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Site characteristics determine the effectiveness of tillage and cover crops on the net ecosystem carbon balance in California vineyard agroecosystems

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Globally, wine grape vineyards cover approximately 7.4 M ha. The potential for carbon (C) storage in vineyards is of great interest to offset greenhouse gas emissions and mitigate the effects of climate change. Sustainable soil management practices such as cover crop adoption and reduced tillage may contribute to soil organic carbon (SOC) sequestration. However, site-specific factors such as soil texture, other soil physicochemical properties, and climate largely influence the range and rate to which SOC may be stored. To measure the potential for C storage in vineyards under varying sustainable soil management practices, we calculated the net ecosystem carbon balance (NECB) of three cover crops [perennial grass (Poa bulbosa hybrid cv. Oakville Blue); annual grass (barley, Hordeum vulgare); resident vegetation (natural weed population)] under conventional tillage (CT) and no-till (NT) management. Results provided evidence that vineyards served as C sinks. In sandy soils, the type of cover crop and tillage may be of little influence on the NECB. While in finer-textured soils, tillage reduced the NECB and higher biomass-producing cover crops enhanced the overall C storage potential of the vineyard agroecosystem. Overall, our results revealed that site characteristics, namely, soil texture and climate, were key determinants of the C storage potential of vineyards in Mediterranean climates such as those found in coastal and inland California wine grape production regions.

KEYWORDS

climate change, carbon sequestration, cover crops, tillage activities, soil conservation

1 Introduction

As temperatures rise and rain events become more unpredictable due to the changing climate, soils are under threat of loss of soil organic matter (SOM), soil nutrient imbalances, loss of soil biodiversity, contamination, and compaction (Panagos et al., 2020). Moreover, it is estimated that almost 36 billion tons of soils are lost annually due to water and wind erosion (Alsina et al., 2014; Borrelli et al., 2017; Wolff et al., 2018). Soil erosion is further exacerbated by tillage, and an overreliance on soil cultivation for weed control and aeration over the past half century has resulted in a significant loss of SOM across agricultural soils (Alsina et al., 2014; Mitchell et al., 2017). Thus, over the last decade, there has been a substantial increase in attention toward rebuilding SOM and using soils as a tool to mitigate climate change (Lal, 2004; Powlson et al., 2011; Lazcano et al., 2020; Nilahyane et al., 2020).

Traditionally, the interrows of vineyards were kept free of vegetation with the use of herbicides and tillage. However, it has been shown that both practices may have detrimental effects on soil quality and the surrounding ecosystem (Patiño-Zúñiga et al., 2009; Ferreira et al., 2020; Gatti et al., 2022). Thus, the adoption of cover crops and reduction of interrow tillage have been proposed as sustainable alternatives to conventional vineyard floor management practices (Figure 1) (Alsina et al., 2013). Research suggested that cover crops may not only reduce soil erosion and water runoff but also improve water infiltration in most soils of temperate regions by increasing SOM, so soils and

water can be better conserved (Alvaro-Fuentes et al., 2008; Steenwerth and Belina, 2008a; Belmonte et al., 2018; Cataldo et al., 2020). In addition, SOM can be further preserved under reduced till or no-till (NT) practices, whereby soil aggregates and accompanying SOM remain undisturbed (Šimon et al., 2009; Peregrina et al., 2010; Seddaiu et al., 2013). The preservation of SOM in turn bolsters soil organic carbon (SOC), which would ameliorate the soil physical, chemical, and biological functions and was identified as a key target carbon (C) pool in mitigating climate change via C sequestration (Jobbágy and Jackson, 2000; Blanco-Canqui et al., 2013; Stockmann et al., 2013). In fact, the Intergovernmental Panel on Climate Change (IPCC) has estimated that by 2030, global SOC sequestration has the potential to mitigate up to about 5.3 Gt CO₂ per year (Porter et al., 2017). Vineyard agroecosystems represent a large potential for agricultural soil carbon sequestration (SCS): grape vineyards, including wine, table, and raisin grapes, make up 341,555 hectares (844,000 acres) of agricultural land in California [California Department of Food and Agriculture (CDFA), 2022]. Also, the large SCS potential can attribute to the grapevine's long life cycle and permanent woody organs allow them to potentially store higher amounts of C compared to annual crops (Alonso et al., 2014; Nistor et al., 2018). However, there are limitations to SCS, including a lack of standardized methods of SOC determination and uncertainty regarding the stability of different soil C pools (Powlson et al., 2011). Furthermore, the effectiveness and rate of long-term SOC sequestration in agricultural soils can be largely influenced by



FIGURE 1

Framework for increasing soil organic carbon (SOC) conversion effiency in vineyards by increasing inputs of cover crops, compost, and perennialization and decreasing outputs from tillage and erosion avoidance.

site-specific conditions including climate, soil texture, other soil physiochemical properties, and management practices (Carlisle et al., 2010; Powlson et al., 2011).

Previous studies have indeed identified vineyards as C sinks. However, the management practices largely influenced the storage potential of the vineyard systems (Brunori et al., 2016; Payen et al., 2021). Subsequently, management practices that increase SOM have been encouraged over the last decade as an SCS strategy, yet their impact on SOC storage rates remains unclear (Novara et al., 2019). While some studies have recorded an increase in SOC sequestration rate due to cover crop adoption (Steenwerth and Belina, 2008a; Alonso et al., 2014), others have reported the opposite (Celette, 2007; Novara et al., 2019; Jian et al., 2020). Also, the effectiveness has shown significant connections with the longevity of the management practices and site characteristics (Krull et al., 2001; Mitchell et al., 2017).

One source of limitation in quantifying C sequestration in previous literature may be the presence or absence of measurement of CO2 efflux from assessing soil respiration (R_s). Soil CO₂ efflux results from the combination of biological and physical processes, both of which are sensitive to edaphic factors and highly variable in space and time (Ryan and Law, 2005). In soils of hot climates, CO_2 efflux may be great enough to potentially offset short-term storage of C from cover crop adoption (Yu et al., 2019). Furthermore, some studies have shown that cover crops enhance microbial activity, thus increasing CO2 efflux (Steenwerth et al., 2010; Maier et al., 2011; Freidenreich et al., 2021). In fact, CO₂ emissions may be further increased under specific types of cover crops due to the lower C:N ratios of certain plant materials, such as legumes (Alluvione et al., 2010; Freidenreich et al., 2021). Since recent work has shown that SOC increases as tillage frequency or intensity is reduced, some studies have investigated the synergic effects of cover crops combined with reduced till or NT management in California vineyards (Conant et al., 2007; Steenwerth and Belina, 2008a; Wolff et al., 2018). However, there is still a lack of information regarding the potential of C storage in vineyard systems under combinations of soil management practices, considering C stored or lost at vineyard scale.

Thus, a net ecosystem carbon balance (NECB) is needed to elucidate C inputs and outputs in commercial production settings at vineyard scale. One method of estimating net C storage within a system is through the net ecosystem production (NEP) methodology (Randerson et al., 2002). The NEP reflects the balance of ecosystem primary production as stored C minus ecosystem respiration as lost C to determine whether there is net gain or loss of C in the monitored ecosystem. On the other hand, the NEP is effectively parallel to the net primary production (NPP), as NPP only quantifies the C input by plants, but NEP includes the consideration of C input by plants in addition to C release by soils. Therefore, by combining the NPP of both cover crop and grapevine with soil CO_2 efflux and C losses as grape clusters and canes being removed during harvest and dormant season, respectively, the NECB can be quantified to reveal the C gain or loss at vineyard scale (Cates and Jackson, 2019).

Thus, the objective of this study was to investigate the synergic effects of the implementation of cover crops and NT practices by quantifying C inputs and losses at the vineyard scale through the NECB determination in two different wine production regions and to investigate the contributions of specific site characteristics toward these vineyard floor management practices in two hyperarid seasons in California. We hypothesized that the vineyard agroecosystem can serve as a C storage pool, and the effectiveness of the ecological functioning of NT and cover crops will be determined by site characteristics, including climate and soil texture.

2 Materials and methods

2.1 Experimental design

Field experiments were conducted at two sites for two consecutive growing seasons (2019-2020 and 2020-2021). The first site was located in Five Points, Fresno County, CA, USA (36.671514, -119.925823), in a Ruby Cabernet/Freedom (27% Vitis vinifera hybrid) vineyard. Grapevines were planted in 2012 with a spacing of 3.0 m \times 1.2 m (row \times vine) with a row orientation of east to west (E-W). The grapevines were head trained and cane pruned. The vineyard was trained to a high quadrilateral trellis with fruiting wire at 1.54 m and catch wires at 1.68 m above the vineyard floor. The grapevines were drip irrigated with two emitters per plant delivering 4.0 L/h each. The second site was located in Oakville, Napa County, CA, USA (38.428, 122.409), planted with Merlot (clone 181)/3309 C (Vitis riparia × Vitis rupestris). Grapevines were planted in 2018 with a spacing of 3.0 m \times 2.0 m (row \times vine) with a row orientation of E-W. The grapevines were spur pruned and trained to quadrilateral cordons 1.54 m above the vineyard floor with catch wires at 1.68 m. The grapevines were drip irrigated with two emitters per plant delivering 2 L/h each.

At both sites, experiments were arranged in a split-plot 3×2 factorial design (three different cover crops subjected to two tillage managements) with four (Napa) and three replications (Fresno). At both vineyards, each treatment replicate consisted of 15 grapevines. Three vines in the middle of each replicate were used for on-site measurements, including the parameters from both grapevines and soils under grapevines, while the distal vines on either end were treated as border plants. Cover crop treatments included 1) perennial grass (PG) (*Poa bulbosa* hybrid cv. Oakville Blue), 2) annual grass (AG) (barley, *Hordeum vulgare*), and 3) resident vegetation (RV) (natural weed population). Tillage management consisted of NT in which interrows were disked 2–3 cm only once in the fall as preparing for seeding and conventional tillage (CT) that was disked to a

depth of 10 cm once in the fall and twice in the spring to incorporate the cover crop residue. Tillage events in Fresno occurred on 1 November, 16 April, and 25 June for 2020 and 12 October, 28 April, and 18 June for 2021. Tillage events in Oakville occurred on 9 October, 21 April, and 24 June for the 2020 season and 2 October, 16 April, and 17 June for 2021. The cover crop seed was drilled to a 1.5-m-wide strip according to seed manufacturer's recommended practices prior to receiving fall/winter rains in 2019 and 2020 at a rate of 605 kg/ha and 84 kg/ha for the PG and AG treatments, respectively. RV was allowed to grow within a 1.5-m strip in the interrow and mowed according to vineyard manager's discretion. Berms were 1.0 m wide and kept free of vegetation using a glyphosate herbicide application in the spring.

2.2 Site conditions

The site conditions in both Fresno County and Napa County vineyards are presented in Table 1. Annual mean daily air and soil temperatures, as well as maximum and minimum air temperatures (1 January to 31 December), from the two sites were obtained from the California Irrigation Management Information System (CIMIS) stations nearest the experimental vineyard (station #77 in Napa County, CA, and station #2 in Fresno County, CA, USA).

At both sites, air temperatures were consistent between years. However, the mean daily maximum temperature was 0.2°C higher in 2021 than 2020, while it was 1.3°C lower in 2021 compared to 2020 in Napa County. Daily maximum soil temperature was 0.9°C and 0.3°C higher in 2021 compared to 2020 in both Fresno County and Napa County, respectively. Daily average soil temperatures were also slightly higher by a degree of 0.1°C–0.2°C at both sites in 2021. Furthermore, average soil temperatures during both years were 0.3°C–0.5°C higher in Fresno County and 0.4°C–0.5°C higher in Napa County compared to the long-term average for the region over the past 10 years (2011–2021) (Table 1).

At the Fresno County vineyard, approximately 199.0 mm of rain was received at the experimental site beginning in October of the preceding year until the harvest in year 1, while 152.5 mm of rain was received during the same period in year 2. In year 1 of

TABLE 1 Site conditions at two commercial vineyards in Fresno County and Napa County from experimental years (2019–2021) and long-term mean values (2011–2021).

Year	Ai	r temperatur	re (°C)	So	il temperatu	re (°C)	Precipitation (mm) ^b	GDD (°C)	
	Daily max	Daily min	Daily average	Daily max	Daily min	Daily average			
Fresno Count	y (Five points)								
2020									
Mean	25.6	9.1	17.3	25.7	10.7	22.6	199	2,358	
Annual max ^a	35.8	18	27.3	-	-	-	-	-	
Annual min ^a	15.6	0.6	7.6	-	-	-	-	-	
2021									
Mean	25.8	9.2	17.6	26.6	10.4	22.8	152.5	2,488	
Annual max	37.8	18.7	28.5	-	-	-	-	-	
Annual min	12.5	3.1	7.7			-	-	-	
2011-2021									
Mean	25.4	9.4	17.2	25.6	10.1	22.3	209.6	2,259	
Napa County	(Oakville)								
2020									
Mean	24.4	7	14.9	22.5	10	16.5	234.2	1,647	
Annual max	31.8	12.3	21.1	-	-	-	-	-	
Annual min	17.1	2	8.5	-	-	-	-	-	
2021									
Mean	23.1	6.3	14.2	22.8	10.2	16.4	278.3	1,519	
Annual max	30	10.8	19.2	-			-	-	
Annual min	12.5 2.6 7.6		-	-	-	-	-		
2011-2021									
Mean	23.2	7.1	14.5	22.8	7.9	16	577.8	1,504	

^a Annual maximum (max) and annual minimum (min) indicate the greatest or lowest value observed during the respective year. ^b Total precipitation occurred during the annual winter rainy season, calculated from October of the preceding year through September of the following year (e.g., 2020 values were calculated from 1 October 2019 to 30 September 2020). ^c-, not applicable and GDD, growing degree days.

the study, the greatest amount of precipitation was received in March (67 mm), while in year 2, the greatest amount (87 mm) was received in January followed by December and October. Therefore, compared to year 1, year 2 received more fall-towinter precipitation. The Napa County vineyard received a greater amount of rainfall compared to that of Fresno County, as 234.2 mm of rain was received in year 1 and 278.3 mm of rain in year 2. December and January received the greatest amount of precipitation in both year 1 and year 2.

Soil texture was assessed using the hydrometer method (S-14.10) from the North American Proficiency Testing (NAPT) program. The SOC content was measured in the interrows of each experimental unit under dry conditions. In July 2020, bulk density was assessed at the centers of interrows (122 cm from the vine rows) and the edges of the berms (61 cm from the grapevine trunk) using brass rings of 10 cm internal diameter and 7.5 cm length. No differences in bulk densities were found in the soil samples between CT and NT interrows, thus all soil samples were taken at the same depth of 30 cm. Three soil cores were randomly collected per experimental unit to a depth of 30 cm and partitioned into 0-15 cm and 15-30 cm subsamples. Subsamples were homogenized and kept in a cool environment until analysis. At analysis, samples were dried, sieved to <2 mm, ball-milled, and analyzed for SOC by combustion method (S-9.30), and soil texture was determined by hydrometer analysis (S-14.10) according to the NAPT program. Soil pH was determined via the saturated paste method as described by Gavlak et al. (1994).

2.3 Soil respiration

To calculate the losses of soil C through R_s, soil CO₂ efflux was measured in situ using a CIRAS-3 (PP Systems, Amesbury, MA, USA) portable gas exchange system coupled with a closed system soil respiration chamber (SRC-2). The SRC-2 chamber consisted of a soil surface area of 78 cm² and a system volume of 1,171 ml and was placed on the vineyard soil surface in the interrows and on the bare soil under the grapevines in each experimental unit to measure Rs at both locations. To minimize leakage, the chamber was fit onto a 10-cm-diameter polyvinyl chloride (PVC ring) placed 5 cm deep in the vineyard interrow and on the bare soil under the grapevines in each experimental unit. The PVC rings remained in the soil throughout the experiment except for removal for mowing and tillage whereby the ring was replaced at least 24 h prior to sampling. Sampling areas were selected with careful consideration for the least plant material to avoid CO₂ contributions from aboveground plant parts. Additionally, as autotrophic and heterotrophic respiration could not be separated in this study, the measured CO₂ efflux included emissions from all soil processes. Upon measurement, the chamber was allowed to stabilize for 1 min before the gas accumulated in the chamber headspace was continuously

sampled in the closed circuit. Efflux rate of $CO_2~(\mu mol~CO_2~m^{-2}~s^{-1})$ was calculated based on linear fit by the CIRAS-3 analyzer as shown in Eq. 1:

Eq. 1. $R_s = \frac{Cn - Co}{Tn} \times \frac{V}{A}$

where R_s is the respiration rate (CO₂ flux, or moles of CO₂ unit area⁻¹ unit time⁻¹), C_o is the CO₂ concentration at T = 0, and C_n is the concentration at a time T_n later. A is the area of soil surface exposed (78 cm²), and V is the total system volume (1,171 ml). The air within the SRC-2 chamber was continuously and automatically mixed during the measurement period to ensure representative samples.

R_s measurements took place no more than 2 h after solar noon and was measured at six time points per experimental unit in each season in Oakville (24 January, 13 April, 21 April, 23 April, 22 May, and 19 June in 2020; 29 January, 14 April, 16 April, 14 May, 15 June, and 9 July in 2021) and five time points in Fresno (2 March, 25 March, 16 April, 17 April, and 12 June in 2020; 14 February, 24 March, 28 April, 29 April, and 1 July in 2021). Measurement time points were selected to represent soil conditions throughout the season, including the day before and after a tillage event and important precipitation events at both sites. Mean Rs values for the season were calculated for each treatment and the bare soil control. In the first year of the study, soil moisture was measured as the volumetric water content (VWC) at the time of each R_s measurement. No significant differences were found between treatments, and thus, soil moisture was not monitored the following season, and measurement dates were targeted before and after tillage and precipitation events.

2.4 Estimates of net primary productivity

2.4.1 Grapevine net primary production

NPP_{grapevine} was estimated as the summation of annual production (harvest yield, leaf biomass, and cane production) and permanent organs (trunk and root biomass). Harvest commenced when the fruit reached approximately 25°Bx in Oakville (25 August 2020 and 1 September 2021) and 21°Bx in Fresno (6 October 2020 and 7 September 2021). At both sites, clusters from three data vines per experimental unit were manually removed, counted, and weighed on a top-loading balance. Subsamples were collected from clusters within each experimental unit at harvest, and C content (% mass) was determined *via* combustion (Western Region Method S-9.30) (Gavlak et al., 1994).

To assess the leaf biomass at the Fresno vineyard, leaf area index (LAI) was measured in late spring to characterize the grapevine canopy growth by a smartphone program, VitiCanopy, *via* iOS system (Apple Inc., Cupertino, CA, USA) (De Bei et al., 2016) and converted into leaf area based on the ground area (3.6 m²). The gap fraction threshold was set to 0.75, extinction coefficient was set to 0.7, and subdivisions were 25. An

extendable mounting device was used to effectively position the device approximately 75 cm underneath the canopy. The device was positioned with the maximum length of the screen being perpendicular to the cordon, and the cordon in line with the middle of the screen according to previous work (De Bei et al., 2016; Yu and Kurtural, 2020). In each experimental unit, three images were taken to capture half canopy of each vine and analyzed by the software. Subsamples of 100 leaves were collected per experimental unit, and leaf area (cm²) was determined by leaf area meter (Li-Cor 3300, Lincoln, NE, USA), dried at 80°C. Dry weight (g) was recorded, and values were extrapolated to determine LAI as previously reported (Torres et al., 2021). At the Oakville vineyard, two vines per treatment were completely defoliated and biomass was measured. Carbon content (% mass) of leaves was estimated as 56% of dry weight (Zhang et al., 2021).

Cane production (pruning wood weights) was measured at dormancy among the three data vines per experimental unit. The C content (%) of pruning wood was estimated based on previous literature, which was 9% as the average percentage fractions of biomass of canes (Morandé et al., 2017). Annual biomass accumulation in permanent organs (trunk, cordon, and roots) was also acquired from literature and included one prior study based at the same experimental site (Oakville, CA) and another at a vineyard of similar age and productivity in the San Joaquin Valley (Williams et al., 2011; Martínez-Lüscher and Kurtural, 2021).

2.4.2 Cover crop NPP

NPP_{cover crop} was calculated by collecting aboveground and belowground biomass at crop physiological maturity as described previously (Steenwerth and Belina, 2008b) or just before termination. A one, 1 m² quadrat was randomly placed in the interrow of each experimental unit and all aboveground biomass was collected. Belowground biomass (roots) was also collected to a depth of 30 cm. Cover crop fresh biomass was determined and then dried at 60°C for 48 h to obtain the dry biomass. The C contribution (%) of the cover crop was estimated as 50% C of the dry biomass (Cates and Jackson, 2019).

2.5 Determination of the net ecosystem carbon balance

The NECB (Mg C ha⁻¹ year⁻¹) was calculated as follows:

Eq. 2. NECB = $NPP_{grapevine} + SOC + NPP_{cover crop} - R_{s:}$ interrow - harvest - $R_{s: under vine}$

where NPP_{grapevine} is the summation of annual (leaves and fruits) and perennial (permanent organs) growth, SOC is the soil organic carbon sequestered to a depth of 30 cm adjusted for the interrow spatial coverage, NPP_{cover crop} is the sum of aboveground and belowground cover crop biomass to 10 cm, $R_{s; interrow}$ is the soil respiration of the portion of the vineyard

where the cover crop is grown, harvest is the amount of C removed through yield, and $R_{s: under vine}$ is the soil respiration of the portion of the soil left bare. Interrow coverage was estimated as 48% of one hectare and bare soil 52% of one hectare. A positive NECB signifies that the system is the net sink of C, and a negative NECB signifies a net source of C to the atmosphere.

2.6 Statistical analyses

Statistical analyses were conducted with R studio version 3.6.1 (RStudio: Integrated Development for R., Boston, MA, USA) for Mac OS. After normality assessment, data were submitted to a two-way analysis of variance (ANOVA) to assess the statistical differences between the different cover crop and tillage treatments and the respective interaction effects. Means \pm standard errors (SEs) were calculated, and when the F value was significant (P \leq 0.05), a Tukey's "honest significant difference" (HSD) *post-hoc* test was executed by using "agricolae" 1.2-8 R package. Figures were made using GraphPad Prism v8.1.2 for Windows (GraphPad Inc., San Diego, CA, USA).

3 Results

3.1 Soil properties

The soil texture at the Fresno County vineyard is classified as a sandy loam with approximately 66% sand, 22% silt, and 12% clay and a bulk density value 1.4 g cm ⁻³. The SOC (% mass C) was not affected by type of cover crop nor tillage system over the course of the experiment at either depth in the Fresno County vineyard (Table 2). Although SOC was not significantly different between years at this site, greater SOC was observed in the upper 0–15-cm portion of soil compared to the 15–30-cm portion. The soil texture at the Napa County (Oakville) vineyard was classified as a loam with approximately 33% sand, 42% silt, and 25% clay and a bulk density value of 1.3 g cm ⁻³. As was observed at the Fresno vineyard, SOC was greater in the upper 0–15-cm portion of the soil in the Napa County vineyard (Table 2). While the type of cover crop again had no influence on SOC at either depth, tillage reduced SOC at both depths.

3.2 Soil respiration

Across the five readings of R_s measured in the interrows at the Fresno County vineyard, when R_s readings were averaged to determine the seasonal mean R_s , no overall differences were observed over the duration of the experiment (Table 3). Likewise, the effect of tillage was inconsistent on R_s (Figure 2). At the Napa County (Oakville) site, tillage also increased R_s

Factors and		Fresno	County (Five p	oints)		Napa County (Oakville)						
treatment	Average SOC (% mass)	p- value ^a	Bulk density (g cm ⁻³)		C (t ha ⁻¹)	<i>p</i> - value	Average SOC (% mass)	P- value	Bulk density (g cm ⁻³)		C (t ha ⁻¹)	<i>p</i> - value
Tillage system (T))											
NT	0.70	ns	1.4	ns	14.68	ns	1.44 a	**	1.3	ns	28.15 a	**
CT	0.69		1.4		14.39		1.37 b		1.3		26.62 b	
Cover crop (CC)												
AG	0.88	ns	1.4	ns	18.56	ns	1.47	ns	1.3	ns	28.67	ns
RV	0.90		1.4		18.93		1.45		1.3		28.33	
PG	0.86		1.4		17.99		1.41		1.3		27.55	
Depth (D)												
0-15 cm	0.88 a	***	1.4	ns	18.50 a	***	1.45 a	**	1.3	ns	28.18 a	**
15-30 cm	0.50 b		1.4		10.57 b		1.36 b		1.3		26.59 b	
Year (Y)												
2020	0.64	ns	1.4	ns	13.42	ns	1.30	ns	1.3	ns	25.37	ns
2021	0.69		1.4		14.53		1.40		1.3		27.29	
$CC \times T$	-	ns	_	ns	-	ns	_	ns	-	ns	-	ns
$CC \times T \times D$	-	ns	-	ns	-	ns	-	ns	-	ns	-	ns
$CC \times T \times D \times Y$	-	ns	-	ns	-	ns	-	ns	-	ns	-	ns

TABLE 2 Soil organic carbon (% by mass), bulk density (g cm⁻³), and total C (Mg ha⁻¹) in Fresno County and Napa County.

^a ANOVA was used to compare data (p-value indicated). Letters within columns indicate significant mean separation according to Tukey's honestly significant difference (HSD) test (at p = 0.05), where *: p-value < 0.005; **: p-value < 0.001, and ***: p-value < 0.0001. ^b NT, no tillage; CT, conventional tillage; AG, annual grass; RV, residual vegetation; PG, perennial grass; SOC, soil organic carbon; ns, not significant; and –, not applicable.

during the first measurement in the Napa County vineyard. When readings were averaged to yield seasonal R_s , tillage displayed a stronger overall effect and increased R_s . There were no interactions observed between the type of cover crop and the tillage system at either site. Finally, under vine, R_s was higher in 2021 than that in 2020 at both sites.

higher in 2021 compared to 2020, thus the same trend was observed among grapevine NPP as well.

3.3 Grapevine and cover crop NPP

At the Fresno County vineyard, there was an effect of cover crop on pruning wood and leaf C contributions (Table 4), whereby the PG reduced C input through the production of annual growth (canes and leaves) compared to the AG and RV. However, this difference did not translate into differences in yield or C input from fruit. Ultimately, no differences in grapevine NPP were observed between different cover crops nor tillage system. Contrary to the Fresno site, at the Napa County (Oakville) vineyard, the annual growth (harvest, pruning wood, leaves) was affected by tillage system rather than cover crop. Grapevines under tillage resulted in higher C contributions from pruning wood and leaves. This contributed to differences in grapevine NPP; whereby tillage increased NPP, RV resulted in the highest NPP compared to AG and then PG. A significant interaction of the two factors was found, and year-to-year differences were also observed at the Napa County site whereby C contributions from harvest and pruning wood were

3.4 Estimates of net carbon balance at the grapevine and vineyard scale

When harvest mass and Rs under vine were subtracted from grapevine NPP to generate the grapevine net carbon balance (NCB), a year-to-year difference was observed at the Fresno County vineyard, as values in 2020 were greater than that of 2021 (Table 4). At the Napa County vineyard, there was an effect of cover crop, tillage, and an interaction of the two factors on grapevine NCB. The same pattern as grapevine NPP was observed whereby NCB values were greatest under RV followed by the AG and then PG. Tillage resulted in greater NCB. When SOC adjusted for spatial coverage of the interrow (+), $R_{s:\ inter-row}$ (-), and $NPP_{cover\ crop}$ (+) were added to the NCB to yield the NECB for the six different cover crop and tillage systems, no treatment effects were observed at the Fresno County site. At the Napa County site, however, some previous statistical trends from the grapevine NCB reversed when the remaining components of the vineyard ecosystem were added. Overall, PG and tillage reduced the NECB, but there was no interaction between the two factors (Figure 3). A greater NECB was also observed in 2020 compared to 2021.

TABLE 3 Average soil respiration (R_s) (t C ha⁻¹ year⁻¹) analyzed over the 2-year study across five (Fresno) and six (Napa) timepoints as measured in each cover crop and tillage treatment combination and under vine (bare soil).

Fresno County (Five points)

	Т	'illage systen	n (T)	Cover crop (CC)						
	NT	СТ	<i>p</i> -value ^a	AG	RV	PG	Under vine	<i>p</i> -value		
2020										
3-Feb	3.45	2.12	ns	2.15	3.43	2.78	1.03	ns		
25-Mar	4.27	5.13	ns	4.26	5.43	4.41	1.23	ns		
16-Apr	4.34	3.60	ns	2.85	3.99	5.07	2.04	ns		
17-Apr	3.80	4.63	ns	4.14	4.20	4.32	2.42	ns		
12-Jun	2.10	2.11	ns	2.12	2.15	2.04	2.46	ns		
Season-long Average R _s	3.59	3.52	ns	3.10	3.84	3.72	1.84	ns		
2021										
14-Feb	2.41	3.24	ns	3.42 a	2.20 b	2.86 b	1.92 c	*		
24-Mar	3.35	5.35	ns	4.02 b	4.02 b	5.01 a	1.38 c	*		
28-Apr	4.99	4.32	ns	5.10	4.12	4.74	1.85	ns		
29-Apr	4.29	7.52	ns	6.48	5.39	5.84	3.34	ns		
1-Jul	3.66	4.32	ns	4.79	3.76	3.42	6.30	ns		
Season-long Average R _s	3.74	4.95	ns	4.76	3.90	4.37	2.96	ns		
$CC \times T$	-	-	ns	-	-	-	-	ns		
Year (Y)	-	-	ns	-	-	-	-	ns		
CC × T × Y Napa County (Oakvi	- lle)	-	ns	-	-	-	-	ns		
	Т	'illage systen	n (T)		Cover crop (CC)					
	NT	CT	<i>p</i> -value ^a	AG	RV	PG	Under vine	<i>p</i> -value		
2020										
24 January	7.14 b	8.52 a	*	10.02 a	6.19 b	7.29 b	5.29 c	*		
13 April	6.70 b	9.61 a	*	8.30	7.02	9.14	8.13	ns		
21 April	6.78	6.90	ns	8.21	6.75	5.56	2.82	ns		
23 April	5.52	5.43	ns	5.09	5.15	6.19	2.05	ns		
22 May	0.91	1.60	ns	0.67	1.02	2.08	1.36	ns		
19 June	5.30	4.27	ns	4.93	4.56	4.86	2.64	ns		
Seasonal Average R _s	5.39 b	6.06 a	**	6.20	5.12	5.85	3.71	ns		
2021										
29 January	14.78 b	15.87 a	*	15.60 b	10.06 c	20.32 a	12.48 b	*		
14 April	7.14	5.42	ns	5.89	6.19	6.76	5.29	ns		
16 April	5.72	8.50	ns	8.74	7.14	5.45	2.40	ns		
14 May	6.78	6.90	ns	8.21	6.75	5.56	2.82	ns		
15 June	7.26	5.43	ns	5.09	7.76	6.19	2.05	ns		
9 July	0.91	1.60	ns	0.67	1.02	2.08	1.36	ns		
Seasonal Average $R_{\rm s}$	7.10 b	7.29 a	*	7.36	6.49	7.73	4.40	ns		
$CC \times T$	-	-	ns	-	-	-	-	ns		
Year (Y)	-	-	ns	-	-	-	-	ns		
$CC\times T\times Y$	-	-	ns	-	-	-	-	ns		

^a ANOVA was used to compare data (p-value indicated). Letters within columns indicate significant mean separation according to Tukey's honestly significant difference (HSD) test (at p = 0.05), where *: p-value < 0.05; **: p-value < 0.001, and ***: p-value < 0.0001. ^b NT, no tillage; CT, conventional tillage; AG, annual grass; RV, residual vegetation; PG, perennial grass; ns, not significant; and -, not applicable.



TABLE 4 Components of the vineyard's net primary production (NPP as t ha⁻¹ of dry matter), the whole vine's net carbon balance (NCB), and the vineyard's net ecosystem carbon balance (NECB) calculated over the 2-year study for six different cover crop and tillage systems.^{a,b}.

Treatment		Harvest	Pruning	Leaves	Permanent	1	R _s	Grapevine	Cover	R _s Inter-	SOC	NECB
Tillage (T)	Cover crop (CC)	(Mg C ha ⁻¹)	wood (Mg C ha ⁻¹)	(Mg C ha ⁻¹)	organs (Mg C ha ⁻¹)	NPP (Mg C ha ⁻¹ year ⁻¹)	Under vine (Mg C ha ⁻¹ year ⁻¹)	NCB (Mg C year ⁻¹)	Crop NPP (Mg C ha ⁻¹)	row (Mg C ha ⁻¹ year ⁻¹)	(Mg C ha ⁻ ¹)	(Mg C ha ⁻¹ year ⁻¹)
Fresno Co	unty (Five	points)										
2020												
NT	AG	1.39	1.76	3.02	11.5	17.7	1.84	15.35	1.80	2.81	17.71	23.36
	PG	1.41	1.30	2.97	11.5	17.2	1.84	14.83	1.70	4.02	19.25	22.96
	RV	1.32	1.73	3.53	11.5	18.1	1.84	15.82	2.30	3.94	17.64	23.5
CT	AG	1.25	1.75	3.22	11.5	17.7	1.84	15.54	3.93	3.39	19.81	25.31
	PG	1.48	1.56	3.27	11.5	17.8	1.84	15.39	1.83	3.42	18.97	23.74
	RV	1.65	2.17	2.93	11.5	18.3	1.84	15.66	1.17	3.74	17.64	22.89
2021												
NT	AG	1.24	1.61	2.98	11.5	17.3	2.96	14.56	2.00	3.71	17.71	22.24
	PG	1.63	1.25	2.92	11.5	17.3	2.96	14.15	2.15	3.69	19.32	22.69
	RV	1.80	1.53	3.48	11.5	18.3	2.96	14.99	1.83	3.81	17.71	22.54
CT	AG	1.11	1.75	3.18	11.5	17.6	2.96	14.91	4.08	5.81	19.88	23.63
	PG	1.21	1.50	3.09	11.5	17.3	2.96	14.58	2.07	5.06	18.97	22.25
	RV	1.34	2.01	2.88	11.5	17.8	2.96	14.88	1.33	3.98	17.36	21.94
CC		ns	*	*	-	ns	-	ns	**	ns	ns	ns
Т		ns	ns	ns		ns		ns	ns	ns	ns	ns
$\mathrm{CC} imes \mathrm{T}$		ns	ns	ns		ns		ns	ns	ns	ns	ns
Year		ns	ns	ns		ns		*	*	ns	ns	ns
Year × CC		ns	ns	ns		ns		ns	ns	ns	ns	ns

(Continued)

Tuestaset

SOC

NECD

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Treatm	Treatment		Pruning	Leaves	Permanent	Grapevine	R _s	Grapevine	Cover	R _s Inter-	SOC	NECB
Tillage (T)	Cover crop (CC)	ha ⁻¹) (wood (Mg C ha ⁻¹)	(Mg C ha ⁻¹)	organs (Mg C ha ⁻¹)	NPP (Mg C ha ⁻¹ year ⁻¹)	Under vine (Mg C ha ⁻¹ year ⁻¹)	NCB (Mg C year ⁻¹)	Crop NPP (Mg C ha ⁻¹)	row (Mg C ha ⁻¹ year ⁻¹)	(Mg C ha ⁻ 1)	(Mg C ha ⁻¹ year ⁻¹)
$\mathrm{Year} \times \mathrm{T}$		ns	ns	ns		ns		ns	ns	ns	ns	ns
Year \times CO	C×T	ns	ns	ns		ns		ns	ns	ns	ns	ns
Napa Cou	ınty (Oakvi	lle)										
2020												
NT	AG	0.15	0.37	0.57	6.7	7.8	3.71	5.71	2.92	5.56	29.25	18.49
	PG	0.17	0.21	0.49	6.7	7.6	3.71	5.47	0.96	5.89	28.28	16.68
	RV	0.14	0.43	0.44	6.7	7.7	3.71	5.63	2.59	4.73	28.76	18.41
CT	AG	0.16	0.51	0.55	6.7	7.9	3.71	5.83	2.96	6.85	29.25	18.00
	PG	0.16	0.32	0.64	6.7	7.8	3.71	5.72	1.19	5.82	27.30	16.61
	RV	0.16	0.77	0.74	6.7	8.4	3.71	6.29	3.20	5.50	27.30	18.29
2021												
NT	AG	0.44	0.56	0.61	7.0	8.6	4.40	5.88	3.20	7.25	28.76	17.74
	PG	0.33	0.39	0.53	7.0	8.3	4.40	5.63	1.62	7.70	27.79	16.05
	RV	0.26	0.62	0.47	7.0	8.3	4.40	5.80	3.60	6.68	30.23	18.83
CT	AG	0.37	0.68	0.58	7.0	8.6	4.40	5.97	3.32	7.48	27.30	17.08
	PG	0.38	0.48	0.67	7.0	8.5	4.40	5.86	1.47	7.75	26.81	15.72
	RV	0.46	0.96	0.77	7.0	9.2	4.40	6.45	3.46	5.64	26.81	18.27
CC		ns	ns	ns	-	*	-	**	**	ns	ns	***
Т		ns	**	*		***		***	ns	**	**	*
$\mathrm{CC}\times\mathrm{T}$		ns	ns	ns		*		*	ns	ns	ns	ns
Year		***	**	ns		***		ns	ns	ns	ns	*
Year × CO	C	ns	ns	ns		ns		ns	ns	ns	ns	ns
$\mathrm{Year} \times \mathrm{T}$		ns	ns	ns		ns		ns	ns	ns	ns	ns
Year × CO	C×T	ns	ns	ns		ns		ns	ns	ns	ns	ns

Casaria

TABLE 4 Continued

Harroot

Devening Lagrage

Da

^a ANOVA was used to compare data (p-value indicated). Letters within columns indicate significant mean separation according to Tukey's honestly significant difference (HSD) test (at p = 0.05), where *: p-value < 0.005; **: p-value < 0.001, and ***: p-value < 0.0001. ^b NT, no tillage; CT, conventional tillage; AG, annual grass; RV, residual vegetation; PG, perennial grass; NPP; net primary production; NCB, net carbon balance; R_s, respiration; SOC, soil organic carbon; NECB, net ecosystem carbon balance; ns, not significant; and –: not applicable.

4 Discussion

4.1 Soil texture plays a key role in the effects of cover crop and tillage on SOC

After 2 years of the adoption of treatments, there were no statistical differences between the type of cover crop on SOC at either experimental vineyard, which agreed with some previous studies when cover crops were implemented and the effects on SOC were monitored only for a short period of time (Belmonte et al., 2018; Novara et al., 2019). However, there were many other studies that have noticed a significant increase in SOM under cover cropping (Morlat and Jacquet, 2003; Steenwerth and Belina, 2008b), and the results might be dependent on the specific climatic, topography, and soil conditions in the experimental sites (Novara et al., 2019; Yu et al., 2019). As previously reported, greater SOC content was generally found in the upper portion of the soil (0–15 cm) compared to the deeper

portion (Steenwerth and Belina, 2008b; Mitchell et al., 2017; Yu et al., 2019). Previous literature has shown that this phenomenon was quite consistent among most soil textures of temperate regions (Krull et al., 2001; Tautges et al., 2019), which might be the reason when significant effects were observed, the effectiveness of cover crops on the soil properties might be higher in the upper layer soils than those of deeper layer soils (Mitchell et al., 2017; Tautges et al., 2019). On the other hand, SOC values at both sites were similar with other studies under similar soil texture and climates (Steenwerth et al., 2010; Wolff et al., 2018; Yu et al., 2019). However, the lack of strong effects of tillage on SOC at the Fresno County vineyard can be attributed to the significantly less SOM with more sandy soil texture in the Fresno County vineyard compared to the Napa County vineyard with finer soil texture, which could have possibly led to excessive aeration and permeability in the soil (Yu et al., 2019; Beltrán et al., 2021), hence a very low SOM baseline for the Fresno site. As for the Napa site with CT significantly deducing soil SOC,



affected by cover crop treatment in the same vineyard. Values represent means \pm SE. Different letters indicate significant differences (p \leq 0.05) between respective treatments according to two-way ANOVA followed by Tukey's honest significant difference (HSD) test. AG, annual grass; PG, perennial grass; RV, resident vegetation.

there has been recent work in the similar Mediterranean climates that is generally in agreement with these observations, reporting increased SOC accumulation under minimum and complete lack of tillage (López-Bellido et al., 2010; Wolff et al., 2018). However, it should be noted that broad conclusions regarding increased SOC under NT management are nuanced, as other studies have discussed the effects of reduced tillage on SOC content and whether they would be truly beneficial toward plant growth, soil biodiversity, or overall soil health, since SOC has to be decayed or immobilized to be usable by soil microbiomes and plants (Janzen, 2006); and whether the long-term effects in deeper soil layers and the whole soil profile would still accredit the capability of NT on C sequestration (Baker et al., 2007; Blanco-Canqui and Lal, 2008); and whether the conflicting results can be solved by standardizing research methodologies and technologies (Derpsch et al., 2014).

$4.2~\text{R}_{\text{s}}$ and cover crop biomass are main drivers of the NECB

At the Fresno County vineyard, while the soils under cover crop displayed differences in R_s in the early season R_s measurement events in 2021, ultimately no seasonal differences were observed. This may be due to the higher sand content in the soil at this experimental site (Bouma and Bryla, 2000) and more active cover crops during the spring measurements, causing higher R_s at the early season (Nilahyane et al., 2020). While tillage indeed increased R_s values due to oxidation of organic matter from exposed soil aggregates and generally higher soil temperature during the growing season. Overall, these effects were minimal in terms of determining the NECB from the vineyard agroecosystem at the Fresno site.

A similar trend was observed at the Napa County vineyard, where R_s interrow showed differences in the early season, despite no significant differences of cover crop type over seasonal average Rs values. CT increased the seasonal average Rs values and the annual production, NPP, and NCB. Previous research showed that the higher SOC can be directly linked soil texture, where finer soils showed greater SOC storing capacity than coarser soils (Bird et al., 2003; Zinn et al., 2007; Wiesmeier et al., 2014). On the contrary, there was evidence shown that climate might have a larger effect on SOC than soil texture (Wang et al., 2010). Nevertheless, these two factors might have contributed to the significant difference in Rs at the Napa site, where the soil texture was finer and the climate was relatively cooler. However, when the remaining components of the vineyard agroecosystem were added, the NECB was enhanced by NT rather than CT. This indicated that Rs and SOC have played a greater role in determining the NECB compared to the other factors, as losses of C through interrow R_s and SOC under tillage were large enough to negate the previous increase in NPP. Moreover, the loss of interaction between the two factors, cover crop and tillage, may further confirm that interrow Rs plays a significant role in NECB determination and thus the C storage potential of vineyards.

On the other hand, high variability in biomass led to a significant effect of cover crop type on the NECB at both experimental sites. This was likely due to the compensation of C input through biomass from the cover crops implemented in this study, as the AG and RV all showed greater biomasses compared to the low-stature PG. AG and RV have generated greater biomasses compared to low-profile *Poa bulbosa* as the PG as higher biomass-producing plants would generally have higher C-storing capacity in many previous studies (Williams et al., 2011; Agostini et al., 2015; Williams et al., 2020). And eventually, the differences in cover crop biomass capacity led to the observed changes in the NECB, where the greatest values were under RV, followed by AG and PG.

5 Conclusion

Our findings provide more evidence of that the vineyard agroecosystem can serve as a C sink for short-term implementation of cover crops with NT practice. Corroborating previous research under sandy soils, tillage and type of cover crop

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had little to no effect on the NECB. However, under the finertextured soils, CT reduced the NECB through a reduction in SOC and increase in R_s , or soil CO₂ efflux. The type of cover crop also impacted the NECB, as cover crops that produced greater biomass increased the NECB. Ultimately, vineyard site characteristics, including soil texture and climate, were key determinants of the effectiveness of C storage potential, as they can determine SOC and R_s of vineyards in Mediterranean vineyard agroecosystems in both Napa and Fresno. Overall, the implementation of NT and cover crop practices should be carefully considered with a thorough understanding of the specific site characteristics to fully maximize their effectiveness.

Data availability statement

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

Author contributions

SK designed the trial and acquired the funding. MZ, NT, RY, and LM executed the trial. MZ and NT curated the data. MZ wrote the first version of the manuscript. All authors read and approved the final submitted version of the paper.

Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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