Check for updates

OPEN ACCESS

EDITED BY Marc Landry, Université de Bordeaux, France

REVIEWED BY Isain Zapata, Rocky Vista University, United States Marilyn Urrutia-Pereira, Federal University of Pampa, Brazil

*CORRESPONDENCE Clara G. Zundel Clara.zundel@wayne.edu

RECEIVED 06 March 2025 ACCEPTED 28 April 2025 PUBLISHED 23 May 2025

CITATION

Zundel CG, Shampine MM, Jakubiec A, Lewis C, Brokamp C, Strawn JR, Jovanovic T, Ryan PH, Woodcock E, Yang Z, Petriello M and Marusak H (2025) Outdoor air pollution and psychiatric symptoms in adolescents: a study of peripheral inflammatory marker associations. *Front. Psychiatry* 16:1588964. doi: 10.3389/fpsyt.2025.1588964

COPYRIGHT

© 2025 Zundel, Shampine, Jakubiec, Lewis, Brokamp, Strawn, Jovanovic, Ryan, Woodcock, Yang, Petriello and Marusak. This is an openaccess article distributed under the terms of the Creative Commons Attribution License (CC BY). The use, distribution or reproduction in other forums is permitted, provided the original author(s) and the copyright owner(s) are credited and that the original publication in this journal is cited, in accordance with accepted academic practice. No use, distribution or reproduction is permitted which does not comply with these terms.

Outdoor air pollution and psychiatric symptoms in adolescents: a study of peripheral inflammatory marker associations

Clara G. Zundel^{1*}, MacKenna M. Shampine¹, Alexander Jakubiec¹, Christine Lewis¹, Cole Brokamp^{2,3}, Jeffrey R. Strawn^{4,2}, Tanja Jovanovic^{1,5}, Patrick H. Ryan^{2,3}, Eric Woodcock¹, Zhao Yang⁶, Michael Petriello^{6,7} and Hilary Marusak^{15,7}

¹Department of Psychiatry and Behavioral Neurosciences, Wayne State University School of Medicine, Detroit, MI, United States, ²Department of Pediatrics, University of Cincinnati College of Medicine, Cincinnati, OH, United States, ³Division of Biostatistics and Epidemiology, Cincinnati Children's Hospital Medical Center, Cincinnati, OH, United States, ⁴Anxiety Disorders Research Program, Department of Psychiatry and Behavioral Neuroscience, University of Cincinnati, Cincinnati, OH, United States, ⁵Merrill Palmer Skillman Institute for Child and Family Development, Wayne State University, Detroit, MI, United States, ⁶Institute of Environmental Health Sciences, Wayne State University, Detroit, MI, United States, ⁷Department of Pharmacology, Wayne State University School of Medicine, Detroit, MI, United States

Introduction: Fine particulate matter ($PM_{2.5}$) air pollution is associated with increased internalizing symptoms (e.g. depressive and anxiety symptoms), particularly during adolescence—a critical period for the emergence of anxiety disorders and vulnerability to neurotoxicants. Preclinical studies suggest that inflammation, including cytokines, reactive proteins, and lipid mediators, may explain the link between $PM_{2.5}$ and psychiatric risk. However, growing evidence suggests that these relationships may differ by sex, with females potentially more vulnerable to the effects of air pollution on psychiatric symptoms, though the underlying mechanisms remain unclear.

Methods: This study examined the relationships among recent (past-month) PM2.5 exposure, peripheral inflammatory markers, and anxiety and depressive symptoms in 78 adolescents (M \pm SD = 13.3 \pm 2.3 years, 48.7% female) from the Detroit, MI area.

Results: Higher PM_{2.5} concentrations were significantly associated with elevated levels of inflammatory lipid mediators: PGE2, 12(S)-HETE, 12(S)-HEPE, and 15(S)-HETE. A significant PM_{2.5}-by-sex interaction was observed for IL-6, with higher PM_{2.5} exposure associated with higher IL-6 concentrations in females but not males. Additionally, higher PM_{2.5} concentrations were significantly associated with greater total anxiety, generalized anxiety, and social anxiety symptoms, but only in females. Higher IL-8 concentrations were associated with greater depressive symptoms, and a significant TNF- α -by-sex interaction was observed for total and social anxiety symptoms, with higher TNF- α concentrations linked to greater symptoms in females but not males.

Discussion: These findings suggest that PM_{2.5} exposure is associated with inflammation and anxiety symptoms in adolescence, with notable sex differences. As a modifiable risk factor, reducing outdoor air pollution exposure may help mitigate psychiatric symptoms in youth.

KEYWORDS

outdoor, air pollution, particulate matter, adolescence, anxiety, inflammation

1 Introduction

Outdoor air pollution is a significant driver of global disease burden, linked to cardiovascular and respiratory diseases, cancers, and other health conditions (1). In 2021, air pollution ranked as the second leading risk factor for mortality worldwide, exceeded only by high blood pressure. This underscores its critical impact on global health and the urgent need for targeted interventions. Among air pollutants, $PM_{2.5}$ —particulate matter with an aerodynamic diameter of less than 2.5 microns—represents the largest contributor to the global disease burden (1). $PM_{2.5}$ is particularly concerning due to its small size, which allows it to bypass the body's natural barriers, including the blood-brain barrier, and embed more deeply in the lungs. This can trigger systemic inflammatory responses, which may indirectly affect the central nervous system, further amplifying its adverse health effects (2–4).

While the adverse effects of $PM_{2.5}$ on physical health are welldocumented, accumulating evidence suggests that this pollutant may also significantly impact mental health. Chronic $PM_{2.5}$ exposures are associated with worsening psychiatric disorders (e.g., anxiety and depression), as well as cognitive dysfunction, with exposure durations varying across studies (e.g., past-month, past-year) (3, 5). However, even acute elevations in air pollution (i.e., days prior) have been associated with increased emergency department visits for psychiatric symptoms in children and adults (6–8), underscoring the importance of investigating the psychological consequences of short-term increases in air pollution.

While the adverse mental health effects of PM_{2.5} have been documented across the lifespan, adolescence is a particularly vulnerable period, when the brain may be particularly sensitive to the adverse effects of PM_{2.5}. During adolescence, the brain undergoes significant pruning and reorganization, particularly in frontolimbic pathways that subserve emotional processing and executive functioning (9). Additionally, youth inhale more air relative to their body size than adults, and their natural barriers such as the blood-brain barrier and the nasal, gut, and lung epithelia —are still maturing. As a result, these immature barriers may be less effective at filtering and trapping air pollutants (10). Adolescence is also a critical period for the onset of psychiatric disorders, with mental health conditions, including anxiety and depressive disorders, typically emerging during this developmental stage (11, 12). As such, understanding the association between PM_{2.5} exposure and mental health during adolescence is crucial for identifying early intervention strategies and mitigating potential long-term effects.

Systemic inflammation is increasingly recognized as a key mechanism linking air pollutants (3), like PM2.5, to physical and mental health outcomes. Investigating inflammatory markers is crucial for understanding how air pollution impacts the developing brain. Exposure to PM_{2.5} can trigger inflammatory responses that disrupt neurobiological processes, further exacerbating psychiatric risk during adolescence. Previous research has demonstrated that recent exposure to air pollution (e.g., days to 1 month prior to sample), including PM_{2.5}, is associated with elevated concentrations of peripheral inflammatory markers (e.g., interleukin-6 (IL-6), tumor necrosis factor alpha (TNF- α)) in children and adolescents (13–15). These elevated inflammatory markers have been linked to various adverse health outcomes, including respiratory diseases, cardiovascular conditions, and metabolic disorders (13-15). However, fewer studies examine the inflammatory pathways that may mediate effects on mental health in youth. This gap is particularly relevant given the emerging evidence suggesting that inflammation plays a central role in the development of psychopathology, including anxiety and depression (9).

A remarkably underexplored area is the role of lipid mediators in the inflammatory response associated with air pollution exposure. Pro-inflammatory (e.g., prostaglandins [PGs]) and antiinflammatory (e.g., hydroxy-PUFAs [HETEs]) lipid mediators play a critical role in regulating inflammatory and immune responses (16). They provide unique insights into the generation and resolution phases of inflammation, which is crucial for understanding how acute and chronic exposure to air pollution may disrupt homeostasis and contribute to psychiatric risk. Both in vivo and in vitro studies have demonstrated that particulate matter exposure can induce the production of these lipid mediators, which may precipitate and/or potentiate systemic inflammatory responses (17-19). Thus, understanding the role of lipid mediators in the inflammatory response to air pollution is essential, especially during adolescence, when dynamic neuro-immune interactions influence brain maturation and behavioral regulation, potentially heightening vulnerability to environmental stressors and psychiatric risk (20).

The current exploratory study aimed to investigate the relationship between recent exposure to PM_{2.5}, peripheral

markers of inflammation, and anxiety and depressive symptoms in adolescents. Specifically, we focused on key peripheral inflammatory markers, including pro-inflammatory cytokines, c-reactive protein (CRP), and lipid mediators, to elucidate their role in the effects of PM_{2.5} on psychiatric risk. Additionally, given evidence of sex differences in inflammatory marker concentrations (21) and the relationship between air pollution exposure and psychiatric symptoms (3, 22), we examined whether associations between PM_{2.5}, inflammation, and psychiatric symptoms differed by sex.

2 Materials and methods

2.1 Participants

Participants (adolescents and one parent/guardian) were recruited from the Detroit, Michigan, USA, metropolitan area (approximately \leq 50 miles from the city) between 2021 and 2024. Recruitment efforts included community-based advertisements and flyers, partnerships with local schools, recreation centers, a local pediatrics practice (Wayne Pediatrics), and a children's health center (The Children's Center).

Eligibility criteria included children aged 10–17 years, righthandedness, and English speaking. Exclusion criteria included current or past pervasive development disorders or other severe psychopathology (e.g., schizophrenia, autism spectrum disorder), history of neurological disorder (e.g., epilepsy, traumatic brain injury), current or potential pregnancy, contraindications for phlebotomy, significant sensory impairments (e.g., hearing loss), alcohol detection with breathalyzer or positive urine drug screen, oral contraceptive use, and non-stimulant medication use. Study visits were scheduled between 10 AM and 4 PM to minimize circadian effects. Height and weight were collected to compute body mass index (BMI).

Parents/guardians provided written informed consent, and adolescents provided assent before any study procedures. This study was approved by the Wayne State University Institutional Review Board. A total of 123 adolescent-caregiver dyads enrolled in a parent study on adolescent neurodevelopment and anxiety risk. Of the 123, 99 (80.5%) provided blood samples. Of 99, 21 adolescents reported medication use (i.e., over-the-counter allergy medications, acetaminophen/ibuprofen) the day prior or the day of plasma sample collection and were excluded from analyses. Thus, the final sample for this study was 78 adolescents.

2.2 Air pollution estimates

Participants' current addresses were geocoded using our previously described DeGAUSS software and geocoder (23). Daily average $PM_{2.5}$ concentrations were estimated at the home address using a generalized random forest model via the Air Pollution Prediction Commons (appc 0.5.0) (24). This model trains weather and atmospheric data, wildfire smoke plumes, elevation, and satellite-based aerosol products on $PM_{2.5}$ concentrations from the EPA's Air

Quality System. Daily estimates were averaged over the month before the participant completed the study as an estimate of recent pastmonth exposure (24). Cross-validated model accuracy for monthly concentrations is highly accurate, with a median absolute error of $0.72 \ \mu g/m^3$ and 0.89 correlation between predicted and observed values. A histogram of past-month PM_{2.5} concentrations in the sample is provided in Supplementary Figure 1.

In addition to air pollution estimates, residential addresses were also used to determine the 2020 national rank of the Area Deprivation Index (ADI) (25), a composite measure of neighborhood socioeconomic disadvantage, to be included as a covariate in all analyses. The 2020 ADI metric was selected as the most recent estimate available at the start of data collection (August 2021) to ensure it reflected socioeconomic conditions closest to participants' enrollment without extending beyond the study period. The ADI has been used in prior research examining associations between PM2.5 and health outcomes in both youth and adult populations (26, 27). To categorize the level of neighborhood disadvantage, the ADI was dichotomized based on the >75th percentile threshold, distinguishing participants from neighborhoods in the highest and lowest quartiles of socioeconomic disadvantage.

2.3 Anxiety and depression symptoms

Anxiety symptoms were assessed using the child-report version of the 41-item Screen for Child Anxiety-Related Disorders (SCARED) (28). The SCARED evaluates symptoms over the past three months and provides a total score, along with subscores for specific types of anxiety: panic, generalized anxiety disorder (GAD), separation anxiety, social anxiety, and school avoidance. We focused on the total score and the GAD and social anxiety disorder subscores, as these have been linked to recent trafficrelated PM_{2.5} air pollution exposure in children and adolescents (29). Depression symptoms were assessed using the Child Depression Inventory-short form (CDI-SF) (30). Sum scores were calculated, with higher scores indicating greater anxiety and depressive symptoms. The CDI total scores were positively skewed and thus were log₁₀ transformed for analysis.

2.4 Peripheral inflammatory marker collection and analysis

Non-fasting blood samples were collected via venipuncture into ethylenediamine tetraacetic acid (EDTA) tubes (BD Vacutainer, GK3E EDTA). Samples were centrifuged at 4°C and 1,500 RPM within 5 min of collection and plasma was stored at -80 C until processing.

Following study completion, plasma samples were analyzed at the Wayne State University Lipidomics Core Facility using liquid chromatography-tandem mass spectrometry (LC-MS/MS). LC-MS/ MS (QTRAP7500with SelexIon system, Sciex) methods were employed for qualitative profiling of the lipidome utilizing Information Dependent Acquisition (IDA) strategies and quantitative analysis using Multiple Reaction Monitoring (MRM) methods, following previously published protocols (31). LipidView was used for lipid identification, and MultiQuant was used to process the chromatographic data. For the current study, we focused on inflammatory lipid mediators derived from cyclooxygenase (COX) and lipoxygenase (LOX) pathways, as these are associated with short-term air pollution exposure in adults (18). These included PGE2, PGE3, 12(S)-HETE, 15(S)-HETE, and 12-HEPE.

For the pro-inflammatory cytokines (IL-6, IL-8, and TNF- α) and CRP analysis, plasma levels were quantified using the U-PLEX Proinflammatory Combo 4 (human) and Human CRP Assay according to the manufacturer's instructions. The assays were run using a MESO QuickPlex SQ 120 instrument, and data analysis was conducted using the Discover Workbench Desktop Analysis Software (version 4.0; Meso Scale Diagnostics).

Plasma lipid mediator and pro-inflammatory cytokine concentrations were positively skewed and were log₁₀ transformed to normalize the distribution. Descriptive statistics (median, IQR, range) for the raw values of the inflammatory markers are provided in Supplementary Table S1.

2.5 Selection of covariates and statistical analysis

Statistical analysis and data visualization were performed using R (version 4.4.2) with R Studio (version 2024.09.1). Descriptive statistics were used to characterize the study sample demographics, psychiatric symptoms, and inflammatory markers. Age, biological sex, BMI, the 2020 national rank of Area Deprivation Index (ADI) with a cutoff at the 75th percentile, and parental smoking in the home were selected as potential covariates based on prior research indicating associations with either PM2.5, inflammatory markers, or psychiatric symptoms (21, 32-34). While PM_{2.5} is only significantly correlated with sex and ADI in the sample, we included additional covariates to increase the precision of our estimates, given their known associations with variability in psychiatric symptoms (age sex), and inflammatory markers (BMI, parental smoking). These covariates were included in analyses where applicable. Spearman correlations were calculated to assess the relationships between PM_{2.5} concentrations, inflammatory markers, psychiatric symptoms, and covariates. The full correlation matrix for these variables can be found in Supplementary Tables S2-S4. Since puberty may be associated with changes in inflammatory markers (21), supplemental analyses were also performed using pubertal status instead of age as a covariate, and these analyses are described in the Supplementary Material.

2.6 Interaction terms and model specification

Linear regression analyses were conducted to assess the associations between past-month $PM_{2.5}$ concentrations and

inflammatory markers and between $PM_{2.5}$ and psychiatric symptom scores.

Sex differences were considered in all analyses, given the welldocumented evidence that both inflammatory markers (21) and psychiatric symptoms vary by sex (35, 36). However, it remains unclear whether $PM_{2.5}$'s effects on inflammation or psychiatric symptoms differ by sex, warranting the inclusion of interaction terms to explore these potential moderating effects.

For inflammatory markers, the main effect of sex and the $PM_{2.5}$ *sex interaction term were included to assess whether $PM_{2.5}$ -related inflammation differed by sex. Similarly, in the psychiatric symptom analyses, an interaction term between $PM_{2.5}$ exposure and sex was included based on prior literature suggesting that air pollution effects on mental health outcomes vary by sex (3, 22). Additionally, in this sample, females had both higher $PM_{2.5}$ exposure and higher levels of anxiety symptoms compared to males. The main effect of sex was omitted from the psychiatric symptom models to prevent overfitting and to isolate the effect of $PM_{2.5}$ within each sex group.

Lastly, the analysis of inflammatory markers in relation to psychiatric symptoms also included both main effects and an interaction term to assess whether inflammation-psychiatric symptom associations differed by sex, given the lack of clarity in existing research.

3 Results

3.1 Demographics

The sample was 48.7% female, with the mean age of 13.3 years (range: 10-17). The sample was racially and ethnically diverse (36% White Non-Hispanic, 33% Black Non-Hispanic, 15% Hispanic, 7% More Than One Race, 1% Other), and 45% and 55% of youth had clinically significant anxiety and depressive symptoms, respectively. The mean past-month PM_{2.5} concentration was 9.11 \pm 2.68 µg/m³, ranging from 5.78 to 16.83 µg/m³. Sex differences were observed in several demographics: females had higher mean past-month PM_{2.5} concentrations (9.79 + 2.68 µg/m³) than males. Additionally, a higher percentage of females (68%) had clinically significant anxiety symptoms compared to males (23%). Females also had a higher mean BMI than males (25.11 + 5.17 kg/m²). Further demographic details are provided in Table 1.

3.2 PM_{2.5} and inflammatory lipid mediators, cytokines, and CRP

Past-month PM_{2.5} concentrations were significantly associated with concentrations of PGE2, 12(S)-HETE, 12(S)-HEPE, and 15(S)-HETE. Specifically, for each 1 μ g/m³ increase in PM_{2.5}, PGE2 levels increased by 20% (β = 1.20). Similar significant increases in concentrations were observed for 12(S)-HETE (29%), 12(S)-HEPE (23%), and 15(S)-HETE (10%). These results are visually displayed in Figures 1A–D). No significant associations were observed

TABLE 1 Participant characteristics (N=78).

Demographic	Full Sample (N=78)	Males (N=40)	Females (N=38)	Males vs. Females comparisons
	N (%) or Mean <u>+</u> SD			P-value
Age	13.28 ± 2.28	13.10 + 2.05	13.47 + 2.51	p = 0.473
Body Mass Index (BMI)	23.51 ± 4.94	22.02 + 4.28	25.11 + 5.17	p = 0.005
Race/Ethnicity				p = 0.203
White, non-Hispanic	28 (35.9)	16 (40.0)	12 (31.6)	
Black, non-Hispanic	26 (33.3)	10 (25.0)	16 (42.1)	
Hispanic	12 (15.4)	5 (12.5)	7 (18.4)	
More than one race	6 (7.7)	5 (12.5)	1 (2.6)	
Other	1 (1.3)	0 (0)	1 (2.6)	
Household Income				p = 0.217
Low (< \$50,000)	33 (42.3)	14 (35.0)	19 (50.0)	
Middle (\$50,000 - \$100,000)	17 (21.8)	8 (20.0)	9 (23.7)	
High (> \$100,000)	28 (35.9)	18 (45.0)	10 (26.3)	
Pubertal Status				p = 0.028
Pre-puberty	11 (14.1)	9 (23.7)	2 (5.6)	
Post-puberty	63 (80.8)	29 (76.3)	34 (94.4)	
Smokers in Home	10 (12.8)	4 (10.0)	6 (15.8)	p = 0.512
Met or Exceeded Clinical Cutoff for Anxiety	35 (44.9)	9 (22.5)	26 (68.4)	p < 0.001
Met or Exceeded Clinical Cutoff for Depression	43 (55.1)	19 (47.5)	24 (63.2)	p = 0.253
Total SCARED Score	25.26 ± 15.37	17.72 + 10.21	33.00 + 16.03	p < 0.001
Total CDI Score	3.91 ± 3.81	3.13 + 3.47	4.71 + 4.01	p = 0.068
Past-month PM _{2.5} concentration (ug/m ³)	9.11 ± 2.68	8.47 + 2.55	9.79 + 2.68	p = 0.029
2020 National Rank of Area Deprivation Index (ADI) 75th Percentile Cutoff	35 (44.9)	12 (31.6)	23 (62.2)	p = 0.008

Anxiety symptoms were measured using the cutoff score (\geq 25) on the Screen for child Anxiety-Related Disorders (SCARED). Depression symptoms were measured using the cutoff score (\geq 3) on the Children's Depression Inventory (CDI) Short Form. Smoking in Home was measured via parental self-report. Differences in demographic characteristics between males and females were tested using independent sample t-tests for continuous variables (e.g., age) and chi-squared tests for categorical variables (e.g., parental smoking in home). Pubertal status was measured using child self-report of Tanner Stages.

Bolded values indicate p < 0.05.

between past-month PM_{2.5} and CRP, IL-8, or TNF- α (p's > 0.05, see Supplementary Figure 2). Past-month PM_{2.5} concentrations were significantly associated with IL-6 concentrations, with a differential effect by sex. For each 1 µg/m³ increase in PM_{2.5}, IL-6 levels decreased by 11% in males and increased by 4% in females (see Figure 1E). There was a main effect of sex on PGE3 levels, with females exhibiting higher PGE3 concentrations than males (β = 1.55), but no significant association was observed between pastmonth PM_{2.5} and PGE3 (p = 0.65). No other significant interactions between PM_{2.5} and sex were observed for the other inflammatory markers. The model results for these analyses can be found in Supplementary Tables S5, S6.

3.3 PM_{2.5} and psychiatric symptoms

Past-month PM_{2.5} concentrations were significantly associated with mental health symptoms, with differential effects observed by sex. Significant interaction between PM_{2.5} and sex were found for the SCARED total score, GAD subscore, and Social AD subscore. These results are visually displayed in Figures 2A–C. Specifically, for each 1 μ g/m³ increase in PM_{2.5}, SCARED total scores increased by 1.36 in females, with no significant effect observed for males. Similarly, the GAD subscore increased by 0.48 in females per 1 μ g/m³ increase in PM_{2.5}, with no significant effect in males, and the Social AD subscore increased by 0.25 in females with no significant



with shaded areas representing the 95% confidence intervals. The y-axis is presented on a logarithmic scale to facilitate visualization across multiple orders of magnitude.

effect in males. No significant associations were observed between past-month $PM_{2.5}$ and the depression total score (p > 0.05). The model results for these analyses can be found in Supplementary Table S7.

3.4 Inflammatory markers and psychiatric symptoms

A significant main effect of IL-8 was observed for the CDI total score, with each 10-fold increase in IL-8 associated with a 6.03-fold

increase in the CDI total score. Significant interactions between TNF- α levels and sex were found for the SCARED total score and SCARED Social AD subscore. For the SCARED total score, higher TNF- α levels were associated with an increase for females (β = 22.92), with no significant effect observed in males. Similarly, for the SCARED Social AD subscore, higher TNF- α levels were linked to an increase for females (β = 6.09), with no significant effect of sex was observed across all inflammatory models, with females generally exhibiting higher anxiety and depression symptoms than males. No significant associations were observed for PGE2, PGE3, 12(S)-HETE, 12(S)

HEPE, 15(S)HETE, IL-6, or CRP with psychiatric symptoms. These results are presented in, Figure 3 and Supplementary Tables S8–S16.

4 Discussion

To our knowledge, this is the first study to investigate the relationships between recent $PM_{2.5}$ exposure, peripheral inflammatory markers, and psychiatric symptoms in adolescents. We found that adolescents with higher past-month $PM_{2.5}$ concentrations had significantly higher inflammatory markers, specifically PGE2 (prostaglandin E2), 12(S)-HETE (12-hydroxy-eicosatetraenoic acid), 12(S)-HEPE (12-hydroxy-eicosapentaenoic acid), and 15(S)-HETE (15-hydroxy-eicosatetraenoic acid). Additionally, we observed a significant main effect of sex on PGE3, with females exhibiting higher concentrations than males; however, since PGE3 was not associated with $PM_{2.5}$ exposure or psychiatric symptoms, its role in linking air pollution to psychiatric symptoms appears limited. Of note, the past-month $PM_{2.5}$ concentrations observed in this study ranged from 5.78 to 16.83 µg/m³. While the EPA's standards for $PM_{2.5}$ are set

at an annual average of 9 μ g/m³ (EPA, 2024), two-thirds of our sample had exposures below this threshold, indicating that most participants experienced levels considered to be lower risk. Despite this, we observed significant associations with inflammatory markers, suggesting that even lower levels of PM_{2.5} exposure may contribute to adverse health effects in adolescents.

Our findings reveal that $PM_{2.5}$ exposure is associated with alterations in lipid mediators involved in inflammatory responses, including PGE2, 12(S)-HETE, 12(S)-HEPE and 15(S)-HETE. PGE2 is derived from the COX (cyclooxygenase) pathway and plays a complex role in inflammation. Further, PGE2 acts as a key initiator of the early inflammatory response and also contributes to resolution in later phases (37). In contrast, 12(S)-HETE, 12(S)-HEPE, and 15(S)-HETE originate from the 5-LOX (5-lipoxygenase) pathway (18), which is involved in resolving acute inflammation. The LOX-derived lipid mediators are widely expressed in the cardiovascular and central nervous systems, with localization within the primary olfactory cortex and hippocampus (38, 39).

Similar - lipid mediator profiles have been observed in prior studies of air pollution. Anti-inflammatory lipid mediators,



(GAD) Subscore, (C): SCARED Social Anxiety Disorder (AD) Subscore. In each panel, the regression lines are fitted to the data for each sex group (indicated by color). The shaded area represents the 95% confidence intervals of the regression estimates.

including 12(S)-HETE, 12(S)-HEPE, and 15(S)-HETE have been shown to increase in lower animals and in adults following recent/ acute (e.g., days to months prior) air pollution exposure (18, 19, 40). Conversely, significant decreases of 12(S)-HETE, 12(S)-HEPE, and 15(S)-HETE were reported when air pollution decreased during the 2018 Beijing, China Olympics (41, 42). Interestingly, a recent study in adults observed a spike in urinary PGE2 concentrations 8 hours after exposure to a wildfire event ceased, which is characterized by drastic increases in $PM_{2.5}$ concentrations due to smoke (43). However, the exact functional role of this increase – whether proinflammatory or contributing to inflammation resolution – remains unclear (43), highlighting the complex and context-dependent role of PGE2 in the inflammatory process, particularly in the context of air pollution exposure.

To our knowledge, this is the first study to report elevations of specific lipid mediators associated with air pollution in adolescents, highlighting the need to understand the developmental impacts of air pollution on the immune and inflammatory systems. Such a disruption in the equilibrium between anti- and pro-inflammatory mediators (referred to as "unalamation") may contribute to chronic low-grade inflammation and disease risk (44). Notably, despite these associations, we did not observe significant associations between $PM_{2.5}$ exposure and canonical pro-inflammatory

markers, such as CRP and TNF- α . This contrasts with three prior studies linking PM_{2.5} exposure to increased levels of these markers in children and adolescents (13, 14). However, two of these studies were conducted outside the US, potentially involving different PM_{2.5} compositions, and all three focused on narrower age ranges —specifically, two studies included younger children (8–11 years old) (13–15), and the third only included 17-year-olds (15). In contrast, our study specifically examined a broader adolescent age range (10–17 years), a developmental stage marked by significant biological and physiological changes. Differences in exposure windows and biological measures (e.g., plasma vs. exhaled breath) may also account for these discrepancies.

However, we did find a significant $PM_{2.5}$ -by-sex interaction on IL-6 concentrations. In females, increased $PM_{2.5}$ exposure was associated with higher IL-6 concentrations, whereas in males, the relationship was reversed, with lower IL-6 concentrations. These divergent patterns suggest that inflammatory responses to $PM_{2.5}$ may differ between sexes, potentially reflecting variation in immunoregulatory mechanisms and hormonal influences that may shape cytokine production (20, 21). While the clinical implications of higher versus lower IL-6 concentrations remains to be fully understood, these findings underscore the complexity of the inflammatory response to environmental pollutants.



Relationship between inflammatory markers and psychiatric symptoms. (A) IL-8 and Child Depression Inventory (CDI) Total Score, (B) TNF- α and SCARED Total Score, (C) TNF- α and SCARED Social Anxiety Disorder(AD) Subscore. In panels (B, C), the regression lines are fitted to the data for each sex group (indicated by color). The shaded area represents the 95% confidence intervals of the regression estimates. The x-axis is presented on a logarithmic scale to facilitate visualization across multiple orders of magnitude.

We also found that adolescents with higher past-month $PM_{2.5}$ exposure had significantly more total, generalized anxiety, and social anxiety symptoms, highlighting a potential link between air pollution exposure and mental health in adolescents - a population uniquely vulnerable to environmental stressors. Importantly, our $PM_{2.5}$ by sex interaction analyses revealed that these associations were driven primarily by females. In our sample, females not only exhibited significantly higher anxiety symptoms in general but also had higher past-month $PM_{2.5}$ concentrations. $PM_{2.5}$ exposure was assessed via residential address, indicating that these exposure differences are not due to personal characteristics per se but rather reflect the fact that, in our sample, females tended to reside in more deprived neighborhoods that likely have higher outdoor $PM_{2.5}$ concentrations.

These findings raise an important question: Could reducing exposure to fine particulate matter decrease the burden of anxiety disorders in youth? Our findings are also consistent with our earlier longitudinal study of unmedicated adolescents with GAD in which we observed that daily fluctuations in ambient $PM_{2.5}$ concentrations predicted changes in anxiety symptom severity two to four days after exposure (45). Taken together, these data suggest that air pollution not only correlates with anxiety symptoms but may also actively contribute to their persistence and exacerbation over time. This finding highlights the need for greater awareness of environmental risk factors in anxiety disorders and underscores the potential clinical value of integrating air quality considerations into prevention and intervention strategies.

Furthermore, significant interactions between TNF- α concentrations and sex were found for both total anxiety and social anxiety symptoms. Specifically, higher TNF- α was associated with higher anxiety symptoms in females only. These findings are consistent with recent work in adults, where elevated TNF- α concentrations were not only significantly higher in those with GAD but also positively correlated with anxiety symptoms (46). Moreover, a recent preclinical study has demonstrated that sustained TNF signaling is associated with the development of anxiety-like behaviors (47). Together, these studies support the notion that TNF- α may play a critical role in the pathophysiology of anxiety, underscoring the importance of further investigating inflammatory pathways as potential targets for intervention.

Beyond its associations with anxiety, higher IL-8 concentrations were also significantly linked to increased depressive symptoms. This finding aligns with a recent study in a community-based sample of adolescents aged 13–19, which similarly reported an association between elevated IL-8 levels and greater depressive symptoms (48). Taken together, these findings suggest that specific inflammatory pathways may differentially contribute to distinct psychiatric outcomes in adolescence.

Our findings have significant public health implications. PM_{2.5} may represent a novel and modifiable exposure and an actionable and measurable target for interventions that could reduce psychiatric morbidity on a population level. This is particularly important given that anxiety symptoms and anxiety disorders in this population are associated with significant morbidity including academic underachievement, increased risk of mood disorders,

suicidality and impairment in peer and family relationships. As such, by addressing environmental risk factors such as air pollution (49), we could potentially mitigate the prevalence and severity of anxiety disorders in children and adolescents, shifting the focus from reactive treatment to proactive prevention.

While this study provides valuable initial insights into the relationship between air pollution exposure and peripheral markers of inflammation-potentially relevant to the risk of psychopathology-several limitations warrant discussion. One key limitation is that inflammatory markers were assessed at a single timepoint, which prevents the evaluation of changes in these markers over time. Next, PM2.5 exposure was averaged over the past month, which may obscure acute fluctuations due to its high temporal variability (50). Future studies should assess inflammatory markers at multiple time points and consider exposure windows that better capture short-term fluctuations. Additionally, children's inflammatory systems may become more sensitive to challenges or stresses following PM_{2.5} exposure (51). To better capture these dynamic responses, future research should incorporate stressinducing tasks to examine changes in real-time inflammatory responses in those exposed to higher concentrations of air pollution. Further, participants were primarily drawn from the Detroit, MI area, which limits the generalizability of these findings. Next, our study focuses on a specific period, late childhood through adolescence. Importantly, while anxiety symptoms may begin earlier, this is the period of maximal increases in anxiety symptoms and the emergence of anxiety disorders. Thus, it remains possible that the effects of PM25 on symptoms may emerge later in development or interact with other risk factors over time, which could contribute to variability in findings across studies. Further, self-report measures have inherent limitations and may differ from symptoms assessed by clinicians. Finally, whether these symptoms represent psychiatric disorders (e.g., GAD) cannot be determined in this sample as we did not include assessment by a clinician.

In conclusion, this exploratory study represents a critical first step in understanding the complex interplay between air pollution exposure, systemic inflammation, and mental health in adolescents. To our knowledge, this is the first study to examine these relationships during adolescence, a sensitive period for neurodevelopment and heightened psychiatric risk. Our findings suggest that youth with higher recent $PM_{2.5}$ exposure have elevated concentrations of antiinflammatory lipid mediators and anxiety symptoms, highlighting the need for further research into the pathways linking environmental exposures to mental health outcomes.

5 Conflict of interest

Dr. Strawn has received research support from the National Institutes of Health NIMH/NCATS/NIEHS/NICHD, Patient-Centered Outcomes Research Institute PCORI, MindMed and material support from Myriad Genetics. He receives royalties from Springer and Cambridge University Press, honoraria from the Neuroscience Education Institute, and is an author for UpToDate. He has consulted to MindMed, Abbvie Cerevel, Otsuka, and Genomind. All other co-authors have no conflicts of interest to disclose.

Data availability statement

The datasets presented in this article are not readily available because they are not publicly shared. However, data may be made available upon reasonable request to the corresponding and senior authors for collaboration or further analysis. Requests to access the datasets should be directed to Clara Zundel, clara.zundel@ wayne.edu and Hilary Marusak, hmarusak@med.wayne.edu.

Ethics statement

The studies involving human participants were approved by Wayne State University Institutional Review Board and conducted in accordance with local legislation and institutional requirements. Written informed consent was obtained from participants' legal guardians, and adolescents provided both oral and written assent.

Author contributions

CZ: Conceptualization, Data curation, Formal Analysis, Funding acquisition, Investigation, Methodology, Project administration, Visualization, Writing – original draft, Writing – review & editing. MS: Investigation, Writing – review & editing. AJ: Investigation, Writing – review & editing. CL: Writing – review & editing. CB: Methodology, Writing – review & editing. JS: Writing – review & editing. TJ: Writing – review & editing. PR: Writing – review & editing. EW: Writing – review & editing. ZY: Writing – review & editing. MP: Writing – review & editing. HM: Conceptualization, Funding acquisition, Methodology, Supervision, Writing – review & editing.

Funding

The author(s) declare that financial support was received for the research and/or publication of this article. This study was supported by National Institutes of Health Award K01MH119241 to HM and a Wayne State University Department of Psychiatry and Behavioral Neurosciences New Investigator Award to CZ. CZ was partially supported by F32 MH133274. HM was partially supported by R01MH132830 and R21HD105882. EW was partially supported by R00 DA048125. MP and ZY are supported in part by R01ES035692-01A1, P42ES030991, 1R01ES034407, and P30ES036084. This study was supported in part by grants from National Institutes of Health S10RR027926 and S10OD032292 to the Lipidomics Core Facility of Wayne State University, as well as a CURES Pilot Project Grant (from P30ES036084).

Acknowledgments

The authors thank the following individuals and organizations for their assistance with study design, recruitment, data collection, and/or analysis: Samantha Ely, Julia, Evanski, Leah Gowatch, Shelley Paulisin, Mariya Matsko, Jovan Jande, Emily Crisan, Shravya Chanamolu, Sneha Bhargava, Iveta Kopil, Shreya Desai, Amanpreet Bhogal, Dr. Krishna Rao Maddipati, Dr. Laura Benjamins, Dr. Sharon Marshall, Dr. Christopher Youngman, Wayne Pediatrics, The Children's Center, and The Wayne State University Lipidomics Core. Finally, we appreciate copy-editing assistance from ChatGPT.

Conflict of interest

Authors JS received research support from the National Institutes of Health NIMH/NCATS/NIEHS/NICHD, Patient-Centered Outcomes Research Institute PCORI, MindMed and material support from Myriad Genetics. He receives royalties from Springer and Cambridge University Press, honoraria from the Neuroscience Education Institute, and is an author for UpToDate. He has consulted to MindMed, Abbvie Cerevel, Otsuka, and Genomind.

The remaining authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

The author(s) declared that they were an editorial board member of Frontiers, at the time of submission. This had no impact on the peer review process and the final decision.

Generative AI statement

The author(s) declare that Generative AI was used in the creation of this manuscript. ChatGPT was utilized to assistant with copy-editing of the final manuscript.

Publisher's note

All claims expressed in this article are solely those of the authors and do not necessarily represent those of their affiliated organizations, or those of the publisher, the editors and the reviewers. Any product that may be evaluated in this article, or claim that may be made by its manufacturer, is not guaranteed or endorsed by the publisher.

Supplementary material

The Supplementary Material for this article can be found online at: https://www.frontiersin.org/articles/10.3389/fpsyt.2025. 1588964/full#supplementary-material.

References

1. Health Effects Institute. *State of Global Air 2024. Special Report.* Boston, MA: Health Effects Institute (2024). Available online at: https://www.stateofglobalair.org/resources/report/state-global-air-report-2024.

2. Costa LG, Cole TB, Dao K, Chang YC, Coburn J, Garrick JM. Effects of air pollution on the nervous system and its possible role in neurodevelopmental and neurodegenerative disorders. *Pharmacol Ther.* (2020) 210:107523. doi: 10.1016/j.pharmthera.2020.107523

3. Zundel CG, Ryan P, Brokamp C, Heeter A, Huang Y, Strawn JR, et al. Air pollution, depressive and anxiety disorders, and brain effects: A systematic review. *Neurotoxicology*. (2022) 93:272–300. doi: 10.1016/j.neuro.2022.10.011

4. Genc S, Zadeoglulari Z, Fuss SH, Genc K. The adverse effects of air pollution on the nervous system. J Toxicol. (2012) 2012:782462. doi: 10.1155/2012/782462

5. Chandra M, Rai CB, Kumari N, Sandhu VK, Chandra K, Krishna M, et al. Air pollution and cognitive impairment across the life course in humans: A systematic review with specific focus on income level of study area. *Int J Environ Res Public Health.* (2022) 19. doi: 10.3390/ijerph19031405

6. Brokamp C, Strawn JR, Beck AF, Ryan P. Pediatric psychiatric emergency department utilization and fine particulate matter: A case-crossover study. *Environ Health Perspect.* (2019) 127:97006. doi: 10.1289/EHP4815

7. Kim SH, Shin SD, Song KJ, Ro YS, Kong SY, Kim J, et al. Association between ambient PM(2.5) and emergency department visits for psychiatric emergency diseases. *Am J Emerg Med.* (2019) 37:1649–56. doi: 10.1016/j.ajem.2018.11.034

8. Muhsin HA, Steingrimsson S, Oudin A, Astrom DO, Carlsen HK. Air pollution and increased number of psychiatric emergency room visits: A case-crossover study for identifying susceptible groups. *Environ Res.* (2022) 204:112001. doi: 10.1016/j.envres.2021.112001

9. Arain M, Haque M, Johal L, Mathur P, Nel W, Rais A, et al. Maturation of the adolescent brain. *Neuropsychiatr Dis Treat*. (2013) 9:449–61. doi: 10.2147/NDT.S39776

10. Brockmeyer S, D'Angiulli A. How air pollution alters brain development: the role of neuroinflammation. *Transl Neurosci*. (2016) 7:24–30. doi: 10.1515/tnsci-2016-0005

11. Merikangas KR, He JP, Burstein M, Swanson SA, Avenevoli S, Cui L, et al. Lifetime prevalence of mental disorders in U.S. adolescents: results from the National Comorbidity Survey Replication–Adolescent Supplement (NCS-A). J Am Acad Child Adolesc Psychiatry. (2010) 49:980–9. doi: 10.1016/j.jaac.2010.05.017

12. Kessler RC, Berglund P, Demler O, Jin R, Merikangas KR, Walters EE. Lifetime prevalence and age-of-onset distributions of DSM-IV disorders in the national comorbidity survey replication. *Arch Gen Psychiatry*. (2005) 62:593-603. doi: 10.1001/archpsyc.62.6.593

13. Parenteau AM, Alen NV, La J, Luck AT, Teichrow DJ, Daang EM, et al. Associations of air pollution with peripheral inflammation and cardiac autonomic physiology in children. *New Dir Child Adolesc Dev.* (2022) 2022:125–54. doi: 10.1002/cad.v2022.181-182

14. Calderon-Garciduenas I, Villarreal-Calderon R, Valencia-Salazar G, Henriquez-Roldan C, Gutierrez-Castrellon P, Torres-Jardon R, et al. Systemic inflammation, endothelial dysfunction, and activation in clinically healthy children exposed to air pollutants. *Inhal Toxicol.* (2008) 20:499–506. doi: 10.1080/08958370701864797

15. Sabeti Z, Ansarin A, Ansarin K, Zafari V, Seyedrezazadeh E, Shakerkhatibi M, et al. Sex-specific association of exposure to air pollutants and Nrf2 gene expression and inflammatory biomarkers in exhaled breath of healthy adolescents. *Environ Pollut.* (2023) 326:121463. doi: 10.1016/j.envpol.2023.121463

16. Dennis EA, Norris PC. Eicosanoid storm in infection and inflammation. Nat Rev Immunol. (2015) 15:511-23. doi: 10.1038/nri3859

17. Beck-Speier I, Karg E, Behrendt H, Stoeger T, Alessandrini F. Ultrafine particles affect the balance of endogenous pro- and anti-inflammatory lipid mediators in the lung: *in-vitro* and *in-vivo* studies. *Particle Fibre Toxicology*. (2012) 9. doi: 10.1186/1743-8977-9-27

18. Wang T, Han Y, Li H, Wang Y, Xue T, Chen X, et al. Changes in bioactive lipid mediators in response to short-term exposure to ambient air particulate matter: A targeted lipidomic analysis of oxylipin signaling pathways. *Environ Int.* (2021) 147:106314. doi: 10.1016/j.envint.2020.106314

 Li R, Navab K, Hough G, Daher N, Zhang M, Mittelstein D, et al. Effect of exposure to atmospheric ultrafine particles on production of free fatty acids and lipid metabolites in the mouse small intestine. *Environ Health Perspect.* (2015) 123:34–41. doi: 10.1289/ehp.1307036

20. Brenhouse HC, Schwarz JM. Immunoadolescence: Neuroimmune development and adolescent behavior. *Neurosci Biobehav Rev.* (2016) 70:288–99. doi: 10.1016/ j.neubiorev.2016.05.035

21. Stumper A, Moriarity DP, Coe CL, Ellman LM, Abramson LY, Alloy LB. Pubertal status and age are differentially associated with inflammatory biomarkers in female and male adolescents. J Youth Adolesc. (2020) 49:1379–92. doi: 10.1007/s10964-019-01101-3

22. Szyszkowicz M, Zemek R, Colman I, Gardner W, Kousha T, Smith-Doiron M. Air pollution and emergency department visits for mental disorders among youth. *Int J Environ Res Public Health.* (2020) 17. doi: 10.3390/ijerph17124190

23. Brokamp C, Wolfe C, Lingren T, Harley J, Ryan P. Decentralized and reproducible geocoding and characterization of community and environmental exposures for multisite studies. *J Am Med Inform Assoc.* (2018) 25:309-14. doi: 10.1093/jamia/ocx128

24. Brokamp C, Manning E, Duan Q. *appc: Air Pollution Predictor Commons R package version 0.5.0.* (2024). Available online at: https://geomarker.io/appc/, https://github.com/geomarker-io/appc.

25. Kind AJH, Buckingham WR. Making neighborhood-disadvantage metrics accessible - the neighborhood atlas. *N Engl J Med.* (2018) 378:2456–8. doi: 10.1056/ NEJMp1802313

26. Marshall AT, Adise S, Cardenas-Iniguez C, Hippolyte OK, Parchment CA, Villalobos TI, et al. Family- and neighborhood-level environmental associations with physical health conditions in 9- and 10-year-olds. *Health Psychol.* (2023) 42:878–88. doi: 10.1037/hea0001254

27. Bowe B, Xie Y, Yan Y, Al-Aly Z. Burden of cause-specific mortality associated with PM2.5 air pollution in the United States. *JAMA Netw Open.* (2019) 2:e1915834. doi: 10.1001/jamanetworkopen.2019.15834

28. Birmaher B, Brent DA, Chiappetta L, Bridge J, Monga S, Baugher M. Psychometric properties of the Screen for Child Anxiety Related Emotional Disorders (SCARED): a replication study. J Am Acad Child Adolesc Psychiatry. (1999) 38:1230-6. doi: 10.1097/00004583-199910000-00011

29. Yolton K, Khoury JC, Burkle J, LeMasters G, Cecil K, Ryan P. lifetime exposure to traffic-related air pollution and symptoms of depression and anxiety at age 12 years. *Environ Res.* (2019) 173:199–206. doi: 10.1016/j.envres.2019.03.005

30. Kovacs M. The children's depressive, inventory (CDI). Psychopharmacol Bulletin. (1985) 21:995-8.

31. Markworth JF, Vella L, Lingard BS, Tull DL, Rupasinghe TW, Sinclair AJ, et al. Human inflammatory and resolving lipid mediator responses to resistance exercise and ibuprofen treatment. *Am J Physiol Regul Integr Comp Physiol.* (2013) 305:R1281–96. doi: 10.1152/ajpregu.00128.2013

32. Khanna D, Khanna S, Khanna P, Kahar P, Patel BM. Obesity: A chronic lowgrade inflammation and its markers. *Cureus*. (2022) 14:e22711. doi: 10.7759/ cureus.22711

33. Mahabee-Gittens EM, Matt GE, Mazzella MJ, Doucette JT, Ratnani P, Merianos AL. Inflammatory marker levels in children with tobacco smoke exposure. *Cytokine*. (2024) 173:156448. doi: 10.1016/j.cyto.2023.156448

34. Kilpatrick DJ, Hung P, Crouch E, Self S, Cothran J, Porter DE, et al. Geographic variations in urban-rural particulate matter (PM(2.5)) concentrations in the United States, 2010-2019. *Geohealth*. (2024) 8:e2023GH000920. doi: 10.1029/2023GH000920

35. Hale WW 3rd, Raaijmakers Q, Muris P, van Hoof A, Meeus W. Developmental trajectories of adolescent anxiety disorder symptoms: a 5-year prospective community study. *J Am Acad Child Adolesc Psychiatry*. (2008) 47:556–64. doi: 10.1097/CHI.0b013e3181676583

36. Yoon Y, Eisenstadt M, Lereya ST, Deighton J. Gender difference in the change of adolescents' mental health and subjective wellbeing trajectories. *Eur Child Adolesc Psychiatry*. (2023) 32:1569–78. doi: 10.1007/s00787-022-01961-4

37. Ricciotti E, FitzGerald GA. Prostaglandins and inflammation. Arterioscler Thromb Vasc Biol. (2011) 31:986-1000. doi: 10.1161/ATVBAHA.110.207449

38. Chu J, Pratico D. The 5-lipoxygenase as a common pathway for pathological brain and vascular aging. *Cardiovasc Psychiatry Neurol.* (2009) 2009:174657. doi: 10.1155/2009/174657

39. Lammers CH, Schweitzer P, Facchinetti P, Arrang JM, Madamba SG, Siggins GR, et al. Arachidonate 5-lipoxygenase and its activating protein: prominent hippocampal expression and role in somatostatin signaling. *J Neurochem.* (1996) 66:147–52. doi: 10.1046/j.1471-4159.1996.66010147.x

40. Wang T, Han Y, Chen X, Chen W, Li H, Wang Y, et al. Particulate air pollution and blood pressure: signaling by the arachidonate metabolism. *Hypertension*. (2023) 80:2687–96. doi: 10.1161/HYPERTENSIONAHA.123.21410

41. Zhu K, Browne RW, Blair RH, Bonner MR, Tian M, Niu Z, et al. Changes in arachidonic acid (AA)- and linoleic acid (LA)-derived hydroxy metabolites and their interplay with inflammatory biomarkers in response to drastic changes in air pollution exposure. *Environ Res.* (2021) 200:111401. doi: 10.1016/j.envres.2021.111401

42. Mu L, Niu Z, Blair RH, Yu H, Browne RW, Bonner MR, et al. Metabolomics Profiling before, during, and after the Beijing Olympics: A Panel Study of Within-Individual Differences during Periods of High and Low Air Pollution. *Environ Health Perspect.* (2019) 127:57010. doi: 10.1289/EHP3705 43. Lin Y, Wang X, Chen R, Weil T, Ge Y, Stapleton HM, et al. Arachidonic acid metabolites in self-collected biospecimens following campfire exposure: exploring non-invasive biomarkers of wildfire health effects. *Environ Sci Technol Lett.* (2024) 11:201–7. doi: 10.1021/acs.estlett.3c00923

44. Maddipati KR. Distinct etiology of chronic inflammation - implications on degenerative diseases and cancer therapy. *Front Immunol.* (2024) 15:1460302. doi: 10.3389/fimmu.2024.1460302

45. Vancil A, Strawn JR, Rasnick E, Levine A, Schroeder HK, Specht AM, et al. Pediatric anxiety and daily fine particulate matter: A longitudinal study. *Psychiatry Res Commun.* (2022) 2. doi: 10.1016/j.psycom.2022.100077

46. Wang WY, Liu N, Qi XX, Han B, Sun JN, Chen ZL, et al. Predictive effect of lipopolysaccharide-stimulated inflammatory cytokines on symptoms of generalized anxiety disorder. *World J Psychiatry.* (2024) 14:1308–18. doi: 10.5498/wjp.v14.i9.1308

47. Kemp GM, Altimimi HF, Nho Y, Heir R, Klyczek A, Stellwagen D. Sustained TNF signaling is required for the synaptic and anxiety-like behavioral response to acute stress. *Mol Psychiatry*. (2022) 27:4474–84. doi: 10.1038/s41380-022-01737-x

48. Leung CY, Weiss SJ. Cytokines and depressive symptoms among adolescents. Biol Res Nurs. (2025), 10998004251318385. doi: 10.1177/10998004251318385

49. Warner EN, Strawn JR. Risk factors for pediatric anxiety disorders. *Child Adolesc Psychiatr Clin N Am.* (2023) 32:485–510. doi: 10.1016/j.chc.2022.10.001

50. Wei J, Li Z, Lyapustin A, Wang J, Dubovik O, Schwartz J, et al. First close insight into global daily gapless 1 km PM(2.5) pollution, variability, and health impact. *Nat Commun.* (2023) 14:8349. doi: 10.1038/s41467-023-43862-3

51. Miller GE, Passarelli V, Chen E, Kloog I, Wright RJ, Amini H. Ambient PM(2.5) and specific sources increase inflammatory cytokine responses to stimulators and reduce sensitivity to inhibitors. *Environ Res.* (2024) 252:118964. doi: 10.1016/j.envres.2024.118964