

OPEN ACCESS

EDITED BY
Shani Tiwari,
Council of Scientific and Industrial
Research (CSIR), India

REVIEWED BY
Vishnu Murari,
Indian Institute of Technology
Kanpur, India
Amit Kumar Mishra,
Jawaharlal Nehru University, India
Arti Choudhary,
Utkal University, India

*CORRESPONDENCE
Abhay Kumar Singh
singhak@bhu.ac.in

SPECIALTY SECTION
This article was submitted to
Climate Change and Cities,
a section of the journal
Frontiers in Sustainable Cities

RECEIVED 31 March 2022
ACCEPTED 06 July 2022
PUBLISHED 02 August 2022

CITATION
Chauhan PK, Kumar A, Pratap V and
Singh AK (2022) Seasonal
characteristics of PM₁, PM_{2.5}, and
PM₁₀ over Varanasi during 2019–2020.
Front. Sustain. Cities 4:909351.
doi: 10.3389/frsc.2022.909351

COPYRIGHT
© 2022 Chauhan, Kumar, Pratap and
Singh. This is an open-access article
distributed under the terms of the
[Creative Commons Attribution License
\(CC BY\)](https://creativecommons.org/licenses/by/4.0/). The use, distribution or
reproduction in other forums is
permitted, provided the original
author(s) and the copyright owner(s)
are credited and that the original
publication in this journal is cited, in
accordance with accepted academic
practice. No use, distribution or
reproduction is permitted which does
not comply with these terms.

Seasonal characteristics of PM₁, PM_{2.5}, and PM₁₀ over Varanasi during 2019–2020

Prashant Kumar Chauhan¹, Akhilesh Kumar², Vineet Pratap¹
and Abhay Kumar Singh^{1*}

¹Department of Physics, Institute of Science, Banaras Hindu University, Varanasi, India, ²Kashi Naresh Government Post Graduate College Gyanpur, Bhadohi, India

Particulate matter (PM) concentrations and aerosol optical depth (AOD) are measured and correlated simultaneously using a high-volume sampler and a MICROTUPS-II Sunphotometer, respectively. The present work deals with the characteristics of particulate matter (PM₁, PM_{2.5}, and PM₁₀) over Varanasi, from April 2019 to March 2020. Daily variation, as well as seasonal variation, reveals the dominance of fine-mode particles over the Varanasi region in the winter season and the dominance of coarse-mode particles in the summer season, which was further confirmed by calculating the ratio between particulate matter (PM₁/PM₁₀ and PM_{2.5}/PM₁₀). This ratio was discovered to be lowest in the summer and highest in the winter. Annual mean concentrations of PM₁, PM_{2.5}, and PM₁₀ are found to be 93.91, 111.34, and 180.70 μgm⁻³, respectively. The seasonal variation shows relatively a higher concentration of PM₁, PM_{2.5}, and PM₁₀ in the winter season, which may be due to stable meteorological conditions and increased biomass burning in winter. Diurnal and seasonal variations in AOD were also studied during this period. A large and small value of AOD represents the dominance of fine particles over coarse particles. At 500 nm, maximum (1.17) and minimum (0.44) AODs were measured in December and August of 2019, respectively. There was a statistically significant correlation between PM particles (PM₁, PM_{2.5}, and PM₁₀) and AOD. Elemental analysis shows that fluorine and carbon are the major elements that were observed in selected samples during the post-monsoon and winter season using SEM-EDX analysis.

KEYWORDS

particulate matter, aerosols, AOD, Indo-Gangetic Basin, MICROTUPS II Sunphotometer

Introduction

A primary atmospheric component of air pollution, i.e., particulate matter (PM), has become a major source of concern all around the world due to its negative impacts on air quality, human health, and the earth's ecosystem (Chowdhury and Dey, 2016; Ghude et al., 2016; Seinfeld et al., 2016; Singh et al., 2016a). Air pollution is one of

the leading causes of sickness. Due to severely poor air quality, several cities in India have been named among the top 20 most polluted cities in the world during the previous few decades, with eleven of them being on the Indo-Gangetic Plain (IGP) (Garaga et al., 2018). Furthermore, it is the world's fourth leading cause of premature mortality, according to the reports (Cohen et al., 2017). PM₁₀, PM_{2.5}, and PM₁ refer to airborne particulate matter having aerodynamic diameters of $\leq 10 \mu\text{m}$, $\leq 2.5 \mu\text{m}$, and $\leq 1 \mu\text{m}$, respectively, and can also be considered as coarse-mode, fine-mode, and accumulation-mode particles, respectively (Spandana et al., 2021). PM_{2.5} and PM₁ are significantly more harmful than PM₁₀ due to their complexity and smaller size (Miri et al., 2016, 2017; Li et al., 2017). Industry, power plants, three-wheelers, and other combustion activities are all responsible for the production of PM, which comes from both primary emission and chemical alteration of precursor gases, which results in secondary particles (Rahman et al., 2020). In the recent years, urbanization, industrialization, and anthropogenic activities have led to an increase in PM concentrations in the surrounding atmosphere. Due to the proliferation of these types of activities, major cities throughout the world have faced severe air quality concerns, since the middle of the twentieth century (Cheng et al., 2013; Elbayoumi et al., 2013). Particles produced from both natural and anthropogenic sources make up PM in the atmosphere (Kaufman et al., 2005). PM can travel a long distance from its source if the atmospheric conditions are favorable (Ancelet et al., 2015; Tiwari et al., 2018, 2019). Increased quantities of particulate matter (PM) pose a regional and global environmental threat (Delfino et al., 2005; Obaidullah et al., 2012). Particulate matter (PM) plays a significant role in disrupting the Earth's radiative budget by absorption and scattering of incoming solar radiation and outgoing long wave radiation. They also play a role in cloud formation, through which they also have an impact on cloud lifespan and the precipitation process (Wang and Penner, 2009; Seinfeld et al., 2016; Tiwari et al., 2018). Depending on the dominant activity, such as light absorption or scattering, PM can operate as both a cooling and warming agent in the atmosphere (Myhre, 2009). The scattering and absorption coefficients of a particle are determined by its physical and chemical characteristics. As a result of this coefficient, air particulates generate regional smog, discoloration, texture loss, and invisibility in a specific area (Malm and Day, 2000). PM_{2.5} and PM₁ have a harmful impact on human health because they can enter into alveoli through the respiratory system, causing lung illnesses, heart attacks, neurological disorders, and other health problems, while PM₁₀ affects the atmosphere's radiation balance as well as visibility (Salma et al., 2002; Pope et al., 2008). According to the research, high PM_{2.5} exposures can also damage brain development in new born babies and children (Egondi et al., 2018). About 1.24 million deaths happened in India out of the 5 million deaths that happen globally because of air pollution, 54% of which are attributed to ambient particulate

matter (Balakrishnan et al., 2019). Due to westerly/south-westerly winds in conjunction with dry weather conditions, IGP experiences a sudden and intense dust storm during the pre-monsoon season. These dust storms carry coarse-mode particles from Southeast Asia and the Thar Desert (Tiwari and Singh, 2013; Kumar et al., 2015a; Singh et al., 2016b; Tiwari et al., 2019). Due to the increasing aggregation of coarse-mode particles, air quality and visibility have deteriorated (Smith et al., 2019; Taneja et al., 2020). Paddy residue is widely burnt in Haryana and Punjab every year during the post-monsoon season, which is a major source of air pollution (Kumar et al., 2016; Singh et al., 2016a; Ojha et al., 2020). IGP has a higher particulate matter concentration due to the low height of the atmospheric boundary layer, low wind speeds, and increased amount of wood burning during the winter time, which reduces visibility (Massie et al., 2004; Rajesh and Ramachandran, 2017; Ali et al., 2019). In 2016, an analysis of haze episodes in Delhi using applied carbon tracers revealed that the burning of agricultural waste was the dominant source of pollution (Sawhani et al., 2019). Aerosol optical depth (AOD) is a measure that is used to assess PM concentrations and to describe air quality and atmospheric conditions, and the extinction of incoming solar radiation by air particles distributed in a vertical column of the atmosphere can also be evaluated by this quantity (Kompalli et al., 2010; Dey et al., 2012; Srivastava et al., 2014). The concentrations of PM₁₀ and PM_{2.5} have a significant connection with AOD (Srivastava et al., 2012a; Xin et al., 2014). Apart from this, the Moderate Resolution Imaging Spectroradiometer (MODIS) satellite data are widely utilized to investigate atmospheric processes and climate change because they have good spatial and temporal coverage ($1^\circ \times 1^\circ$) (Kaufman et al., 2002). Tiwari and Singh (2013) used MODIS Terra data for 12 months above Varanasi in 2011. MODIS data are accessible on the NASA Giovanni website free of cost. The research was conducted utilizing a standard MICROTUPS-II Sunphotometer during the year 2019–2020, and it included a comparison of the ground-based measurement of AOD to those obtained by satellite using level 3 MODIS data. To observe the morphology, structure, and chemical composition of PM, some collected samples have been characterized through scanning electron microscope-energy dispersive X-ray spectrometry (SEM-EDX). Over Varanasi, only a few studies have been reported on the chemical characteristics of particulate matter and the sources of its emissions (Murari et al., 2016; Tiwari et al., 2016; Pratap et al., 2020a). Prior research has found that the PM_{2.5}/PM₁₀ ratio has a substantial temporal variability and a fine correlation with climatic factors such as temperature, wind speed, and relative humidity (Akinlade et al., 2015; Speranza et al., 2016; Mukherjee and Agrawal, 2017). Although few researchers have evaluated the properties of pollutants over the IGP, they have been only limited to a specific proportion of particles (either PM₁₀ or PM_{2.5}) through which the evolution of anthropogenic activities could have been understood (Saxena et al., 2017; Jain et al., 2019).

Thus, this study analyses the variation in the concentration of PM_1 , $PM_{2.5}$, and PM_{10} and its correlation with aerosol optical depth during pre-monsoon and post-monsoon season at BHU, Varanasi, situated in central Ganges Valley. Using an NOAA Hybrid Single-Particle Lagrangian Integrated Trajectory (HYSPLIT), 5-day air mass back trajectories at three different heights such as 500, 1,000, and 1,500 m have been displayed for the selected day every month to identify the regional or transboundary origin of particles (Draxler and Rolph, 2003).

Experimental observations and methodology

Site description and meteorological condition

During the study period (April 2019–March 2020), particulate matter monitoring is carried out on the rooftop of the Department of Physics at Banaras Hindu University (BHU), Varanasi. Varanasi (25.26° N, 82.99° E, 83 m above mean sea level) is situated in the eastern part of Uttar Pradesh, which is a semi-urban city in the central Gangetic Plain and has various sources of contaminated pollutants and suffers from high particulate loading (Ram and Sarin, 2011; Tiwari and Singh, 2013; Kumar et al., 2015b; Tiwari et al., 2018). The Indo-Gangetic Plain (IGP) runs parallel to the Himalayan range, having a stretch of 2400 km from Jammu and Kashmir in the west to the Assam in the east, and has a significant aerosol hotspot of air contaminants due to its unique geomorphology, weather patterns, and climatic susceptibility (Dey et al., 2004; Ramachandran and Kedia, 2010; Srivastava et al., 2012b; Kumar et al., 2020; Pratap et al., 2020a). During the pre-monsoon season in which hot and dry summer exists, coarse-mode aerosol prevails, and extreme rainfall occurs during monsoon whereas an excess of fine-mode aerosols in extremely cold weather during winter and post-monsoon was reported (Satheesh et al., 2006; Tiwari et al., 2018).

Meteorological conditions over Varanasi during this study period from 01 April 2019 to 31 March 2020 are shown in Figure 1. Varanasi is extremely hot in the summer and humid in the monsoon while being outstandingly cold in the winter season. During the study period, temperature ($^\circ\text{C}$), relative humidity (%), and wind speed (ms^{-1}) data are downloaded from the CPCB portal. The relative humidity, temperature, and wind speed were plotted jointly. The average relative humidity is found to be 61.58%, while the minimum and maximum relative humidities are 17.20 and 90% in May and September, respectively. The mean temperature is found to be 25.23°C while the minimum and maximum temperatures are 4.04 and 40.49°C in December and April, respectively. The average wind speed is found to be 1.34 ms^{-1} while the minimum and maximum wind speeds are 0.30 and 5.64 ms^{-1} in August, respectively.

Instrumentation and methodology

PM sampler

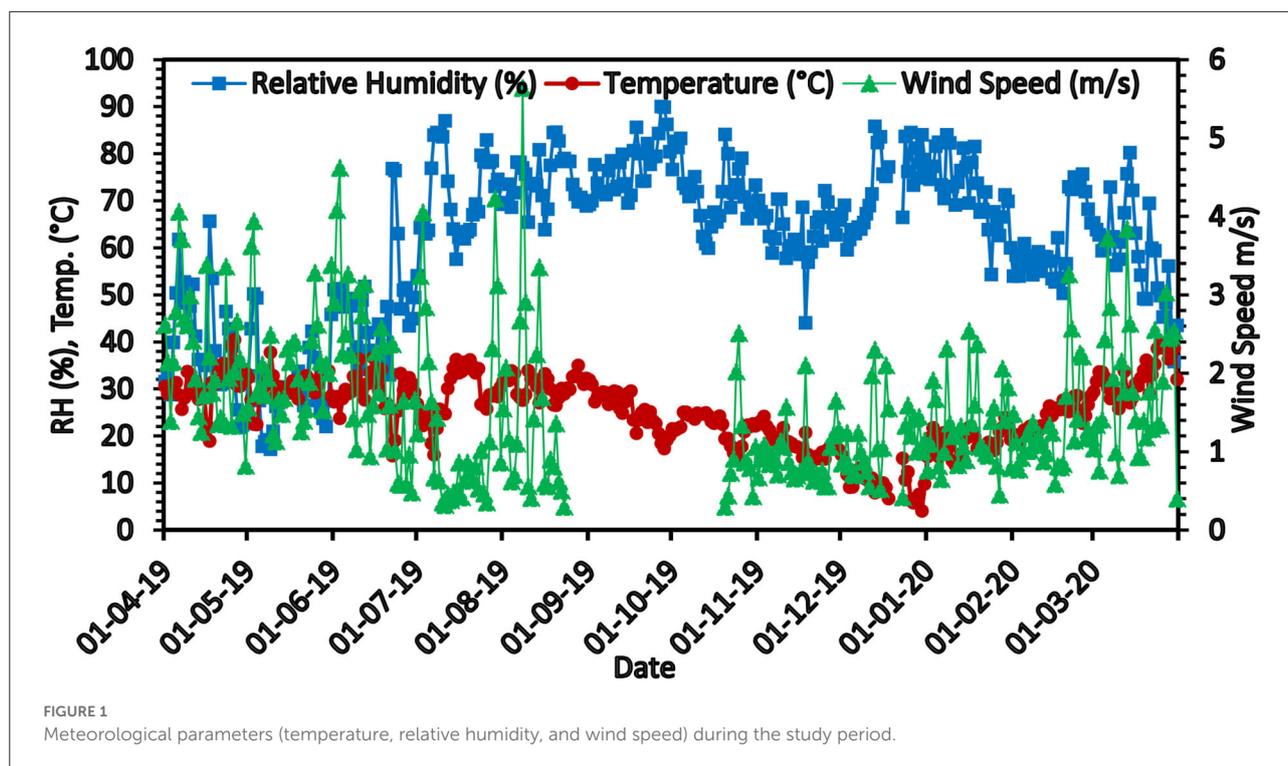
Particulate matter samples were collected two times a week for 24 h using three Envirotech Pvt Ltd respirable dust samplers: Model no. APM 460 NL respirable dust samplers for collecting PM_{10} , APM 550 for $PM_{2.5}$ with a sampling rate of 16.7 lpm, and APM 577 with a sampling rate of 10 lpm for PM_1 dust samples for further analysis of their mass concentrations and elemental analysis during the study period (Bansal et al., 2019; Sah et al., 2019). Coarse-mode particulates (PM_{10}) are aggregated on a GF/A 8×10 in glass microfiber filter paper whereas fine particles ($PM_{2.5}$ and PM_1) are collected on a $2\text{-}\mu\text{m}$ PTFE microfiber filter paper of 46.2 mm diameter. The gravimetric method is used to calculate particulate matter concentration. For moisture removal, the filter papers were kept in desiccators for 24 h before and after sampling.

MODIS

The Moderate Resolution Imaging Spectroradiometer is used to monitor the ocean, land, and atmosphere. This instrument is onboard Terra and Aqua satellites, launched in December 1999 and May 2002 (Alam et al., 2014). MODIS provides daily global dust data in 36 spectral bands from visible to thermal infrared ($0.41\text{--}14.38 \mu\text{m}$) (Sharma et al., 2012). The MODIS global gridded level 3 aerosol products is derived from the level 2 MODIS product with 10 km of spatial resolution (Remer et al., 2005). The MODIS equipment does a full scan of the earth's surface once every 1 to 2 days. At the local solar time, 10:30 in the morning and 01:30 in the afternoon, respectively, the Terra and Aqua satellites pass over the Indian Territory. The daily AOD satellite data from MODIS on-board Terra satellite level 3 AOD (mean_MOD08_D3_6_1_Deep_Blue_Aerosol_Optical_Depth_550_Land_Mean) for $1^\circ \times 1^\circ$ grid from April 2019 to March 2020 over Varanasi were collected at 550 nm wavelength.

MICROTUPS-II

Ground-based AOD observations were made about three times a week during clear sky conditions from 10:00 to 15:00 h using a portable multi-band MICROTUPS-II (MT-II) Sunphotometer (Solar Light Company, USA). It is a five-channel handheld sunphotometer that gives instantaneous columnar AODs at five wavelengths: 380, 440, 500, 675, and 870 nm. The MT-II Sunphotometer has a 2.5° field of view. The MT includes temperature and pressure sensors associated with GPS connectivity for obtaining position and time coordinates (Morys et al., 2001). This instrument is based on the principle of measuring the intensity of incoming solar radiation at specific wavelengths and then converting it to optical depth using the Langley method. In this study, a considerable correlation



of spectral and seasonal AOD with the particulate matter was observed.

SEM-EDX

A computer-controlled SEM coupled with EDX was used to characterize the morphology and elemental analysis of airborne particles. This instrument is made by Carl Zeiss Microscopy in Germany. The main filter paper was randomly sliced into a 1-mm² sample for the investigation of elemental composition and surface morphology (Pratap et al., 2020a). To achieve more favorable secondary images and to make them electrically conductive, the samples were sputtered with gold. The EDX spectrum of airborne particulate was recorded at 15–30 KV (Murari et al., 2016). For qualitatively spectra, elements were measured at different points.

Results and discussion

Variation in mass concentration of particulate matters

Using a high-volume sampler, the mass concentration of particulate matter was measured over Varanasi from April 2019 to March 2020. In Figure 2, the daily variation in mass concentrations of PM₁, PM_{2.5}, and PM₁₀ is shown. Mean mass concentration from April 2019 to March 2020 was found to

be 89.9 ± 44.4 for PM₁, 106.5 ± 67.2 for PM_{2.5}, and $180.8 \pm 71.4 \mu\text{g}/\text{m}^3$ for PM₁₀, respectively. Most of the time mass concentrations of particulate matter were found to be exceeded the NAAQS limit ($40 \mu\text{g}/\text{m}^3$ for PM_{2.5} and $60 \mu\text{g}/\text{m}^3$ for PM₁₀). A similar result was reported by Kumar et al. (2020) in Varanasi. Murari et al. (2017) reported that annual mass concentrations of PM₁₀ and PM_{2.5} were found to be 161.3 and $81.8 \mu\text{g}/\text{m}^3$, respectively, in 2014 over Varanasi. Jain et al. (2021) found the average concentration of PM₁₀ and PM_{2.5} in Varanasi was 257.90 and $99.33 \mu\text{g}/\text{m}^3$, respectively from January 2015 to December 2016. Singh et al. (2021) recorded mean mass concentrations of daily PM₁₀ and PM_{2.5} of 239 ± 128 and $123 \pm 89 \mu\text{g}/\text{m}^3$, respectively, during the period from July 2014 to June 2018 over Varanasi. The seasonal behavior of PM₁, PM_{2.5}, and PM₁₀ reported in the previous studies in IGP is also shown in Table 1. Mass concentrations of particulate matter depend on different seasons which are shown in Figure 3. As per data available, we have considered October as a post-monsoon season, November–February as the winter season, March as a transition period named as the spring season, and April–June took as the summer season (Prasad and Singh, 2009; Kumar et al., 2021). Mass concentration of PM₁ for post-monsoon, winter and spring was found to be, 104.6 ± 43.2 , and $57.5 \pm 31.8 \mu\text{g}/\text{m}^3$, respectively. Mass concentration of PM₁ for the summer season was not available. Mass concentration of PM_{2.5} during post-monsoon, winter, spring, and summer was found to be 79.0 ± 19.6 , 156.1 ± 59.3 , 82.0 ± 31.3 , and $43.5 \pm$

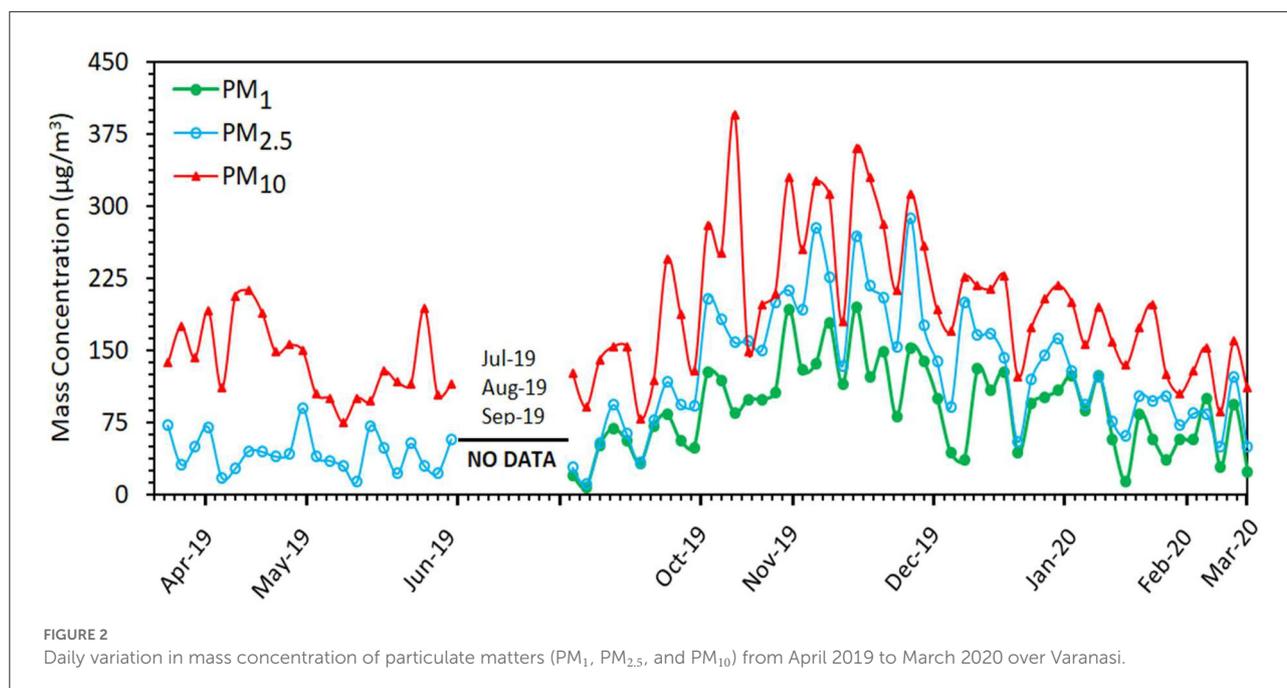
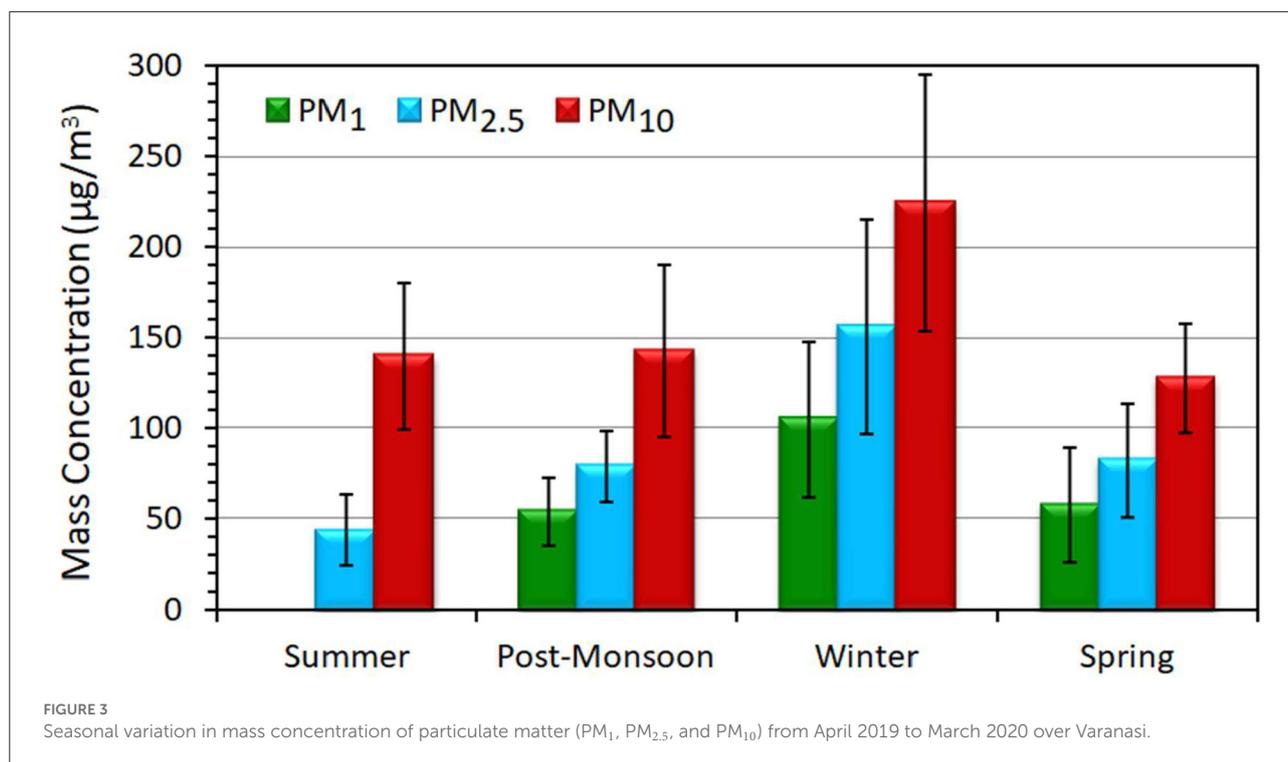


TABLE 1 Previous study of PM₁, PM_{2.5}, and PM₁₀ samples at different locations in IGP.

Location	Study period	Season	PM ₁ (µg m ⁻³)	PM _{2.5} (µg m ⁻³)	PM ₁₀ (µg m ⁻³)	References
Varanasi	2016	Winter		129.808 ± 36.348	207.992 ± 66.861	Kumar et al., 2020
		Post-monsoon		132.434 ± 51.752	173.784 ± 37.671	
	2017	Pre-monsoon		39.211 ± 22.872	142.618 ± 40.034	
		Post-monsoon		121.46 ± 65.374	153.577 ± 82.595	
	Winter		110.732 ± 42.581	177.344 ± 68.971		
2018	Pre-monsoon		59.413 ± 17.933	160.54 ± 54.866		
Ghaziabad	2016–17	Winter		134 ± 48	213 ± 80	Pratap et al., 2020b
		monsoon		76 ± 35	189 ± 149	
	2018–19	Pre-monsoon		126 ± 33	231 ± 92	Gupta et al., 2021
		Post-monsoon		222 ± 76	384 ± 142	
Kanpur	2015–16	Winter		191 ± 81	262 ± 114	Rajput et al., 2019
		Summer		13.7–84	28.2–158.1	
		monsoon		16.1–95.5	21.6–174.1	
		Post-monsoon		39.4–308.8	95.6–401.8	
Gurugram	2017	Summer		68.2–157.5	88.2–206.3	Rahman et al., 2020
		Winter		114	202	
Delhi	2015–16	Whole year		261	440	Jain et al., 2021
				135 ± 64	242 ± 95	
Varanasi				99 ± 33	257 ± 89	
Kolkata				115 ± 29	179 ± 77	
IIT Kanpur	2013	Post-monsoon	132.87 ± 27.97	–	–	Singh and Gupta, 2016
		Pre-winter				
Varanasi	2019–20	Post-monsoon	53.6 ± 18.8	79 ± 19.6	142.5 ± 47.9	Present study
		Winter	104.6 ± 43.2	156.1 ± 59.3	224 ± 70.8	
		Spring	57.5 ± 31.8	82 ± 31.3	127.6 ± 30	
		Summer		43.5 ± 19.7	139.6 ± 40.2	



19.7 $\mu\text{g}/\text{m}^3$ respectively. Additionally, the mass concentration of PM₁₀ during post-monsoon, winter, spring, and summer was found to be 142.5 ± 47.9 , 224.0 ± 70.8 , 127.6 ± 30.0 , and 139.6 ± 40.2 $\mu\text{g}/\text{m}^3$, respectively. It can be observed that the higher concentration of PM₁, PM_{2.5}, and PM₁₀ was found to be in the winter season. Winter is a season of stable meteorological conditions, calm wind, cold weather, and increased biomass burning as well as increased vehicular emissions (Guttikunda and Calori, 2013; Banerjee et al., 2015). Most importantly, the atmospheric boundary layer shifted downward and particulate matter gets trapped in the lower atmosphere (Li et al., 2020).

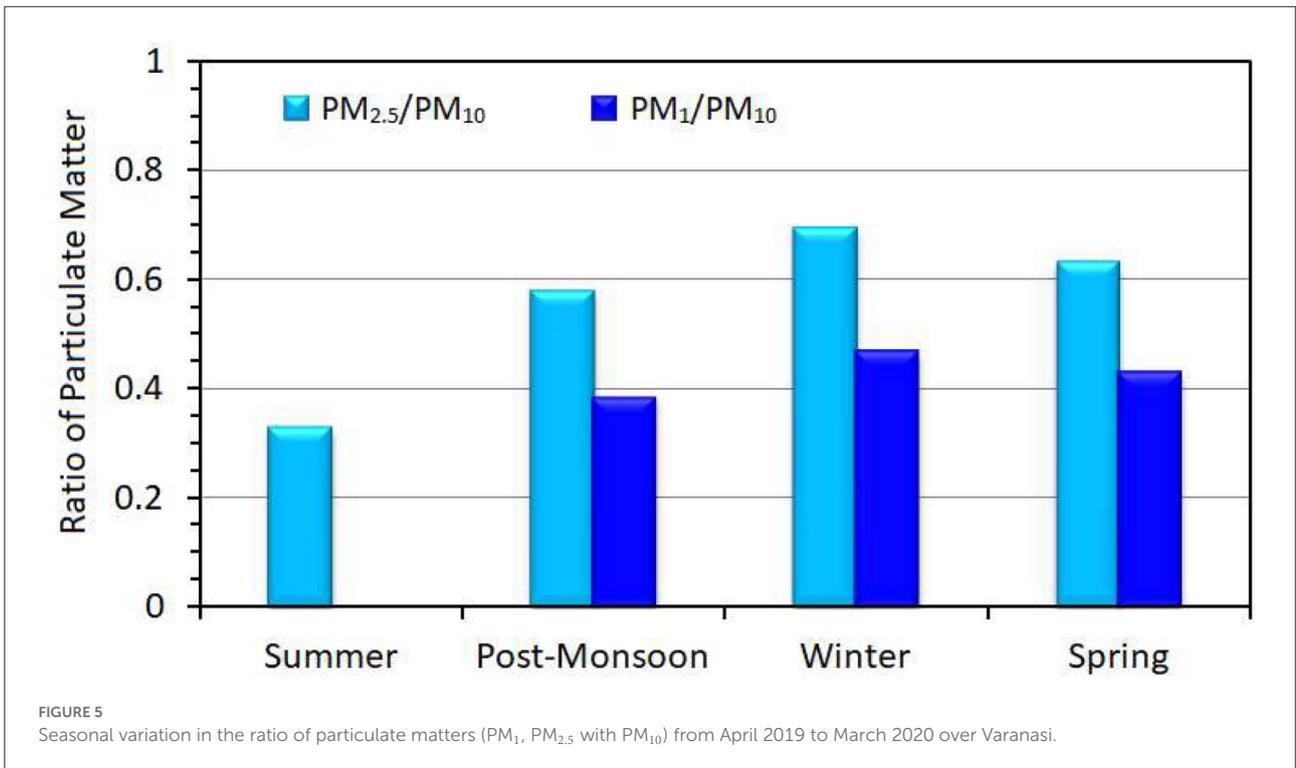
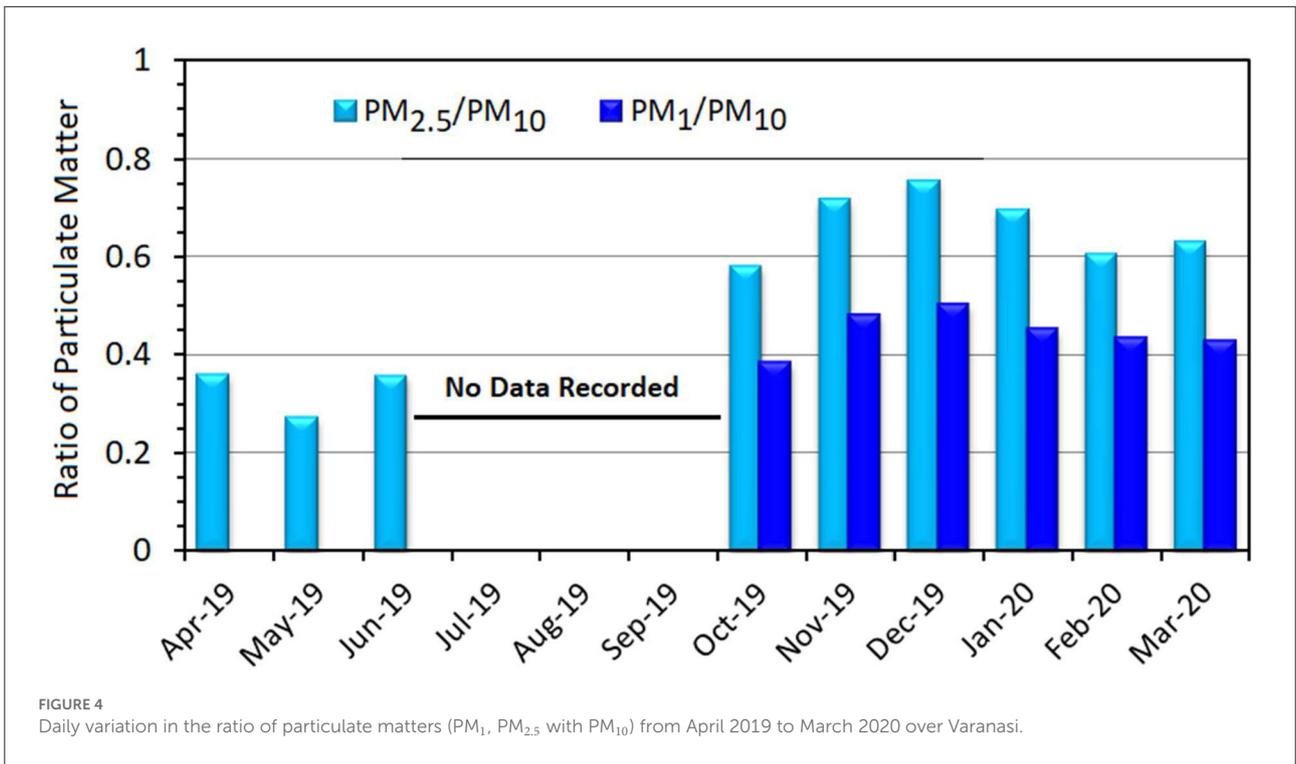
The ratio of mass concentration of particulate matters

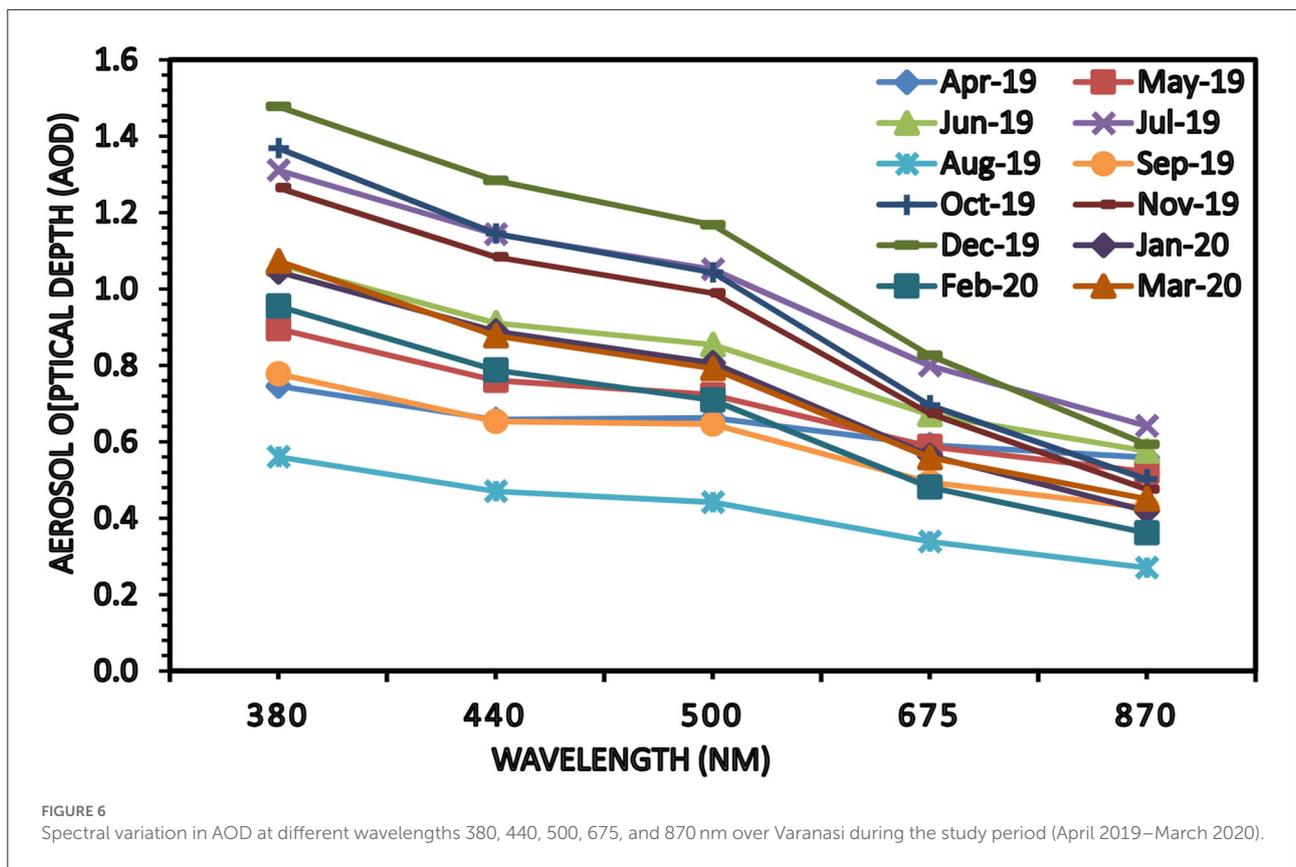
To identify the dominant contribution of mass concentration in different months and seasons, the ratio of PM₁ and PM_{2.5} to PM₁₀ was calculated. The monthly variation in the ratio of particulate matter is shown in Figure 4, and the seasonal variation in the ratio of particulate matter is shown in Figure 5. From Figure 4, the ratio of PM_{2.5}/PM₁₀ was found to be <0.4 in April, May, and June 2019, which suggests the dominance of coarse-mode aerosol particles. On the other hand, this ratio was found to be >0.5 in the other months and was highest in November and December (>0.7), which suggests an increase in fine-mode aerosol particles. Similar results for the ratio PM₁/PM₁₀ were found (this ratio is not shown in April, May, and June 2019 due to data unavailability). Concentrations

of particulate matter depend on the different seasons (Pratap et al., 2020a,b), which are shown in Figure 5. The ratio of PM_{2.5}/PM₁₀ was found to be lower (0.4) in the summer season, which suggests the dominance of coarse-mode particles. It may be due to the dust transportation from mid-arid regions and the Thar Desert over the Indo-Gangetic Basin (Kumar et al., 2015b; Murari et al., 2015; Tiwari et al., 2019). On the other hand, ratios of PM₁/PM₁₀ and PM_{2.5}/PM₁₀ were found to be higher in the winter and spring seasons, which suggests the enhancement of fine-mode particles. Punjab and Haryana states are considered the largest regions of crop residue burning, which leads to fine aerosol loading during post-monsoon as well as during winter seasons. These aerosol particles have transport properties and hence get transported long distances, which may cover the whole Indo-Gangetic Plain (Kaskaoutis et al., 2014). The ratio of PM_{2.5}/PM₁₀ and PM₁/PM₁₀ was found to be relatively higher in the post-monsoon as compared to the summer season. It may be due to the washout of coarser particles from the atmosphere while smaller particles remain suspended in the atmosphere. It is also reported that crop residue burning is intense over the Indo-Gangetic Plain during the post-monsoon season, causing an increase in fine-mode particles (Sarkar et al., 2018).

Variation of AOD(τ)

During the sampling period from April 2019 to March 2020, the monthly mean spectral fluctuation of AOD at five distinct wavelengths (380, 440, 500, 675, and 870 nm) derived





from the ground-based MICROTOPS II Sunphotometer over Varanasi is shown in Figure 6. The spectral fluctuation of AOD demonstrates rather plainly that the AOD was found to be substantially greater at shorter wavelengths, which can be due to the presence of fine particles, while the AOD was observed to be comparatively lower at longer wavelengths, which can be attributed to the presence of coarse particles (Reddy et al., 2011). A seasonal variation in AOD at five distinct wavelengths was also measured by the MICROTOPS-II Sunphotometer as shown in Figure 7. AOD at all five distinct wavelengths is found to be maximum during the post-monsoon season followed by winter, while it is minimum in the monsoon except for the summer season. The reason for the low value of AOD during the monsoon season could be the washout of aerosol particles from the atmosphere during rainfall (Dey and Di Girolamo, 2010; Lodhi et al., 2013; Tiwari and Singh, 2013), and crop residue burning during the winter season could be the reason for higher AOD (Kaskaoutis et al., 2014; Sarkar et al., 2018). In the summer season, AOD has an increasing trend and found maximum at 870 nm. Generally, dust storms are formed in the summer, which can be the reason for the presence of coarse particles. The daily average of AOD over Varanasi was found to be ranged from 0.2 to 2.0. AODs from MODIS and MICROTOPS-II are measured at 550 and 500 nm, respectively (Tiwari and Singh, 2013). Using the Angstrom power law (equation 1), the MICROTOPS-II

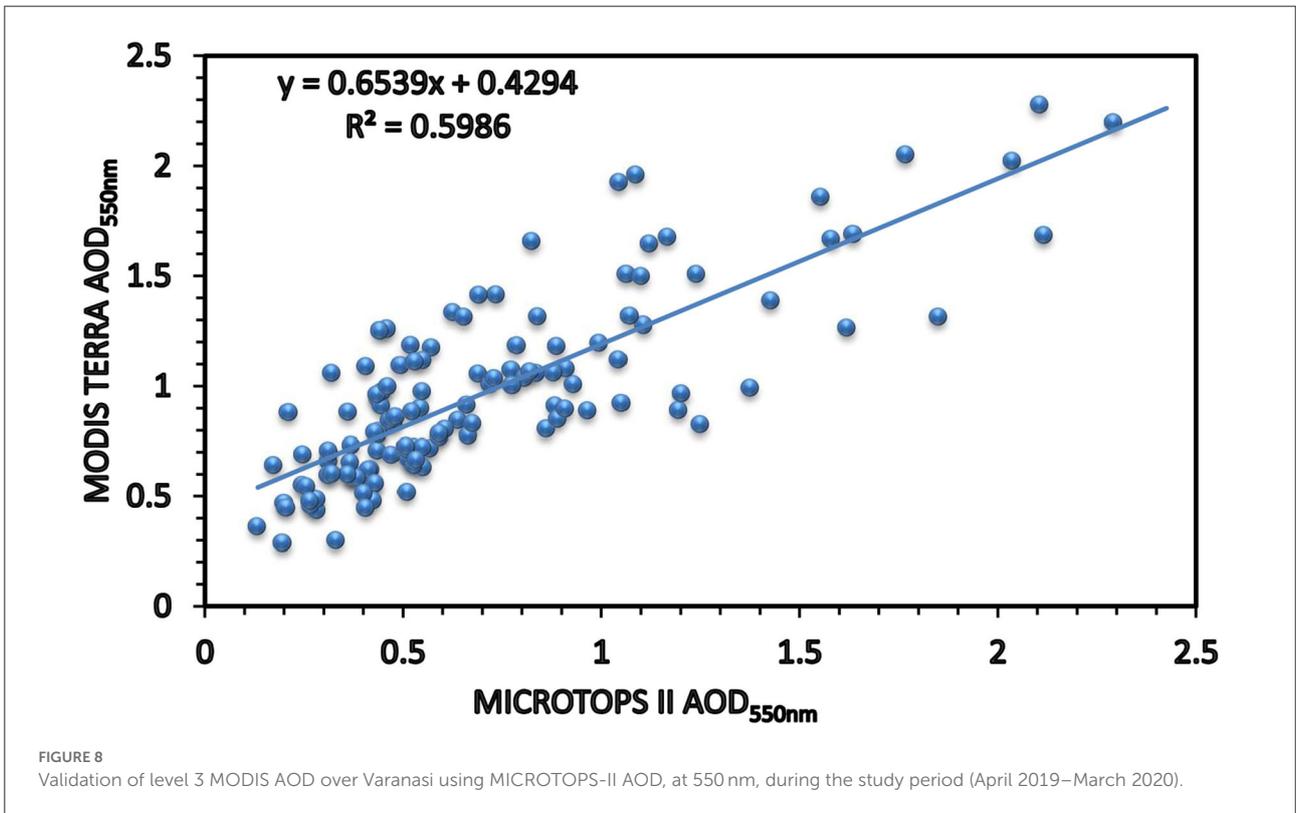
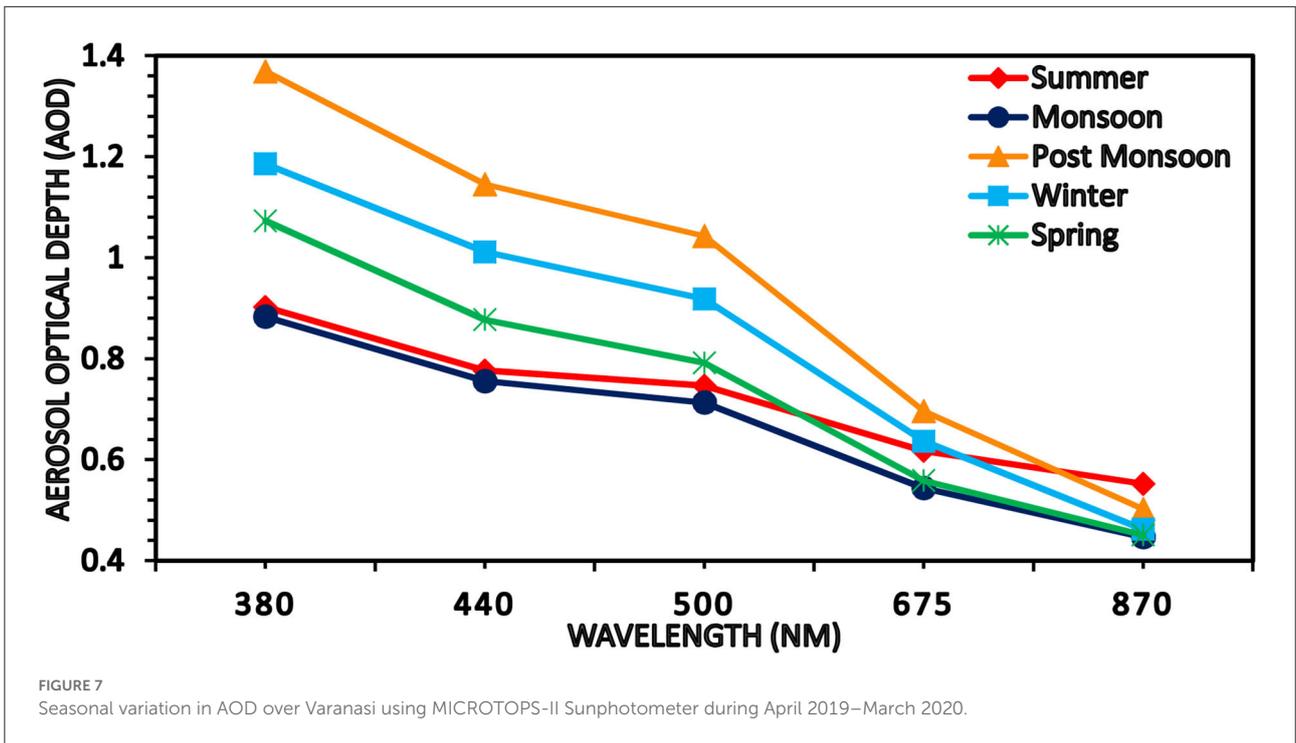
AOD at 500 nm was calculated at a wavelength of 550 nm to validate the MODIS and MICROTOPS-II Sunphotometer. Angstrom’s power law is given as Ångström (1964).

$$AOD_{550nm} = AOD_{500nm} \times (550/500)^{-\alpha} \tag{1}$$

where α is angstrom exponent (AE). For two different wavelengths (λ_1 and λ_2), AE is given by

$$\alpha = -\ln(\tau_{\lambda_1}/\tau_{\lambda_2})/\ln(\lambda_2/\lambda_1) \tag{2}$$

Angstrom exponent (wavelength range 380–870 nm) is a measure of the aerosol particle size and a fraction of fine to coarse-mode aerosols (Tiwari et al., 2018). AOD loading is observed to have large seasonal fluctuations, which are obtained higher in the winter season (November–February) and lower during the monsoon season (July–September). Over Varanasi, we found that satellite and ground-measured AOD data have an excellent one-to-one correlation, as shown in Figure 8. It has been observed that the total correlation coefficient between MODIS and MICROTOPS-II AOD data is 0.59. Later, diurnal variation in MODIS and MICROTOPS-II Sunphotometer AOD data at 550 nm is observed during this study period, as shown in Figure 9.



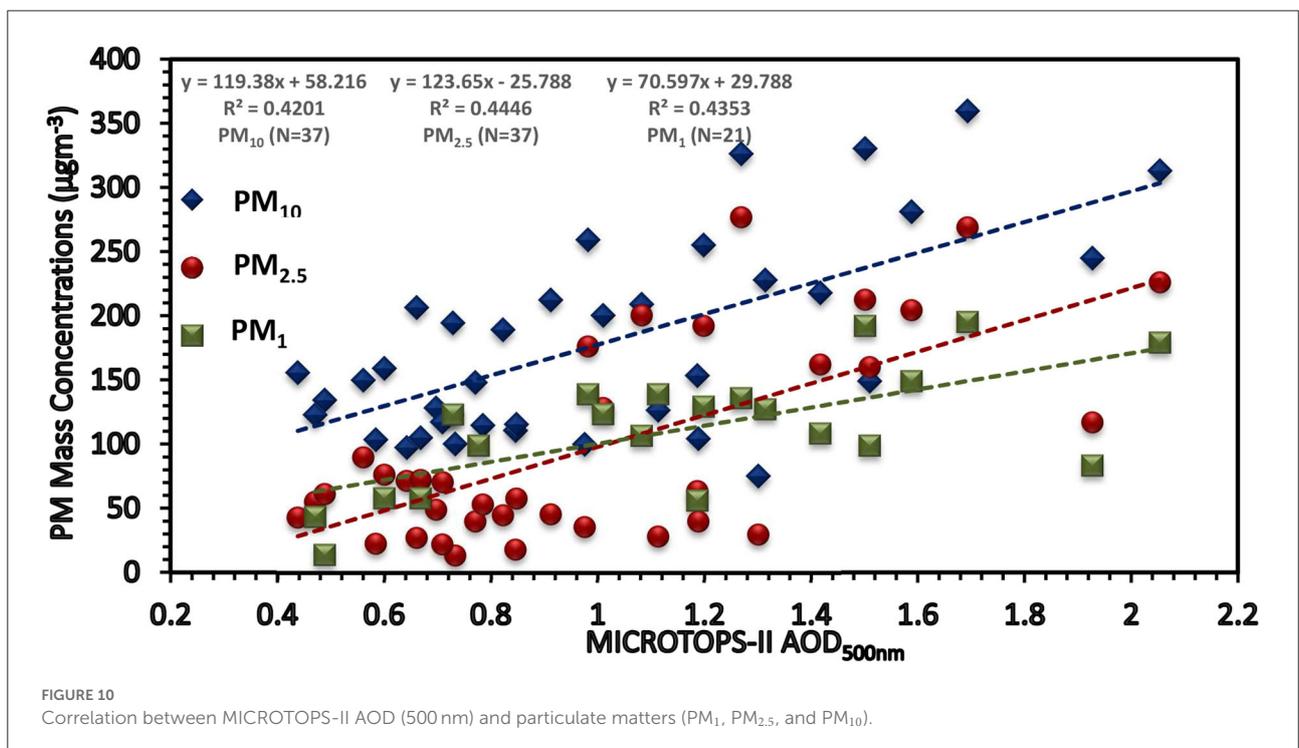
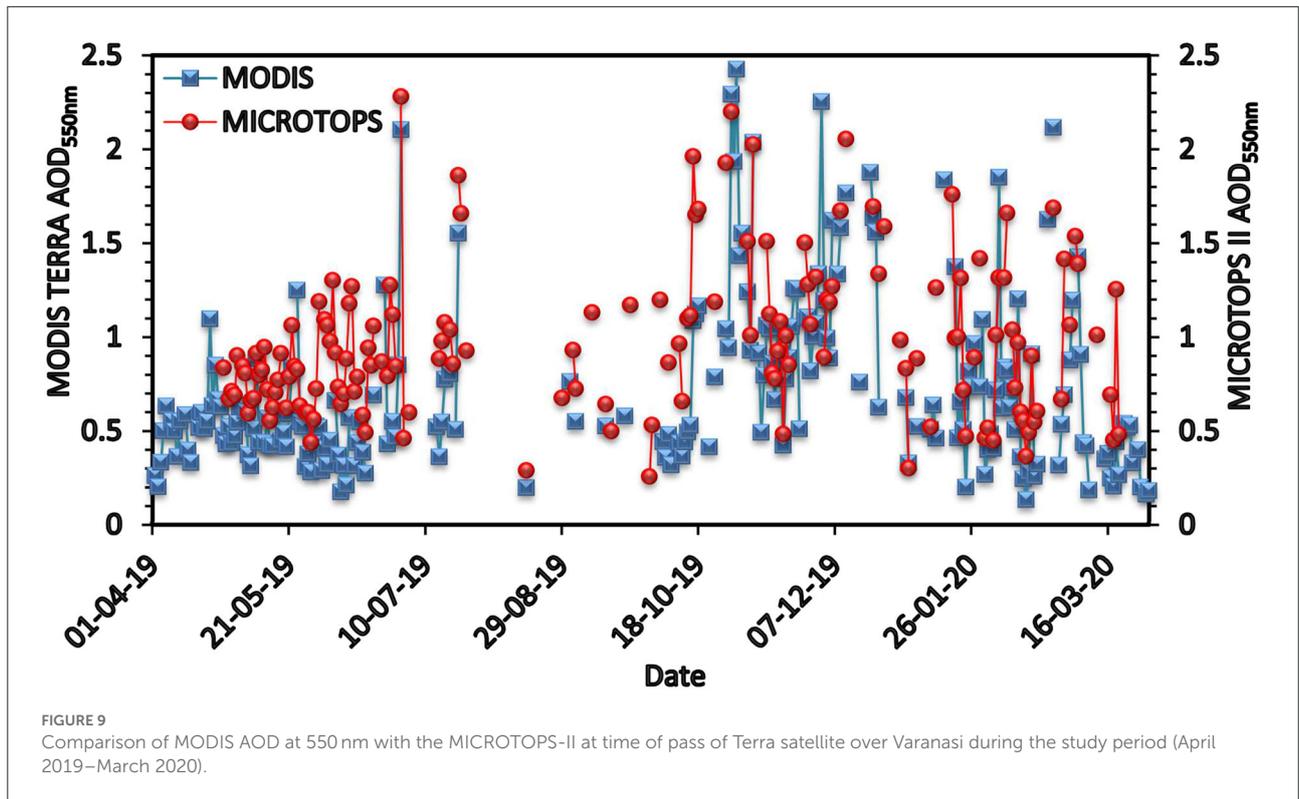
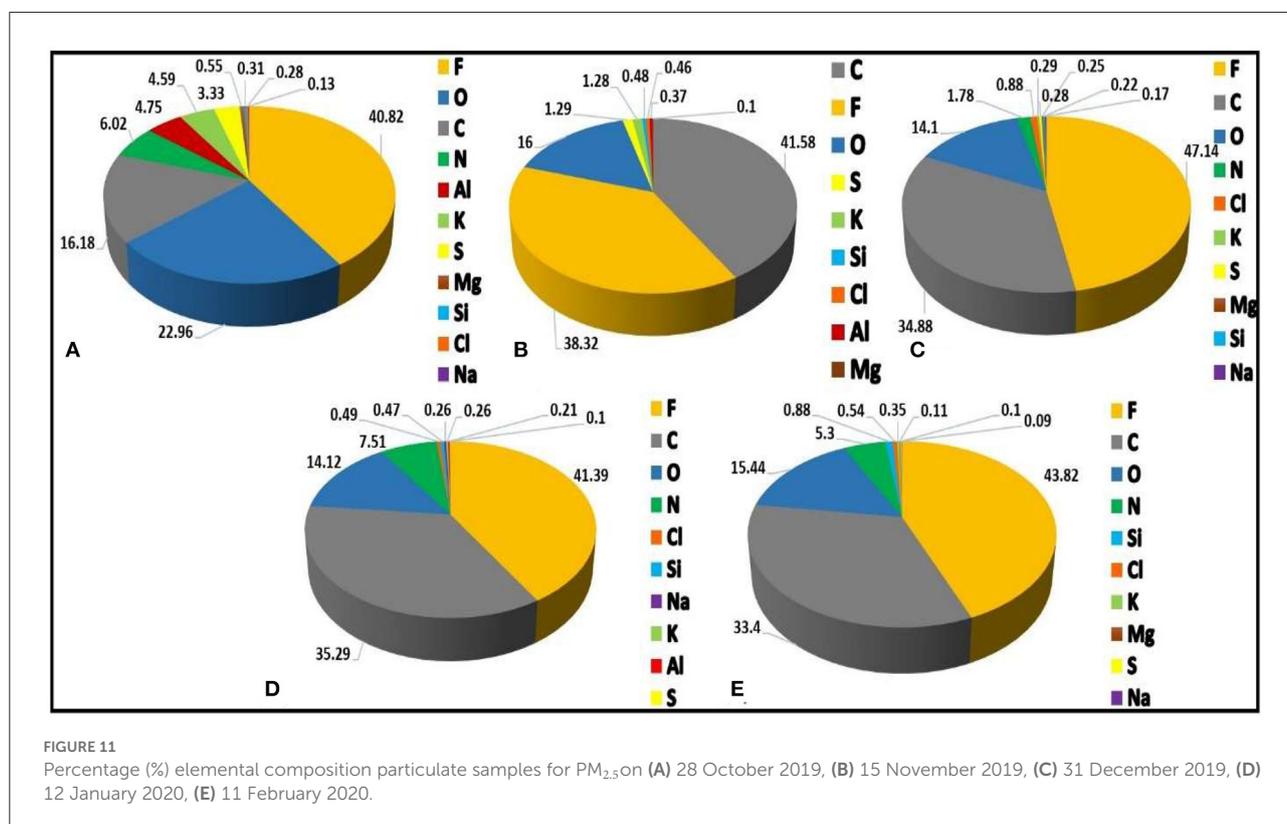


TABLE 2 Weight % (\pm standard deviation) of the elements quantified in the PM_{2.5} sample during the study period.

S. No.	Elements	Weight % (Oct 19)	Weight % (Nov 19)	Weight % (Dec 19)	Weight % (Jan 20)	Weight % (Feb 20)
1.	F	40.82	38.32	47.14	41.39	43.82
2.	C	16.18	41.58	34.88	35.29	33.4
3.	O	22.96	16	14.1	14.12	15.44
4.	N	6.02	–	1.78	7.51	5.3
5.	Na	0.13	–	0.17	0.26	–
6.	Mg	0.55	0.1	0.25	–	0.11
7.	Al	4.75	0.37	–	0.21	–
8.	Si	0.31	0.48	0.22	0.47	0.88
9.	S	3.33	1.29	0.28	0.1	0.1
10.	Cl	0.28	0.46	0.88	0.49	0.54
11.	K	4.59	1.28	0.29	0.26	0.35



Correlation between MICROTOPS-II AOD (500 nm) and mass concentration of particulate matters

Figure 10 depicts the relationship between MICROTOPS—II AOD (500 nm) and particulate matter (PM₁₀, PM_{2.5}, and PM₁). AOD data were chosen for the day when PM data were available. Sometimes, AOD was not available for

the day on which PM data were available. So, one-to-one corresponding data are only correlated. A total of 37 samples of PM₁₀ and PM_{2.5} have been chosen, whereas 21 samples of PM₁ have been chosen for this correlation. PM₁₀ has a positive correlation with $R^2 = 0.42$, PM_{2.5} has a positive correlation with $R^2 = 0.44$, and PM₁ has a positive correlation with $R^2 = 0.43$, respectively. Concentrations of particulate matter have a significant effect on AOD since an increment

in the concentration of PM leads to an enhancement of AOD, which suggests a positive relationship. Similarly, during the monsoon season, when particles are washed out of the atmosphere, the PM concentrations decreased, and hence, the AOD also decreased.

Elemental composition of particulate matter

The elemental analysis of PM is also a significant factor that impacts the scattering nature of particulates. The various constituents' weight percentages of PM_{2.5} samples were measured throughout the investigation, which is summarized in Table 2. We have studied the chemical concentration of five PM_{2.5} samples. The order of these compositions is F > O > C > N > Al > K > S > Mg > Si > Cl > Na, C > F > O > S > K > Si > Cl > Al > Mg > Na, F > C > O > N > Cl > K > S > Mg > Si > Na > Al, F > C > O > N > Cl > Si > Na > K > Al > S, and F > C > O > N > Si > Cl > K > Mg > S > Na > Al in October 2019, November 2019, December 2019, January 2020, and February 2020, respectively, as shown in Figure 11. These samples are found to be rich in F, O, C, and N. The concentration of fluorine is higher in all months except November. The use of firecrackers during Diwali may contribute to carbon emissions. Carbon is the second highest contributor to PM_{2.5} samples during the winter season due to crop residual burning and anthropogenic activities. The lesser contributions of Mg, Si, Na, S, and Al have been found during the study period due to other anthropogenic activities (Liu et al., 2019).

Source identification of particulate matter and their transportation

The National Oceanic and Atmospheric Administration (NOAA) Hybrid Single Particle Lagrangian Integrated Trajectories (HYSPPLIT) model, which uses NCEP reanalysis wind data as input, computes 5 days of air mass back trajectories at three different altitudes of 500, 1,000, and 1,500 m above ground level at the receptor site (Draxler and Rolph, 2003). Back trajectories are crucial for identifying source regions, transport channels, and aerosol characteristics. About 5-day air mass back trajectory for nine selected days during the study period on 30 April 2019, 30 May 2019, 09 June 2019, 28 October 2019, 15 November 2019, 31 December 2019, 12 January 2020, 11 February 2020, and 10 March 2020 is shown in Figure 12. On the 30th of April 2019, air masses appeared to be coming from Saudi Arabia and Pakistan at high altitudes (1,500 and 1,000 m) and from Rajasthan at lower altitudes. On 30 May

2019, air masses seemed to come from high-altitude regions such as Russia, Kazakhstan, and Turkey. On 9 June 2019, an air mass at a high altitude comes from Georgia and the area around Kazakhstan. Before it gets to the sampling site at a low altitude, it goes through the Bay of Bengal. On the 28th of October 2019, the high-altitude wind comes from the Caspian Sea and Pakistan, while Uttar Pradesh is the source region for low-level wind. On 15 November 2019, all three-level winds (1,500, 1,000, and 500 m) came from Pakistan. On 31 December 2019, high-altitude wind at 1,500 m comes from Kazakhstan, 1,000 m wind comes from China, and low-level wind comes from Madhya Pradesh. On the 12th of January 2020, air masses appeared to be coming from Iraq and Saudi Arabia at high altitudes (1,500 and 1,000 m) and from Uzbekistan at lower altitudes. On 11 February 2020, the high-altitude wind seems to come from Nepal; the wind at 1,000 m comes from Rajasthan, and the low-level wind seems to come from Kazakhstan. On 10 March 2020, a low-altitude air mass arrives from Georgia, while high-altitude wind arrives from the Arabian Gulf and Syria.

Conclusions

MICROTOS-2 AOD (500 nm) and particulate matters (PM₁, PM_{2.5}, and PM₁₀) over Varanasi, from April 2019 to March 2020, were observed and analyzed daily, monthly as well as seasonally. Elemental analysis of five samples was also done through SEM-EDX analysis. The primary findings and outcomes of the study are presented here:

During the given period, mass concentrations of PM₁, PM_{2.5}, and PM₁₀ were found to be higher as compared to the NAAQS standard limit. In addition, mass concentration was found to be higher in the winter season while lower in the summer season. By calculating the ratio of PM₁ and PM_{2.5} to PM₁₀, the dominance of fine-mode particles in the winter season due to crop residue burning and vehicular emissions was found. While in the summer season, dominance of coarse-mode particles was found due to dust storms.

Maximum AOD is obtained at a low wavelength while AOD is found to be minimum at a higher wavelength, which shows the presence of a greater concentration of fine particles than coarse particles. The seasonal variation in AOD was found to be maximum during the post-monsoon season and minimum during the monsoon season. Higher AOD at a higher wavelength shows the dominance of coarse-mode particles during the summer season.

Aerosol optical depth produced by the MODIS Terra satellite and AOD measured by MICROTOS II have a good correlation with $R^2 = 0.5986$. The mean values of MODIS AOD and MICROTOS AOD were 0.72 and 0.97, respectively. The correlation between AOD and particulate

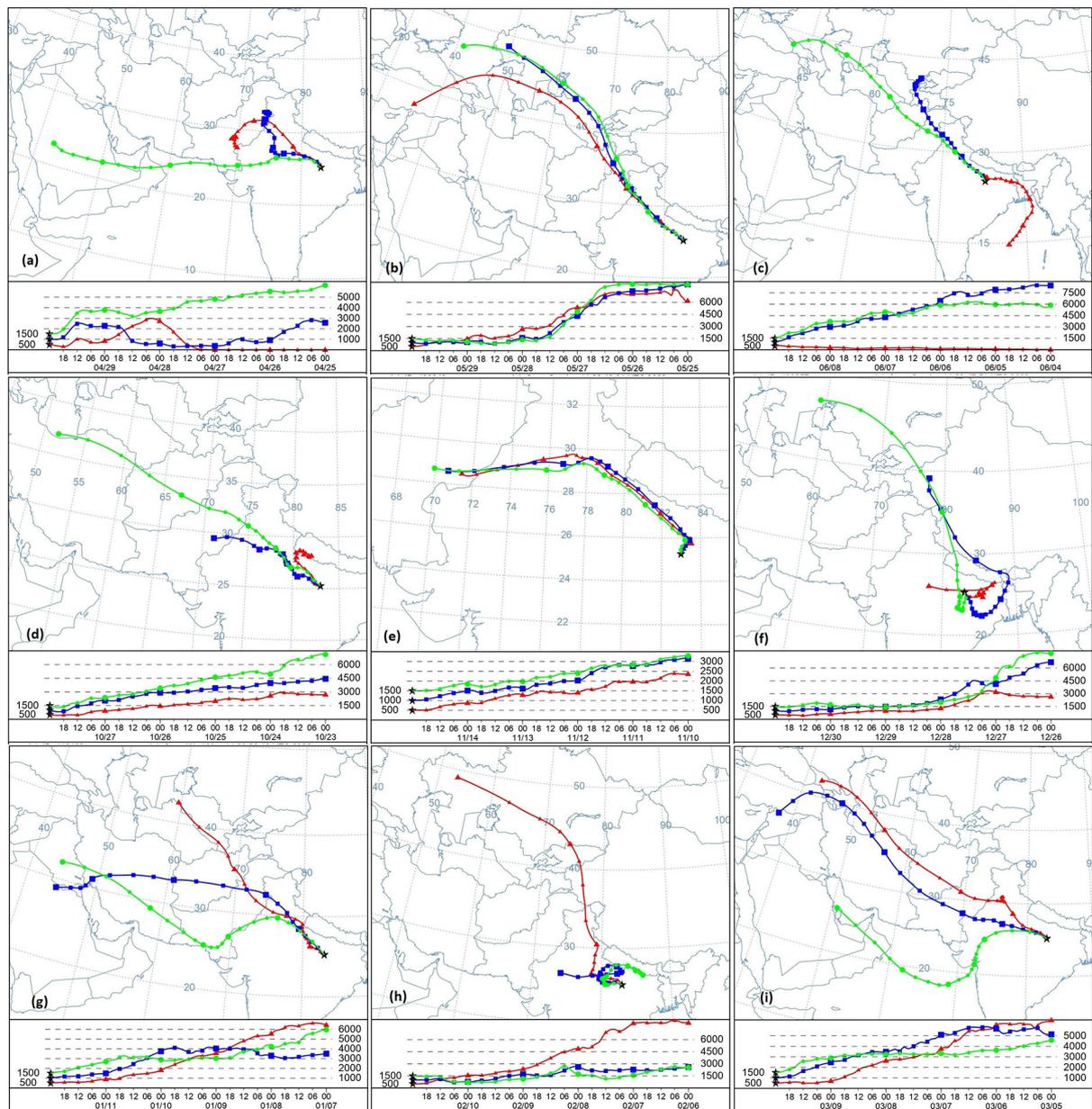


FIGURE 12 About 5 days back trajectories derived from HYSPLIT model for Varanasi for nine typical days having large aerosol loading on (a) 30 April 2019, (b) 30 May 2019, (c) 09 June 2019, (d) 28 October 2019, (e) 15 November 2019, (f) 31 December 2019, (g) 12 January 2020, (h) 11 February 2020, (i) 10 March 2020.

matter (PM₁₀, PM_{2.5}, and PM₁) was found to be positive and very good.

The dominance of different elements such as F, C, O, and N occurred over five samples (Teotia and Teotia, 1994) indicates that fluorine (F) is the highest contributor to PM_{2.5} particles except in November, indicating the impact of continental crust and soil contamination. Because of burning biomass in the winter, the concentration of carbon was the highest in the

sample in November, and it was the second highest for the other 4 months.

The analysis of the backward trajectory of the air mass reveals that there are multiple routes for the conveyance of air masses originating from various source locations on different days, which indicates the presence of a variety of aerosols over Varanasi (Singh et al., 2014). Fine particles are a major concern for human health. So, a higher

concentration of PM₁ particles may lead to heavy health risks in the future.

Data availability statement

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

Author contributions

PKC: writing, original draft preparation, analysis of data, and plotting. AK: visualizing, conceptualizing, and analysis and editing. VP: conceptualizing and plotting. AKS: reviewing, supervising, and editing. All authors contributed to the article and approved the submitted version.

Funding

The work was partially supported by the Institute of Eminence (IoE) to BHU (Scheme no: 6031).

References

- Ångström, A. (1964). The parameters of atmospheric turbidity. *Tellus* 16, 64–75. doi: 10.3402/tellusa.v16i1.8885
- Akinlade, G. O., Olaniji, H. B., Olise, F. S., Owoade, O. K., Almeida, S. M., Almeida-Silva, M., et al. (2015). Spatial and temporal variations of the particulate size distribution and chemical composition over Ibadan, Nigeria. *Environ. Monit. Assess.* 187, 1–14. doi: 10.1007/s10661-015-4755-4
- Alam, K., Trautmann, T., Blaschke, T., and Subhan, F. (2014). Changes in aerosol optical properties due to dust storms in the Middle East and Southwest Asia. *Remote Sens. Environ.* 143, 216–227. doi: 10.1016/j.rse.2013.12.021
- Ali, M. U., Liu, G., Yousaf, B., Ullah, H., Irshad, S., Ahmed, R., et al. (2019). Evaluation of floor-wise pollution status and deposition behavior of potentially toxic elements and nanoparticles in air conditioner dust during urbanistic development. *J. Hazard. Mater.* 365, 186–195. doi: 10.1016/j.jhazmat.2018.11.005
- Ancelet, T., Davy, P. K., and Trompeter, W. J. (2015). Hourly variations in particulate matter source contributions in New Zealand's southernmost city. *Air Qual. Clim. Change* 49, 26–33. doi: 10.3316/informit.466648262606092
- Balakrishnan, K., Dey, S., Gupta, T., Dhaliwal, R. S., Brauer, M., Cohen, A. J., et al. (2019). The impact of air pollution on deaths, disease burden, and life expectancy across the states of India: the Global Burden of Disease Study 2017. *Lancet Planet. Health.* 3, e26–e39. doi: 10.1016/S2542-5196(18)30261-4
- Banerjee, T., Murari, V., Kumar, M., and Raju, M. P. (2015). Source apportionment of airborne particulates through receptor modeling: Indian scenario. *Atmos. Res.* 164, 167–187. doi: 10.1016/j.atmosres.2015.04.017
- Bansal, O., Singh, A., and Singh, D. (2019). Short-term perturbation in aerosol characteristics over Northwestern India: a case study during Diwali festival. *J. Earth Syst. Sci.* 128, 1–13. doi: 10.1007/s12040-019-1223-5
- Cheng, Z., Jiang, J., Fajardo, O., Wang, S., and Hao, J. (2013). Characteristics and health impacts of particulate matter pollution in China (2001–2011). *Atmos. Environ.* 65, 186–194. doi: 10.1016/j.atmosenv.2012.10.022
- Chowdhury, S., and Dey, S. (2016). Cause-specific premature death from ambient PM_{2.5} exposure in India: estimate adjusted for baseline mortality. *Environ. Int.* 91, 283–290. doi: 10.1016/j.envint.2016.03.004
- Cohen, A. J., Brauer, M., Burnett, R., Anderson, H. R., Frostad, J., Estep, K., et al. (2017). Estimates and 25-year trends of the global burden of disease attributable to ambient air pollution: an analysis of data from the Global Burden of Diseases Study 2015. *Lancet.* 389, 1907–1918. doi: 10.1016/S0140-6736(17)30505-6
- Delfino, R. J., Sioutas, C., and Malik, S. (2005). Potential role of ultrafine particles in associations between airborne particle mass and cardiovascular health. *Environ. Health Perspect.* 113, 934–946. doi: 10.1289/ehp.7938
- Dey, S., and Di Girolamo, L. (2010). A climatology of aerosol optical and microphysical properties over the Indian subcontinent from 9 years (2000–2008) of Multiangle Imaging Spectroradiometer (MISR) data. *J. Geophys. Res. Atmos.* 115:D15204. doi: 10.1029/2009JD013395
- Dey, S., Di Girolamo, L., van Donkelaar, A., Tripathi, S. N., Gupta, T., and Mohan, M. (2012). Variability of outdoor fine particulate (PM_{2.5}) concentration in the Indian Subcontinent: a remote sensing approach. *Remote Sens. Environ.* 127, 153–161. doi: 10.1016/j.rse.2012.08.021
- Dey, S., Tripathi, S. N., Singh, R. P., and Holben, B. N. (2004). Influence of dust storms on the aerosol optical properties over the Indo-Gangetic basin. *J. Geophys. Res. Atmos.* 109:D20211. doi: 10.1029/2004JD004924
- Draxler, R. R., and Rolph, G. D. (2003). *HYSPLIT (HYbrid Single-Particle Lagrangian Integrated Trajectory)*. Model access via NOAA ARL READY. NOAA Air Resources Laboratory, Silver Spring, MD. Available online at: <http://www.arl.noaa.gov/ready/hysplit4.HTML>
- Egondi, T., Ettarh, R., Kyobutungi, C., Ng, N., and Rocklöv, J. (2018). Exposure to outdoor particles (PM_{2.5}) and associated child morbidity and mortality in socially deprived neighborhoods of Nairobi, Kenya. *Atmosphere* 9:351. doi: 10.3390/atmos9090351
- Elbayoumi, M., Ramli, N. A., Yusof, N. F. F. M., and Al Madhoun, W. (2013). Spatial and seasonal variation of particulate matter (PM₁₀ and PM_{2.5}) in Middle Eastern classrooms. *Atmos. Environ.* 80, 389–397. doi: 10.1016/j.atmosenv.2013.07.067
- Garaga, R., Sahu, S. K., and Kota, S. H. (2018). A review of air quality modeling studies in India: local and regional scale. *Curr. Pollut. Rep.* 4, 59–73. doi: 10.1007/s40726-018-0081-0

Acknowledgments

PKC is thankful to UGC, New Delhi for providing JRF.

Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

The handling editor ST declared a past co-authorship with one of the authors AKS.

Publisher's note

All claims expressed in this article are solely those of the authors and do not necessarily represent those of their affiliated organizations, or those of the publisher, the editors and the reviewers. Any product that may be evaluated in this article, or claim that may be made by its manufacturer, is not guaranteed or endorsed by the publisher.

- Ghude, S. D., Chate, D. M., Jena, C., Beig, G., Kumar, R., Barth, M. C., et al. (2016). Premature mortality in India due to PM_{2.5} and ozone exposure. *Geophys. Res. Lett.* 43, 4650–4658. doi: 10.1002/2016GL068949
- Gupta, L., Dev, R., Zaidi, K., Sunder Raman, R., Habib, G., and Ghosh, B. (2021). Assessment of PM₁₀ and PM_{2.5} over Ghaziabad, an industrial city in the Indo-Gangetic Plain: spatio-temporal variability and associated health effects. *Environ. Monitor. Assess.* 193, 1–21. doi: 10.1007/s10661-021-09411-5
- Guttikunda, S. K., and Calori, G. (2013). A GIS-based emissions inventory at 1 km × 1 km spatial resolution for air pollution analysis in Delhi, India. *Atmos. Environ.* 67, 101–111. doi: 10.1016/j.atmosenv.2012.10.040
- Jain, S., Sharma, S. K., Srivastava, M. K., Chatterjee, A., Singh, R. K., Saxena, M., et al. (2019). Source apportionment of PM₁₀ over three tropical urban atmospheres at Indo-Gangetic Plain of India: an approach using different receptor models. *Arch. Environ. Contam. Toxicol.* 76, 114–128. doi: 10.1007/s00244-018-0572-4
- Jain, S., Sharma, S. K., Srivastava, M. K., Chatterjee, A., Vijayan, N., Tripathy, S. S., et al. (2021). Chemical characterization, source apportionment, and transport pathways of PM_{2.5} and PM₁₀ over Indo Gangetic Plain of India. *Urban Clim.* 36:100805. doi: 10.1016/j.uclim.2021.100805
- Kaskaoutis, D. G., Kumar, S., Sharma, D., Singh, R. P., Kharol, S. K., Sharma, M., et al. (2014). Effects of crop residue burning on aerosol properties, plume characteristics, and long-range transport over northern India. *J. Geophys. Res. Atmos.* 119, 5424–5444. doi: 10.1002/2013JD021357
- Kaufman, Y. J., Boucher, O., Tanré, D., Chin, M., Remer, L. A., and Takemura, T. (2005). Aerosol anthropogenic component estimated from satellite data. *Geophys. Res. Lett.* 32:L17804. doi: 10.1029/2005GL023125
- Kaufman, Y. J., Tanré, D., and Boucher, O. (2002). A satellite view of aerosols in the climate system. *Nature* 419, 215–223. doi: 10.1038/nature01091
- Kompalli, S. K., Babu, S. S., and Moorthi, K. K. (2010). Inter-comparison of aerosol optical depth from the multi-wavelength solar radiometer with other radiometric measurements. *Indian J. Radio Spa. Phys.* 39, 364–371.
- Kumar, A., Kumar, S., Pratap, V., and Singh, A. K. (2021). Performance of water vapor retrieval from MODIS and ECMWF and their validation with ground-based GPS measurements over Varanasi. *J. Earth Syst. Sci.* 130, 1–11. doi: 10.1007/s12040-020-01529-3
- Kumar, M., Singh, R. K., Murari, V., Singh, A. K., Singh, R. S., and Banerjee, T. (2016). Fireworks induced particle pollution: a spatio-temporal analysis. *Atmos. Res.* 180, 78–91. doi: 10.1016/j.atmosres.2016.05.014
- Kumar, M., Tiwari, S., Murari, V., Singh, A. K., and Banerjee, T. (2015a). Wintertime characteristics of aerosols at middle Indo-Gangetic Plain: impacts of regional meteorology and long-range transport. *Atmos. Environ.* 104, 162–175. doi: 10.1016/j.atmosenv.2015.01.014
- Kumar, P., Pratap, V., Kumar, A., Choudhary, A., Prasad, R., Shukla, A., et al. (2020). Assessment of atmospheric aerosols over Varanasi: physical, optical and chemical properties and meteorological implications. *J. Atmos. Solar Terrestrial Phys.* 209:105424. doi: 10.1016/j.jastp.2020.105424
- Kumar, S., Kumar, S., Kaskaoutis, D. G., Singh, R. P., Singh, R. K., Mishra, A. K., et al. (2015b). Meteorological, atmospheric and climatic perturbations during major dust storms over Indo-Gangetic Basin. *Aeolian Res.* 17, 15–31. doi: 10.1016/j.aeolia.2015.01.006
- Li, C., Hu, Y., Zhang, F., Chen, J., Ma, Z., Ye, X., Yang, X., Wang, L., Tang, X., Zhang, R., and Mu, M. (2017). Multi-pollutant emissions from the burning of major agricultural residues in China and the related health-economic effects. *Atmos. Chem. Phys.* 17, 4957–4988. doi: 10.5194/acp-17-4957-2017
- Li, Q., Wu, B., Liu, J., Zhang, H., Cai, X., and Song, Y. (2020). Characteristics of the atmospheric boundary layer and its relation with PM_{2.5} during haze episodes in winter in the North China Plain. *Atmos. Environ.* 223:117265. doi: 10.1016/j.atmosenv.2020.117265
- Liu, K., Wang, F., Li, J., Tiwari, S., and Chen, B. (2019). Assessment of trends and emission sources of heavy metals from the soil sediments near the Bohai Bay. *Environ. Sci. Pollut. Res.* 26, 29095–29109. doi: 10.1007/s11356-019-06130-w
- Lodhi, N. K., Beegum, S. N., Singh, S., and Kumar, K. (2013). Aerosol climatology at Delhi in the western Indo-Gangetic Plain: microphysics, long-term trends, and source strengths. *J. Geophys. Res. Atmos.* 118, 1361–1375. doi: 10.1002/jgrd.50165
- Malm, W. C., and Day, D. E. (2000). Optical properties of aerosols at Grand Canyon national park. *Atmos. Environ.* 34, 3373–3391. doi: 10.1016/S1352-2310(00)00108-4
- Massie, S. T., Torres, O., and Smith, S. J. (2004). Total Ozone Mapping Spectrometer (TOMS) observations of increases in Asian aerosol in winter from 1979 to 2000. *J. Geophys. Res. Atmos.* 109:D18211. doi: 10.1029/2004JD004620
- Miri, M., Derakhshan, Z., Allahabadi, A., Ahmadi, E., Conti, G. O., Ferrante, M., et al. (2016). Mortality and morbidity due to exposure to outdoor air pollution in Mashhad metropolis, Iran. The AirQ model approach. *Environ. Res.* 151, 451–457. doi: 10.1016/j.envres.2016.07.039
- Miri, M., Ebrahimi Aval, H., Ehrampoush, M. H., Mohammadi, A., Toolabi, A., Nikonahad, A., et al. (2017). Human health impact assessment of exposure to particulate matter: an AirQ software modeling. *Environ. Sci. Pollut. Res.* 24, 16513–16519. doi: 10.1007/s11356-017-9189-9
- Morys, M., Mims, F. M. III, Hagerup, S., Anderson, S. E., Baker, A., Kia, J., et al. (2001). Design, calibration, and performance of MICROTOPS II handheld ozone monitor and Sun photometer. *J. Geophys. Res. Atmos.* 106, 14573–14582. doi: 10.1029/2001JD900103
- Mukherjee, A., and Agrawal, M. (2017). World air particulate matter: sources, distribution and health effects. *Environ. Chem. Lett.* 15, 283–309. doi: 10.1007/s10311-017-0611-9
- Murari, V., Kumar, M., Barman, S. C., and Banerjee, T. (2015). Temporal variability of MODIS aerosol optical depth and chemical characterization of airborne particulates in Varanasi, India. *Environ. Sci. Pollut. Res.* 22, 1329–1343. doi: 10.1007/s11356-014-3418-2
- Murari, V., Kumar, M., Mhawish, A., Barman, S. C., and Banerjee, T. (2017). Airborne particulate in Varanasi over middle Indo-Gangetic Plain: variation in particulate types and meteorological influences. *Environ. Monitor. Assess.* 189, 1–15. doi: 10.1007/s10661-017-5859-9
- Murari, V., Kumar, M., Singh, N., and Banerjee, T. (2016). Particulate morphology and elemental characteristics: variability at middle Indo-Gangetic Plain. *J. Atmos. Chem.* 73, 165–179. doi: 10.1007/s10874-015-9321-5
- Myhre, G. (2009). Consistency between satellite-derived and modeled estimates of the direct aerosol effect. *Science* 325, 187–190. doi: 10.1126/science.1174461
- Obaidullah, M., Dyakov, I. V., Peeters, L., Bram, S., and De Ruyck, J. (2012). Measurements of particle concentrations and size distributions in three parking garages. *Int. J. Energy Environ.* 6, 508–515.
- Ojha, N., Sharma, A., Kumar, M., Girach, I., Ansari, T. U., Sharma, S. K., et al. (2020). On the widespread enhancement in fine particulate matter across the Indo-Gangetic Plain towards winter. *Sci. Rep.* 10, 1–9. doi: 10.1038/s41598-020-62710-8
- Pope, I. I. C. A., Renlund, D. G., Kfoury, A. G., May, H. T., and Horne, B. D. (2008). Relation of heart failure hospitalization to exposure to fine particulate air pollution. *Am. J. Card.* 102, 1230–1234. doi: 10.1016/j.amjcard.2008.06.044
- Prasad, A. K., and Singh, R. P. (2009). Validation of MODIS Terra, AIRS, NCEP/DOE AMIP-II Reanalysis-2, and AERONET Sun photometer derived integrated precipitable water vapor using ground-based GPS receivers over India. *J. Geophys. Res. Atmos.* 114:D05107. doi: 10.1029/2008JD011230
- Pratap, V., Kumar, A., Kumar, P., and Singh, A. K. (2020b). “Pre-monsoon study of aerosol parameters and particulate matters over Varanasi for 2017,” in *2020 URSI Regional Conference on Radio Science (URSI-RCRS)* (Varanasi: IEEE), 1–2.
- Pratap, V., Kumar, A., Tiwari, S., Kumar, P., Tripathi, A. K., and Singh, A. K. (2020a). Chemical characteristics of particulate matter and their emission sources over Varanasi during the winter season. *J. Atmos. Chem.* 77, 83–99. doi: 10.1007/s10874-020-09405-6
- Rahman, M. H., Sharma, V. P., Kundu, S., and Datta, A. (2020). Seasonal variation of potential source locations of atmospheric particulates over the Indo-Gangetic Plain of India. *Open J. Air Pollut.* 9:1. doi: 10.4236/ojap.2020.91001
- Rajesh, T. A., and Ramachandran, S. (2017). Characteristics and source apportionment of black carbon aerosols over an urban site. *Environ. Sci. Pollut. Res.* 24, 8411–8424. doi: 10.1007/s11356-017-8453-3
- Rajput, P., Izhar, S., and Gupta, T. (2019). Deposition modeling of ambient aerosols in human respiratory system: health implication of fine particles penetration into the pulmonary region. *Atmos. Pollut. Res.* 10, 334–343. doi: 10.1016/j.apr.2018.08.013
- Ram, K., and Sarin, M. M. (2011). Day-night variability of EC, OC, WSOC and inorganic ions in the urban environment of Indo-Gangetic Plain: implications to secondary aerosol formation. *Atmos. Environ.* 45, 460–468. doi: 10.1016/j.atmosenv.2010.09.055
- Ramachandran, S., and Kedia, S. (2010). Black carbon aerosols over an urban region: radiative forcing and climate impact. *J. Geophys. Res. Atmos.* 115:D10202. doi: 10.1029/2009JD013560
- Reddy, B. S. K., Kumar, K. R., Balakrishnaiah, G., Gopal, K. R., Reddy, R. R., Reddy, L. S. S., et al. (2011). Aerosol climatology over an urban site, Tirupati (India) derived from columnar and surface measurements: first-time results obtained from a 30-day campaign. *J. Atmos. Solar Terrestrial Phys.* 73, 1727–1738. doi: 10.1016/j.jastp.2011.03.015
- Remer, L. A., Kaufman, Y. J., Tanré, D., Mattoo, S., Chu, D. A., Martins, J. V., et al. (2005). The MODIS aerosol algorithm, products, and validation. *J. Atmos. Sci.* 62, 947–973. doi: 10.1175/JAS3385.1

- Sah, D., Verma, P. K., Kandikonda, M. K., and Lakhani, A. (2019). Pollution characteristics, human health risk through multiple exposure pathways, and source apportionment of heavy metals in PM10 at Indo-Gangetic site. *Urban Clim.* 27, 149–162. doi: 10.1016/j.uclim.2018.11.010
- Salma, I., Balásházy, I., Hofmann, W., and Zárny, G. (2002). Effect of physical exertion on the deposition of urban aerosols in the human respiratory system. *J. Aerosol Sci.* 33, 983–997. doi: 10.1016/S0021-8502(02)00051-4
- Sarkar, S., Singh, R. P., and Chauhan, A. (2018). Crop residue burning in northern India: an increasing threat to Greater India. *J. Geophys. Res. Atmos.* 123, 6920–6934. doi: 10.1029/2018JD028428
- Satheesh, S. K., Krishna Moorthy, K., Kaufman, Y. J., and Takemura, T. (2006). Aerosol optical depth, physical properties, and radiative forcing over the Arabian Sea. *Meteorol. Atmos. Phys.* 91, 45–62. doi: 10.1007/s00703-004-0097-4
- Sawłani, R., Agnihotri, R., Sharma, C., Patra, P. K., Dimri, A. P., Ram, K., et al. (2019). The severe Delhi SMOG of 2016: a case of delayed crop residue burning, coincident firecracker emissions, and atypical meteorology. *Atmos. Pollut. Res.* 10, 868–879. doi: 10.1016/j.apr.2018.12.015
- Saxena, M., Sharma, A., Sen, A., Saxena, P., Mandal, T. K., Sharma, S. K., et al. (2017). Water-soluble inorganic species of PM10 and PM2.5 at an urban site of Delhi, India: seasonal variability and sources. *Atmos. Res.* 184, 112–125. doi: 10.1016/j.atmosres.2016.10.005
- Seinfeld, J. H., Bretherton, C., Carslaw, K. S., Coe, H., DeMott, P. J., Dunlea, E. J., et al. (2016). Improving our fundamental understanding of the role of aerosol-cloud interactions in the climate system. *Proc. Natl. Acad. Sci. U.S.A.* 113, 5781–5790. doi: 10.1073/pnas.1514043113
- Sharma, D., Singh, D., and Kaskaoutis, D. G. (2012). Impact of two intense dust storms on aerosol characteristics and radiative forcing over Patiala, Northwestern India. *Adv. Meteorol.* 2012:956814. doi: 10.1155/2012/956814
- Singh, A., Rastogi, N., Patel, A., Satish, R. V., and Singh, D. (2016a). Size-segregated characteristics of carbonaceous aerosols over the Northwestern Indo-Gangetic Plain: year-round temporal behavior. *Aerosol. Air Qual. Res.* 16, 1615–1624. doi: 10.4209/aaqr.2016.01.0023
- Singh, A., Srivastava, A. K., Varaprasad, V., Kumar, S., Pathak, V., and Shukla, A. K. (2021). Assessment of near-surface air pollutants at an urban station over the central Indo-Gangetic Basin: role of pollution transport pathways. *Meteorol. Atmos. Phys.* 133, 1127–1142. doi: 10.1007/s00703-021-00798-x
- Singh, A., Tiwari, S., Sharma, D., Singh, D., Tiwari, S., Srivastava, A. K., et al. (2016b). Characterization and radiative impact of dust aerosols over the northwestern part of India: a case study during a severe dust storm. *Meteorol. Atmos. Phys.* 128, 779–792. doi: 10.1007/s00703-016-0445-1
- Singh, A. K., Srivastava, M. K., Singh, M., Srivastava, A. K., Kumar, S., Tiwari, S., et al. (2014). Characterization of atmospheric aerosol by SEM-EDX and Ion-chromatography techniques for eastern Indo-Gangetic plain location, Varanasi, India. *Int. J. Adv. Earth Sci.* 3, 41–51.
- Singh, D. K., and Gupta, T. (2016). Role of transition metals with water soluble organic carbon in the formation of secondary organic aerosol and metallo-organics in PM1 sampled during post monsoon and pre-winter time. *J. Aerosol. Sci.* 94, 56–69. doi: 10.1016/j.jaerosci.2016.01.002
- Smith, C. L., Moores, J. E., Lemmon, M., Guzewich, S. D., Moore, C. A., Ellison, D., et al. (2019). Visibility and line-of-sight extinction estimate in Gale crater during the 2018/MY34 global dust storm. *Geophys. Res. Lett.* 46, 9414–9421. doi: 10.1029/2019GL083788
- Spandana, B., Rao, S. S., Upadhy, A. R., Kulkarni, P., and Sreekanth, V. (2021). PM2.5/PM10 ratio characteristics over urban sites of India. *Adv. Space Res.* 67, 3134–3146. doi: 10.1016/j.asr.2021.02.008
- Speranza, A., Caggiano, R., Margiotta, S., Summa, V., and Trippetta, S. (2016). A clustering approach based on a triangular diagram to study the seasonal variability of simultaneous measurements of PM10, PM2.5, and PM1 mass concentration ratios. *Arab. J. Geosci.* 9, 1–8. doi: 10.1007/s12517-015-2158-z
- Srivastava, A. K., Singh, S., Tiwari, S., Kanawade, V. P., and Bisht, D. S. (2012a). Variation between near-surface and columnar aerosol characteristics during the winter and summer at Delhi in the Indo-Gangetic Basin. *J. Atmos. Sol. Terr. Phys.* 77, 57–66. doi: 10.1016/j.jastp.2011.11.009
- Srivastava, A. K., Soni, V. K., Singh, S., Kanawade, V. P., Singh, N., Tiwari, S., et al. (2014). An early South Asian dust storm during March 2012 and its impacts on ground-based Himalayan foothills: a case study. *Sci. Total Environ.* 493, 526–534. doi: 10.1016/j.scitotenv.2014.06.024
- Srivastava, A. K., Tripathi, S. N., Dey, S., Kanawade, V. P., and Tiwari, S. (2012b). Inferring aerosol types over the Indo-Gangetic Basin from ground-based sunphotometer measurements. *Atmos. Res.* 109, 64–75. doi: 10.1016/j.atmosres.2012.02.010
- Taneja, K., Ahmad, S., Ahmad, K., and Attri, S. D. (2020). Impact assessment of a severe dust storm on atmospheric aerosols over an urban site in India. *Curr. Sci.* 118:737. doi: 10.18520/cs/v118/i5/737-749
- Teotia, S. P. S., and Teotia, M. (1994). Dental caries: a disorder of high fluoride and low dietary calcium interactions (30 years of personal research). *Fluoride* 27, 59–59.
- Tiwari, S., Kaskaoutis, D., Soni, V. K., Dev Attri, S., and Singh, A. K. (2018). Aerosol columnar characteristics and their heterogeneous nature over Varanasi, in the central Ganges valley. *Environ. Sci. Pollut. Res.* 25, 24726–24745. doi: 10.1007/s11356-018-2502-4
- Tiwari, S., Kumar, A., Pratap, V., and Singh, A. K. (2019). Assessment of two intense dust storm characteristics over Indo-Gangetic basin and their radiative impacts: a case study. *Atmos. Res.* 228, 23–40. doi: 10.1016/j.atmosres.2019.05.011
- Tiwari, S., Mishra, A. K., and Singh, A. K. (2016). Aerosol climatology over the Bay of Bengal and the Arabian Sea inferred from space-borne radiometers and lidar observations. *Aerosol Air Qual. Res.* 16, 2855–2868. doi: 10.4209/aaqr.2015.06.0406
- Tiwari, S., and Singh, A. K. (2013). Variability of aerosol parameters derived from ground and satellite measurements over Varanasi located in the Indo-Gangetic Basin. *Aerosol Air Qual. Res.* 13, 627–638. doi: 10.4209/aaqr.2012.06.0162
- Wang, M., and Penner, J. E. (2009). Aerosol indirect forcing in a global model with particle nucleation. *Atmos. Chem. Phys.* 9, 239–260. doi: 10.5194/acp-9-239-2009
- Xin, J., Zhang, Q., Wang, L., Gong, C., Wang, Y., Liu, Z., et al. (2014). The empirical relationship between the PM2.5 concentration and aerosol optical depth over the background of North China from 2009 to 2011. *Atmos. Res.* 138, 179–188. doi: 10.1016/j.atmosres.2013.11.001