50°C. *Geochim. Cosmochim. Acta* 67, 3583–3594. doi: 10.1016/s0016-7037(03)00163-7

Krumbein, W. E. (1983). Microbial geochemistry. Blackwell, London.

Li, G., and Elderfield, H. (2013). Evolution of carbon cycle over the past 100 million years. *Geochim. Cosmochim. Acta* 103, 11–25. doi: 10.1016/j.gca.2012.10.014

Li, Y., Tian, D. S., Wang, J. S., Niu, S. L., Tian, J., Ha, D. H., et al. (2019). Differential mechanisms underlying responses of soil bacterial and fungal communities to nitrogen and phosphorus inputs in a subtropical forest. *PeerJ* 7:e7631. doi: 10.7717/peerj.7631

Liu, G. S. (1996). Soil physical and chemical analysis and profile description. Standards Press of China, Beijing. (In Chinese).

Liu, G. C. (2008). Erosion law and control Technology of Purple Soil. Chengdu, China: Sichuan University press. (In Chinese).

Liu, W. J., Xu, Z. F., Sun, H. G., Zhao, T., Shi, C., and Liu, T. Z. (2018). Geochemistry of the dissolved loads during high-flow season of rivers in the southeastern coastal region of China: anthropogenic impact on chemical weathering and carbon sequestration. *Biogeosciences* 15, 4955–4971. doi: 10.5194/bg-15-4955-2018

Liu, C. J., Zhang, F., Wang, X. P., Chan, N. W., Rahman, H. A., Yang, S. T., et al. (2022). Assessing the factors influencing water quality using environment water quality index and partial least squares structural equation model in the Ebinur Lake watershed, Xinjiang, China. *Environ. Sci. Pollut. Res.* 29, 29033–29048. doi: 10.1007/ \$11356-021-17886-5

Lu, R. K. (1999). *Methods of agrochemical analysis of soils*. China Agricultural Science and Technology Press, Beijing. (In Chinese).

Martin, M. (2011). Cutadapt removes adapter sequences from high-throughput sequencing reads. *EMBnet J* 17, 10–12. doi: 10.14806/ej.17.1.200

Mcardle, B. H., and Anderson, M. J. (2001). Fitting multivariate models to community data: a comment on distance-based redundancy analysis. *Ecology* 82, 290–297. doi: 10.1890/0012-9658(2001)082[0290:FMMTCD]2.0.CO;2

McDonald, D., Price, M., Goodrich, J., Nawrocki, E. P., DeSantis, T. Z., Probst, A., et al. (2012). An improved Greengenes taxonomy with explicit ranks for ecological and evolutionary analyses of bacteria and archaea. *ISME J.* 6, 610–618. doi: 10.1038/ ismej.2011.139

Meidute, S., Demoling, F., and Erland, B. (2008). Antagonistic and synergistic effects of fungal and bacterial growth in soil after adding different carbon and nitrogen sources. *Soil Biol. Biochem.* 40, 2334–2343. doi: 10.1016/j.soilbio.2008.05.011

Mo, B. B., and Lian, B. (2011). Interactions between Bacillus mucilaginosus and silicate minerals (weathered adamellite and feldspar): weathering rate, products, and reaction mechanisms. *Chin. J. Geochem.* 30, 187–192. doi: 10.1007/s11631-011-0500-z

Moore, J., Jacobson, A. D., Holmden, C., and Craw, D. (2013). Tracking the relationship between mountain uplift, silicate weathering, and long-term  $CO_2$  consumption with ca isotopes: southern Alps, New Zealand. *Chem. Geol.* 341, 110–127. doi: 10.1016/j.chemgeo.2013.01.005

Nesbitt, H. W., and Young, G. M. (1982). Early Proterozoic climates and plate motions inferred from major element chemistry of lutites. *Nature* 299, 715–717. doi: 10.1038/299715a0

Nie, Y. X., Wang, M. C., and Wei, Z. (2018). Ammonium nitrogen content is a dominant predictor of bacterial community composition in an acidic forest soil with exogenous nitrogen enrichment. *Sci. Total Environ.* 624, 407–415. doi: 10.1016/j. scitotenv.2017.12.142

Pacheco, F. A. L., Landim, P. M. B., and Szocs, T. (2013). Anthropogenic impacts on mineral weathering: a statistical perspective. *Appl. Geochem.* 36, 34–48. doi: 10.1016/j. apgeochem.2013.06.012

Price, M. N., Dehal, P. S., and Arkin, A. P. (2009). FastTree: Computing Large Minimum Evolution Trees with Profiles instead of a Distance Matrix. *Mol Biol Evol* 26, 1641–1650. doi: 10.1093/molbev/msp077

Rice, K. C., and Herman, J. S. (2012). Acidification of earth an assessment across mechanisms and scales. *Appl. Geochem.* 27, 1–14. doi: 10.1016/j. apgeochem.2011.09.001

Rousk, J., Bååth, E., Brookes, P. C., Lauber, C. L., Lozupone, C., Caporaso, J. G., et al. (2010). Soil bacterial and fungal communities across a pH gradient in an arable soil. *ISME J.* 4, 1340–1351. doi: 10.1038/ismej.2010.58

Shi, Y., Li, Y. T., and Yang, T. (2021). Threshold effects of soil pH on microbial cooccurrence structure in acidic and alkaline arable lands. *Sci. Total Environ.* 800:149592. doi: 10.1016/j.scitotenv.2021.149592

Sokolova, T. A. (2013). Decomposition of clay minerals in model experiments and in soils: possible mechanisms, rates, and diagnostics (analysis of literature). *Eurasian Soil Sci.* 46, 182–197. doi: 10.1134/S1064229313020130

Song, C., Liu, C. L., and Han, G. L. (2017). Impact of fertilization with irrigation on carbonate weathering in an agricultural soil in northern China: a column experiment. *Geochem. J.* 5, 605–616. doi: 10.5194/esurf-5-605-2017

Sorokin, D. Y., Vejmelkova, D. K., Lücker, S., Streshinskaya, G. M., Rijpstra, W. I. C., Sinninghe-Damste, J., et al. (2014). Nitrolancea hollandica gen. Nov., sp. nov., a chemolithoautotrophic nitrite-oxidizing bacterium from a bioreactor belonging to the phylum *Chloroflexi. Int. J. Syst. Evol. Microbiol.* 64, 1859–1865. doi: 10.1099/ js.0.062232-0

Strefler, J., Amann, T., Bauer, N., Kriegler, E., and Hartmann, J. (2018). Potential and costs of carbon dioxide removal by enhanced weathering of rocks. *Environ. Res. Lett.* 13:034010. doi: 10.1088/1748-9326/aaa9c4

Sun, X. J., Cai, P., Sørensen, S. J., Mortimer, M., Gao, C. H., Huang, Q. Y., et al. (2020). Interspecific interactions in dual-species biofilms of soil bacteria: effects of fertilization practices. J. Soils Sediments 20, 1494–1501. doi: 10.1007/s11368-019-02500-6

Tabatabai, M. A., and Olson, R. A. (1985). Effect of acid rain on soils. Crit. Rev. Environ. Sci. Technol. 15, 65–110. doi: 10.1080/10643388509381727

Tang, J. Y., Bullen, N. P., Ahmad, S., and Whitney, J. C. (2018). Diverse NADase effector families mediate interbacterial antagonism via the type VI secretion system. *J. Biol. Chem.* 293, 1504–1514. doi: 10.1074/jbc.RA117.000178

Them, T. R., Gill, B. C., Selby, D., Gröcke, D. R., Friedman, R. M., and Owens, J. D. (2017). Evidence for rapid weathering response to climatic warming during the Toarcian oceanic anoxic event. *Sci. Rep.* 7:5003. doi: 10.1038/s41598-017-05307-y

Tombácz, E., and Szekeres, M. (2006). Surface charge heterogeneity of kaolinite in aqueous suspension in comparison with montmorillonite. *Appl. Clay Sci.* 34, 105–124. doi: 10.1016/j.clay.2006.05.009

Torres, M. A., West, A. J., Clark, K. E., Pairs, G., Bouchez, J., Ponton, C., et al. (2016). The acid and alkalinity budgets of weathering in the Andes–Amazon system: insights into the erosional control of global biogeochemical cycles. *Earth Planet. Sci. Lett.* 450, 381–391. doi: 10.1016/j.epsl.2016.06.012

Trunk, K., Peltier, J., Liu, Y. C., Dill, B. D., Walker, L., Gow, N. A. R. G., et al. (2018). The type VI secretion system deploys antifungal effectors against microbial competitors. *Nat. Microbiol.* 3, 920–931. doi: 10.1038/s41564-018-0191-x

Wang, L. X., Mou, P. P., Huang, J. H., and Wang, J. (2007). Spatial heterogeneity of soil nitrogen in a subtropical forest in China. *Plant Soil* 295, 137–150. doi: 10.1007/s11104-007-9271-z

Wang, Q., Wang, R. R., He, L. Y., and Sheng, X. F. (2017). Location-related differences in weathering behaviors and populations of culturable rock-weathering bacteria along a hillside of a rock mountain. *Microb. Ecol.* 73, 838–849. doi: 10.1007/s00248-016-0921-7

Warby, R. A. F., Johnson, C. E., and Driscoll, C. T. (2009). Continuing acidification of organic soils across the northeastern USA: 1984–2001. *Soil Sci. Soc. Am. J.* 73, 274–284. doi: 10.2136/sssaj2007.0016

Xu, Z. Y. (2007). Rock mechanics. China Water and Power Press, Beijing (In Chinese).

Xu, Z. F., and Liu, C. Q. (2010). Water geochemistry of the Xijiang basin rivers, South China: chemical weathering and  $CO_2$  consumption. *Appl. Geochem.* 25, 1603–1614. doi: 10.1016/j.apgeochem.2010.08.012

Yang, X. X., Zhao, G. X., Li, T., Wan, G., and Li, P. (2016). Characteristics of soil acidification and its influencing factors in Shandong province. *Trans. Chin. Soc. Agric. Eng.* 32, 155–160. (In Chinese. doi: 10.11975/j.issn.1002-6819.2016.z2.021

Yin, L., Xu, X. K., and Jing, Y. S. (2015). Comparison of two measurement methods of dissolved organic carbon and dissolved total nitrogen. *Environ. Sci. Technol.* 389, 93–97. (In Chinese)

Yu, S., He, S. Y., Sun, P. A., Pu, J. B., Huang, J., Luo, H. X., et al. (2016). Impacts of anthropogenic activities on weathering and carbon fluxes: a case study in the Xijiang River basin, Southwest China. *Environ. Earth Sci.* 75:589. doi: 10.1007/s12665-015-5226-5

Yu, C. Q., Huang, X., Chen, H., Godfray, H. C. J., Wright, J. S., Hall, J. W., et al. (2019). Managing nitrogen to restore water quality in China. *Nature* 567, 516–520. doi: 10.1038/ s41586-019-1001-1

Zeng, J., Liu, X. J., Song, L., Liu, X. G., Zhang, H. Z., Shen, C. C., et al. (2015). Nitrogen fertilization directly affects soil bacterial diversity and indirectly affects bacterial community composition. *Soil Biol. Biochem.* 92, 41–49. doi: 10.1016/j.soilbio.2015.09.018

Zhang, D., Chen, A. Q., Zhao, J. X., Lu, C. H., and Liu, G. C. (2017). Quantitative model prediction of the combined effect of moisture content and temperature on purple mudstone decay in South-Western China. *Geomorphology* 295, 656–661. doi: 10.1016/j. geomorph.2017.07.023

Zhang, R. K., and Fan, G. (2003). Quantitative analytic method and experiments of X-ray diffration phase of clay minerals. *Uranium. Geol.* 19, 180–185. (In Chinese). doi: 10.3969/j.issn.1000-0658.2003.03.009

Zhang, Y. T., Shen, H., He, X. H., Thomas, B. W., Lupwayi, N. Z., Hao, X. Y., et al. (2017). Fertilization shapes bacterial community structure by alteration of soil pH. *Front. Microbiol.* 8:1325. doi: 10.3389/fmicb.2017.01325

Zhang, N. L., Wan, S. Q., Li, L. H., Bi, J., Zhao, M. M., and Ma, K. P. (2008). Impacts of urea N addition on soil microbial community in a semi-arid temperate steppe in northern China. *Plant Soil* 311, 19–28. doi: 10.1007/s11104-008-9650-0

Zhang, Y. W., Zeng, J. H., and Yu, B. S. (2009). Experimental study on interaction between simulated sandstone and acidic fluid. *Pet. Sci.* 6, 8–16. doi: 10.1007/s12182-009-0002-3

Zhao, J. X., Deng, L. M., Lu, C. H., and Liu, G. C. (2021). Effects of simulated acid rain leaching on weathering and soil forming characteristics of purple parent rock. *J. Southwest Univ.* 43, 151–161. (In Chinese). doi: 10.13718/j.cnki.xdzk.2021.11.01

Zheng, B. X., Bi, Q. F., Hao, X. L., Zhou, G. W., and Yang, X. R. (2017). *Massilia* phosphatilytica sp. Nov., a phosphate solubilizing bacteria isolated from a long-term fertilized soil. *Int. J. Syst. Evol. Microbiol.* 67, 2514–2519. doi: 10.1099/ijsem.0.001916

Zhou, Z. H., Wang, C. K., Zheng, M. H., Jiang, L. F., and Luo, Y. Q. (2017). Patterns and mechanisms of responses by soil microbial communities to nitrogen addition. *Soil Biol. Biochem.* 115, 433–441. doi: 10.1016/j.soilbio.2017.09.015

Zhou, Z. H., Zheng, M. H., Xia, J. Y., and Wang, C. K. (2022). Nitrogen addition promotes soil microbial beta diversity and stochastic assembly. *Sci. Total Environ.* 806:150569. doi: 10.1016/j.scitotenv.2021.150569

Zhu, H. Y., Wu, L. J., Xin, G. L., Yu, S., Guo, Y. S., and Wang, J. J. (2019). Impact of anthropogenic sulfate deposition via precipitation on carbonate weathering in a typical industrial city in a karst basin of Southwest China: a case study in Liuzhou. *Appl. Geochem.* 110:104417. doi: 10.1016/j.apgeochem.2019.104417

Zhu, B., Yu, J., Qin, X., Rioual, P., Zhang, Y., Liu, Z., et al. (2013). Identification of rock weathering and environmental control in arid catchments (northern Xinjiang) of Central Asia. *J. Asian Earth Sci.* 66, 277–294. doi: 10.1016/j.jseaes.2013.02.005