



Varied Growth Response of Cogongrass Ecotypes to Elevated CO₂

G. Brett Runion^{1*}, Stephen A. Prior¹, Ludovic J. A. Capo-chichi², H. Allen Torbert¹ and Edzard van Santen³

¹ National Soil Dynamics Laboratory, Agricultural Research Service, United States Department of Agriculture, Auburn, AL, USA, ² Alberta Innovates Technology Futures, Vegreville, AB, Canada, ³ Department of Crop, Soil and Environmental Sciences, Auburn University, Auburn, AL, USA

Cogongrass [*Imperata cylindrica* (L.) P. Beauv] is an invasive C₄ perennial grass which is listed as one of the top ten worst weeds in the world and is a major problem in the Southeast US. Five cogongrass ecotypes [Florida (FL), Hybrid (HY), Louisiana (LA), Mobile (MB), and North Alabama (NA)] collected across the Southeast and a red-tip (RT) ornamental variety were container grown for 6 months in open top chambers under ambient and elevated (ambient plus 200 ppm) atmospheric CO₂. Elevated CO₂ increased average dry weight (13%) which is typical for grasses. Elevated CO₂ increased height growth and both nitrogen and water use efficiencies, but lowered tissue nitrogen concentration; again, these are typical plant responses to elevated CO₂. The HY ecotype tended to exhibit the greatest growth (followed by LA, NA, and FL ecotypes) whiles the RT and MB ecotypes were smallest. Interactions of CO₂ with ecotype generally showed that the HY, LA, FL, and/or NA ecotypes showed a positive response to CO₂ while the MB and RT ecotypes did not. Cogongrass is a problematic invasive weed in the southeastern U.S. and some ecotypes may become more so as atmospheric CO₂ continues to rise.

Keywords: carbon dioxide, global change, Imperata cylindrica, invasive weed, nitrogen use efficiency, water use efficiency

INTRODUCTION

Invasive plants are estimated to cost U.S. agricultural and forest producers 34 billion dollars annually from decreased productivity and increased cost of weed control and are considered to be a major threat to the Earth's biodiversity (Pimentel, 2002). Elevated CO₂ stimulates plant photosynthesis, resource use efficiency, and biomass production (Amthor, 1995) which may affect the physiology and competitiveness of invasive plants. However, the effects of elevated CO_2 on invasive plants remains an understudied aspect of global change research. Bright (1998) summarizes, "Fast-growing, highly invasive plants may also be able to profit directly from the atmosphere's increased carbon content...any slower-growing natives would tend to lose out to the invaders." It has been suggested that the increase in the atmospheric concentration of CO_2 since

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*Correspondence:

G. Brett Runion brett.runion@ars.usda.gov

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Abbreviations: FL, Florida cogongrass ecotype; HY, Louisiana/Mobile hybrid cogongrass ecotype; LA, Louisiana cogongrass ecotype; MB, Mobile cogongrass ecotype; NA, North Alabama cogongrass ecotype; NUE, nitrogen use efficiency; OTC, Open Top Chamber; RT, red-tip ornamental cogongrass variety; WUE, water use efficiency.

the beginning of the 20th century may be a primary factor affecting the establishment and spread of some invasive species (Ziska, 2003).

Invasive plants can disrupt terrestrial ecosystems, particularly in the southeastern U.S. with its numerous ports of entry and mild climate. One example that has become a serious problem is cogongrass [Imperata cylindrica (L.) P. Beauv], a perennial grass native to Southeast Asia which was introduced into the southeastern U.S. in the early 1900s (Tabor, 1949) for forage, erosion control and as packing material (Bryson and Carter, 1993). It is a widespread invader to warmer regions (>500 million ha worldwide), is tolerant of shade, poor soils, and drought and naturalizes aggressively in dense monocultures which displace native plants (Bryson and Carter, 1993). Cogongrass is one of the top ten worst weeds in the world (Holm et al., 1991) and is a Federal Noxious Weed (Miller et al., 2010). Cogongrass is a major problem in the Southeast on disturbed lands such as forest plantations and roadsides and may become problematic on agricultural lands (Patterson et al., 1980). It is present in five or more varieties including a commercially available red-tip (RT; 'Red Baron') ornamental sold by nurseries in some states (Capo-chichi et al., 2008). Sale of this RT variety is prohibited in some southern states and removal of prior plantings has been recommended since it has viable pollen that might spread to invasive cogongrass plants and has been known to revert back to the green aggressive type (Miller et al., 2010).

It has been suggested that populations introduced from several origins over an extended period of time should have higher genetic diversity than populations that were introduced only a few times or from a single source (Pappert et al., 2000). For example, cogongrass was first introduced to Alabama from Japan (Tabor, 1952), but has likely also arrived from other locations to different ports of entry. Genetic characterization of differing populations of cogongrass may help explain spread dynamics and means of establishment (Capo-chichi et al., 2008). These investigators determined that genetic variation within and between cogongrass sites in the southern U.S. was quite large given how recently it was introduced. Spread dynamics were found to be greatly influenced by anthropogenic activities (e.g., soil disturbance, canopy removal) compared to natural factors which may accelerate opportunities for bringing together cross-compatible species previously isolated by ecology and/or geography. The objective of this study was to evaluate the response of five cogongrass ecotypes collected across the southeastern U.S. plus the RT variety to ambient and elevated atmospheric CO₂. This is the first study to examine the effects of elevated CO₂ on cogongrass and the first for any weed species to look at potential differences among ecotypes.

MATERIALS AND METHODS

The study was conducted at the soil bin facilities at the USDA-ARS National Soil Dynamics Laboratory, Auburn, Alabama. The bin used for the experimental setup is 6 m wide and 76 m long and has been modified for container studies; modifications consisted of installing a geomembrane liner (20 mL) and gravel drain system to ensure a good working surface and drainage for container studies. Open top field chambers (OTC; Rogers et al., 1983a), encompassing 7.3 m×7.3 m of ground surface area, were used to continuously deliver target CO₂ concentrations of ambient or ambient plus 200 μ mol mol⁻¹ (elevated) using a delivery and monitoring system described by Mitchell et al. (1995).

The six cogongrass ecotypes used in this study were from a collection maintained at Auburn University and included Louisiana, North Alabama, Florida, Mobile, a LA-MB HY, and RT. The MB ecotype was from the suspected point of introduction near Grand Bay in Mobile County, AL as described by Tabor (1952). The RT represents a commercially available variety that can be found in ornamental nurseries. Rhizomes were collected, stored in plastic bags, and transported to glasshouses at the Plant Sciences Research Center of the Alabama Agricultural Experiment Station on the campus of Auburn University for molecular analysis. An out-crossing involving the non-native and native species led to different genotypes such as the LA-MB HY. Amplified Fragment Length Polymorphism (AFLP) confirmed that the cogongrass ecotype LA-MB HY was derived from nonnative (wild type cogongrass ecotype) and native (non-invasive cogongrass ecotype) species (Capo-chichi et al., 2008). The atpBrbcL non-coding spacer of chloroplast DNA revealed that only a few nucleotide substitutions contributed to the variation among the wild cogongrass MB ecotype and the RT (data not shown).

Plants were grown in a peat-based general purpose growing medium (PRO-MIX Bx, Premier Horticulture Inc., Quakertown, PA 18951, USA) in 1.65 L tree-pots (Short One Tree-pot, 10 cm \times 23 cm, Stuewe and Sons Inc., Corvallis, OR 97333, USA) in a glasshouse for establishment (\sim 3 wk). Plants were then transplanted into 10.65 L tree pots (TPOT4 Round Treepot, 22 cm \times 39 cm, Stuewe and Sons Inc., Corvallis, OR 97333, USA) containing the same standard growth medium described above. Forty-eight containers of each ecotype were selected for use in the study. These plants were ranked, according to size and placed into four groups of 12 containers each, representing the largest 12 first in declining order down to the smallest 12; one container from each group was randomly assigned to each of the 12 OTCs (i.e., four containers of each plant ecotypes in each chamber). The study was conducted as a randomized complete block design with the six blocks occurring along the length of the soil bin. Plants were fertilized monthly with Miracle-Gro (15:30:15, N:P:K; Scotts Products Inc., Marysville, OH, USA) according to manufacture recommendations by mixing 600 g Miracle-Gro in 130 L deionized water; each plant received 500 mL of this solution. Containers were subjected to ambient rainfall and watered once or twice a week (1 L per container) to prevent drought-induced plant mortality.

Prior to harvest, WUE was calculated from LI-6400 Portable Photosynthesis System (LI-COR, Inc., Lincoln, NE, USA) measurements. After 6 months, height was measured and the aboveground portions were harvested by severing the plant at the ground-line. Roots were separated from the growing medium using the sieve method (Bohm, 1979). Above- and belowground plant components were then dried separately in a forced-air oven at 55°C to a constant weight, and dry weights recorded. Subsamples of above- and below ground biomass (1 mm sieve) were analyzed separately for N by dry combustion using a LECO TruSpec analyzer (LECO Corp., St. Joseph, MI, USA). Prior to analyses, data from the four containers of each ecotype within each chamber were averaged making the chamber the experimental unit (N = 6). Statistical analyses were conducted using the Mixed Models procedure (Proc Mixed) from SAS (Littell et al., 1996). In all cases, differences were considered significant at $P \le 0.05$ and trends were recognized at $0.05 > P \le 0.10$.

RESULTS

Cogongrass height was significantly increased (P < 0.001) under elevated atmospheric CO₂ (111.3 cm) compared to ambient conditions (105.4 cm) when averaged across all ecotypes (**Figure 1**). When averaged across CO₂ concentrations, significant differences in height (P < 0.001) were noted among the ecotypes (i.e., LA = NA > FL > HY > MB > RT). Further, a significant interaction (P = 0.001) showed that height was increased by elevated CO₂ for LA, NA, FL, and HY only (**Figure 1**).

Similarly, above ground dry weight was significantly increased (P = 0.001) 13% under elevated (98.5 g) compared with ambient (87.4 g) CO₂ (**Figure 2**). A significant main effect of ecotype (P < 0.001) was also observed (i.e., HY > LA > NA = FL > MB > RT) as was a trend for an interaction (P = 0.095), where dry weight was increased by elevated CO₂ for HY, LA, and FL only (**Figure 2**).

Although elevated CO₂ resulted in a slight increase (8.9%) in belowground dry weight (elevated = 155.6 g vs. ambient = 142.9 g; **Figure 3**), this effect was not statistically significant (P = 0.118). However, the main effect of ecotype was



FIGURE 1 | Height of cogongrass ecotypes under ambient and elevated atmospheric CO₂ (HY, Hybrid; LA, Louisiana; NA, North Alabama; FL, Florida; MB, Mobile; RT, Red-tip). N = 6. Means with standard errors are shown; standard errors reflect the variability in the data and are not a means separation technique. Main effects of CO₂ (P < 0.001), ecotype (P < 0.001), and their interaction (P = 0.001) were significant. Bars with different letters show a significant effect of CO₂ for each ecotype; main effect ecotype means shown at the top of the graph (means followed by the same letter are not significantly different).



FIGURE 2 | Aboveground biomass of cogongrass ecotypes under ambient and elevated atmospheric CO₂ (HY, Hybrid; LA, Louisiana; NA, North Alabama; FL, Florida; MB, Mobile; RT, Red-tip). N = 6. Means with standard errors are shown; standard errors reflect the variability in the data and are not a means separation technique. Main effects of CO₂ (P = 0.001) and ecotype (P < 0.001) were significant and their interaction showed a trend (P = 0.095). Bars with different letters show a significant effect of CO₂ for each ecotype; main effect ecotype means shown at the top of the graph (means followed by the same letter are not significantly different).



FIGURE 3 | Belowground biomass of cogongrass ecotypes under ambient and elevated atmospheric CO₂ (HY, Hybrid; LA, Louisiana; NA, North Alabama; FL, Florida; MB, Mobile; RT, Red-tip). N = 6. Means with standard errors are shown; standard errors reflect the variability in the data and are not a means separation technique. Main effect of ecotype (P < 0.001) was significant while the main effect of CO₂ showed a trend (P = 0.118) and their interaction (P = 0.487) was not significant. The omission of letters above bars indicates no effect of CO₂ for any ecotype; main effect ecotype means shown at the top of the graph (means followed by the same letter are not significantly different).

significant (P < 0.001; HY > NA > FL = LA > MB > RT. No significant CO₂ by ecotype interaction (P = 0.487) was noted for this measure.

Unlike growth measurements, the main effect of CO₂ showed significantly lower (P < 0.001) tissue nitrogen concentration [N] under elevated (6.06 mg N/g) than ambient (6.74 mg N/g) CO₂ (**Figure 4**). A significant main effect of ecotype (P < 0.001) indicated that RT > FL = HY = NA > MB = LA. A significant interaction of CO₂ with ecotype (P = 0.008) showed that [N] was decreased by elevated CO₂ for the FL, HY, and NA ecotypes



FIGURE 4 | Total plant nitrogen concentration of cogongrass ecotypes under ambient and elevated CO₂ (HY, Hybrid; LA, Louisiana; NA, North Alabama; FL, Florida; MB, Mobile; RT, Red-tip). N = 6. Means with standard errors are shown; standard errors reflect the variability in the data and are not a means separation technique. Main effects of CO₂ (P < 0.001), ecotype (P < 0.001), and their interaction (P = 0.008) were significant. Bars with different letters show a significant effect of CO₂ for each ecotype; main effect of ecotype means shown at the top of the graph (means followed by the same letter are not significantly different).



FIGURE 5 | Nitrogen use efficiency of cogongrass ecotypes under ambient and elevated CO₂ (HY, Hybrid; LA, Louisiana; NA, North Alabama; FL, Florida; MB, Mobile; RT, Red-tip). N = 6. Means with standard errors are shown; standard errors reflect the variability in the data and are not a means separation technique. Main effects of CO₂ (P = 0.001) and ecotype (P < 0.001) significant; their interaction (P = 0.118) showed a trend. The omission of letters above bars indicates no effect of CO₂ for any ecotype; main effect ecotype means shown at the top of the graph (means followed by the same letter are not significantly different).

only (**Figure 4**). Calculations of NUE as g plant biomass produced per g plant N were significant (P = 0.001) for the main effect of CO₂ (ambient = 153.1 vs. elevated = 175.2; **Figure 5**). Ecotype, averaged across both CO₂ treatments, significantly (P < 0.001) affected NUE (i.e., LA > MB = NA = HY = FL > RT). The CO₂ by ecotype interaction was not significant for NUE (P = 0.118); however, NUE showed a similar response as other variables in that FL, HY, LA, and NA were numerically higher under elevated CO₂ while MB and RT were actually slightly lower (**Figure 5**).

Water use efficiency, calculated from LICOR gas exchange measurements as mmol CO₂ per mol H₂O, was significantly



FIGURE 6 | Water use efficiency of cogongrass ecotypes under ambient and elevated CO₂ (HY, Hybrid; LA, Louisiana; NA, North Alabama; FL, Florida; MB, Mobile; RT, Red-tip). N = 6. Means with standard errors are shown; standard errors reflect the variability in the data and are not a means separation technique. Main effect of CO₂ (P = 0.001) was significant while ecotype (P = 0.513) and their interaction (P = 0.226) were not. The omission of letters above bars indicates no effect of CO₂ for any ecotype; main effect ecotype means shown at the top of the graph.

increased (P = 0.001) 96% by growth in elevated CO₂ (ambient = 6.8 vs. elevated = 13.3; **Figure 6**). The main effect of ecotype (P = 0.513) and its interaction with CO₂ (P = 0.226) did not affect WUE (**Figure 6**).

DISCUSSION

Cogongrass growth parameters were increased when exposed to elevated CO_2 , which is typical of most plants (Rogers et al., 1994; Amthor, 1995). Although height was only slightly higher (5.6%; **Figure 1**), aboveground dry weight increase (12.7%; **Figure 2**) was in a range (10–15%) typical of C₄ plant response to CO_2 enrichment (Kimball, 1983; Prior et al., 2003). The fact that responsive ecotypes showed growth responses to elevated CO_2 typical for C₄ plants suggests that their invasive potential will not be altered as atmospheric CO_2 continues to rise, but does indicate that some ecotypes of this serious invasive weed may become more problematic.

Cogongrass ecotype also affected both height and aboveground dry weight; in general, the MB and RT were smaller than the other ecotypes. The significant interactions for these variables further showed that MB and RT were not responsive to CO_2 concentration, while the other ecotypes tended to be larger under high CO_2 . It is interesting to note that HY tended to exhibit the greatest response to elevated CO_2 among ecotypes even though it was a cross from the MB ecotype which was not responsive. It is not uncommon for HYs to exhibit growth responses that either differ from or exceed their progenitors (Capo-chichi et al., 2008). This is, after all, why plant breeding programs exist for virtually all important crop species.

Despite the importance of root systems in attaining essential soil resources (i.e., water and nutrients), their response to CO_2 has received less attention than aboveground parts; however, it has been reported that roots often exhibit a larger response

to elevated CO_2 than other plant organs (Rogers et al., 1994). In contrast, our study showed no increase in root dry weight under elevated CO_2 and no significant interaction with cogongrass ecotype. It is interesting to note that, despite the lack of significance, the belowground response pattern (**Figure 3**) was similar to that seen for aboveground growth (**Figure 2**). Further, ecotype effect also followed the same general pattern as aboveground dry weight in that MB and RT were smaller than the other ecotypes.

Given that weed species are more likely to have greater genetic diversity and physiological plasticity (compared to crops), they are more likely to be able to adapt to a changing environment (Ziska and Runion, 2007). This may have significant implications for developing effective weed control stategies given that elevated CO_2 may increase herbicide tolerance in some weeds due to a herbicide dilution effect caused by increased growth, as well as other potential CO_2 -induced changes in plant morphology, biochemistry, and physiology (Ziska et al., 1999; Ziska and Teasdale, 2000; Archambault et al., 2001). However, increased herbicide tolerance under elevated CO_2 is not always observed (Marble et al., 2015). How cogongrass herbicide efficacy will be impacted by elevated CO_2 is not known and deserves futher study.

Total plant nitrogen concentration was reduced under elevated CO_2 (Figure 4). As with plant growth, this is a common response to high CO_2 (Rogers et al., 1994; Norby et al., 2001; Prior et al., 2008; Runion et al., 2009). This is a result of increased plant growth under elevated CO_2 causing a dilution effect on nutrient concentrations (Rogers et al., 1994, 1999). The RT ecotype had the highest [N] and LA was lowest (Figure 4). RT had the smallest growth which likely resulted in the high [N]; it is unclear why MB did not exhibit this pattern given it also had less growth. A significant CO_2 by ecotype interaction indicated that [N] was lowered by elevated CO_2 in FL, HY, and NA only. In general, these ecotypes had a larger dilution effect due to greater growth (Figures 1 and 2); however, why LA did not follow this pattern is not known.

Another common response to elevated CO_2 is increased NUE (Rogers et al., 1994) as observed in this study (**Figure 5**). Nutrient use efficiency (unit of biomass produced per unit of nutrient) generally increases under elevated CO_2 as plants are able to produce more biomass with available nutrients. Ecotype also affected NUE with LA being highest and RT lowest. Although the CO_2 by ecotype interaction was not significant, the response pattern was similar to other variables in that NUE was numerically higher for FL, HY, LA, and NA but not MB and RT under elevated CO_2 .

As with NUE, it is also common for plants grown under elevated CO₂ to exhibit increases in water use effiency (Rogers and Dahlman, 1993). In general, C₃ plants exhibit increased photosynthesis and decreased stomatal conductance under elevated CO₂ (Amthor and Loomis, 1996), leading to increased WUE. However, the CO₂-concentrating mechanism used by C₄ species limits their photosynthetic response to CO₂ enrichment (Amthor and Loomis, 1996), but they do tend to show decreased stomatal conductance which often increases WUE (Rogers et al., 1983b). This was observed in the current study, in that photosynthesis was not significantly affected by CO₂ concentration (ambient = 2.13, elevated = 2.61 μ mol CO₂ m⁻² s⁻¹; *P* = 0.14), while stomatal conductance tended to be decreased (ambient = 0.018, elevated = 0.014 mol H₂O m⁻² s⁻¹; *P* = 0.06) under elevated CO₂ (full data not shown). These responses led to a large increase in WUE (96%) under elevated CO₂ (Figure 6). Ecotype and its interaction with CO₂ did not affect WUE.

CONCLUSION

Cogongrass is one of the top ten worst weeds in the world and is listed as a Federal Noxious Weed. Since its introduction to the Southeastern U.S. it has become a major problem in forest plantations, roadsides, and agricultural systems due to its aggressive ability to develop dense monocultures which can compete with and displace desirable species. This is the first study to examine the effects of elevated CO2 on cogongrass and the first for any weed species to look at potential differences among ecotypes. In our study, elevated CO₂ (averaged across ecotypes) increased height, biomass, and both nitrogen and water use efficiencies, but lowered tissue nitrogen concentration; again, these are typical C₄ plant responses to elevated CO₂. In general, the HY ecotype tended to exhibit the greatest growth (followed by LA, NA, and FL) while RT and MB ecotypes were smallest. Interactions of CO₂ with ecotype showed that MB and RT ecotypes did not respond to CO₂. This lack of response to CO2 for RT (sold commerically as 'Red Baron') is significant since concerns over its potential to spread into the native landscape should not be exacerbated by the rising atmospheric CO₂ concentration. Nevertheless, it is still prohibited for sale in some states and its removal has been recommended. However, HY, LA, FL, and/or NA ecotypes responded positively to elevated CO₂, suggesting some ecotypes of this serious invasive weed may become more problematic in a future CO₂-enriched environment. These findings may influence development of future cogongrass control strategies, a subject area requiring further investigation.

AUTHOR CONTRIBUTIONS

GB conceived and conducted the research, collected and analyzed the data and co-wrote the manuscript; SP assisted with conducting the research, collecting the data, and co-wrote the manuscript; LC provided the cogongrass ecotypes, co-wrote and reviewed the manuscript; HT assisted with conducting the research and reviewed the manuscript; ES assisted with providing the cogongrass ecotypes and reviewed the manuscript.

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Conflict of Interest Statement: The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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