

# Fish behaviour and welfare

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# Fish behaviour and welfare

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# Editorial: Fish behaviour and welfare

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## KEYWORDS

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## Editorial on the Research Topic Fish behaviour and welfare

World fish production from both aquaculture and fisheries increased in 2022, with a record number of around 223.2 million tons, from which 185.4 million tons consist of aquatic animals and 37.8 million tons are algae (1). Fish culture is an important source for many households in different parts of the world. One of its principal objectives is to effectively apply dry feed (pelleted feed) for promoting fish growth. In 2022, global aquaculture surpassed the 130.9 million tons, of which 94.4 million tons were aquatic animals, 51 percent of the total aquatic animal production. The principal source of fish, as one could see, is aquaculture. Number of fish farms has increased considerably so that the demand of higher population and the need for water related products are met. This has resulted in public concern on how to produce aquaculture products for human consumption, principally on the area of fish welfare. Fish (food) can be produced in an environmentally correct way without causing undue suffering to cultivated organisms. Captivity conditions must contribute to the health and welfare of each animal. Fish must be cared for by trained and experienced personnel, including providing veterinary care, and researchers must be strictly qualified and have sufficient experience and training before undertaking animal research (2). Different approach to different welfare topics of different species are presented in this book. All the different topics have in common the wellbeing of the fish that is being cultured for our benefit: Fish life cycle; from egg to slaughter: monitoring the welfare of Nile tilapia, *Oreochromis niloticus*, throughout its entire life cycle in aquaculture (Pedrazzani et al.). Welfare indicators under farm conditions: welfare indicators in Tilapia: an Epidemiological Approach (Flores-García et al.), and Qualitative Behavioral Assessments, a welfare indicator for farmed Atlantic Salmon (*Salmo salar*) in response to a stressful challenge (Wiese et al.). Reproduction: Use of male-to-female sex reversal as a welfare scoring system in the protandrous farmed gilthead sea bream (*Spaurus aurata*) (Holhorea et al.). Farm conditions: Experimental study on the effect of sound stimulation on hearing and behavior of juvenile black rockfish (*Sebastes schleglii*) (Wang et al.); The effects of aerator noise on the swimming, feeding, and growth of *Micropterus salmoides* (Zhang et al.); and Behavior analysis of juvenile steelhead trout under blue and red light color conditions based on multiple object tracking (Li et al.). Sedation: Does sedation with AQUI-S<sup>®</sup> mitigate transport stress and post transport mortality in ballan wrasse (*Labrus bergyltae*)? (Calabrese et al.). Killing methods: Farmed fish welfare



during slaughter in Italy: survey on stunning and killing methods and indicators of unconsciousness (Clemente et al.); and Humane slaughter in Mediterranean Sea bass and bream aquaculture: farm characteristics, stakeholder views, and policy implications (van Pelt et al.).

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## References

1. FAO (2024). *In Brief to The State of World Fisheries and Aquaculture 2024. Blue Transformation in Action*. Rome: FAO. doi: 10.4060/cd0690en
2. CCAC (2024). *Canadian Council on Animal Care. Redefining our view of animal welfare in science*. Available at: [https://ccac.ca/Documents/Standards/Guidelines/CCAC\\_guidelines-Categories\\_of\\_welfare\\_impact.pdf](https://ccac.ca/Documents/Standards/Guidelines/CCAC_guidelines-Categories_of_welfare_impact.pdf) (accessed August 26, 2024).

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# Welfare Indicators in Tilapia: An Epidemiological Approach

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Interest and concern about rearing methods and their impact on animal welfare have increased. Production evaluation is population-based, and animal welfare analysis should be similar. In fish, the most common welfare indicators are gill state, fin damage, and body condition. The objective of this study was to evaluate the feeding rate effect on the welfare indicators of *Oreochromis niloticus* using an epidemiological approach. Five growth stages (from 1.2 to 360 g) were studied using four feeding rates as treatments: underfeeding (80%), recommended feeding (100%), and two levels of overfeeding (120% and 140%). The evaluated welfare indicators include the presence of lesions in different body areas and fins, the decrease in body condition index, and their impact on biomass production. Incidence and relative risk were determined for each indicator. Statistically significant associations were found in the indicators of mortality, weight, body condition (K), and presence of evident damage in the caudal and anal fin in all stages. The results showed that the feed rate directly affects the welfare indicators and production. Mortality, weight reduction, K reduction, and caudal and anal fin damage incidence showed to be relevant indicators in all *O. niloticus* growing stages. As a result of this study, the epidemiological approach seems to be a valuable tool for production. A risk traffic light method is a proposal that could have great potential, with the suggested limits for WI's concerning the individuals present in the culture pond, allowing progressive evaluation and decision-making to correct risky situations.

**Keywords:** welfare indicator, *Oreochromis niloticus*, populational risk, incidence, epidemiological approach

## INTRODUCTION

The main goal of animal production is to obtain protein for human consumption. As in terrestrial animals, in fish, the diet must consider various factors such as age and growth stage, since incorrect management will cause individuals to have few opportunities to develop correctly, regardless of the species (1). The demand for food is growing as a response to the increase in the human population, as it is estimated that by the year 2050 it will reach around nine billion people (2); this implies an enormous challenge for the primary production sectors, which are increasingly under pressure to satisfy this need. Consumption of aquatic animal protein has increased 15% in the last 10 years; in 2018, aquaculture contributed 82.1 million tons of biomass for human consumption,

54.3 million of these came from fish, tilapia (*Oreochromis spp.*) being one of the most important species for freshwater aquaculture (2). On the other hand, the decrease in water availability means that animal production systems must be substantially modified, to make effective and efficient use of water resources and carry out sustainable production. Recirculating aquaculture systems (RAS) are frequent in semi-intensive and intensive aquaculture, based on a cyclical water movement. They consist of taking the water from a pond, passing it through filters, and returning it to the pond already clean. Such systems mean an enormous advantage in saving water, especially in regions where it is scarce. Physicochemical parameters do not have significant variations and prevent diseases from spreading to the entire production unit (3). This technique has been proposed as a sustainable alternative for the efficient use of water and the environmental impact reduction associated with aquaculture.

The public interest and concern about the raising methods and their impact on the welfare of production animals has increased globally, and fish production is no exception. Welfare is defined as the dynamic state of an individual concerning the biological mechanisms used to adapt positively and successfully to changes in the environment, involving health (4), comfort (5), and the emotional state of the animals (4–6). The farmer must be responsible to provide calmness, comfort, protection, and safety to the farmed animals, during their breeding, maintenance, production, transport, and slaughter (7). Like terrestrial production species, aquatic animals require specific management and growth conditions according to the species and life stage. To provide adequate welfare levels, fish farmers must observe, measure, and control various variables such as water quality, population density, and feeding practices (8). According to Mellor et al. (9), the five domains model for the evaluation of animal welfare describes the sum and interaction of the variables related to survival (1: nutrition, 2: environment, and 3: health) and the situational variables (4: behavior), it directly infers on the mental state of the individuals (domain no. 5), which, in turn, allows qualifying the welfare state of the animals at a given moment.

The behavior and welfare of fish have been the subject of debate for years. In 2002, the United Kingdom implemented laws on the management of salmon farm production to improve the living conditions of the animals, resulting in better-quality products (10). Norway in 2005 started regulating aquaculture production with guidelines like those implemented by the UK (11). Recent research has shown that fish welfare is strongly related to fish physiology, which impacts production and considers animal welfare as a key element for the expression of the full genetic production potential of farmed fish (8, 12, 13). Fish possess homeostatic mechanisms that allow individuals to adapt to their environment, through physiological changes both internal and external (14). To know the welfare state in a fish, welfare indicators (WIs) can be used. These can be determined directly on the animals, such as fins condition body deformations, or indirectly, which are mainly environmental conditions. Once WIs are used as standard on laboratories or farms, they become laboratory welfare indicators or operational welfare indicators (10). Most studies coincide that the best and most widely used

WIs are individual-based (or direct) welfare indicators, which can be determined both at the farm and laboratory level (10). The most common of these indicators are operculum beating rate, reflex behavior, gill status, condition factor, fin damage, and body integrity (13, 14). Group based welfare indicators most used are the mortality rate (15), swimming behavior (16), appetite (17), growth rate (18), presence of diseases (10), presence of scales or blood in the water (10), the state of the fins, the integrity of the body, and the body condition (19). Fins are anatomical structures that help in the mobility of the fish, therefore, the integrity of these – mainly the dorsal, lateral, and caudal fins – are indicators of health and welfare (10, 16, 19). In fish, the condition factor (K) is a well-accepted tool for assessing the nutritional status (18), overall quality (20), and feeding management (16, 18). Body condition is variable throughout the lives of fish; thus, it is difficult to define exact values that are indicative of reduced welfare. However,  $<0.9$  is usually indicative of emaciation (21). Feed management in aquaculture farms requires a significant amount of resources, and labor, consequently, represents an important production cost. The feeding rate (amount of feed supplied) is determined in relation to the biomass contained in a fish tank, cage, or pond (16). In the case of tilapia production, as well as in other farmed fish, the feeding is given using standardized feeding tables according to the growth stage (22). Likewise, the companies that manufacture balanced fish feed issue tables of feeding programs, where the suggested handling rate is indicated according to the weight of the organisms and the product. Feeding practices that affect fish welfare also include feeding schedules (23, 24). During production, erroneous management such as underfeeding and overfeeding can occur with negative effects on the welfare of fish, either due to lack of nutrients (25) or deterioration in water quality (25–27).

Epidemiology is the study of disease in populations and of factors that determine its occurrence, the keyword being populations (28). Veterinary epidemiology additionally includes research and assessment of other health-related events, notably productivity (29). All this research involves observing animal populations and making inferences from the observations (29). Epidemiological tools are useful when studying the general status of a certain group, establishing diagnostic criteria to carry out evaluations that allow the prevention, detection, correction, and control of problems, particularly health problems. Epidemiological indicators are calculations used to determine the exposure of a population to a disease or any damage, such as body areas with descaling, hemorrhages, or broken fins (lesion), that is, the probability of the presence of a specific event in a defined time. A cohort study (prospective) is based on the evaluation of the occurrence of an event (in terms of presence/absence) as a result of the follow-up over time of a group, as a consequence of having been exposed or not (comparison groups) to a certain exposure (risk factor) (28, 29). The analysis of the probability of an event occurrence, using epidemiological indicators such as *odds ratio*, to identify risk factors for the presence of bodily injuries and their impact on welfare, has recently been reported in terrestrial animals (30). The incidence (I) represents the number of cases (events) that appear in a population and in each period of time (28). The relative

**TABLE 1** | Experimental design: initial values of the number of fish per pond, weight, length, and K.

	Experiments				
	Fingerlings	Juvenile	On-growing 1	On-growing 2	On-growing 3
<b>Fish</b>					
n per pond	180	100	100	100	60
Weight (g) <sup>a</sup>	1.26 ± 0.03	9.26 ± 0.19	35.13 ± 0.82	66.89 ± 2.07	144.84 ± 10.14
Length (cm) <sup>a</sup>	3.86 ± 0.31	7.84 ± 0.07	12.00 ± 0.21	14.99 ± 0.22	19.14 ± 0.49
K <sup>1</sup>	2.32 ± 0.53	1.92 ± 0.02	2.04 ± 0.07	1.99 ± 0.04	2.07 ± 0.03
<b>Pond</b>					
Biomass weight (kg) <sup>b</sup>	0.22 ± 0.006	0.92 ± 0.02	3.51 ± 0.08	6.68 ± 0.20	8.78 ± 0.55
	2.91%	2.07%	2.32%	3.10%	6.32%
<b>Feeding rates treatments, %</b>					
Underfeeding	6.4	4.0	3.2	3.2	2.4
Control*	8.0	5.0	4.0	4.0	3.0
Overfeeding A	9.6	6.0	4.8	4.8	3.6
Overfeeding B	11.2	7.0	5.6	5.6	4.2
<b>Feed</b>					
Characteristics	DM: 95.04 CP: 49.18 GE: 4.51 PS: <0.35	DM: 95.26 PC: 47.66 GE: 4.84 PS: 1.5	DM: 94.21 PC: 44.38 GE: 4.36 PS: 2.4	DM: 94.04 PC: 40.27 GE: 4.14 PS: 3.5	DM: 92.47 PC: 36.51 GE: 3.99 PS: 4.8
Servings a day	6	5	5	5	4
<b>Duration</b>					
Days	24	27	21	29	60

<sup>a</sup>Values ± SD; <sup>b</sup>Values ± SD, % CV. DM, Dry Matter (%); PC, Protein Crude (%); GE, Gross energy (Kcal/g); PS, Particle Size (mm); balanced specific for the species. \*Recommended (22), feeding rate in relation to the pond biomass. Replicates per treatment = 3 ponds.

Initial biomass per pond, feeding rates treatments (%), feed characteristics and management, and duration (days) of each experiment.

risk (RR) is a measure of the relationship existing between the probability that an event occurs in the exposed group with the same risk factor. The RR is calculated from the cases (events) observed in the group of animals exposed to the risk factor in relation to the cases (events) observed in the group of animals not exposed to the risk factor. It is essential to calculate the corresponding confidence intervals (CIs), which allows for giving greater statistical weight to the calculated RR value. The CI values allow a correct interpretation of the RR result obtained because they approximate to the real value in the population under study since this is inaccessible but it is located inbetween the CI range, with a degree of uncertainty that we can determine (95%) (29, 31). When the RR is equal to or >1.0, then there is a negative effect on the risk factor for the incidence of the event; when it is <1.0, there is no such negative effect on the population (28, 31).

Aquaculture production is generally evaluated considering the population, losing the richness of individual values (8, 13, 17, 32). Therefore, the advantages of analyzing WIs with epidemiological statistical tools could allow measuring the state of a determined group of fish, advancing the knowledge of the animal welfare and the application of WIs in farms or laboratories. Mortality, body condition, damage to the eyes, mouth, opercula, skin, degree of scaling bleeding lesions, and damage such as tears, fraying or bleeding that affects the integrity of the dorsal, lateral, anal, and caudal fins have been part of proposals for the evaluation of welfare indicators in fish (12–14), which have been put into operational practice by applying in *Salmo salar* (10, 33), *Perca*

*fluviatilis* (34), and *Oreochromis niloticus* (35), along with other indicators such as water quality and production rates. Some studies report WIs in the proportion of the damage in an evaluated population (10, 33–35), but in these researches, an analysis is not carried out to determine if the degree of damage registered is considered a situation of population risk, as it would be done in analysis with epidemiological statistical tools. Therefore, the objective of this study was to evaluate the effect of the feeding rate on the welfare indicators of tilapia (*Oreochromis niloticus*) cultivated in recirculating aquaculture systems, using an epidemiological approach.

## MATERIALS AND METHODS

### Location

The study was carried out at the facilities of the Aquaculture Laboratory of the Veterinary and Zootechnical Department of the Life Sciences Division, Campus Irapuato-Salamanca of the University of Guanajuato, located according to Geo Locator (2021) at 20 ° 44'34.65"N and 101 ° 19'51.78"W, at 1,745 meters above sea level. The study was carried out from April to November 2019.

### Experimental Systems

Twelve individual and independent experimental conventional aquaculture recirculation systems (RAS) were used ( $n = 3$  per treatment), which were located inside a greenhouse and were



**TABLE 2 |** Description of the welfare indicator (index description) evaluated.

INCIDENCE		(nO/nTx)*100
Welfare Indicator (Index description)	Formula application	Visual
<b>Mortality incidence:</b> Number of dead fish	<b>nO:</b> no. of individuals from the treatment who died <b>nTx:</b> no. of individuals per treatment	Alive   Death 0   1
<b>Weight reduction incidence:</b> Number of fish weighing less than the average of the treatments Control	<b>nO:</b> no. of individuals in the treatment who presented lower body weight than the average of the control <b>nTx:</b> no. of individuals per treatment	0 Higher that 1 Lower that Average weight, control treatment
<b>K reduction incidence:</b> Number of fish with lower K than the initial one per pond	<b>nO:</b> No. of individuals in the treatment who presented lower K than the initial value <b>nTx:</b> no. of individuals per treatment	0 Higher that 1 Lower that Initial average K per pond
<b>Caudal/Anal fin damage incidence:</b> Number of fish with presence of damage to caudal and anal fins	<b>nO:</b> No. of treatment individuals who presented damage to the caudal and / or anal fin <b>nTx:</b> no. f individuals per treatment	Without damage   Damaged 0   1
Caudal fin without damage	Damaged caudal fin (numbers from ichthyometer)	Damaged caudal fin
		
Anal fin without damage	Hemorrhagic anal fin	Damaged anal fin
		

built based on the modified design of Timmons and Ebeling (3). At the beginning of each experiment, a total water change, and deep cleaning of the components were carried out. Each system consisted of a 1.5 m<sup>3</sup> fish tank with an integrated filter of four elements with a capacity of 0.2 m<sup>3</sup> each (1 settler-clarifier, 2 physical particle separations with filter material inside, and 1 biological with biospheres). Water lines were 2" hydraulic PVC pipes. The internal movement of water was carried out with

a submersible pump (RESUN Model SP3800, Q = 2,000 L/h) placed inside the biological filter. To maintain a constant and suitable temperature for the species, each tank was covered with a dome made of ¼" plastic and PVC, with a small opening. Air was injected at a rate of 40 L/min to each tank and biological filter, using aerating stones connected to a general distribution line to all systems and to a compressor (RESUN GF-750). The RAS were filled to their maximum capacity the same day with water from a

**TABLE 3 |** Suggested risk limits (%) and color for welfare indicator in relation to the individuals present in the culture pond.

*Welfare indicator %	No risk	Moderate risk	High risk
<b>Fingerlings, from 1 g</b>			
Mortality	<14	14–25	>25
Decreased body weight	<34	34–57	>57
Decreased body condition	<11	11–20	>20
Obvious damage to caudal fin	<6	6–12	>12
Obvious damage to anal fin	<8	8–15	>15
<b>Juveniles, from 10 g</b>			
Mortality	<4	4–5	>5
Decreased body weight	<35	35–60	>60
Decreased body condition	<11	11–21	>21
Obvious damage to caudal fin	<6	6–12	>12
Obvious damage to anal fin	<8	8–14	>14
<b>On-growing 1, from 30 g</b>			
Mortality	<2	2–13	>13
Decreased body weight	<47	47–58	>58
Decreased body condition	<24	24–53	>53
Obvious damage to caudal fin	<11	11–18	>18
Obvious damage to anal fin	<17	17–19	>19
<b>On-growing 2, from 65 g</b>			
Mortality	<2	2–4	>4
Decreased body weight	<50	50–57	>57
Decreased body condition	<34	34–38	>38
Obvious damage to caudal fin	<21	21–30	>30
Obvious damage to anal fin	<15	15–19	>19
<b>On-growing 3, from 130 g</b>			
Mortality	<2	2–7	>7
Decreased body weight	<36	36–66	>66
Decreased body condition	<45	45–53	>53
Obvious damage to caudal fin	<29	29–55	>55
Obvious damage to anal fin	<23	23–72	>72

\*Welfare Indicator was determined based on the percentage of affected individuals. The ranges of the risk traffic light were determined considering the lowest data (in whole numbers) of % I Tx, according to the stage and the evaluated welfare indicator (Tables 4–8); if this was too low then the next higher value was used. To consider the lower limit indicator in a situation of moderate risk, the data had to be associated with a RR lower than 0.85 and a present CI in ranges lower than 0.99. As for the upper limit indicator, it was calculated = (% I Tx lower limit \* RR constant) / RR, and said RR constant was determined as 0.999.

single well. The movement and oxygenation of the water began, after 24 hours, adding lyophilized bacteria (AZOO-NitriPro, Nitrosomonas, and Nitrobacter) at a rate of 3 g per 250 L of water, following the protocol described by Espinoza-Moya et al. (36).

## Fish and Food

The project was evaluated and approved by the Institutional Committee of Bioethics in Research of the University of Guanajuato (code: CIBIUG-A59-2020). The specimens of *O. niloticus* (males obtained by sexual reversion) were purchased from a commercial farm located in Chupícuaro, Guanajuato. The fish were transported with the farm's water, inside plastic

bags with oxygen injection, and placed in a reception tank (quarantine) for their acclimatization (14 days). Five experiments were carried out, in 6 months. Each growth stage of the tilapia (5 stages in total) was considered a separate experiment. The growth stages (experiments) according to the initial weight of the fish were in the following ranges: fingerlings (1.2–1.3 g), juveniles (9.0–9.5 g), on-growing 1 (34–37 g), on-growing 2 (65–70 g) and on-growing 3 (130–150 g). Fish size and density were considered for fish management (37). The experimental design: initial values of the number of fish per pond per experiment (n) and their characteristics (average initial values  $\pm$  SD of weight, length, and K), initial biomass per pond (average initial weight  $\pm$  SD and % CV), feeding rates treatments (%), feed characteristics and management (species-specific commercial balance), and duration (days) of each experiment can be seen in Table 1.

Data were obtained from a total of 6,480 tilapias. At the beginning and end of each growth stage total fish length was recorded using an ichthyometer (Pentair Aquatic Ecosystems Inc.). Wet live weight (g) was recorded using a digital scale (RHINO, model BAPRE-3). At the end of each experiment, the same measurements were recorded and the same equipment was used to obtain all animal-related data. Due to the high number of specimens used per treatment, metabolism was decreased (lethargy and tranquillization) using cold water to weigh, measure, and check the animals externally. For both initial and end experiment data collection we followed the same pathway. First, we proceeded individually to reduce the individual's metabolism with cold water, making a sudden change in the maintenance temperature to 5°C (where swimming pattern change is visible), this being a recommended method to calm and immobilize fish (38, 39). The data collection and photographs did not exceed 90 seconds for each animal. Immediately after this, tilapias were put in a container with air injection for their recovery, and awakened fish that showed normal respiration and swimming were transported to a community 10,000 L RAS pond (under similar management conditions to the control treatment). It should be noted that the same set of animals did not participate in successive experiments. Four feeding rates were used as treatments: underfeeding (Ufe, 80%), recommended feeding (22) (Control, 100%), and two levels of overfeeding (OfA: 120% and OfB: 140%). Balanced feed samples were taken to determine crude protein (CP) (AOAC) and gross energy (GE) by combustion (IKA Calorimeter System C 2000 Basic). To adjust the amount of feed to be supplied, a sample was taken each week from the animals which were weighed, mortality was also considered for each experimental system.

## Production Variables and Water Quality

For each pond, total biomass production (g) was determined as the final harvested weight minus the initial weight. Water quality was measured, at 9:00 am every other day, variables were temperature (°C), dissolved oxygen ( $\text{mg L}^{-1}$ ), electrical conductivity ( $\text{mS/cm}^3$ ),  $\text{NH}_4^+$  ( $\text{mg L}^{-1}$ ), and pH, using a multiparameter recorder (YSI Mod. Professional Plus, Cable Quatro). Measurements were carried out directly in the tank in the middle of the water column.

## Welfare Indicators (WIs) and Epidemiological Approach

A thorough visual inspection was performed to determine the presence or absence of visible damage to the eyes, mouth, opercula, skin, descaling, bleeding, and dorsal, lateral, anal, and caudal fin damage measured as tears, fraying, or bleeding (16) (without damage: 0, damaged: 1). Mortality (alive: 0, death: 1), weight reduction and body condition index (Fulton's K) were calculated (higher: 0, lower: 1) (Table 2).

$$\text{Mortality} = \text{no. of harvested fish} - \text{no. of fish stocked}$$

To determine the number of fish with an increase or decrease in weight or K, at the end of each experiment, per pond per treatment, the average value of control treatment per fish per pond was taken, as follows:

$$\text{Final weight per fish} - \bar{x} \text{ weight control treatment}$$

where the positive values are taken as an increase (0) and the negative values as a decrease (1) in weight per fish (Table 2).

$$\text{Fultons } K = (W/L^3)$$

where, W is the individual wet weight (g), and L is the total length (cm).

To determine the number of fish with an increase or decrease in K at the end of each experiment, per pond per treatment, the average value of initial K per fish per pond was taken (constant of K per pond), as follows:

$$\text{Final } K \text{ per fish} - \text{constant of } K \text{ per pond}$$

where the positive values are taken as an increase (0) and the negative values as a decrease (1) in K per fish (Table 2).

The following calculations were performed to determine (29) (Table 2):

$$\text{Incidence Welfare Indicator in the treatment (I Tx)} = \left( \frac{nO}{nTx} \right) * 100$$

the incidence of the welfare indicator corresponding to each section according to the treatment, where nO: no. individuals that presented the event and, nTx: no. individuals by treatment; observed treatment individuals presenting the event (0 or 1) accordingly to the corresponding index description.

$$\text{Incidence of the same Welfare Indicator in the rest (I)}$$

$$= \left( \frac{\sum nO \text{ other treatments}}{\sum nTx \text{ other treatments}} \right) * 100$$

$$\text{Relative Risk (RR)} = \left( \frac{\% I Tx}{\% I} \right)$$

The relative risk was taken with 95% confidence intervals (CI). It is important to note that a RR with a value >1.0 indicates that the factor, in this case, the treatment, represents a risk on the WI evaluated, otherwise, a RR <1.0 indicates a NO risk. If the RR value is equal to 1, then the risk is the same between the groups. The CIs allow determining the lower and upper limits where the real RR value is located, therefore, a lower CI with a value >1.0 indicates a risk at a 95% confidence level.

## WIs Risk Traffic Light

To facilitate the use and practical application of the results reported in this study, a risk traffic light for WIs was designed (Table 3). The ranges of the risk traffic light were determined considering the lowest data of % I Tx, according to the stage and the evaluated welfare indicator (Tables 4–8); if this was too low then the next higher value was used. To consider the lower limit indicator in a situation of moderate risk, the data had to be associated with a RR lower than 0.85 and a present CI in ranges lower than 0.99. As for the upper limit indicator, it was calculated = (% I Tx lower limit \* RR constant)/RR; the RR constant was determined as 0.999.

## Statistical Analysis

To determine the probability of occurrence or not of an event (0 or 1), according to the welfare indicator and index description, being death or, where appropriate, the visible damage observed in the experimental organisms (each dead fish or fish with visible damage is considered an event: nO), the presence/absence data of the welfare indicator were analyzed from an epidemiological approach through a prospective, longitudinal, analytical, experimental study (29). Contingency tables were used to determine the % incidence of the welfare indicators (events) presented in a particular treatment (% I Tx), the % incidence of the welfare indicators presented in the rest of the treatments (% I), the relative risk (RR), and 95% confidence intervals. The independence tests were performed using the Chi-square test (which measures the degree of association or relationship between the treatment and the welfare indicator) (28). The total number of animals per treatment was used to develop the contingency table analysis. A one-way ANOVA was used to analyze the biomass production values, followed by a Duncan test for, bias and kurtosis previously reviewed. To jointly analyze the data of the growth stages, the values were transformed into %, considering the highest value obtained from biomass production as 100%. Statistical program Statgraphics XVI, was used.

## RESULTS

Statistically significant associations (Chi-square test,  $p < 0.05$ ) were observed between the treatments and the welfare indicators (WIs) of mortality, weight reduction, K reduction, and presence of evident damage in caudal and anal fins. In the rest of the WIs evaluated, no statistically significant relationships were observed. In all the stages studied, statistically significant associations were observed between the treatments and the five WIs mentioned

**TABLE 4 |** Contingency table, incidence of mortality, weight reduction, K reduction and caudal and anal fin damage of *O. niloticus* fingerlings according to the feeding rate.

Records				Epidemiological analysis				
		nO	(%)	% I Tx	% I	RR	(CI 95%)	p value
Tx	n Tx	Total, n = 2,160 (100%)		Welfare Indicator				
		Death fish		Mortality				
Ufe	540	112	(5.19)	20.74	25.56	0.81	(0.67–0.97)	0.0240
Control	540	74	(3.43)	13.70	27.90	0.49	(0.39–0.61)	0.0000
OfA	540	130	(6.02)	24.07	24.44	0.98	(0.82–1.17)	ns
OfB	540	210	(9.72)	38.89	19.51	1.99	(1.72–2.30)	0.0000
Total, n = 1,634 (100%)								
		Weight lost cases		Weight reduction				
Ufe	428	211	(12.9)	49.30	56.97	0.86	(0.77–0.96)	0.0062
Control	466	160	(9.79)	34.33	63.18	0.54	(0.47–0.62)	0.0000
OfA	410	256	(15.6)	62.44	52.45	1.19	(1.08–1.30)	0.0004
OfB	330	271	(16.6)	82.12	48.05	1.70	(1.58–1.84)	0.0000
		K lost cases		K reduction				
Ufe	428	56	(3.43)	13.08	21.56	0.60	(0.46–0.79)	0.0001
Control	466	52	(3.18)	11.16	22.60	0.49	(0.37–0.65)	0.0000
OfA	410	117	(7.16)	28.54	16.26	1.75	(1.43–2.14)	0.0000
OfB	330	91	(5.57)	27.58	17.25	1.59	(1.29–1.97)	0.0000
		Damaged caudal fin cases		Caudal fin damage				
Ufe	428	53	(3.24)	12.38	10.20	1.21	(0.89–1.64)	ns
Control	466	28	(1.71)	6.01	12.67	0.47	(0.32–0.69)	0.0001
OfA	410	45	(2.75)	10.98	10.70	1.02	(0.74–1.41)	ns
OfB	330	50	(3.06)	15.15	9.66	1.56	(1.15–2.12)	0.0041
		Damaged anal fin cases		Anal fin damage				
Ufe	428	36	(2.20)	8.41	16.00	0.52	(0.37–0.73)	0.0001
Control	466	62	(3.79)	13.30	14.30	0.93	(0.70–1.22)	ns
OfA	410	68	(4.16)	16.59	13.15	1.26	(0.97–1.63)	ns
OfB	330	63	(3.86)	19.09	12.73	1.50	(1.15–1.95)	0.0029

Tx, treatment; n Tx, no. fish by treatment; nO, no. observed fish who presented the event (cases); %, proportion of individuals who presented the event (cases) in relation to the total n; p-value calculated by means of  $\chi^2$ ; I Tx: incidence of the variable corresponding to each section according to the treatment, calculated =  $(nO / nTx) * 100$ ; I: incidence of the same variable in the rest of the individuals, calculated =  $(\text{sum } nO \text{ of the other treatments} / \text{sum } nTx \text{ of the other treatments}) * 100$ ; RR: relative risk, calculated =  $\% I Tx / \% I$ ; CI, confidence intervals (95%).

Treatments: Feeding rate in relation to the pond biomass for underfeeding (Ufe, 80%), control (100%), overfeeding A (OfA, 120%), and overfeeding B (OfB, 140%), respectively.

(Tables 4–8). The first section of Table 4 is described in detail below to facilitate the reading and understanding of the results of this study. Total n = number of fish that were part of the corresponding WI analysis (mortality), which for on-growing 1 was 2,160 fingerlings; n Tx = number of fish per treatment (540 for each); nO = number of observed fish that presented the WI, in this case death, being 112, 74, 130, and 210 dead fish throughout the experimental period, for the Uf, Control, OfA and OfB treatments, respectively; % = number of WIs in relation to Total n (2,160 = 100%), being 5.19, 3.43, 6.02 and 9.72% according to each treatment; p value = statistical significance of the Chi-square test analysis; % I Tx = Incidence of the WI in a particular treatment, it is obtained by performing the following operation:  $(nO / n Tx) * 100$  resulting in 20.74, 13.70, 24.07 and 38.89% for the Ufe, Control, OfA and OfB treatments, respectively; % I = Incidence of the WI in the rest of the treatments, therefore, for the treatment of Ufe, the following operation is obtained:  $(74 +$

$130 + 210 / 540 + 540 + 540) * 100 = (414 / 1,620) * 100 = 25.56$ . The RR is the relationship between % I Tx and % I, and  $RR = 20.74 / 25.56 = 0.81$  are obtained for the Ufe treatment,  $RR = 13.70 / 27.90 = 0.49$  for the Control,  $RR = 24.07 / 24.44 = 0.98$  for the treatment OfA, and  $RR = 38.89 / 19.51 = 1.99$  for OfB, respectively. In the following sections of Table 4, which correspond to the rest of the WIs evaluated (weight reduction, K reduction, damage to caudal and anal fins), Total n is the number of fish that survived at the end of the experiment, in this case, and accordingly to the calculation made corresponding to the number of surviving fish (1,634 fish), distributed in the treatments (428, 466, 410, and 330, respectively).

In fingerlings, a RR of 1.99 for the WI mortality is observed in the OfB treatment, this means that the probability of observing dead fish under this risk condition is 1.99 times greater than the mortality probability, in the rest of the treatments (Table 4). Compared to the Control treatment, where a lower RR was



**TABLE 5 |** Contingency table, incidence of mortality, weight reduction, K reduction, and caudal and anal fin damage of *O. niloticus* juvenile according to the feeding rate.

Records				Epidemiological analysis				
		nO	(%)	% I Tx	% I	RR	(CI 95%)	p value
Tx	n Tx	Total, n = 1,200 (100%)		Welfare Indicator				
		Death fish		Mortality				
Ufe	300	8	(0.67)	2.67	4.11	0.64	(0.30–1.37)	ns
Control	300	8	(0.67)	2.67	4.11	0.64	(0.30–1.37)	ns
OfA	300	11	(0.92)	3.67	3.78	0.97	(0.49–1.89)	ns
OfB	300	18	(1.50)	6.00	3.00	2.00	(1.11–3.57)	0.0179
Total, n = 1,155 (100%)								
Weight lost cases				Weight reduction				
Ufe	292	251	(21.7)	85.96	48.09	1.78	(1.64–1.94)	0.0000
Control	292	140	(12.1)	47.95	60.95	0.78	(0.69–0.89)	0.0001
OfA	289	100	(8.66)	34.60	65.36	0.52	(0.44–0.62)	0.0000
OfB	282	175	(15.1)	62.06	56.24	1.10	(0.99–1.22)	ns
K lost cases				K reduction				
Ufe	292	83	(7.19)	28.42	17.15	1.65	(1.31–2.09)	0.0000
Control	292	37	(3.20)	12.67	22.48	0.56	(0.40–0.78)	0.0003
OfA	289	33	(2.86)	11.42	22.86	0.49	(0.35–0.70)	0.0000
OfB	282	78	(6.75)	27.66	17.53	1.57	(1.24–2.00)	0.0002
Damaged caudal fin cases				Caudal fin damage				
Ufe	292	34	(2.94)	11.64	10.78	1.08	(0.74–1.56)	ns
Control	292	17	(1.47)	5.82	12.75	0.45	(0.27–0.74)	0.0011
OfA	289	36	(3.12)	12.46	10.51	1.18	(0.82–1.70)	ns
OfB	282	40	(3.46)	14.18	9.97	1.42	(1.00–2.01)	0.0490
Damaged anal fin cases				Anal fin damage				
Ufe	292	24	(2.08)	8.22	15.76	0.52	(0.34–0.78)	0.0013
Control	292	38	(3.29)	13.01	14.14	0.92	(0.65–1.29)	ns
OfA	289	46	(3.98)	15.92	13.16	1.20	(0.88–1.65)	ns
OfB	282	52	(4.50)	18.44	12.37	1.49	(1.10–2.01)	0.0103

Tx, treatment; n Tx, no. fish by treatment; nO, no. observed fish who presented the event (cases); %, proportion of individuals who presented the event (cases) in relation to the total n; p-value calculated by means of  $\chi^2$ ; I Tx, incidence of the variable corresponding to each section according to the treatment, calculated =  $(nO / nTx) * 100$ ; I, incidence of the same variable in the rest of the individuals, calculated =  $(\text{sum } nO \text{ of the other treatments} / \text{sum } nTx \text{ of the other treatments}) * 100$ ; RR: relative risk, calculated =  $\% I Tx / \% I$ ; CI, confidence intervals (95%).

Treatments: Feeding rate in relation to the pond biomass for underfeeding (Ufe, 80%), control (100%), overfeeding A (OfA, 120%), and overfeeding B (OfB, 140%), respectively.

observed, the probability of death is 0.49 times the probability of observing dead fish in the other treatments. This means that the treatment factor OfB represents a mortality risk of practically four times, compared to the mortality risk of the Control treatment. With the WI of weight reduction, a RR of 0.54 was observed in Control, and in the OfB of 1.70, which means that there is 3.49 times more risk that the fish lose weight than with the control treatment. In the WI of K reduction, the OfA treatment obtained a RR of 1.75, while the Control treatment a RR of 0.49, with which in the OfA treatment it is up to 3.5 times more likely that the fish will see their body condition reduced than in the Control treatment. The WI of caudal fin damage showed that in the OfB treatment the RR was 1.56, while in the control treatment the RR was 0.47, with which in OfB the fish have up to 3.32 times the risk of injury to the caudal fins than in the control treatment. The anal fin damage WI presented that the RR of OfB was 1.50 and the Ufe treatment of 0.52, which indicates

that in OfB the fish have a risk of 2.88 more lesions in the anal fin than in the Ufe treatment.

In treatment OfB a RR of 2.0 for the WI mortality is observed for the juvenile stage (Table 5), meaning that the probability of observing dead fish in the OfB treatment is 2.0 times greater than the probability of mortality in the rest of the treatments. Compared with the Ufe and control treatments, where a lower RR is observed, the probability of death is 0.64 times the probability of observing dead fish in the other treatments. The OfB treatment factor represents a mortality risk of practically 3.12 more, compared to the mortality risk of the Ufe and control treatments. With the WI of weight reduction, the RR of 0.52 was observed in the OfA, and in the Ufe of 1.78, with which there is 3.42 more risk that the fish lose weight than with the OfA. In the WI of K reduction, the Ufe treatment obtained a RR of 1.65, while the OfA treatment a RR of 0.49, with which in the Ufe treatment it is up to 3.36 more likely that the fish will see their

**TABLE 6 |** Contingency table, incidence of mortality, weight reduction, K reduction, and caudal and anal fin damage of on-growing 1 of *O. niloticus* according to the feeding rate.

Records				Epidemiological analysis				
		nO	(%)	% I Tx	% I	RR	(CI 95%)	p value
Tx	n Tx	Total, n = 1,200 (100%)		Welfare Indicator				
		Death fish		Mortality				
Ufe	300	1	(0.08)	0.33	4.78	0.07	(0.01–0.50)	0.0004
Control	300	1	(0.08)	0.33	4.78	0.07	(0.01–0.50)	0.0004
OfA	300	2	(0.17)	0.67	4.67	0.14	(0.03–0.58)	0.0014
OfB	300	40	(3.33)	13.33	0.44	30.0	(10.8–83.1)	0.0000
Total, n = 1,156 (100%)								
Weight lost cases				Weight reduction				
Ufe	299	246	(21.2)	82.27	50.99	1.61	(1.48–1.75)	0.0000
Control	299	140	(12.1)	46.82	63.36	0.73	(0.64–0.84)	0.0000
OfA	298	147	(12.7)	49.33	62.47	0.79	(0.69–0.89)	0.0001
OfB	260	150	(12.9)	57.69	59.49	0.97	(0.86–1.09)	ns
K lost cases				K reduction				
Ufe	299	72	(6.23)	24.08	58.5	0.41	(0.33–0.50)	0.0000
Control	299	178	(15.4)	59.53	46.2	1.28	(1.14–1.44)	0.0001
OfA	298	177	(15.3)	59.40	46.2	1.28	(1.14–1.44)	0.0001
OfB	260	147	(12.7)	56.54	47.6	1.18	(1.04–1.34)	0.0117
Damaged caudal fin cases				Caudal fin damage				
Ufe	299	54	(4.67)	18.06	16.5	1.09	(0.82–1.44)	ns
Control	299	32	(2.77)	10.70	19.1	0.55	(0.39–0.79)	0.0008
OfA	298	56	(4.84)	18.79	16.3	1.15	(0.87–1.52)	ns
OfB	260	54	(4.67)	20.77	15.8	1.31	(0.98–1.73)	ns
Damaged anal fin cases				Anal fin damage				
Ufe	299	54	(4.67)	18.06	19.37	0.93	(0.70–1.23)	ns
Control	299	65	(5.62)	21.74	18.09	1.20	(0.92–1.55)	ns
OfA	298	56	(4.84)	18.79	19.11	0.98	(0.74–1.29)	ns
OfB	260	45	(3.89)	17.31	19.53	0.88	(0.65–1.19)	ns

Tx, treatment; n Tx, no. fish by treatment; nO, no. observed fish who presented the event (cases); %, proportion of individuals who presented the event (cases) in relation to the total n; p-value calculated by means of  $\chi^2$ ; I Tx, incidence of the variable corresponding to each section according to the treatment, calculated =  $(nO / nTx) * 100$ ; I, incidence of the same variable in the rest of the individuals, calculated =  $(\text{sum } nO \text{ of the other treatments} / \text{sum } nTx \text{ of the other treatments}) * 100$ ; RR: relative risk, calculated =  $\% I Tx / \% I$ ; CI, confidence intervals (95%).

Treatments: Feeding rate in relation to the pond biomass for underfeeding (Ufe, 80%), control (100%), overfeeding A (OfA, 120%), and overfeeding B (OfB, 140%), respectively.

body condition reduced than in the OfA treatment. The WI of caudal fin damage showed that in the OfB treatment a RR of 1.42, while in the Control treatment the RR was 0.45, with which in OfB the fish have up to 3.32 times more risk of injury to the caudal fin than in Control. Anal fin damage WI in the OfB treatment presented a RR of 1.49 and in Ufe of 0.52, which indicates that fish in OfB conditions have a 2.88 higher risk of lesions in the anal fin than in Ufe.

For on-growing 1 (Table 6), a RR of 30 for WI mortality is observed in the OfB treatment, this means that the probability of observing dead fish for the OfB treatment is 30 times greater than the probability of mortality in the rest of the treatments. Compared with the Ufe and control treatments, where a lower RR is observed, the probability of death is 0.07 times the probability of observing dead fish in the other treatments. This means that the OfB treatment factor represents a mortality risk of practically 428.57 times more, compared to the mortality risk of the Ufe and

Control treatments. With the WI of weight reduction, the RR of 0.73 was observed in the Control, and in the Ufe it was 1.61, with which there is 2.20 more risk that the fish lose weight than in the Control. In the WI of K reduction, the control and OfA treatments obtained a RR of 1.28, while the Ufe treatment had a RR of 0.41, with which in the Ufe treatment they are up to 3.12 times more likely that the fish will suffer reduced bodily condition than in the OfA treatment. The WI of caudal fin damage showed that in the control treatment a RR of 0.55, with which in Control the fish have up to 3.32 times less risk of injury to the caudal fin than in the rest of the treatments, which did not present differences. Anal fin damage WI did not present significant RR in any treatment.

For on-growing 2 (Table 7), we observed a RR of 8.50 for the WI mortality in the OfB treatment meaning that the probability of observing dead fish for the OfB treatment is 8.50 times greater than the probability of mortality in the rest of the treatments,

**TABLE 7 |** Contingency table, incidence of mortality, weight reduction, K reduction, and caudal and anal fin damage of on-growing 2 of *O. niloticus* according to the feeding rate.

Records				Epidemiological analysis				
		nO	(%)	% I Tx	% I	RR	(CI 95%)	p value
Tx	n Tx	Total, n = 1,200 (100%)		Welfare Indicator				
		Death fish		Mortality				
Ufe	300	3	(0.25)	1.00	2.22	0.45	(0.13–1.50)	ns
Control	300	3	(0.25)	1.00	2.22	0.45	(0.13–1.50)	ns
OfA	300	0	(0.00)	0.00	2.56	0.00	n/a	0.0052
OfB	300	17	(1.42)	5.67	0.67	8.50	(3.38–21.8)	0.0000
Total, n = 1,177 (100%)								
Weight lost cases				Weight reduction				
Ufe	297	175	(14.8)	58.94	60.00	0.98	(0.88–1.09)	ns
Control	297	149	(12.6)	50.17	62.95	0.79	(0.70–0.90)	0.0001
OfA	300	198	(16.8)	66.00	57.58	1.14	(1.03–1.26)	0.0103
OfB	283	181	(15.3)	63.96	58.39	1.09	(0.98–1.21)	ns
K lost cases				K reduction				
Ufe	297	135	(11.4)	45.45	38.0	1.19	(1.02–1.38)	0.0246
Control	297	102	(8.67)	34.34	41.8	0.82	(0.68–0.97)	0.0229
OfA	300	117	(9.94)	39.00	40.2	0.96	(0.82–1.14)	ns
OfB	283	116	(9.86)	40.99	36.6	1.03	(0.88–1.21)	ns
Damaged caudal fin cases				Caudal fin damage				
Ufe	297	113	(9.60)	38.05	27.0	1.40	(1.17–1.68)	0.0003
Control	297	91	(7.73)	30.64	29.5	1.03	(0.84–1.26)	ns
OfA	300	62	(5.27)	20.67	32.9	0.62	(0.49–0.79)	0.0001
OfB	283	85	(7.22)	30.04	29.7	1.00	(0.82–1.23)	ns
Damaged anal fin cases				Anal fin damage				
Ufe	297	46	(3.91)	15.49	20.8	0.74	(0.55–1.00)	0.0457
Control	297	82	(6.97)	27.61	16.7	1.65	(1.30–2.09)	0.0000
OfA	300	49	(4.16)	16.33	20.5	0.79	(0.59–1.06)	ns
OfB	283	52	(4.42)	18.37	19.8	0.92	(0.70–1.22)	ns

Tx, treatment; n Tx, no. fish by treatment; nO, no. observed fish who presented the event (cases); %, proportion of individuals who presented the event (cases) in relation to the total n; p-value calculated by means of  $\chi^2$ ; I Tx, incidence of the variable corresponding to each section according to the treatment, calculated =  $(nO / nTx) * 100$ ; I, incidence of the same variable in the rest of the individuals, calculated =  $(\text{sum } nO \text{ of the other treatments} / \text{sum } nTx \text{ of the other treatments}) * 100$ ; RR: relative risk, calculated =  $\% I Tx / \% I$ ; CI, confidence intervals (95%).

Treatments: Feeding rate in relation to the pond biomass for underfeeding (Ufe, 80%), control (100%), overfeeding A (OfA, 120%), and overfeeding B (OfB, 140%), respectively.

where no significant RR was reported. With the WI of weight reduction, the RR of 0.79 was observed in the Control, and in the OfA of 1.14, with which there is 1.44 more risk that the fish lose weight than in the control. In the WI of K reduction, the Ufe treatment obtained a RR of 1.19, while the control treatment a RR of 0.82, with which in the Ufe treatment they are up to 1.45 more likely that the fish will see their body condition reduced than in the OfA treatment. The WI of caudal fin damage showed that in the Ufe treatment a RR of 1.40, while in the OfA treatment the RR was 0.62, with which in OfB the fish have up to 2.25 times the risk of injury to the caudal fins than in control. Anal fin damage WI showed that the control RR was 1.65 and that of Ufe was 0.74, which indicates that in Control the fish have a risk of 2.22 more of presenting lesions in the anal fin than in Ufe.

For the on-growing 3 (Table 8), a RR of 9.27 for WI mortality is observed in the OfB treatment meaning that the probability of observing dead fish for the OfB treatment is 9.27 times greater

than the probability of mortality in the rest of the treatments. Compared with the control treatment, where a lower RR is observed, the probability of death is 0.21 times that of observing dead fish in the other treatments. This means that the treatment factor OfB represents a mortality risk of practically 44.1 more, compared to the mortality risk of the control treatment. With the WI of weight reduction, a RR of 0.49 was observed in the Ufe, and the OfA and OfB both observed a 1.46 RR, with which there is 2.97 more risk that the fish lose weight than in the Ufe. In the WI of K reduction, the OfA treatment obtained a RR of 1.48, while the Ufe treatment a RR of 0.76, which means the OfA is up to 1.94 more likely that the fish will see their body condition reduced than in the Ufe treatment. The WI of caudal fin damage showed that in the OfA treatment a RR of 1.39, while in the OfB treatment the RR was 0.48, with which in OfA the fish have up to 2.89 times the risk of tail fin injury than in OfB. Anal fin damage WI showed that the RR of OfB was 0.29, while in the others the

**TABLE 8 |** Contingency table, incidence of mortality, weight reduction, K reduction, and caudal and anal fin damage of on-growing 3 of *O. niloticus* according to the feeding rate.

Records				Epidemiological analysis				
		nO	(%)	% I Tx	% I	RR	(CI 95%)	p value
Tx	n Tx	Total, n = 720 (100%)		Welfare Indicator				
		Death fish		Mortality				
Ufe	180	7	(0.97)	3.89	7.04	0.55	(0.25–1.21)	ns
Control	180	3	(0.42)	1.67	7.78	0.21	(0.06–0.68)	0.0034
OfA	180	1	(0.14)	0.56	8.15	0.06	(0.009–0.49)	0.0003
OfB	180	34	(4.72)	18.89	2.04	9.27	(4.79–17.91)	0.0000
Total, n = 675 (100%)								
Weight lost cases				Weight reduction				
Ufe	173	62	(9.19)	35.84	72.91	0.49	(0.39–0.60)	0.0000
Control	177	95	(14.0)	53.67	66.87	0.80	(0.69–0.93)	0.0017
OfA	179	148	(21.9)	82.68	56.45	1.46	(1.32–1.62)	0.0000
OfB	146	123	(18.2)	84.25	57.66	1.46	(1.32–1.61)	0.0000
K lost cases				K reduction				
Ufe	173	77	(11.4)	44.51	58.37	0.76	(0.63–0.91)	0.0016
Control	177	87	(12.8)	49.15	56.83	0.86	(0.73–1.02)	ns
OfA	179	129	(19.1)	72.07	48.59	1.48	(1.30–1.68)	0.0000
OfB	146	77	(11.4)	52.74	55.39	0.95	(0.80–1.13)	ns
Damaged caudal fin cases				Caudal fin damage				
Ufe	173	51	(7.56)	29.48	30.28	0.97	(0.74–1.27)	ns
Control	177	60	(8.89)	33.90	28.71	1.18	(0.92–1.51)	ns
OfA	179	68	(10.0)	37.99	27.22	1.39	(1.10–1.76)	0.0071
OfB	146	24	(3.56)	16.44	33.84	0.48*	(0.33–0.71)	0.0000*
Damaged anal fin cases				Anal fin damage				
Ufe	173	40	(5.93)	23.12	18.72	1.23	(0.89–1.71)	ns
Control	177	41	(6.07)	23.16	18.67	1.24	(0.89–1.71)	ns
OfA	179	43	(6.37)	24.02	18.35	1.30	(0.95–1.80)	ns
OfB	146	10	(1.48)	6.85	23.44	0.29*	(0.15–0.54)	0.0000*

Tx, treatment; n Tx, no. fish by treatment; nO, no. observed fish who presented the event (cases); %, proportion of individuals who presented the event (cases) in relation to the total n; p-value calculated by means of  $\chi^2$ ; I Tx, incidence of the variable corresponding to each section according to the treatment, calculated =  $(nO / nTx) * 100$ ; I, incidence of the same variable in the rest of the individuals, calculated =  $(\text{sum } nO \text{ of the other treatments} / \text{sum } nTx \text{ of the other treatments}) * 100$ ; RR: relative risk, calculated =  $\% I Tx / \% I$ ; CI, confidence intervals (95%). \*False positive, as an effect of the high mortality that occurred in said treatment.

Treatments: Feeding rate in relation to the pond biomass for underfeeding (Ufe, 80%), control (100%), overfeeding A (OfA, 120%), and overfeeding B (OfB, 140%), respectively.

RR was the same. It should be noted that in these last two WIs where OfB came out with a lower RR than the other treatments, it was due to the poor quality of the water, which affected the fish behavior.

The registered water quality values during the study are in **Table 9**. The biomass production data per pond concerning the treatments are in **Figure 1**. In the fingerling stage, we observed highly significant differences between the treatments ( $p < 0.01$ ). The Control treatment showed the highest value (95.47%), followed by Ufe (79.90%), and the lowest value in OfB (40.33%). In the juvenile phase, there were significant differences between feeding rates ( $p < 0.05$ ). OfA being the maximum value recorded (94.03%) after the Control treatment (83.63%), the lowest biomass production was observed in OfB. In the group of fish with on-growing 1, significant statistical differences were found between the treatments ( $p < 0.05$ ), with OfA where the highest production occurred and the lowest in the overfeeding

treatments (OfA and OfB). In on-growing 2, highly significant differences were observed between treatments ( $p < 0.01$ ). In the control, Ufe and OfA, the highest values were observed (88.50, 82.33, and 80.33%, respectively), and the lowest value was in the OfB treatment (45.60%). Regarding on-growing 3 phase, the ANOVA analysis did not show significant statistical differences between the treatments, however, in **Figure 1** it can be observed that the highest productions were recorded in the Ufe and control treatments, and the lowest in OfA and OfB. In the lower part of **Figure 1**, the application of the proposed epidemiological traffic light and the qualification of welfare by treatment according to the evaluated indicators can be observed. To determine the degree of welfare, the color of the traffic light was categorized as follows:

Green = 2 points  
 Yellow = 1 point  
 Red = 0 points



**TABLE 9 |** Water quality values during the study.

Treatments	Underfeeding	Control	Overfeeding A	Overfeeding B
<b>Temperature (°C)</b>				
Fingerlings	26.4 ± 1.8	26.2 ± 2.1	26.4 ± 1.8	28.7 ± 2.1
Juveniles	25.3 ± 1.5	25.0 ± 1.8	25.2 ± 1.4	25.6 ± 1.6
On-growing 1	24.9 ± 1.2	24.5 ± 1.7	24.6 ± 1.3	25.1 ± 1.6
On-growing 2	26.7 ± 1.5	25.3 ± 1.8	25.4 ± 1.4	25.8 ± 2.2
On-growing 3	23.6 ± 1.6	23.2 ± 1.6	23.4 ± 1.4	23.1 ± 2.1
<b>Dissolved Oxygen (mg L<sup>-1</sup>)</b>				
Fingerlings	5.5 ± 0.5	4.9 ± 0.6	4.7 ± 0.7	3.5 ± 0.6
Juveniles	5.4 ± 0.6	4.9 ± 0.7	4.7 ± 0.8	3.3 ± 1.1
On-growing 1	5.8 ± 0.8	5.2 ± 0.8	5.1 ± 0.7	3.7 ± 0.9
On-growing 2	5.5 ± 0.6	4.9 ± 0.6	4.7 ± 0.8	3.5 ± 0.9
On-growing 3	5.7 ± 0.8	5.4 ± 0.9	4.6 ± 1.0	3.7 ± 1.0
<b>Electrical Conductivity (mS/cm<sup>3</sup>)</b>				
Fingerlings	699.3 ± 31.2	725.6 ± 39.5	863.4 ± 44.7	896.7 ± 50.8
Juveniles	792.4 ± 39.2	790.5 ± 46.0	815.9 ± 45.6	813.3 ± 48.2
On-growing 1	777.0 ± 37.9	760.9 ± 107.9	791.2 ± 44.7	808.2 ± 52.2
On-growing 2	919.7 ± 137.2	952.6 ± 153.9	978.0 ± 182.4	1070.9 ± 198.1
On-growing 3	947.9 ± 221.0	943.1 ± 209.8	921.7 ± 198.4	951.7 ± 235.8
<b>NH<sub>4</sub><sup>+</sup> (mg L<sup>-1</sup>)</b>				
Fingerlings	0.1 ± 0.1	0.3 ± 0.2	0.3 ± 0.2	0.7 ± 0.1
Juveniles	0.3 ± 0.2	0.4 ± 0.4	0.8 ± 0.6	0.9 ± 0.3
On-growing 1	0.5 ± 0.1	0.7 ± 0.2	0.8 ± 0.7	0.9 ± 0.1
On-growing 2	0.5 ± 0.1	0.6 ± 0.4	0.8 ± 0.4	0.9 ± 0.3
On-growing 3	0.2 ± 0.1	0.2 ± 0.2	0.3 ± 0.2	0.3 ± 0.2
<b>pH</b>				
Fingerlings	7.2 ± 0.1	7.3 ± 0.1	7.1 ± 0.2	7.5 ± 0.2
Juveniles	7.5 ± 0.2	7.3 ± 0.2	7.3 ± 0.1	7.3 ± 0.2
On-growing 1	7.3 ± 0.1	7.2 ± 0.1	7.2 ± 0.1	7.2 ± 0.1
On-growing 2	7.3 ± 0.1	7.2 ± 0.1	7.3 ± 0.1	7.3 ± 0.1
On-growing 3	7.4 ± 0.1	7.3 ± 0.1	7.3 ± 0.1	7.3 ± 0.1

Means ± SD.

Subsequently, the value of the evaluated indicators was added to obtain a comprehensive qualification (fingerlings to on growing 3) of the welfare state according to the following:

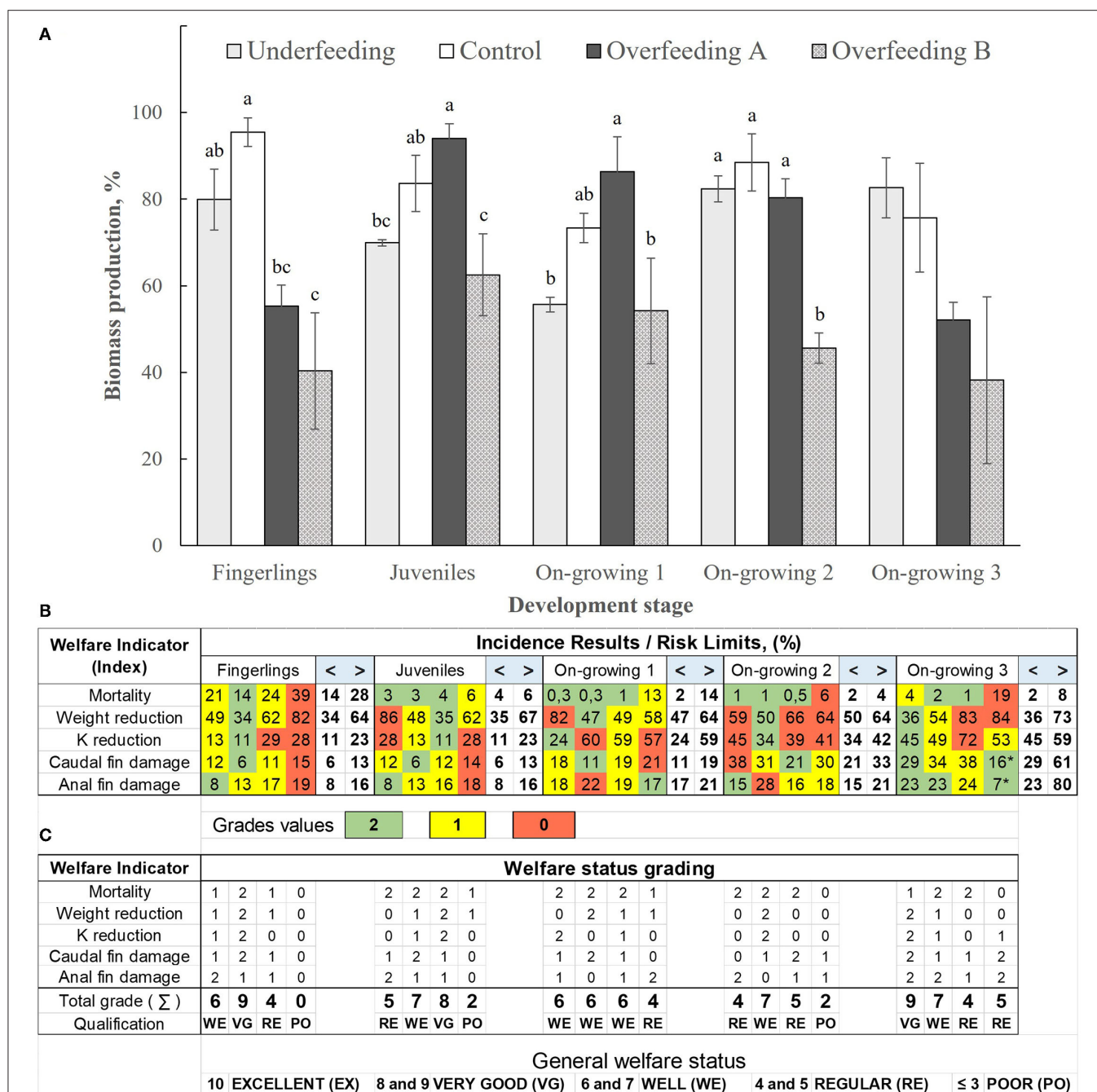
- 10 points = Excellent (EX)
- 8 and 9 points = Very Good (VG)
- 6 and 7 points = Well (WE)
- 4 and 5 points = Regular (RE)
- ≤ 3 points = Poor (PO)

## DISCUSSION

To the authors' knowledge, this is the first report on the analysis of welfare indicators (WIs) using an epidemiological population approach tools at various stages of the tilapia life cycle. To determine if the production conditions that arise because of feed management, in particular the feeding rate used, can be a risk factor on the welfare status of a fish population in a defined time. Fish welfare is an issue that concerns fish farmers in daily practice, whose objective is to be able to produce

organisms properly and follow the correct guidelines for growing this species (8, 14), carrying out responsible and sustainable management. Currently, the consumer is willing to pay an additional price for fish products from farms that are identified with quality and welfare standards (40), therefore, it is important to identify according to the species, which WIs can be used as a standard on laboratories and farms, that is, as laboratory welfare indicators or operational welfare indicators (10).

In the fingerling stage, the water quality indicators were observed within the optimal ranges for tilapia cultivation, except NH<sub>4</sub><sup>+</sup> in the OfB that already presented levels higher than those recommended for the species (41, 42). In the juvenile stages, on-growing 1 and on-growing 2, NH<sub>4</sub><sup>+</sup> was recorded above the optimal values lower than 0.70 mg L<sup>-1</sup> (41, 42) in the OfA and OfB treatments and the control treatment in the on-growing 1. In on-growing 3, all the water indicators were optimal. High feeding rates rapidly affect water quality. As a fish health indicator, this is important, as ponds exceeding the limits for ammonium, nitrates, and nitrites compromise the development and survival of the organisms (1, 43–45). In addition, a higher concentration of



**FIGURE 1 | (A)** Biomass production per pond per treatment (means  $\pm$  SE) in the 5 experiments carried out according to the treatments: underfeeding (80%), control (100%), overfeeding A (120%), and overfeeding B (140%). **(B)** Each cell in color is the value resulting from the % I Tx of the epidemiological analysis (integer) according to the welfare indicator of each treatment (from Tables 4–8). White cells are risk limits (%) for welfare indicators in relation to the individuals present in the culture pond of each treatment (from Table 3; < and >). **(C)** The application of the proposed epidemiological traffic light and the qualification of the welfare state by treatment according to the evaluated indicators. Color traffic lights are categorized as Green (2 points), Yellow (1 point), and Red (0 points). The value of the evaluated indicators added (summative) to obtain a comprehensive/overall qualification (fingerlings to on growing 3) of the welfare state according: Excellent (EX: 10 points), Very Good (VG: 8 and 9 points), Well (WE: 6 and 7 points), Regular (RE: 4 and 5 points) and Poor (PO:  $\leq$  3 points). \*False positive, as an effect of the high mortality occurred in said treatment. Means  $\pm$  SE.

nitrogenous products in the system can favor higher productivity in the pond, which is usually accompanied by a decrease in dissolved oxygen in the water during the early morning hours

(46). According to the results, in fingerlings a greater growth was observed using the control feeding rate (8% with 49% CP), in addition to this, that treatment, it was where the best values of

WIs were observed. It is important to mention that the higher mortality and lower production occurred in the overfeeding treatments, showing that the fingerlings are very sensitive to water conditions (41). Feed management directly impacts the quality of the culture water (47) as can be seen in the results presented here, the concentration of  $\text{NH}_4^+$  and dissolved solids in pond water increases as more food is given to the fish. Contrarily, in the following two stages (juveniles and on-growing 1), the highest biomass obtained was registered when the fish had an overfeeding of 20% greater than that indicated for the species. It should be noted that it is recommended that tilapia diets decrease the concentration of crude protein as the size of the fish increases (48, 49) because young animals need nutrients steadily, the feeding rate is also higher, as well as the frequency of feeding the diet. As the fish grow, the feeding rates and frequencies decrease, since their metabolism is more efficient to take advantage of the nutrients in the feed (20, 50, 51). Although the commercial diets used in the present study, for the juvenile and on-growing 1, are similar in the contribution of protein and energy, the recommended rate (5 and 4%, respectively) (22), could not be covering the nutritional requirements of the species, for which, and as a result of this study, a rate between 5 and 6% (with 47–48% of CP) and 4 and 4.8% is suggested from the point of view of obtaining biomass (with 44–45% CP), respectively. These values are like those studied by El-Dakar et al. (52) in fingerlings and juveniles of hybrids of *O. niloticus* and *O. mossambicus*. In addition, it was in these two growth stages where greatest the values in % of fish with lesions were recorded, in a moderate risk range of welfare indicators evaluated.

In on-growing 2, the best growths were observed in the control treatment, as were the WIs, except in the incidence of lesions in the anal fin, which suggests a change in behavior concerning age. Tilapia is a cichlid, so aggressiveness is part of its normal behavior, and one of the targets of social injuries is the anal fins (53). Social lesions also occur in other types of fish, such as Siluridae, where the objective is to attack the flanks (54). The purpose of these injuries is to warn other members of their species about the hierarchy and occupation of space, not to hurt (53), however, these injuries can be aggravated if the environment of culture is not suitable (55). Regarding on-growing 3, the highest biomass obtained was recorded in the Ufe treatment, indicating a better use of the feed nutrients (20, 50, 51). The survival of tilapia depends on the type of facilities available, the temperature, the feeding, and the age of the culture (42). In RAS, the survival of the fish is greater, since the conditions are more stable than in open-air cultivation (earthen ponds), as there is a lower proportion of incidences due to the action of predators and diseases (1, 56). In this type of system, 0 to 20% mortality has been observed (17, 32), results like those observed in the OfB treatments in the present study.

The decrease in fish weight is one of the main concerns among aquaculturists; it is very difficult to observe growth with the naked eye but, feasible to follow the development of the pond over time, during cultivation, in the unfolds and pond changes. According to Rey et al. (57), the weight and the comparison of this value must be done by pond because each cage or pond behaves differently. They have their environment. The most

advisable action is periodically to monitor the population and consider the average weight of the pond, instead of making comparisons against databases, especially if they are from regions with different conditions. Therefore, using indicators such as the incidents proposed in this study will allow monitoring of the physiological condition and welfare of the organisms during the production cycle. To identify possible higher-risk situations and make timely and consistent decisions with each stage of growth. Body condition (K) is a variable factor and fluctuates throughout the life cycle of the fish. It is difficult to define exact values that are indicative of reduced welfare, but  $<0.9$  is usually indicative of emaciation (21). Body condition in fish has been used as a productive indicator, making it possible to measure fish growth over time. It has also been mentioned that it is an indicator of welfare in salmon (10), and in perch has been compared between manual and automatic feeding methods (34). What is absent in the previous studies is that there is no average initial body condition. So, it is not known how much it increased or decreased over the experiment time, a fact that is obtained in the present study and that was used as a comparative in the final body condition of the individuals. Thus, body condition can be used as a measurable indicator of welfare, particularly if it is based on the premise that fish entering a culture stage are expected to increase or at least maintain their body condition.

In the present study, was determined that the relationship between the incidence of tilapia caudal fin damage arises because of the used feeding rate. When feed is restricted, frequent and significant damage to fins caused by aggression from dominant individuals is observed, in addition to a decrease in body condition (33). It has been reported that feeding management directly affects the integrity of the fins of *Perca fluviatilis* L. (34) and *Salmo salar* (33), depending on the stage of the culture. However, these studies evaluated the relationship between the feeding management factor and the damage caused to the fins but, did not determine whether feeding management is a risk factor related to the proportion of animals that present evident damage to the fins. The fish fins are part of their anatomy since these limbs allow them to move in their environment (58). The fins are prone to damage their integrity can be affected by various elements, both biotic and abiotic, which can be detrimental to welfare. These factors include population density (59, 60), presence of disease (61), abrasions from pond and cage surfaces (62), aggressions among members of the population (63, 64), feeding management (65–67) and poor water quality, which is related to directly to various circumstances such as low dissolved oxygen concentration, and high concentration of  $\text{NH}_4^+$  and dissolved solids (55, 68). One of the most important fins for fish is the caudal fin since they are propelled by this member and allow them to measure the force with which they move (58).

According to a study in salmon (33), where two experimental groups were used: control and with food restriction, it is reported that 7.5 and 12.5% of the population presented damage, for the control and food restriction, respectively. It should be mentioned that, at first glance, it can infer that there is an effect of the treatment on the number of animals that presented damage. Analyzing the reported data using the Chi-square test, there is no statistically significant relationship between each treatment

and the number of observed events. In this case, fish with dorsal fin damage ( $p = 0.4561$ ). It is important to emphasize that the authors mentioned that, in the food restriction treatment, fins with a greater degree of damage were observed than those that were presented in the control treatment. Although in the present study, damage to the dorsal fin was not recorded in the tilapias. According to the risk traffic light proposed here (Table 8), fish weighing 66 g (on-growing 2 in tilapia), similar to those in the study by Cañon-Jones et al. (33) (initial weight  $61.7 \pm 6.4$  g), is considered a risky situation when from 15 and 21% of the population presented damage to the anal and caudal fins, respectively, values higher than those observed in salmon (7.5 and 12%), which is probably why the Chi-square test did not show a statistically significant relationship, which is consistent with the present study.

WIs have been determined in various species (10, 15), and are usually specific to the physiological or life stage (15) and the type of culture system used (10). The results obtained in the present study indicate that the 5 experiments allow us to see, as a whole, the effect of the feeding rate in the different growth stages of *O. niloticus*. By integrating the risk limits (%) in a final assessment (Excellent, Very Good, Well, Regular and Poor; Figure 1), a congruence is observed in the condition or general welfare status of the fish analyzed through the monitoring of survival, weight, Fulton index and damage to fins (anal and caudal). Although, in all the stages these same indicators were found where statistically significant relationships were observed, the level of risk changes due to the interaction between the growth stage and the feeding treatment. Mortality, for example, reaches a high risk (cells in red Figure 1), in the overfeeding treatment B for the phases of fingerlings, and on-growing 2 and 3, and only alert or moderate risk (cells yellow), for the juvenile and on-growing 1 phase. This indicates that, despite being the same species, the fish in the early growth phases behave differently in their development and physiology of the amount, quality, and characteristics of the food provided (20, 25, 47, 50).

Gutierrez-Rabadan et al. (15), established WIs in the lumpfish (*Cyclopterus lumpus*), using two experimental groups: animals that were in the hatchery and those in cages, with 60 and 35 animals, respectively. Of the fish in the first group ( $n = 60$ ), 31 presented caudal fin damage, 32 with pelvic fin damage, and in the second group ( $n = 35$ ), 6 fish with tail fin damage and 6 with fin injuries. Proportionally, it is reported that 52 and 53% of the population present damage to the caudal and pelvic fin respectively. In the hatchery, and 6 and 6% to the caudal and anal fin respectively, when fish are kept in cages. When analyzing these same data using the Chi-square test, we observe that there is a statistically significant relationship between the treatment (hatchery) and the number of observed events (fish with damage to the fins) ( $p = 0.0009$  for caudal fin;  $p = 0.0005$  for pelvic fin). Fin damage can result from aggression but also stress (60). Also, can cause detrimental effects on growth and survival by increasing susceptibility to opportunistic infections. Although in the tilapias of the present study, damage to the pelvic fins was not recorded. According to the risk traffic light proposed here (Table 8), in fish weighing 5 to 152 g (growth stages 2 to 5 in tilapias in this study), like those from Gutierrez-Rabadan et al.

(15) (weight range 5–152 g), is considered a risky situation when 12% of the population show damage to the anal and caudal fins, respectively, a lower value than those observed in lumpfish (52 and 53%), which is probably why the Chi-square test showed a statistically significant relationship. When calculating the RR (5.16 for tail fin with CI of 1.87–14.24; 3.11 for pelvic fin with CI of 1.44–6.69), it is observed that culture management directly impacts fish welfare indicators, that is, it is a risk factor to the population.

The integration of the indicators allows qualifying of the general condition of wellbeing obtained in each treatment. Thanks to this analysis tool are possible to observe that the most appropriate feeding protocol depends on the phase of the life cycle of the tilapia. In the present study, the control treatment followed the recommendation proposed for the species (22). Using this regimen, the best growth was obtained in fingerlings and on-growing 2 phases which coincide with the best general welfare state rating. (Very well and well, respectively). However, for the juvenile and on-growing 1 phase, the best biomass production, and general welfare rating were obtained with the overfeeding treatment A (120%), indicating a higher nutritional requirement, particularly protein, in these phases (49). In practice, this form of data visualization and analysis could be useful for making and correcting decisions in crop management and planning an adequate feeding regimen, thanks to the evaluation of the general welfare status of the animals with operational indicators.

In a high-risk productive activity such as aquaculture, increasing efficiency and reducing risks are permanent goals. Usage of operational welfare indicators offers the use of easily obtained biological information. Allowing the identification of “red flags” promptly to avoid breaks in the production cycles. The epidemiological analysis showed its potential application and the methodology to obtain specific alarm values related to the characteristics of each farm (growth stage, feeding amount, the season of the year, etc.). Generating animal welfare programs proposes the opportunity to systematically test the inputs for production (genetic lines, type of feed, tolerance to environmental factors, etc.) and increase efficiency with a focus on constant improvement.

## CONCLUSIONS AND ANIMAL WELFARE IMPLICATIONS

To have information that allows people responsible for tilapia production, either on the farm or in experiments to make prompt decisions and evaluations is a must. Mortality incidence, weight reduction, K reduction, and damage to caudal and anal fins could be used as laboratory welfare indicators or operational welfare indicators in the cultivation of *O. niloticus* applying an epidemiological approach. The feeding rate used directly affects production and the welfare indicators, and in a different way depending on the growth phase. As a result of this study, the epidemiological approach seems to be a valuable tool for production. The proposed risk traffic light method could have great



potential, with the suggested limits for WI's concerning the individuals present in the culture pond, allowing progressive evaluation and decision-making to correct risky situations that arise.

## DATA AVAILABILITY STATEMENT

The datasets presented in this article are not readily available because The dataset is part of ongoing research and is still being used to complete it. Requests to access the datasets should be directed to ar.martinez@ugto.mx.

## ETHICS STATEMENT

The animal study was reviewed and approved by Institutional Committee of Bioethics in Research of the University of Guanajuato (code: CIBIUG-A59-2020).

## AUTHOR CONTRIBUTIONS

LF-G: investigation, data obtaining, data analysis, and writing—original draft preparation. JC-C: investigation and

data collection. CP-J: data analysis, writing, and partial economical support. PA-R: data analysis and writing—review and editing. JM-L: data analysis. PA-A: design, construction of experimental systems, and writing—review and editing. RM-Y: conceptualization, methodology, formal analysis, investigation, writing—original draft preparation, and project administration. All authors contributed to the article and approved the submitted version.

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## REFERENCES

- Mota VC, Limbu P, Martins CIM, Eding EH, Verreth JAJ. The effect of nearly closed RAS on the feed intake and growth of Nile tilapia (*Oreochromis niloticus*), African catfish (*Clarias gariepinus*) and European eel (*Anguilla anguilla*). *Aquac Eng.* (2015) 68:1–5. doi: 10.1016/j.aquaeng.2015.06.002
- Food and Agriculture Organization (FAO). *The State of World Fisheries and Aquaculture*. Rome: FAO (2020).
- Timmons MB, Ebeling JM. *Recirculating Aquaculture*. Cayuga Aqua Ventures. Ithaca, NY: Northeastern Regional Aquaculture Center (2010).
- Broom DM. Animal welfare: concepts and measurement. *J Anim Sci.* (1991) 69:4167–75. doi: 10.2527/1991.69104167x
- Carenzi C, Verga M. Animal welfare: review of the scientific concept and definition. *Ital J Anim Sci.* (2009) 8(sup1):21–30. doi: 10.4081/ijas.2009.s1.21
- World Organization of Animal Health (OIE). *Animal Welfare*. (2020). Available online at: <https://www.oie.int/en/what-we-do/animal-health-and-welfare/animal-welfare/> (accessed October 30, 2020).
- Arvizu LO, Téllez ER. *Animal Welfare in México: A Normative View*. Universidad Nacional Autónoma de México. Mexico City: UNAM (2016). p. 145.
- De Brlyne N. *Veterinary Aspects of Aquatic Animals Health and Welfare, Aquaculture and Ornamental Fish Trade*. Brussels: Federation of Veterinarians of Europe (2014).
- Mellor DJ, Beausoleil NJ, Littlewood K, McLean AN, McGreevy PD, Jones B, et al. The 2020 five domains model: Including human – animal interactions in assessments of animal welfare. *Animals.* (2020) 10:1–24. doi: 10.3390/ani10101870
- Noble C, Gismervik K, Iversen MH, Kolarevic J, Nilsson J, Stien LH, et al. *Welfare Indicators for Farmed Atlantic Salmon: Tools for Assessing Fish Welfare*. Tromsø: Norwegian Seafood Research Fund (2018).
- Ellingsen K, Grimsrud K, Nielsen HM, Mejdell C, Olesen I, Honkanen P, et al. Who cares about fish welfare? A Norwegian study. *Br Food J.* (2015) 117:257–73. doi: 10.1108/BFJ-08-2013-0223
- Broom DM. Cognitive ability and sentience: Which aquatic animals should be protected? *Dis Aquac Org.* (2007) 75:99–108. doi: 10.3354/dao075099
- Damsgard B, Juell JE, Braastad BO. *Welfare in Farmed Fish*. Tromsø: Norwegian Institute of Fisheries and Aquaculture Research (2006).
- Martins CLM, Galhardo L, Noble C, Damsgard D, Spedicato MT, Zupa W, et al. Behavioral indicators of welfare in farmed fish. *Fish Physiol Biochem.* (2012) 38:17–41. doi: 10.1007/s10695-011-9518-8
- Gutierrez-Rabadán C, Spreadbury C, Consuegra S, García-de-Leaniz C. Development, validation and testing of an Operational Welfare Score Index for farmed lumpfish *Cyclopterus lumpus* L. *Aquaculture.* (2021) 531:1–12. doi: 10.1016/j.aquaculture.2020.735777
- Branson EJ. *2008 Fish Welfare*. Monmouthshire: Blackwell Publishing.
- Luo G, Gao Q, Wang C, Liu W, Sun D, Li L, et al. Growth, digestive activity, welfare, and partial cost-effectiveness of genetically improved farmed tilapia (*Oreochromis niloticus*) cultured in recirculating aquaculture system and an indoor biofloc system. *Aquaculture.* (2014) 423:1–7. doi: 10.1016/j.aquaculture.2013.11.023
- Nash RDM, Valencia AH, Jeffrey AJ. The Origin of Fulton's Condition Factor: Setting the Record Straight. *Fisheries.* (2006) 31:236–8.
- Roberts RJ. *Fish Pathology*. Oxford: Wiley-Blackwell (2012).
- Volkoff H, Peter RE. Feeding behavior of fish and its control. *Zebrafish.* (2006) 3:1–10. doi: 10.1089/zeb.2006.3.131
- Stien LH, Bracke MBM, Folkedal O, Nilsson J, Oppedal F, Torgersen T, et al. Salmon Welfare Index Model (SWIM 10): a semantic model for overall welfare assessment of caged Atlantic salmon: review of the selected welfare indicators and model presentation. *Rev Aquac.* (2013) 5:33–57. doi: 10.1111/j.1753-5131.2012.01083.x
- Food and Agriculture Organization (FAO). *Aquaculture Feed and Fertilizer Resources Information System*. (2019). Available online at: <http://www.fao.org/fishery/affris/species-profiles/nile-tilapia/tables/en/> (accessed June 30, 2019).
- Sánchez JA, López-Olmeda JF, Blanco-Vives B, Sánchez-Vázquez FJ. Effects of feeding schedule on locomotor activity rhythms and stress response in sea bream. *Physiol Behav.* (2009) 98:125–9. doi: 10.1016/j.physbeh.2009.04.020
- Vera LM, De Pedro N, Gómez-Milán E, Delgado MJ, Sánchez-Muros MJ, Madrid JA, et al. Feeding entrainment of locomotor activity rhythms, digestive enzymes and neuroendocrine factors in goldfish. *Physiol Behav.* (2007) 90:518–24. doi: 10.1016/j.physbeh.2006.10.017
- Gao Y, Wang Z, Hur JW, Lee JY. Body composition and compensatory growth in Nile tilapia *Oreochromis niloticus* under different feeding intervals. *Chin J Oceanol Limnol.* (2015) 33:945–56. doi: 10.1007/s00343-015-4246-z
- Roh HJ, Park J, Kim A, Kim N, Lee Y, Kim BS, et al. Overfeeding-induced obesity could cause potential immuno-physiological disorders



- in rainbow trout (*Oncorhynchus mykiss*). *Animals*. (2020) 10:1–15. doi: 10.3390/ani10091499
27. Sriyasa P, Chitmanat C, Whangchai N, Promya J, Lebel L. Effect of water de-stratification on dissolved oxygen and ammonia in tilapia ponds in Northern Thailand. *Int Aqua Res*. (2015) 7:287–99. doi: 10.1007/s40071-015-0113-y
  28. Katz DL, Elmore JG, Wild DMG, Lucan SC. *Jekel's Epidemiology, Biostatistics, Preventive Medicine and Public Health*. Philadelphia, United States: Elsevier Saunders (2014).
  29. Thrusfield M. *Veterinary Epidemiology*. 4th ed. Edinburg: Wiley Blackwell (2018).
  30. Leishman EM, van Staaveren N, Osborne VR, Wood BJ, Baes CF, Harlander-Mataushek A. The prevalence of integument injuries and associated risk factors among Canadian Turkeys. *Front Vet Sci*. (2022) 8:757776. doi: 10.3389/fvets.2021.757776
  31. Schmidt CO, Kohlmann T. When to use the odds ratio or the relative risk? *Int J Public Health*. (2008) 53:165–7. doi: 10.1007/s00038-008-7068-3
  32. Hisano H, Barbosa PTL, Hayd LA. Evaluation of Nile tilapia in monoculture and polyculture with giant fresh water prawn in biofloc technology system and in recirculation aquaculture system. *Int Aqua Res*. (2019) 11:335–46. doi: 10.1007/s40071-019-00242-2
  33. Cañon-Jones HA, Noble C, Damsgard B, Pearce GP. Investigating the influence of predictable and unpredictable feed delivery schedules upon the behaviour and welfare of Atlantic salmon parr (*Salmo salar*) using social network analysis and fin damage. *Appl Anim Behav Sci*. (2012) 138:132–140. doi: 10.1016/j.applanim.2012.01.019
  34. Stejskal V, Matoušek J, Prokešová M, Podhorec B, Křišťan J, Polícar T, et al. Fin damage and growth parameters relative to stocking density and feeding method in intensively cultured European perch (*Perca fluviatilis* L.). *J Fish Dis*. (2019) 43:1–10. doi: 10.1111/jfd.13118
  35. Pedrazzani AS, Quintiliano MH, Bolfe F, Sans ECO, Molento CFM. Tilapia on-farm welfare assessment protocol for semi-intensive production systems. *Front Vet Sci*. (2020) 7:1–16. doi: 10.3389/fvets.2020.606388
  36. Espinoza-Moya EA, Angel-Sahagún CA, Mendoza-Carrillo JM, Albertos-Alpuche PJ, Alvarez-González CA, Martínez-Yáñez AR. Herbaceous plants as part of biological filter for aquaponic system. *Aquac Res*. (2016) 47:1716–26. doi: 10.1111/are.12626
  37. García F, Romera DM, Gozi KS, Onaka EM, Fonseca FS, Schallch SHC, et al. Stocking density of Nile tilapia in cages placed in a hydroelectric reservoir. *Aquaculture*. (2013) 411:51–6. doi: 10.1016/j.aquaculture.2013.06.010
  38. Coyle SD, Durbin RM, Tidwell JH. *Anesthetics in Aquaculture*. Southern Regional Aquaculture Center. Lexington: Kentucky State University (2004).
  39. Matthews M, Varga ZM. Anesthesia and Euthanasia in Zebrafish. *J Inst Lab Anim Res*. (2012) 53:192–204. doi: 10.1093/ilar.53.2.192
  40. Olesen I, Alfnes F, Røra MB, Kolstad K. Eliciting consumers' willingness to pay for organic and welfare-labelled salmon in a non-hypothetical choice experiment. *Livest Sci*. (2010) 127:218–26. doi: 10.1016/j.livsci.2009.10.001
  41. Caglan A, Benli K, Köksal G. The acute toxicity of ammonia on tilapia (*Oreochromis niloticus* L) larvae and fingerlings. *Turkish J Vet Anim Sci*. (2003) 29:339–44. doi: 10.3906/vet-0306-44
  42. Mengitsu SB, Mulder HA, Benzie JAH, Komen H. A systematic literature review of the major factors causing yield gap by affecting growth, feed conversion ratio and survival in Nile tilapia (*Oreochromis niloticus*). *Rev Aquac*. (2020) 12:1–18. doi: 10.1111/raq.12331
  43. Abdelghany AE, Ahmad MH. Effects of feeding rates on growth and production of Nile tilapia, common carp and silver carp polycultured in fertilized ponds. *Aquac Res*. (2002) 33:415–23. doi: 10.1046/j.1365-2109.2002.00689.x
  44. Dolomatov S, Zukow W, Dzierzanowski M, Mieszkowski J, Muszkieta R, Klimezyk M. Roles of nitrates in the adaptation of fish to hypoxic conditions. *Water Resour*. (2016) 43:177–83. doi: 10.1134/S0097807816120046
  45. Foss A, Siikavuopio SI, Saether BS, Evensen TH. Effect of chronic ammonia exposure on growth in juvenile Atlantic cod. *Aquaculture*. (2004) 237:179–89. doi: 10.1016/j.aquaculture.2004.03.013
  46. Oberle M, Salomon S, Ehrmaier B, Richter P, Lebert M, Strauch SM. Diurnal stratification of oxygen in shallow aquaculture ponds in central Europe and recommendations for optimal aeration. *Aquaculture*. (2019) 501:482–7. doi: 10.1016/j.aquaculture.2018.12.005
  47. Liang JY, Chien YH. Effects of feeding frequency and photoperiod on water quality and crop production in a tilapia-water spinach raft aquaponics system. *Int Biodeterior Biodegradation*. (2013) 85:693–700. doi: 10.1016/j.ibiod.2013.03.029
  48. Castillo JDA, do Nascimento TMT, Mansano CFM, Sakomura NK, da Silva NP, Fernandes JDK. Determining the daily digestive protein intake for Nile tilapia at different growth stages. *Boletim do Instituto de Pesca*. (2017) 44:54–66. doi: 10.20950/1678-2305.2017.54.63
  49. Van Trung D, Dui NT, Hao NT, Glencross B. Development of a nutritional model to define the energy and protein requirements of tilapia, *Oreochromis niloticus*. *Aquaculture*. (2011) 320:69–75. doi: 10.1016/j.aquaculture.2011.07.029
  50. Houlihan D, Boujard T, Jobling M. *Food Fish Intake*. Oxford: Blackwell Science (2001).
  51. Saravan S, Geurden I, Figueiredo-Silva AC, Kaushik SJ, Haidar MN, Verreth JAJ, et al. Control of voluntary feed intake in fish: a role for dietary oxygen demand in Nile tilapia (*Oreochromis niloticus*) fed diets with different macronutrient profiles. *Br J Nutr*. (2012) 108:1519–29. doi: 10.1017/S0007114511006842
  52. El-Dakar AY, Shlaby SM, Mostafa EE, Abdel-Aziz MF. The optimum level of dietary protein and feeding for improving the growth performance and feed efficiency of juveniles Hybrid tilapia (*Oreochromis niloticus* × *Oreochromis aurea*) reared in brackish water. *Egypt J Aquac Biol Fish*. (2021) 25:839–56. doi: 10.21608/ejaf.2021.195285
  53. Gonçalves-de-Freitas G, Bolognesi MC, dos Santos Gauy AC, Brandão ML, Gaiquinto PC, Fernandes-Castilho M. Social Behavior and Welfare in Nile Tilapia. *Fishes*. (2019) 4:2–14. doi: 10.3390/fishes4020023
  54. Almazán-Rueda P, Schrama, JW, Verreth, JAJ. Behavioural responses under different methods and light regimes of the African catfish (*Clarias gariepinus*) juveniles. *Aquaculture*. (2004) 231:349–57. doi: 10.1016/j.aquaculture.2003.11.016
  55. Hoyle I, Oidmann B, Ellis T, Turnbull J, North B, Nikolaidis J, Nowles TG. A validated macroscopic key to assess fin damage in farmed rainbow trout (*Oncorhynchus mykiss*). *Aquaculture*. (2007) 270:142–8. doi: 10.1016/j.aquaculture.2007.03.037
  56. Tesfahun A, Tesfegen M. Food and feeding habits of Nile tilapia *Oreochromis niloticus* (L) in Ethiopian water bodies: a review. *Int J Fish Aquat Stud*. (2018) 6:43–7. Available online at: <https://www.fisheriesjournal.com/archives/2018/vol6issue1/PartA/5-6-54-506.pdf>
  57. Rey S, Tresaurer J, Pattillo C, McAdam BJ. Using model selection to choose a size-based condition index that is consistent with operational welfare indicators. *J Fish Biol*. (2021) 99:782–95. doi: 10.1111/jfb.14761
  58. Bemis WE, Hilton EJ, Brown B, Arrindell R, Richmond AR, Little CD, et al. Methods for Preparing Dry, Partially Articulated Skeletons of Osteichthyans, with Notes on Making Ridewood Dissections of the Cranial Skeleton. *Copeia*. (2004) 3:603–9. doi: 10.1643/CI-03-054R1
  59. North BP, Turnbull JF, Ellis T, Porter MJ, Migaud H, Bron J, et al. The impact of stocking density on the welfare of rainbow trout (*Oncorhynchus mykiss*). *Aquaculture*. (2006) 255:466–79. doi: 10.1016/j.aquaculture.2006.01.004
  60. Turnbull J, Bell A, Adams C, Bron J, Huntingford F. Stocking density and welfare of cage farmed Atlantic salmon: application of a multivariate analysis. *Aquaculture*. (2005) 243:121–32. doi: 10.1016/j.aquaculture.2004.09.022
  61. Pelis RM, McCormick SD. Fin development in stream- and hatchery-reared Atlantic salmon. *Aquaculture*. (2003) 220:525–36. doi: 10.1016/S0044-8486(02)00625-7
  62. St.Hilaire S, Ellis T, Cooke A, North BP, Turnbull JF, Knowles T, et al. Fin erosion on rainbow trout on commercial trout farms in the United Kingdom. *Vet Rec*. (2006) 159:446–51. doi: 10.1136/vr.159.14.446
  63. Latremouille DN. Fin erosion in aquaculture and natural environments. *Rev Fish Sci*. (2003) 11:315–35. doi: 10.1080/10641260390255745
  64. Van de Nieuwegiesse PG, Boerlage AS, Verreth JAJ, Schrama JW. Assessing the effects of a chronic stressor, stocking density, on welfare indicators of juvenile African catfish, *Clarias gariepinus* Burchell. *Appl Anim Behav Sci*. (2008) 115:233–43. doi: 10.1016/j.applanim.2008.05.008
  65. Noble C, Kadri S, Mitchell DF, Huntingford FA. Growth, production and fin damage in cage-held 0 + Atlantic salmon pre-smolts (*Salmo salar* L) fed either a) on demand, or b) to a fixed satiation restriction regime: data from a commercial farm. *Aquaculture*. (2008) 275:163–8. doi: 10.1016/j.aquaculture.2007.12.028

66. Noble C, Kadri S, Mitchell DF, Huntingford FA. The effect of feed regime on the growth and behaviour of 1 + Atlantic salmon postsmolts (*Salmo salar* L.) in semi-commercial sea cages. *Aquac Res.* (2007) 38:1686–91. doi: 10.1111/j.1365-2109.2007.01833.x
67. Noble C, Mizusawa K, Suzuki K, Tabata M. The effect of differing self-feeding regimes on the growth, behaviour and fin damage of rainbow trout held in groups. *Aquaculture.* (2007) 264:214–22. doi: 10.1016/j.aquaculture.2006.12.028
68. Person-Le Ruyet J, Le Bayo N, Gros S. How to assess fin damage in rainbow trout, *Onchorynchus mykiss*? *Aquat Living Resour.* (2007) 20:191–5. doi: 10.1051/alr:2007031

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# Use of male-to-female sex reversal as a welfare scoring system in the protandrous farmed gilthead sea bream (*Sparus aurata*)

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Gilthead sea bream is a highly cultured marine fish throughout the Mediterranean area, but new and strict criteria of welfare are needed to assure that the intensification of production has no negative effects on animal farming. Most welfare indicators are specific to a given phase of the production cycle, but others such as the timing of puberty and/or sex reversal are of retrospective value. This is of particular relevance in the protandrous gilthead sea bream, in which the sex ratio is highly regulated at the nutritional level. Social and environmental factors (e.g., contaminant loads) also alter the sex ratio, but the contribution of the genetic component remains unclear. To assess this complex issue, five gilthead sea bream families representative of slow/intermediate/fast growth were grown out with control or a plant-based diet in a common garden system from early life to the completion of their sexual maturity in 3-year-old fish. The plant-based diet highly enhanced the male-to-female sex reversal. This occurred in parallel with the progressive impairment of growth performance, which was indicative of changes in nutrient requirements as the result of the different energy demands for growth and reproduction through development. The effect of a different nutritional and genetic background on the reproductive performance was also assessed by measurements of circulating levels of sex steroids during the two consecutive spawning seasons, varying plasma levels of 17 $\beta$ -estradiol (E<sub>2</sub>) and 11-ketotestosterone (11-KT) with age, gender, diet, and genetic background. Principal component analysis (PCA) of 3-year-old fish displayed a gradual increase of the E<sub>2</sub>/11-KT ratio from males to females with the improvement of nutritional/genetic background. Altogether, these results support the use of a reproductive tract scoring system for leading farmed fish toward their optimum welfare condition, contributing to improving the productivity of the current gilthead sea bream livestock.

## KEYWORDS

protandrous fish, sex reversal, sex steroids, welfare scoring, plant-based diets, nutrition and genetics interactions

## Introduction

Fish farming has evolved as one of the most sustainable production sectors because of its high feed conversion efficiency and its lower carbon footprint when compared with other animal production systems (1). Nonetheless, aquaculture production is becoming more intensified to meet the increased global demand for fish protein aquaculture (2). It is, thereby, important to encompass the development of aquaculture with novel and stricter criteria of welfare for the simultaneous improvement of aquaculture productivity and welfare of farmed fish (3, 4). Certainly, important research efforts are now conducted within the AquaIMPACT H2020 project for integrating information from fish breeding and nutrition to promote the production of healthier and more robust fish with higher phenotypic plasticity to cope with a challenging environment. This includes the use of gut microbiota as a reliable criterion to evaluate the success of selective breeding for improving the performance and competitiveness of European aquaculture (5, 6). At the same time, novel fish feed formulations, and epigenetic and behavioral approaches are widely applied to assure a more ethical and sustainable aquaculture production with the increase of water temperature and hypoxia as major environmental problems in coastal marine ecosystems (7–9). Thus far, fish welfare assessment is still in the infancy state due to the limited understanding of the diverse fish species' welfare-relevant biology (10). However, several benchmarking systems on key performance indicators (KPIs) based on growth performance, survival rates, and external tissue damage (skin/fin erosion) have been currently validated in salmon, but also in Mediterranean fish species, to ensure that farmed fish are not far from their optimum welfare (11–14). Otherwise, behavioral indicators are becoming especially useful for alerting farmers that something is potentially wrong and warrants investigation before significant welfare issues can occur (15, 16). In any case, the best monitoring solution, especially those based on telemetry techniques and bio-loggers for tracking swimming activity and/or heart or breathing rates, highly depends on the asked question, species biology, and culture system (17).

A common feature of welfare indicators is their accuracy for a given time and culture condition, reflecting immediacy rather than a historical background. However, the success of reproductive performance, measured by means of fecundity, puberty onset, and sex reversal, can also have a high value from a retrospective point of view, as it is the end point of a complex cascade of developmental events that encompass a wide range of biotic and abiotic factors (18–22). In particular, the sex ratio in gonochoristic fish tends to be balanced in optimal culture conditions (23), although it can be affected by chemicals (24) and other environmental factors such as rearing density, temperature, pH, oxygen, and diet composition (25–29). Similarly, sex reversal in hermaphrodite species, such

as in the protandrous gilthead sea bream, is socially controlled and endocrine-regulated by the circulating levels of estradiol ( $E_2$ ) and 11-ketotestosterone (11-KT) (30), and intriguingly the exposure to synthetic estrogens (e.g., 17 $\alpha$ -ethynylestradiol) prevents the male-to-female sex reversal (31). Stress may also influence the onset of sex change through the mediation of cortisol, although the exact mechanisms in which it may act as a mediator in sex change remain to be fully established (32). Otherwise, puberty onset is determined by genetic factors and controlled by the nutritional status and/or the body's growth (33). Thus, similar to what occurs in humans, better welfare conditions for fish entail an increase in their growth before reaching their first sexual maturation (34, 35). However, early puberty, in particular in males, occurs in several species kept under aquaculture conditions and is often associated with a final growth retardation or health risks (36, 37). Moreover, the age of puberty can be controlled in farmed fish by selective breeding and feeding level (38–40), and a recent gilthead sea bream study stated that plant-based diets have the potential to alter the sex steroid profile during the pre-spawning and spawning period, promoting the enhanced male-to-female sex reversal when the presence of powerful functional females is compromised by the diet (41).

Taking into account all the above findings, we had herein a double objective: (i) to assess how the male-to-female sex reversal is affected by nutrition and genetics in the protandrous gilthead sea bream and (ii) to provide new insights into the use of male-to-female sex reversal and population sex ratio as a reliable best practice framework for animal welfare certification of a highly cultured farmed fish in all the Mediterranean basin. The rationale for this procedure is that the complex balance of environmental variables that regulate animal welfare conditions can also affect sex change in sequential hermaphrodites. To pursue this issue, sex reversal was monitored in fish families with different nutritional backgrounds and different heritable growth within the PROGENSA<sup>®</sup> selection program (42), which co-selected among other traits with changes in gut microbiota composition and metabolic plasticity (5, 43), as well as swimming performance and aerobic scope (44, 45).

## Materials and methods

### Diets

Two extruded diets were formulated and produced by BioMar (BioMar Process Innovation Technical Center, Brande, Denmark), at a range of pellet sizes corresponding to the respective fish size as fish grew (i.e., 1.9, 3, 4.5, and 6.5 mm). Both diets were isonitrogenous, isolipidic, and isoenergetic and met all known nutritional requirements of gilthead sea bream. Fish meal (FM) was included at 23% in the control diet (D1) and at 3% in the experimental diet (D2). The addition of

fish oil (FO) was 14.1% for D1, and 3.9% for D2 with the replacement of rapeseed oil, decreasing EPA+DHA content from 3.8 to 1.02%. Lysine, methionine, choline, lecithin, and monocalcium phosphate were added to D2 to reach D1 levels ([Supplementary Table S1](#)).

## Experimental setup and sample collection

Broodstock crossings of eight (two females and six males) and five (three females and two males) fish from the gilthead sea bream PROGENSA<sup>®</sup> selection program rendered sixteen families with differences in heritable growth, as described elsewhere ([42](#)). Briefly, juvenile fish of these families, previously genotyped by DNA fin analysis, were individually tagged (dorsal muscle) with passive integrated transponders (PITs) (ID-100A 1.25 Nano Transponder, Trovan, Madrid, Spain) and maintained in a common garden system fed D1 or D2 diets in replicate 3,000-L tanks under the natural photoperiod and temperature conditions (latitude 40° 5' N; 0° 10' E) at the Institute of Aquaculture Torre de la Sal (IATS), over the course of a 12-month feeding trial (September 2017 to September 2018). At this end, five families were selected by their growth trajectories during this period as a representative of fast growth (e5e2, e6e2; 158 and 49 individuals, respectively), intermediate growth (c2c7, e4e1; 91 and 174 individuals, respectively), and slow growth (c4c3; 60 individuals), and were distributed at similar family in the common garden system. Growth performance and reproductive status were assessed in these families until the completion of sexual maturation in 3-year-old fish (December 2020).

Over the course of the entire trial, the concentration of water oxygen was always higher than 80% saturation. Fish were fed by automatic feeders 1–2 times per day and 3–7 days per week according to fish size and season, with the ratio adjusted weekly to a level close to satiation. The final rearing density was 19–20 kg/m<sup>3</sup>. Fish body weight and body length were measured individually using an FR-200 FishReader W (Trovan, Madrid, Spain) at different monthly intervals during the first (Age +1), second (Age +2), and third year (Age +3) of the production cycle. At the time of a maximum number of spermiating fish (December), overnight-fasted fish were anesthetized with 100 mg/L MS-222 (Sigma, Saint Louis, MO, USA) for blood extraction and sexing by stripping. It is a non-lethal and accurate sexing method at this time and developmental stage as almost all males are fluent by stripping, in coincidence with the annual peak of E<sub>2</sub> in females and 11-KT in males that resulted in a minimum presence (<5%) of intersex fish ([41](#)). Blood was collected (100 fish/diet for Age +2 fish, 150 fish/diet for Age +3 fish) from caudal vessels using heparinized syringes and centrifuged at 3,000 × g for 20 min

at 4°C, and plasma aliquots were stored at −20°C until sex steroid analyses.

All procedures were approved by the Ethics and Animal Welfare Committee of IATS and CSIC. They were carried out in the IATS's registered aquaculture infrastructure facility (code ES120330001055) in accordance with the principles published in the European Animal Directive (2010/63/EU) and Spanish Laws (Royal Decree RD53/2013) for the protection of animals used in scientific experiments.

## Sex steroids

Quantification of plasma sex steroids was performed by enzyme immunoassays (EIAs) as described by Rodríguez et al. ([46](#)) for 11-KT and by Molés et al. ([47](#)) for E<sub>2</sub>. Briefly, steroids were extracted from 100 µl plasma in 1 ml methanol and supernatants were dried and reconstituted in EIA buffer (0.1 M potassium phosphate, pH 7.4 containing 0.01% sodium azide, 0.4 M NaCl, 0.001 M EDTA, and 0.1% BSA). Steroid standards were purchased from Sigma-Aldrich. Mouse anti-rabbit immunoglobulin monoclonal antibody (Ab), rabbit steroid Abs (T-Ab, 11-KT-Ab, and E<sub>2</sub>-Ab), and enzymatic tracers [steroid acetylcholinesterase (AChE) conjugates: T-AChE, 11-KT-AChE, and E<sub>2</sub>-AChE] were obtained from Vitro S.A. (Sevilla, Spain). Samples and standard curves of 11-KT (0.0001–1.0 ng/ml) and E<sub>2</sub> (0.005–9.0 ng/ml) were run in duplicate. Optical density was read at 405 nm using a microplate reader (Bio-Rad 3550). The inter-assay coefficients of variation at 50% of binding were 5.02% ( $n = 10$ ) with a 0.88 slope for 11-KT and 5.97% ( $n = 10$ ) with a 0.68 slope for E<sub>2</sub>.

## Statistical analysis

Statistical analysis was performed using SigmaPlot version 14.0 (Systat Software, San Jose, CA, USA) with all  $P$ -values set to 0.05 for significance determination. Body weight and sex ratio differences between both dietary groups were assessed by means of the Student's  $t$ -test. Male and female body weight differences within each age, diet, and family (or grouped families) were determined by means of the Student's  $t$ -test. One-way ANOVA, followed by a Holm-Sidak *post-hoc* test, was conducted in order to assess significant differences in male-to-female sex steroids (11-KT and E<sub>2</sub>) concentration between families of the same diet and age. Sex steroid differences between diets within each family were assessed by means of a Student's  $t$ -test. For evidencing gradation in sex steroids and body weight between diets, families, and males/females of each family, a principal component analysis (PCA) was performed using EZinfo version 3.0 (Umetrics, Umea, Sweden). Differences in E<sub>2</sub>/11-KT quotient between families and diets were assessed



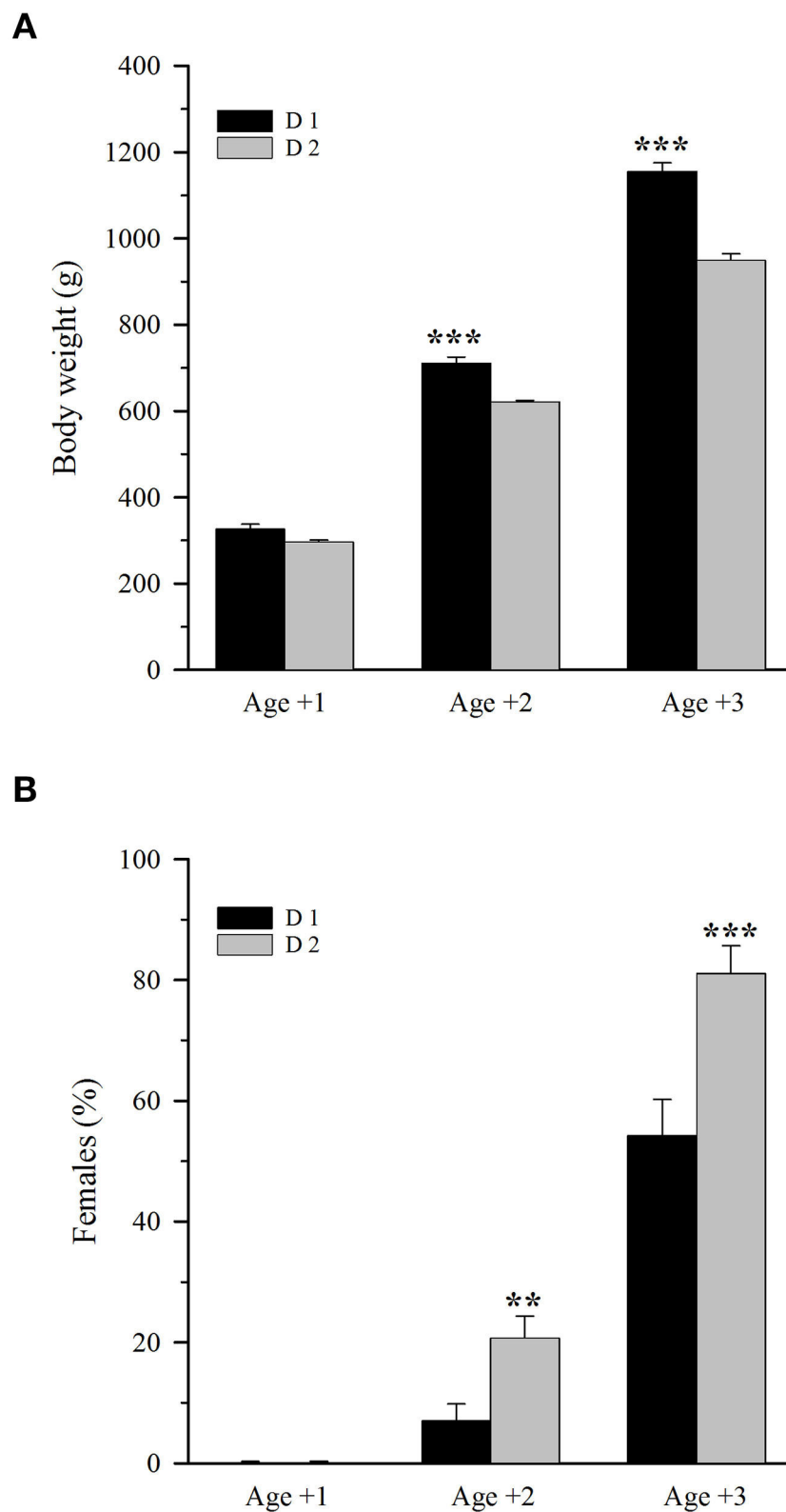


FIGURE 1

(A) Body weight and (B) female percentage in 1-, 2-, and 3-year-old gilthead sea bream fed an FM/FO diet (D1) or a plant-based diet (D2). Values are the mean  $\pm$  SEM of three tanks per diet ( $n = 261$  (D1) –271 (D2) fish/diet at Age +1, 202 (D1) –207 (D2) fish/diet at Age +2, 138 (D1) –146 (D2) fish/diet at Age +3). Asterisks indicate significant differences between the experimental diets within each age (Student's  $t$ -test, \*\* $P < 0.01$ , \*\*\* $P < 0.001$ ).



by means of a one-way ANOVA, followed by a Holm-Sidak *post-hoc* test.

## Results

### Growth and sex ratio progression

Dietary treatment had a clear effect on fish size regardless of their genetic background. Thus, considering all fish families as a whole under each dietary treatment, fish fed D2 consistently showed a lower body weight than that fed D1. Differences in body weight were not statistically significant at Age +1 (Figure 1A). However, at Age +2 and Age +3, the body weight difference between both dietary groups was 12.5 and 18%, respectively, with the growth performance negatively affected ( $P < 0.001$ ) by the plant-based diet. Regarding sex, all fish were males at Age +1 regardless of diet (Figure 1B). However, the plant-based diet largely enhanced the male-to-female sex reversal, resulting in a significantly ( $P < 0.01$ ) higher female percentage at both Age +2 (20.9 vs. 7.2%) and Age +3 (81.2 vs. 54.4 %).

### Sexual dimorphism: Body weight and sex steroids

Data on body weight at Age +2 showed no significant differences between males and females fed D1, neither when considering all fish families as a whole nor analyzing each family separately (Figure 2A). However, in fish fed D2, a significant body weight sexual dimorphism toward larger females was evidenced comparing male and female populations (Figure 2B). It must be noted that this feature was significant only for the fast-growing family e6e2 (Figure 2B).

At Age +3, gilthead sea bream females had approximately 15% more body weight than their male counterparts regardless of diet (Figure 3). In fish fed D1, this clear sexual dimorphism was mostly observed in families c2c7, e5e2, and e6e2 (Figure 3A), whereas for fish fed D2, it was supported by a significantly higher body weight of females of families e4e1, e5e2, and e6e2 (Figure 3B).

E<sub>2</sub> plasma levels in Age +2 males were quite similar, with the only significant difference between families e5e2 and e4e1 when fed the D1 diet (Figure 4A). For Age +2 and Age +3, plasma levels of E<sub>2</sub> increased around 130% in males fed D1, whilst those of males fed D2 decreased to around 50% (Figure 4B). Age +3 male fish fed D1 showed higher E<sub>2</sub> plasma levels than fish fed D2, with no differences among families for the same dietary group. Male plasma levels of 11-KT at Age +2 showed a gradually decreasing trend from slow- to fast-growing families, with e5e2 and e6e2 families significantly different from the slow and intermediate families within both diets (Figure 4C). Comparison

between diets showed that fish fed D2 had significantly higher levels of 11-KT in the case of slow- and intermediate-growing families, and the same trend, although non-significant, was maintained in fast-growing families (Figure 4C). Male 11-KT levels generally increased from Age +2 to Age +3, keeping the gradual decrease of 11-KT from slow- to fast-growing families (Figure 4D). Four of five families of Age +3 males displayed higher 11-KT levels when fed D2.

Results of female sex steroids at Age +2 were not conclusive due to the low number of gilthead sea bream that underwent sex change from male to female (Figures 5A, C). Nonetheless, at Age +3, female E<sub>2</sub> plasma levels showed a clear diet effect, with fish fed D1 having significantly higher levels within all families (Figure 5B). A genetic effect was reduced to fish fed D2, with higher circulating levels of E<sub>2</sub> in slow-growing fish than in fast- and intermediate-growing families (Figure 5B). 11-KT plasma levels of all 3-year-old female families were below 0.05 ng/ml, and no differences were observed between families and dietary groups (Figure 5D).

### Sex steroids ratio

The PCA of plasma sex steroid levels at Age +3 showed that >86% of the total variance was explained by the two first components (Figure 6A). Each fish was categorized according to its diet, sex, and family group. For better representation, e<sub>x</sub>e<sub>y</sub> families were joined as a unique fast-growth family group, while c<sub>x</sub>c<sub>y</sub> families were joined as a unique slow-growth family group. Movement along the X-axis (60.32% of total variance) accounted for plasma sex steroid levels, with the highest values of E<sub>2</sub> on the left and the maximum values of 11-KT on the right. This sex steroid distribution clearly discriminated females (black and orange boxes) on the left and males (green and blue boxes) on the right. The Y-axis (25.85% of total variance) accounted for body weight changes, separating the fast-growth families at the top from the slow-growth families at the bottom. In other words, the resulting plasma E<sub>2</sub>/11-KT ratio was affected by both diet and genetics, increasing this hormonal quotient with the improvement of both the nutritional and genetic background (Figure 6B). In males, a genetic effect was not seen, but the diet effect persisted with a decreased E<sub>2</sub>/11-KT ratio in fish fed the plant-based diet. This occurred in parallel with a genetically regulated male-to-female sex reversal, displaying fast-growing families fed D1 a significantly lower percentage of phenotypic females (54%) than slow-growing families (65%) fed the same diet (Figure 6C). This percentage of phenotyped females reached a plateau (79–81%) in fish fed D2 regardless of their genetic background, which is indicative of a genetic and nutrition interaction according to which the plasma E<sub>2</sub>/11-KT ratio becomes more fine-regulated in females than in males, and in fish fed D1 diet rather than in fish fed D2 diet. Moreover, it is noteworthy that the highest plasma E<sub>2</sub>/11-KT

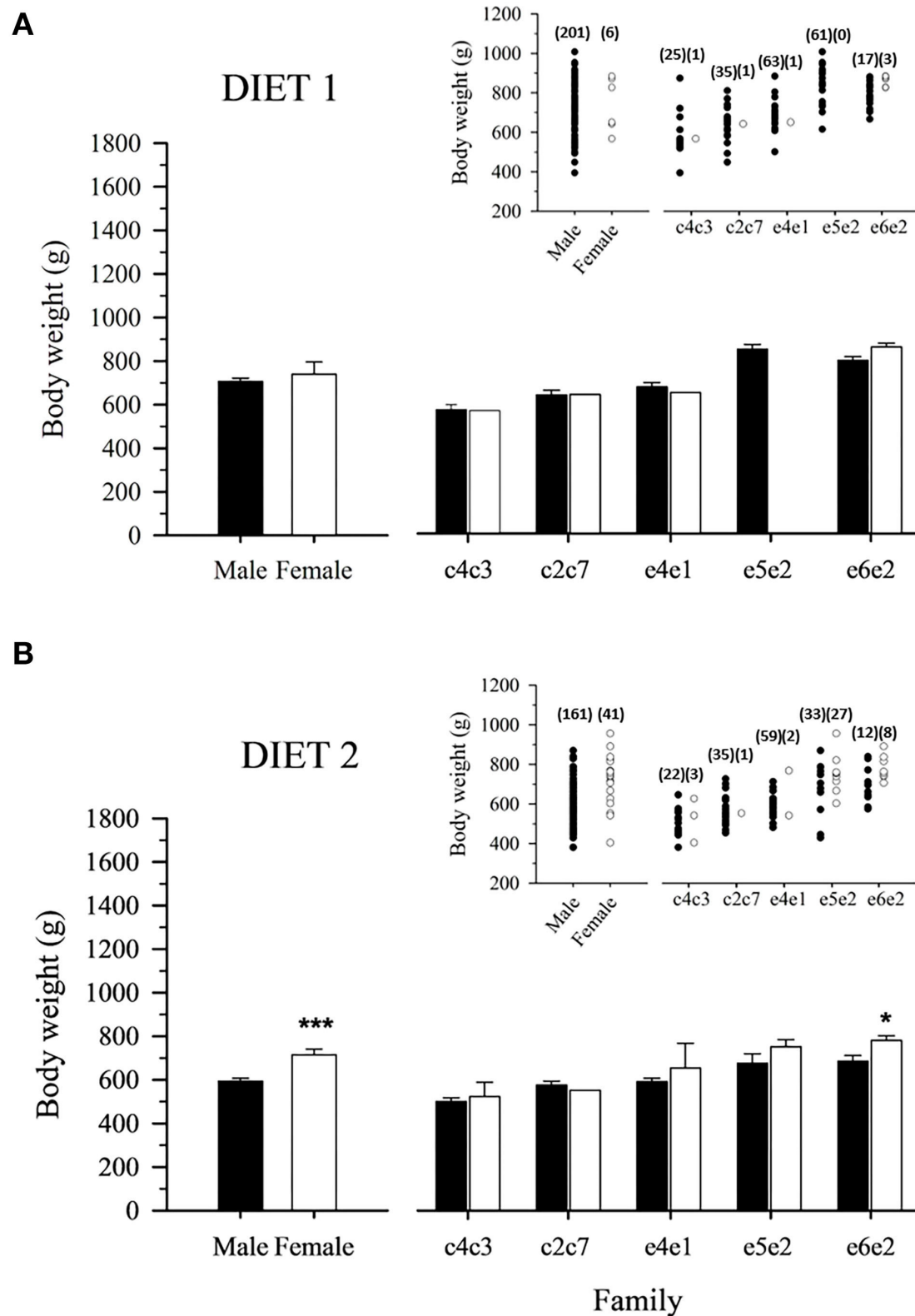


FIGURE 2

Mean male and female body weight of 2-year-old gilthead sea bream-fed control (A) or experimental diet (B) as a whole population and separated by family. Values are the mean  $\pm$  SEM of three tanks per diet ( $n = 207$  fish for diet 1, 202 for diet 2). Asterisks indicate significant differences (Student's  $t$ -test,  $P < 0.05$ ) between males and females body weight. Inserts indicate individual body weight in each population. The number of males and females for the populations and families is indicated in parenthesis.

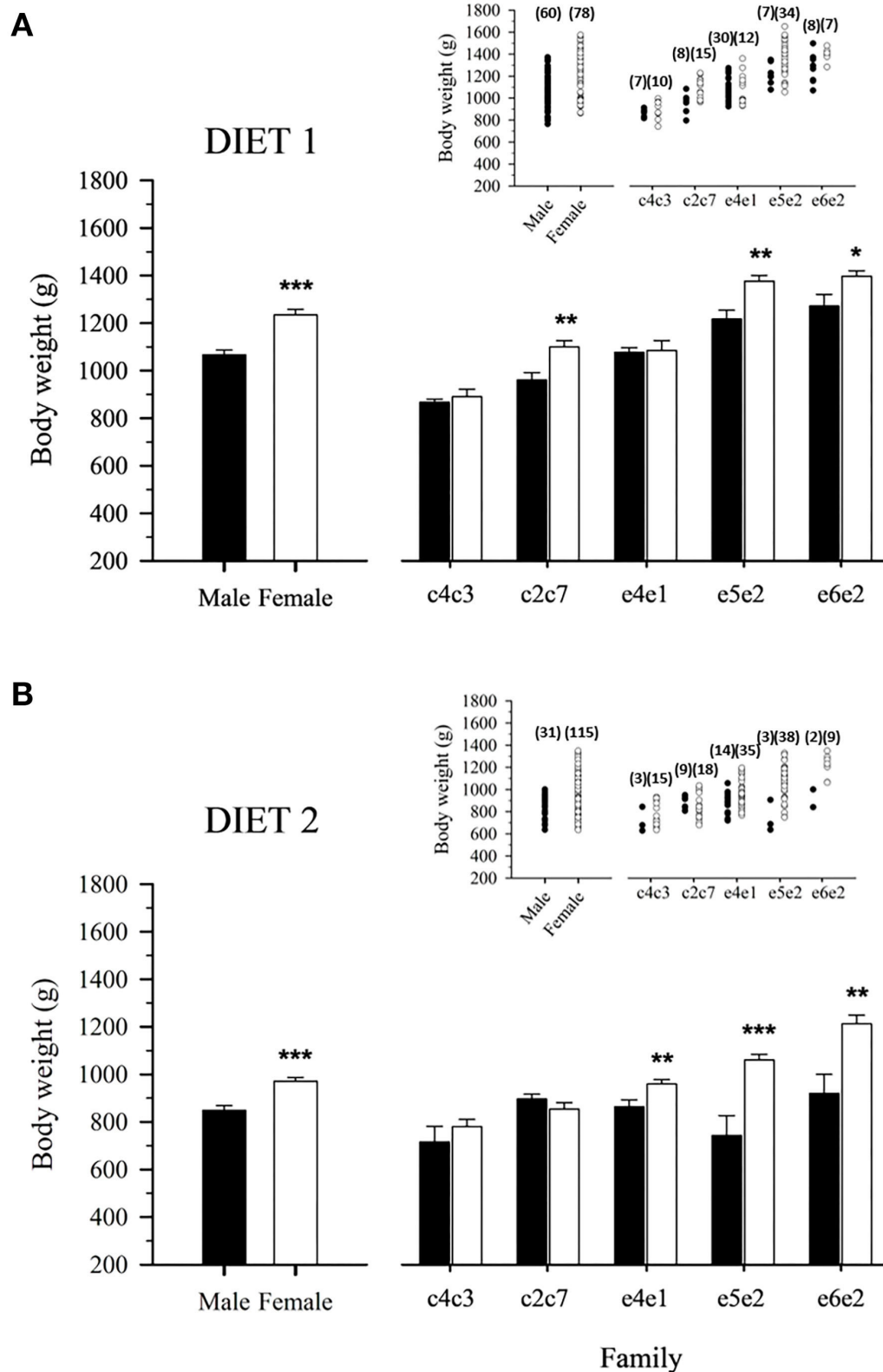
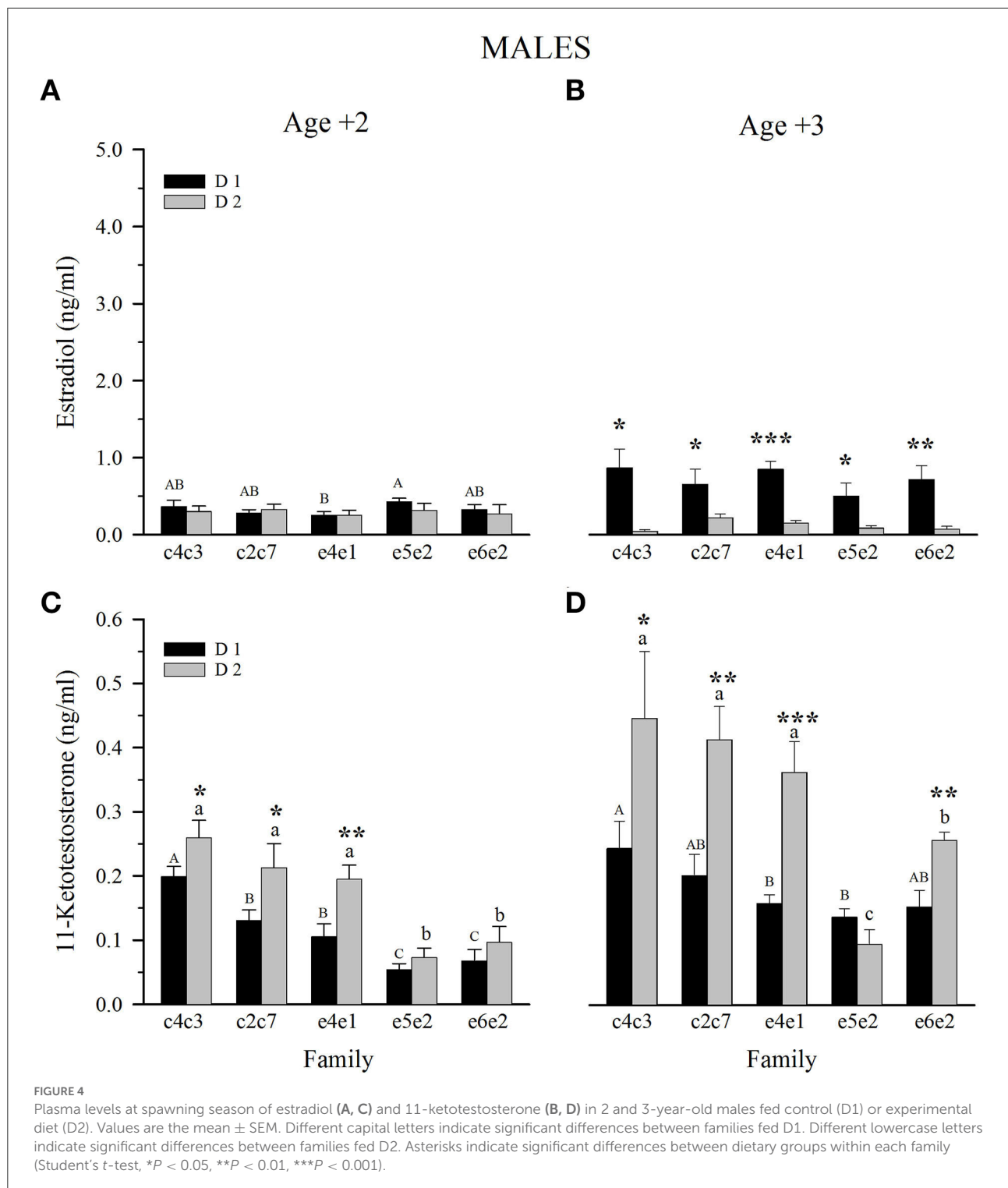


FIGURE 3

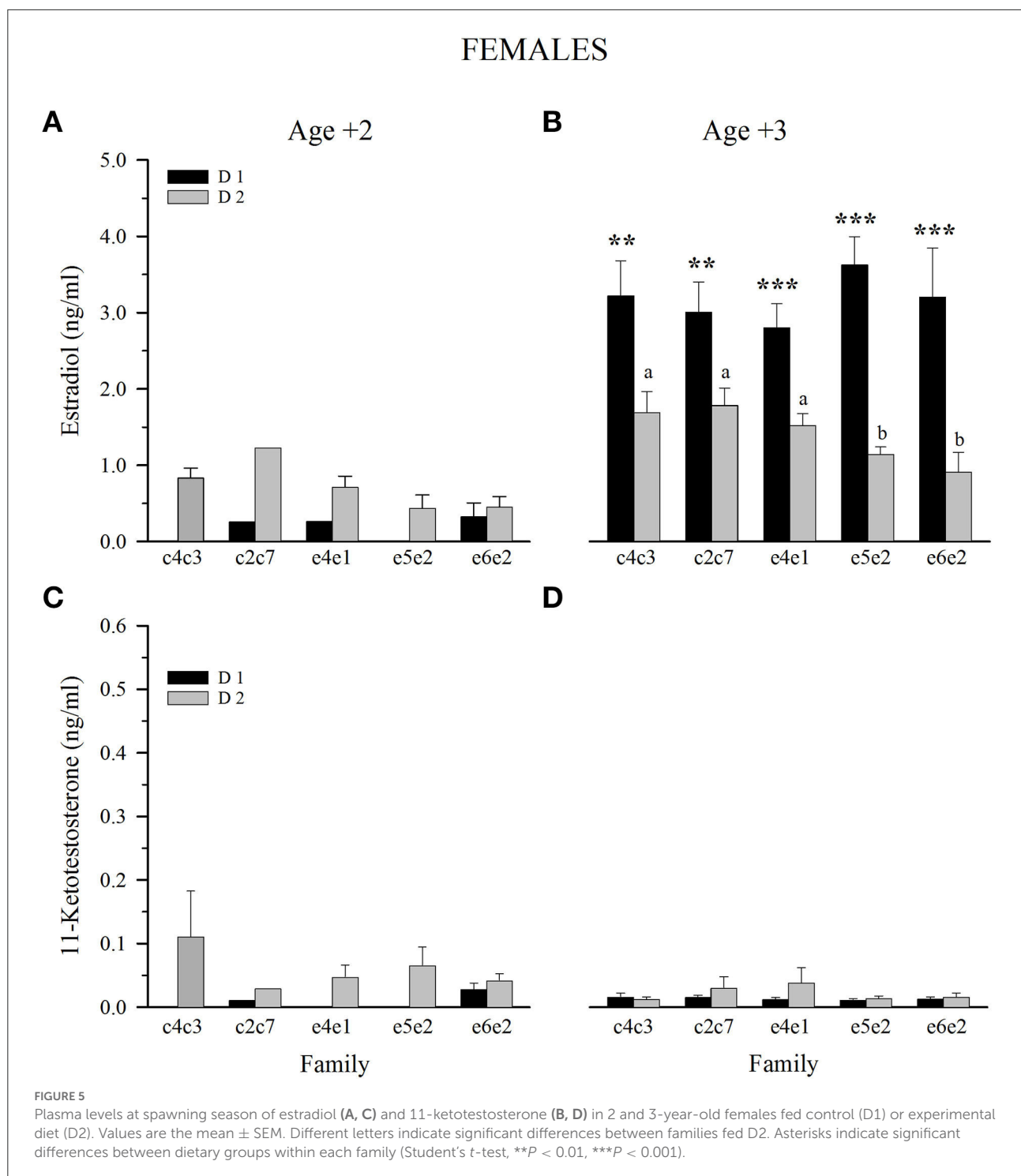
Mean male and female body weight of 3-year-old gilthead sea bream-fed control (A) or experimental diet (B) as a whole population and separated by family. Values are the mean  $\pm$  SEM of three tanks per diet ( $n = 138$  fish for diet 1, 146 for diet 2). Asterisks indicate significant differences (Student's  $t$ -test,  $*P < 0.05$ ,  $**P < 0.01$ ,  $***P < 0.001$ ) between males and females body weight. Inserts indicate individual body weight in each population. The number of males and females for the populations and families is indicated in parenthesis.



ratio (powerful sex female steroid profile) was concurrent with a lower abundance of functional/powerful females when fish from fast- ( $e_x e_y$ ) or slow-growth ( $c_x c_y$ ) fish families were grouped and analyzed together as two different experimental groups in our common garden rearing system.

## Discussion

The present study underlines the effect of different nutrition and genetic backgrounds in the plasma sex steroids profile and male-to-female sex reversal in gilthead sea bream with



differences in heritable growth and a sexual growth dimorphism, which was exacerbated in fast-growing fish by feeding a plant-based diet. Indeed, when fish attained 2 years of age, a sexual growth dimorphism was only observed in fish fed D2. However, body weight differences intensified as fish grew, and were equally significant and visible with both diets,

especially in fast-growing families. Besides, previous studies have highlighted that families with a fast growth phenotype within the PROGENSA<sup>®</sup> selection program displayed a plastic gut microbiota to cope better with changes in diet composition, also contributing to a better disease progression of parasitic enteritis in fish challenged with the myxozoa *Enteromyxum*

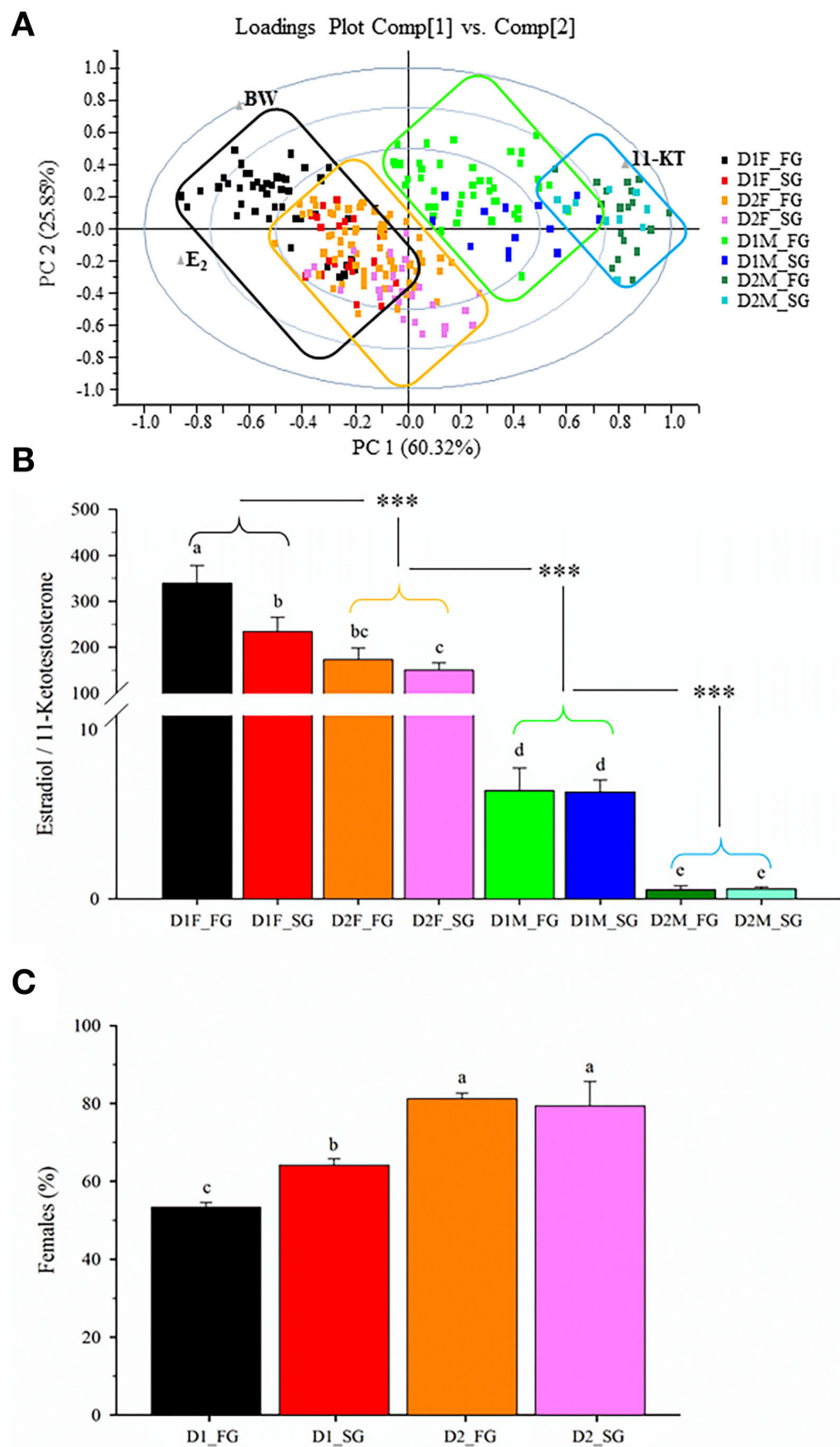


FIGURE 6

(A) Principal component analysis of body weight and plasma sex steroids levels during the second spawning period (3-year-old fish). (B) Sex steroids quotient of females (F) and males (M) of each dietary group (D1–D2), and fast (FG) or slow (SG) growth family groups. Values are the mean  $\pm$  SEM of the  $E_2/11\text{-KT}$  quotient. Different letters indicate significant differences between each group. Asterisks indicate significant differences between males and female groups with different dietary regimes (Holm-Sidak *post-hoc* test,  $***P < 0.001$ ). (C) Female percentage of fast and slow growth family groups.



*lei* (5, 43). Available studies also evidenced that such selective breeding for growth and skeletal deformities has an impact on humoral immune markers (48), carcass and morphometric traits (49, 50), energy partitioning between growth and swimming activity (44), and even more in reproductive success as recently evidenced in fish with a normal phenotype but with a genetic background of skeletal deformities (51). It appears, thereby, that selection for growth and deformity traits co-selects in the PROGENSA<sup>®</sup> selection program for a number of relevant traits, including the sex ratio, which is becoming nutritionally and genetically regulated through the life cycle. The ultimate physiological mechanism remains elusive, but this study aimed to provide new insights into the sex ratio in a protandrous fish for its use as an operational welfare scoring system in a challenging environment.

Sex reversal in fish is defined as a mismatch between the phenotypic and the genetic sex (52). Thus, in gilthead sea bream, in particular, the general thinking is that individuals of this fish species act as functional males by the end of the first-second year of life, and sex reversal generally takes place 1 year after the first male sexual maturation (53). The percentage of male-to-female sex reversal can vary from 15 to 80% during the second year of life (53), but the possibility of a later sex reversal was reported by Brusléa-Sicard and Fourcalt (54) and further corroborated by Chaoui et al. (55). Herein, the male-to-female sex reversal was accomplished by a relatively high percentage of individuals (50–85%) regardless of diet, which confirms the notion that a sex change in gilthead sea bream is cued by social and environmental factors when a critical age or size is attained (30, 56, 57). Indeed, the removal of functional females from the population drives the feminization of the remaining males (53, 58), probably *via* the production and release of specific pheromones that can activate or block some sex-specific networks. The aquatic environment may also contain a wide range of endocrine-disrupting chemicals that reduce reproductive performance and can even inhibit the sex reversal in gilthead sea bream (31, 59, 60). However, xenobiotic-induced sex reversal did not appear to be our case, because fish grew from early life stages in an eco-friendly environment where a wide-screening of undesirable compounds in fish edible matter revealed bio-contaminant loads to be much lower than the maximum established residue level (61).

The net balance between gonadal estrogen and androgen production directs sexual differentiation and gonadal development in fish (62). Indeed, E<sub>2</sub> and 11-KT are typically considered the predominant steroids in the regulation of sex change in most fish species (63). Thus, we found herein that both E<sub>2</sub> and 11-KT increased over time, with males and females displaying the highest plasma levels of 11-KT or E<sub>2</sub> at Age +3, respectively. A nutritionally mediated effect was also reported, although it is difficult to deconvolute the extent to which this observation is due to a specific nutrient or to a different loading of plant phytoestrogens with both estrogenic and antiestrogenic

effects on vertebrates (64). In tilapia farming, in particular, herbal extracts could be used as safe alternative agents to control precocious tilapia maturity and prolific breeding in production (65). In the present study, the inclusion level of soy protein concentrate, a rich source of phytoestrogens, varied between 16% in D1 and 25% in D2, although it is within the tolerance range for gilthead sea bream (65). In any case, plasma levels of fish fed D2 were lower for E<sub>2</sub> and higher for 11-KT, displaying these fish a masculinized sex steroid profile that would promote the male-to-female sex reversal in the absence of high powerful functional females that ensure reproduction success, as stated before by Simó-Mirabet et al. (41). Masculinization of gonochoristic fish populations also occurs as a result of elevated temperatures and other environmental stressors. This would be mediated, at least in part, by the increase of circulating cortisol, which is now recognized as a universal mediator of sex reversal in fish due to its implication in delaying ovarian meiosis and increasing 11-KT (52). For instance, in the protogynous three-spot wrasse, cortisol treatment had a masculinizing effect (66). This feature was also reported by us in the protandrous gilthead sea bream-fed plant-based diets, regardless of the well-known hypocholesterolemic effect of plant ingredients in most farmed fish (67). In fact, since cholesterol is the precursor of cortisol, its reduced dietary supply or intestinal absorption could initially lead to a female-biased sex ratio. Nonetheless, there are more factors at play in this process, and Nile tilapia fry fed with saponin-supplemented diets (hypocholesterolemic diets) displayed a significant male-biased population (68). In other words, the sex ratio can be influenced by a number of nutritional factors, including changes in the dietary fatty acid composition as a result of a high replacement of marine feedstuffs by vegetable oils (69). However, all this is the result of a complex trade-off, which is also indicative of the amazing diversity and evolution of sex determination in vertebrates, and fish in particular (70). Similarly, global warming due to climate change would affect the offspring quality of a wide range of animals (71–73). This is especially important for aquatic ectotherms, where temperature values above the optimal for each species and fish strain can shift the sex ratio toward either the male or female phenotype in gonochoristic fish (74–76) and perhaps hermaphroditic fish.

Moreover, in the present study, discriminant analysis of sex steroids and body weight displayed a clear separation of phenotypic males and females, with also a differentiation of slow- and fast-growth families. Thus, families of fast heritable growth displayed a more mature/advanced sex steroid phenotype, resulting in a higher E<sub>2</sub>/11-KT ratio (feminization phenotype) that would trigger the inhibition of male-to-female sex reversal of the remaining males. This hormonal quotient, rather than the circulating amount of a given sex steroid, would determine what sex-specific network is activated or suppressed, leading or not to the sex reversal in the individual. Indeed, the rise of the E<sub>2</sub>/11-KT ratio was associated with

a female bias in painted turtles (77), and we found herein that this plasma sex steroid ratio was progressively increased from males to females with both the nutritional and genetic improvement. However, the increased plasma E<sub>2</sub>/11-KT ratio in female fish fed D1 was negatively associated with the rate of male-to-female sex reversal in genetically improved fish, whereas this genetically mediated response was mostly masked in fish fed D2. This finding highlighted a nutrition and genetic interaction on the progression of sexual maturation, as reported for gut microbiota composition and function. Indeed, fast-growing fish families in the PROGENSA<sup>®</sup> program are more resilient to changes in gut microbiota composition, but at the same time, metatranscriptomic analyses confirm and extend the notion that the core microbiota of genetically improved fish is able to modulate their metabolic activity to cope better with changes in diet composition (5, 43). Otherwise, there is no evidence in gilthead sea bream that the modulating effects of gut microbiota by feed additives are a specific feature of each additive and genetic background (unpublished results).

As indicated earlier, reproductive physiology and spawning are sensitive processes to changes in environmental conditions and physiological stress, and how and to what extent external and internal factors have an impact on broodstock welfare is of relevance to assure reproduction success, but also the offspring plasticity and quality (78, 79). Concretely, in gilthead sea bream, several attempts at nutritional programming are made through changes in the parental nutrition (80–83). Such an approach is not unique to fish, as parental nutrition and hormonal status of humans and terrestrial animals directly impact all stages of gamete maturation, fetal development, and long-term offspring health (84, 85). Therefore, one of the challenges of modern aquaculture is to assure the welfare of breeders, which could also inform the welfare condition from a retrospective point of view. Certainly, we can conclude that fast-growing fish families fed a control diet became powerful females, whilst fish of slow-growth families and/or fish fed plant-based diets experienced a pseudo-feminization effect (i.e., fish with a weaker female signal, which enhances the ratio of sex reversal). Therefore, it appears that the progression of sex reversal is directly regulated by both nutritional and genetic background among many other environmental and social factors. Thus, the study of sex reversal as a biological endpoint is becoming a reliable tool of relevance for the animal welfare certification of a highly cultured protandrous fish such as gilthead sea bream. Several items support this assumption. First, fish growing with plant-based diets from early life stages shared an enhanced onset of puberty and sex reversal in concurrence with some growth impairment of sexually mature fish. Second, genetically improved fish for growth are more resilient to the progression of male-to-female sex reversal with the use of alternative fish feed formulations, but further research should be directed toward the effect of specific nutrients on reproductive performance and

maturation as a means to enhance the offspring quality. Finally, a decreased plasma E<sub>2</sub>/11-KT ratio is becoming indicative of a negative welfare status in the long term, supporting this finding the use of such reproductive tract scoring systems for leading a protandrous farmed fish toward their optimum welfare condition.

## Conclusion

Results of this long-term dietary and genetics trial disclose that sex steroids profile and male-to-female sex reversal are nutritionally and genetically regulated in the protandrous gilthead sea bream. Moreover, the sex ratio is proposed as a reliable welfare indicator alerting of disturbances in reproductive performance, and perhaps overall growth and offspring quality. Such a scoring system is becoming, thereby, an exploitable finding for the certification of animal welfare in a given gilthead sea bream production system.

## Data availability statement

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

## Ethics statement

The animal study was reviewed and approved by Ethics and Animal Welfare Committee of the Institute of Aquaculture Torre de la Sal and CSIC Ethics Committee.

## Author contributions

JP-S: conceptualization. PH, AF, JC-G, and JP-S: formal analysis. AF, JC-G, and JP-S: funding acquisition. PH, AF, and JP-S: methodology. AF, JA, and JP-S: resources. PH, JC-G, and JP-S: draft writing. PH, AF, JC-G, JA, and JP-S: draft review and edition. All authors contributed to the article and approved the submitted version.

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## Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

## References

- Béné C, Barange M, Subasinghe R, Pinstrip-Andersen P, Merino G, Hemre GI, et al. Feeding 9 billion by 2050 – Putting fish back on the menu. *Food Secur.* (2015) 7:261–74. doi: 10.1007/s12571-015-0427-z
- Ahmad A, Abdullah SR, Hasan HA, Othman AR, Ismail NI. Aquaculture industry: Supply and demand, best practices, effluent and its current issues and treatment technology. *J Environ Manage.* (2021) 287:112271. doi: 10.1016/j.jenvman.2021.112271
- Jennings S, Stentiford GD, Leocadio AM, Jeffery KR, Metcalfe JD, Katsiadaki I, et al. Aquatic food security: insights into challenges and solutions from an analysis of interactions between fisheries, aquaculture, food safety, human health, fish and human welfare, economy and environment. *Fish Fish.* (2016) 17:893–938. doi: 10.1111/faf.12152
- Ahmed N, Thompson S, Glaser M. Global aquaculture productivity, environmental sustainability, and climate change adaptability. *Environ Manage.* (2019) 63:159–72. doi: 10.1007/s00267-018-1117-3
- Piazzon MC, Naya-Català F, Perera E, Palenzuela O, Sitjà-Bobadilla A, Pérez-Sánchez J. Genetic selection for growth drives differences in intestinal microbiota composition and parasite disease resistance in gilthead sea bream. *Microbiome.* (2020) 8:1–17. doi: 10.1186/s40168-020-00922-w
- Terova G, Naya-Català F, Rimoldi S, Piazzon MC, Torrecillas S, Toxqui MS, et al. Highlights from gut microbiota survey in farmed fish - European sea bass and gilthead sea bream case studies. *Aquaculture Eu.* (2022) 47:1.
- Naya-Català F, do Vale Pereira G, Piazzon MC, Fernandes AM, Caldach-Giner JA, Sitjà-Bobadilla A, et al. Cross-talk between intestinal microbiota and host gene expression in gilthead sea bream (*Sparus aurata*) juveniles: Insights in fish feeds for increased circularity and resource utilization. *Front Physiol.* (2021) 12:8265. doi: 10.3389/fphys.2021.748265
- Naya-Català F, Simó-Mirabet P, Caldach-Giner J, Pérez-Sánchez J. Transcriptomic profiling of Gh/Igf system reveals a prompted tissue-specific differentiation and novel hypoxia responsive genes in gilthead sea bream. *Sci Rep.* (2021) 11:1–13. doi: 10.1038/s41598-021-95408-6
- Naya-Català F, Martos-Sitcha JA, de Las Heras V, Simó-Mirabet P, Caldach-Giner JA, Pérez-Sánchez J. Targeting the mild-hypoxia driving force for metabolic and muscle transcriptional reprogramming of gilthead sea bream (*Sparus aurata*) juveniles. *Biology.* (2021) 10:416. doi: 10.3390/biology10050416
- Sánchez-Suárez W, Franks B, Torgerson-White L. From land to water: taking fish welfare seriously. *Animals.* (2020) 10:1585. doi: 10.3390/ani10091585
- Noble C, Gismervik K, Iversen MH, Kolarevic J, Nilsson J, Stien LH, et al. *Welfare Indicators for Farmed Rainbow Trout: Tools for Assessing Fish Welfare.* Tromsø: Nofima (2020). Available online at: <https://nofima.com/results/new-handbook-on-welfare-indicators-for-farmed-rainbow-trout/>
- Saraiva JL, Arechavala-Lopez P, Castanheira MF, Volstorff J, Studer BH, A. global assessment of welfare in farmed fishes: the FishEthoBase. *Fishes.* (2019) 4:30. doi: 10.3390/fishes4020030
- Arechavala-Lopez P, Diaz-Gil C, Saraiva JL, Moranta D, Castanheira MF, Nuñez-Velázquez S, et al. Effects of structural environmental enrichment on welfare of juvenile seabream (*Sparus aurata*). *Aquac Reports.* (2019) 15:100224. doi: 10.1016/j.aqrep.2019.100224
- Yavuzcan Yildiz H, Chatzifotis S, Anastasiadis P, Parisi G, Papandroulakis N. Testing of the Salmon Welfare Index Model (SWIM 10) as a computational welfare assessment for sea-caged European sea bass. *Ital J Anim Sci.* (2021) 20:1423–30. doi: 10.1080/1828051X.2021.1961106
- Martins CI, Galhardo L, Noble C, Damsgård B, Spedicato MT, Zupa W, et al. Behavioral indicators of welfare in farmed fish. *Fish Physiol Biochem.* (2012) 38:17–41. doi: 10.1007/s10695-011-9518-8
- Sharma M, Thakur J, Verma S, Sharma P. Behavioural responses in effect to chemical stress in fish: a review. *Int J Fish Aquat Stud.* (2019) 7:1–5.
- Caldach-Giner J, Holhorea PG, Ferrer Ballester MÁ, Naya-Català F, Rosell-Moll E, Vega García C, et al. Revising the impact and prospects of activity and ventilation rate bio-loggers for tracking welfare and fish-environment interactions in salmonids and Mediterranean farmed fish. *Front Mar Sci.* (2022) 9:1–18. doi: 10.3389/fmars.2022.854888
- Okuzawa K. Puberty in teleosts. *Fish Physiol Biochem.* (2002) 26:31–41. doi: 10.1023/A:1023395025374
- Harries JE, Runnalls T, Hill E, Harris CA, Maddix S, Sumpter JP, et al. Development of a reproductive performance test for endocrine disrupting chemicals using pair-breeding fathead minnows (*Pimephales promelas*). *Environ Sci Technol.* (2000) 34:3003–11. doi: 10.1021/es991292a
- Martínez P, Viñas AM, Sánchez L, Díaz N, Ribas L, Piferrer F. Genetic architecture of sex determination in fish: applications to sex ratio control in aquaculture. *Front Genet.* (2014) 5:340. doi: 10.3389/fgene.2014.00340
- Ohga H, Adachi H, Matsumori K, Kodama R, Nyuji M, Selvaraj S, et al. mRNA levels of kisspeptins, kisspeptin receptors, and GnRH1 in the brain of chub mackerel during puberty. *Comp Biochem Physiol Part A Mol Integr Physiol.* (2015) 179:104–12. doi: 10.1016/j.cbpa.2014.09.012
- Alonso-Fernández A, Otero J, Bañón R, Campelos JM, Santos J, Mucientes G. Sex ratio variation in an exploited population of common octopus: ontogenic shifts and spatio-temporal dynamics. *Hydrobiologia.* (2017) 794:1–16. doi: 10.1007/s10750-016-3065-3
- Penman DJ, Piferrer F. Fish gonadogenesis. Part I: genetic and environmental mechanisms of sex determination. *Rev Fish Sci.* (2008) 16:16–34. doi: 10.1080/10641260802324610
- Dang Z, Kienzler A. Changes in fish sex ratio as a basis for regulating endocrine disruptors. *Environ Int.* (2019) 130:104928. doi: 10.1016/j.envint.2019.104928
- Baroiller JF, D'Cotta H. Environment and sex determination in farmed fish. *Comp Biochem Physiol Part - C Toxicol Pharmacol.* (2001) 130:399–409. doi: 10.1016/S1532-0456(01)00267-8

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## Supplementary material

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fvets.2022.1083255/full#supplementary-material>

26. Geffroy B, Bardonnet A. Sex differentiation and sex determination in eels: consequences for management. *Fish Fish.* (2016) 17:375–98. doi: 10.1111/faf.12113
27. Goikoetxea A, Todd EV, Gemmell NJ. Stress and sex: does cortisol mediate sex change in fish? *Reproduction.* (2017) 154:R149–60. doi: 10.1530/REP-17-0408
28. Gholampour TE, Raieni RF, Pouladi M, Larijani M, Pagano M, Faggio C. The dietary effect of Vitex agnus-castus hydroalcoholic extract on growth performance, blood biochemical parameters, carcass quality, sex ratio and gonad histology in Zebrafish (*Danio rerio*). *Appl Sci.* (2020) 10:1402. doi: 10.3390/app10041402
29. Renn SC, Hurd PL. Epigenetic regulation and environmental sex determination in cichlid fishes. *Sex Dev.* (2021) 15:93–107. doi: 10.1159/000517197
30. Todd EV, Liu H, Muncaster S, Gemmell NJ. Bending genders: the biology of natural sex change in fish. *Sex Dev.* (2016) 10:223–41. doi: 10.1159/000449297
31. García Hernández MP, Cabas I, Rodenas MC, Arizcun M, Chaves-Pozo E, Power DM, et al. 17 $\alpha$ -ethynylestradiol prevents the natural male-to-female sex change in gilthead seabream (*Sparus aurata* L.). *Sci Rep.* (2020) 10:1–15. doi: 10.1038/s41598-020-76902-9
32. Gemmell NJ, Todd EV, Goikoetxea A, Ortega-Recalde O, Hore TA. Natural sex change in fish. *Curr Top Dev Biol.* (2019) 134:71–117. doi: 10.1016/bs.ctdb.2018.12.014
33. Weltzien FA, Andersson E, Andersen Ø, Shalchian-Tabrizi K, Norberg B. The brain – pituitary – gonad axis in male teleosts, with special emphasis on flatfish (Pleuronectiformes). *Comp Biochem Physiol A.* (2004) 137:447–77. doi: 10.1016/j.cbpa.2003.11.007
34. Gluckman PD, Hanson MA. Evolution, development and timing of puberty. *Trends in Endocrinology & Metabolism* (2006) 17: 7–12. doi: 10.1016/j.tem.2005.11.006
35. Espigares F, Rocha A, Molés G, Gómez A, Carrillo M, Zanuy S. New insights into the factors mediating the onset of puberty in sea bass. *Gen Comp Endocrinol.* (2015) 224:176–85. doi: 10.1016/j.ygcen.2015.08.013
36. Felip A, Zanuy S, Carrillo M. Comparative analysis of growth performance and sperm motility between precocious and non-precocious males in the European sea (*Dicentrarchus labrax*, L.). *Aquaculture.* (2006) 256:570–8. doi: 10.1016/j.aquaculture.2006.02.014
37. Taranger GL, Carrillo M, Schulz RW, Fontaine P, Zanuy S, Felip A, et al. Control of puberty in farmed fish. *Gen Comp Endocrinol.* (2010) 165:483–515. doi: 10.1016/j.ygcen.2009.05.004
38. Kause A, Ritola O, Paananen T, Mäntysaari E, Eskelinen U. Selection against early maturity in large rainbow trout *Oncorhynchus mykiss*: the quantitative genetics of sexual dimorphism and genotype-by- environment interactions. *Aquaculture.* (2003) 228:53–68. doi: 10.1016/S0044-8486(03)00244-8
39. Barson NJ, Aykanat T, Hindar K, Baranski M, Bolstad GH, Fiske P, et al. Sex-dependent dominance at a single locus maintains variation in age at maturity in salmon. *Nature.* (2015) 528:405–8. doi: 10.1038/nature16062
40. Escobar-Aguirre S, Felip A, Mazon MJ, Ballester-Lozano G, Pérez-Sánchez J, Björsson BT, et al. Long-term feeding of a maintenance ration affects the release of Igf-1 and leptin, and delays maturation in a male teleost fish, *Dicentrarchus labrax* L. *Aquaculture.* (2020) 527:735467. doi: 10.1016/j.aquaculture.2020.735467
41. Simó-Mirabet P, Felip A, Estensoro I, Martos-Sitcha JA, de las Heras V, Caldich-Giner JÀ, et al. Impact of low fish meal and fish oil diets on the performance, sex steroid profile and male-female sex reversal of gilthead sea bream (*Sparus aurata*) over a 3-year production cycle. *Aquaculture.* (2018) 490:64–74. doi: 10.1016/j.aquaculture.2018.02.025
42. Perera E, Simó-Mirabet P, Shin HS, Rosell-Moll E, Naya-Catalá F, de las Heras V, et al. Selection for growth is associated in gilthead sea bream (*Sparus aurata*) with diet flexibility, changes in growth patterns and higher intestine plasticity. *Aquaculture.* (2019) 507:349–60. doi: 10.1016/j.aquaculture.2019.04.052
43. Naya-Catalá F, Piazzon MC, Caldich-Giner JÀ, Sitjà-Bobadilla A, Pérez-Sánchez J. Diet and host genetics drive the bacterial and fungal intestinal metatranscriptome of gilthead sea bream. *Front Microbiol.* (2022) 13:883738. doi: 10.3389/fmicb.2022.883738
44. Perera E, Rosell-Moll E, Martos-Sitcha JA, Naya-Catalá F, Simó-Mirabet P, Caldich-Giner JÀ, et al. Physiological trade-offs associated with fasting weight loss, resistance to exercise and behavioral traits in farmed gilthead sea bream (*Sparus aurata*) selected by growth. *Aquac Reports.* (2021) 20:100645. doi: 10.1016/j.aqrep.2021.100645
45. Perera E, Rosell-Moll E, Naya-Catalá F, Simó-Mirabet P, Caldich-Giner JÀ, Pérez-Sánchez J. Effects of genetics and early-life mild hypoxia on size variation in farmed gilthead sea bream (*Sparus aurata*). *Fish Physiol Biochem.* (2021) 47:121–33. doi: 10.1007/s10695-020-00899-1
46. Rodríguez L, Begtashi I, Zanuy S, Shaw M, Carrillo M. Changes in plasma levels of reproductive hormones during first sexual maturation in European male sea bass (*Dicentrarchus labrax* L.) under artificial day lengths. *Aquaculture.* (2001) 202:235–48. doi: 10.1016/S0044-8486(01)00774-8
47. Molés G, Gómez A, Rocha A, Carrillo M, Zanuy S. Purification and characterization of follicle-stimulating hormone from pituitary glands of sea bass (*Dicentrarchus labrax*). *General Comp Endocrinol.* (2008) 158:68–76. doi: 10.1016/j.ygcen.2008.05.005
48. Vallecillos A, Chaves-Pozo E, Arizcun M, Pérez R, Afonso JM, Berbel C, et al. Genetic parameters for *Photobacterium damsela* subsp. piscicida resistance, immunological markers and body weight in gilthead seabream (*Sparus aurata*). *Aquaculture.* (2021) 543:736892. doi: 10.1016/j.aquaculture.2021.736892
49. Lee-Montero I, Navarro A, Negrín-Báez D, Zamorano MJ, Berbel C, Sánchez JA, et al. Genetic parameters and genotype-environment interactions for skeleton deformities and growth traits at different ages on gilthead seabream (*Sparus aurata* L.) in four Spanish regions. *Anim Genet.* (2015) 46:164–74. doi: 10.1111/age.12258
50. León-Bernabeu S, Shin HS, Lorenzo-Felipe Á, García-Pérez C, Berbel C, Elalfy IS, et al. Genetic parameter estimations of new traits of morphological quality on gilthead seabream (*Sparus aurata*) by using IMAFISH\_ML software. *Aquacult Reports.* (2021) 21:100883. doi: 10.1016/j.aqrep.2021.100883
51. Lorenzo-Felipe A, Shin HS, León-Bernabeu S, Pérez-García C, Zamorano MJ, Pérez-Sánchez J, et al. The effect of the deformity genetic background of the breeders on the spawning quality of gilthead seabream (*Sparus aurata* L.). *Front Mar Sci.* (2021) 8:656901. doi: 10.3389/fmars.2021.656901
52. Baroiller JF, d'Cotta H. The reversible sex of gonochoristic fish: insights and consequences. *Sex Dev.* (2016) 10:242–66. doi: 10.1159/000452362
53. Zohar Y, Abraham M, Gordin H. The gonadal cycle of the captivity-reared hermaphroditic teleost *Sparus aurata* (L.) during the first two years of life. *Ann Biol Anim Biochim Biophys.* (1978) 18:877–82. doi: 10.1051/rnd:19780519
54. Brusléa-Sicard S, Fourcalt B. Recognition of sex-inverting protandric *Sparus aurata*: ultrastructural aspects. *J Fish Biol.* (1997) 50: 1094–1103.
55. Chaoui L, Kara MH, Faure E, Quignard JP. Growth and reproduction of the gilthead seabream *Sparus aurata* in Mellah lagoon (north-eastern Algeria). *Sci Mar.* (2006) 70:545–52. doi: 10.3989/scimar.2006.70n3545
56. Kobayashi Y, Nagahama Y, Nakamura M. Diversity and plasticity of sex determination and differentiation in fishes. *Sex Dec.* (2013) 7:115–25. doi: 10.1159/000342009
57. Liu H, Todd EV, Lokman PM, Lamm MS, Godwin JR, Gemmell NJ. Sexual plasticity: a fishy tale. *Mol Reprod Develop.* (2017) 84:171–94. doi: 10.1002/mrd.22691
58. Happe A, Zohar Y. Self-fertilization in the protandrous hermaphrodite *Sparus aurata*: development of the technology. *Colloq l'INRA.* (1988) 44:177–80.
59. Forner-Piquer I, Fakriadis I, Mylonas CC, Piscitelli F, Di Marzo V, Maradonna F, et al. Effects of dietary bisphenol A on the reproductive function of gilthead sea bream (*Sparus aurata*) testes. *Int J Mol Sci.* (2019) 20:5003. doi: 10.3390/ijms20205003
60. Forner I, Constantinos P, Ioannis CM, Maria F, Fabiana P. Effects of diisononyl phthalate (DiNP) on the endocannabinoid and reproductive systems of male gilthead sea bream (*Sparus aurata*) during the spawning season. *Arch Toxicol.* (2019) 93:727–41. doi: 10.1007/s00204-018-2378-6
61. Nàcher-Mestre J, Ballester-Lozano GF, Garlito B, Portolés T, Caldich-Giner JÀ, Serrano R, et al. Comprehensive overview of feed-to-fillet transfer of new and traditional contaminants in Atlantic salmon and gilthead sea bream fed plant-based diets. *Aquacult Nutri.* (2018) 24:1782–95. doi: 10.1111/anu.12817
62. Piferrer F, Guiguen Y. Fish gonadogenesis. Part II: molecular biology and genomics of sex differentiation. *Rev Fish Sci.* (2008) 16:35–55. doi: 10.1080/10641260802324644
63. Tenugu S, Senthilkumaran B. Sexual plasticity in bony fishes: Analyzing morphological to molecular changes of sex reversal. *Aquacult Fisher.* (2022) 7:525–39. doi: 10.1016/j.aaf.2022.02.007
64. Brzezinski A, Debi A. Phytoestrogens: the “natural” selective estrogen receptor modulators?. *Eu J Obst Gynecol Reprod Biol.* (1999) 85:47–51. doi: 10.1016/S0301-2115(98)00281-4
65. Kokou F, Sarropoulou E, Cotou E, Kentouri M, Alexis M, Rigos G. Effects of graded dietary levels of soy protein concentrate supplemented with methionine and phosphate on the immune and antioxidant responses of gilthead sea bream (*Sparus aurata* L.). *Fish Shellfish Immunol.* (2017) 64:111–21. doi: 10.1016/j.fsi.2017.03.017
66. Nozu R, Nakamura M. Cortisol administration induces sex change from ovary to testis in the protogynous wrasse, *Halichoeres trimaculatus*. *Sex Dev.* (2015) 9:118–24. doi: 10.1159/000373902
67. Gómez-Requeni P, Mingarro M, Caldich-Giner JÀ, Médale F, Martin SA, Houlihan DE, et al. Protein growth performance, amino acid utilisation and



somatotropic axis responsiveness to fish meal replacement by plant protein sources in gilthead sea bream (*Sparus aurata*). *Aquaculture*. (2004) 232:493–510. doi: 10.1016/S0044-8486(03)00532-5

68. Francis G, Levavi-Sivan B, Avitan A, Becker K. Effects of long term feeding of Quillaja saponins on sex ratio, muscle and serum cholesterol and LH levels in Nile tilapia (*Oreochromis niloticus* (L)) *Comp Biochem Physiol Part – C. Toxicol Pharmacol*. (2002) 133:593–603. doi: 10.1016/S1532-0456(02)00167-9

69. Izquierdo MS, Fernández-Palacios H, Tacon AGJ. Effect of broodstock nutrition on reproductive performance of fish. *Aquaculture*. (2001) 197:25–42. doi: 10.1016/S0044-8486(01)00581-6

70. Piferrer F, Ribas L, Díaz N. Genomic approaches to study genetic and environmental influences on fish sex determination and differentiation. *Mar Biotechnol*. (2012) 14:591–604. doi: 10.1007/s10126-012-9445-4

71. Donelson JM, Munday PL, McCormick MI. Climate change may affect fish through an interaction of parental and juvenile environments. *Coral Reefs*. (2012) 31:753–62. doi: 10.1007/s00338-012-0899-7

72. Jonsson B, Jonsson N. Early environment influences later performance in fishes. *J Fish Biol*. (2014) 85:151–88. doi: 10.1111/jfb.12432

73. Schwanz LE. Parental thermal environment alters offspring sex ratio and fitness in an oviparous lizard. *J Exp Biol*. (2016) 219:2349–57. doi: 10.1242/jeb.139972

74. Ospina-Álvarez N, Piferrer F. Temperature-dependent sex determination in fish revisited: prevalence, a single sex ratio response pattern, and possible effects of climate change. *PLoS ONE*. (2008) 3:e2837. doi: 10.1371/journal.pone.0002837

75. Budd AM, Banh QQ, Domingos JA, Jerry DR. Sex control in fish: approaches, challenges and opportunities for aquaculture. *J Mar Sci Eng*. (2015) 3:329–55. doi: 10.3390/jmse3020329

76. Geffroy B, Besson M, Sánchez-Baizán N, Clota F, Goikoetxea A, Sadoul B, et al. Unraveling the genotype by environment interaction in a thermosensitive fish with a polygenic sex determination system. *Proc Natl Acad Sci USA*. (2021) 3:118. doi: 10.1073/pnas.2112660118

77. Bowden RM, Ewert MA, Nelson CE. Environmental sex determination in a reptile varies seasonally and with yolk hormones. *Proc R Soc B Biol Sci*. (2000) 267:1745–9. doi: 10.1098/rspb.2000.1205

78. Duncan NJ, Sonesson AK, Chavanne H. *Principles of Finfish Broodstock Management in Aquaculture: Control of Reproduction and Genetic Improvement*. Advances in Aquaculture Hatchery Technology. Cambridge: Woodhead Publishing Limited (2013). doi: 10.1533/9780857097460.1.23

79. Migaud H, Bell G, Cabrita E, McAndrew B, Davie A, Bobe J, et al. Gamete quality and broodstock management in temperate fish. *Rev Aquac*. (2013) 5:S194–223. doi: 10.1111/raq.12025

80. Izquierdo MS, Turkmen S, Montero D, Zamorano MJ, Afonso JM, Karalazos V, et al. Nutritional programming through broodstock diets to improve utilization of very low fishmeal and fish oil diets in gilthead sea bream. *Aquaculture*. (2015) 449:18–26. doi: 10.1016/j.aquaculture.2015.03.032

81. Turkmen S, Zamorano MJ, Fernández-Palacios H, Hernández-Cruz CM, Montero D, Robaina L, et al. Parental nutritional programming and a reminder during juvenile stage affect growth, lipid metabolism and utilisation in later developmental stages of a marine teleost, the gilthead sea bream (*Sparus aurata*). *Br J Nutr*. (2017) 118:500–12. doi: 10.1017/S0007114517002434

82. Perera E, Turkmen S, Simó-Mirabet P, Zamorano MJ, Xu H, Naya-Català E, et al. Stearoyl-CoA desaturase (scd1a) is epigenetically regulated by broodstock nutrition in gilthead sea bream (*Sparus aurata*). *Epigenetics*. (2020) 15:536–53. doi: 10.1080/15592294.2019.1699982

83. Xu H, Turkmen S, Rimoldi S, Terova G, Zamorano MJ, Afonso JM, et al. Nutritional intervention through dietary vegetable proteins and lipids to gilthead sea bream (*Sparus aurata*) broodstock affects the offspring utilization of a low fishmeal/fish oil diet. *Aquaculture*. (2019) 513:734402. doi: 10.1016/j.aquaculture.2019.734402

84. Sinclair KD, Watkins AJ. Parental diet, pregnancy outcomes and offspring health: metabolic determinants in developing oocytes and embryos. *Reprod Fertil Dev*. (2014) 26:99–114. doi: 10.1071/RD13290

85. Chavatte-Palmer P, Velázquez MA, Jammes H, Duranthon V. Review: epigenetics, developmental programming and nutrition in herbivores. *Animal*. (2018) 12:363–71. doi: 10.1017/S1751731118001337



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# From egg to slaughter: monitoring the welfare of Nile tilapia, *Oreochromis niloticus*, throughout their entire life cycle in aquaculture

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The primary aim of this study was to comprehensively evaluate the welfare of Nile tilapia (*Oreochromis niloticus*) throughout their entire life cycle within aquaculture, spanning from reproduction to slaughter. The methodology was structured to identify welfare indicators closely aligned with the principles of animal freedoms defined by the Farm Animal Council, encompassing environmental, health, nutritional, behavioral, and psychological freedom. Notably, psychological freedom was inherently considered within the behavioral and physical analyses of the animals. To accomplish this, an integrative systematic literature review was conducted to define precise indicators and their corresponding reference values for each stage of tilapia cultivation. These reference values were subsequently categorized using a scoring system that assessed the deviation of each indicator from established ideal (score 1), tolerable (score 2), and critical (score 3) ranges for the welfare of the target species. Subsequently, a laboratory experiment was executed to validate the pre-selected health indicators, specifically tailored for the early life stages of tilapia. This test facilitated an assessment of the applicability of these indicators under operational conditions. Building on the insights gained from this experimentation, partial welfare indices (PWIs) were computed for each assessed freedom, culminating in the derivation of a general welfare index (GWI). Mathematical equations were employed to calculate these indices, offering a quantitative and standardized measure of welfare. This approach equips tilapia farmers and processors with the tools necessary for the continuous monitoring and enhancement of their production systems and stimulate the adoption of more sustainable and ethical practices within the tilapia farming.

## KEYWORDS

grow-out, larviculture, fish welfare, welfare indices, protocols, sustainable aquaculture



# 1. Introduction

The international scientific community's recent recognition of fish as sentient beings (1–4) has encouraged various countries to implement norms and regulations and enact laws to protect these animals when commercially farmed in captivity (5–8). Simultaneously, multiple actors in the food sector—such as importers, retail chains, restaurants, and their respective representative entities—began to request sustainability and animal welfare certificates as a prerequisite for purchasing processed or unprocessed aquaculture products (9, 10). This transformation has occurred in harmony with the aquaculture sector's recognition of the relevance of launching marketing and advertising campaigns focussed on animal welfare and integrating them into the industry's main collective corporate social responsibility commitments (11, 12). As a result, albeit at an early stage, the issue of “fish welfare” is gradually being incorporated into economic actors' social attitudes and practices (13).

However, many challenges must be overcome for welfare to become an inseparable element of farmed fish production. The international recommendations and guidelines currently focus on animal transport and slaughter stages and establish only the minimum animal protection standards (7, 8, 14, 15). In this way, the strict aspects of each species and the critical points in the welfare of these animals end up being neglected. The lack of scientifically based information on tested, standardized, and validated instruments for each farmed fish species and restrictions on their applicability in field situations are often cited as some of the reasons to explain these gaps (13, 16).

Amongst the ~350 species of fish farmed in aquaculture worldwide (17), tilapia, in particular, have shown significant volume growth. According to recent FAO data (18), Nile tilapia, *Oreochromis niloticus*, is currently amongst the three most farmed fish species globally, with China, Indonesia, Egypt, Bangladesh, and Brazil emerging as the largest producers (19). Aquaculture sectors have been under pressure to produce more with fewer resources—increasing production using less feed, water, and space to meet the growing global demand for fish proteins (20). However, this pressure tends to affect the environment and the welfare of farmed fish (21). In this context, the assessment of the health and welfare of tilapia becomes an increasingly relevant challenge in a global scenario. That is why these issues emerge as a central focus in the search for the development of management alternatives aimed at improving the quality of life of the animals and the quality of the final product made available by the industries to their consumers (9, 22).

The first protocol for assessing tilapia welfare, developed by Pedrazzani et al. (23) was limited to the grow-out phase of Nile tilapia. In the present study, our goal is to review the indicators proposed in that protocol, in addition to identifying and validating specific health, environmental, nutritional, and behavioral indicators for all other phases of the development cycle of *O. niloticus* in captivity, including the stages of reproduction, nursery, and transport. This approach will, for the first time, enable a more comprehensive, accurate, and personalized analysis of the evaluation and promotion of tilapia welfare throughout its entire production cycle rather than just the final grow-out

stage. Moreover, in the present study, we also applied a method to establish quantitative and standardized welfare indices for each tilapia cultivation phase.

## 2. Materials and methods

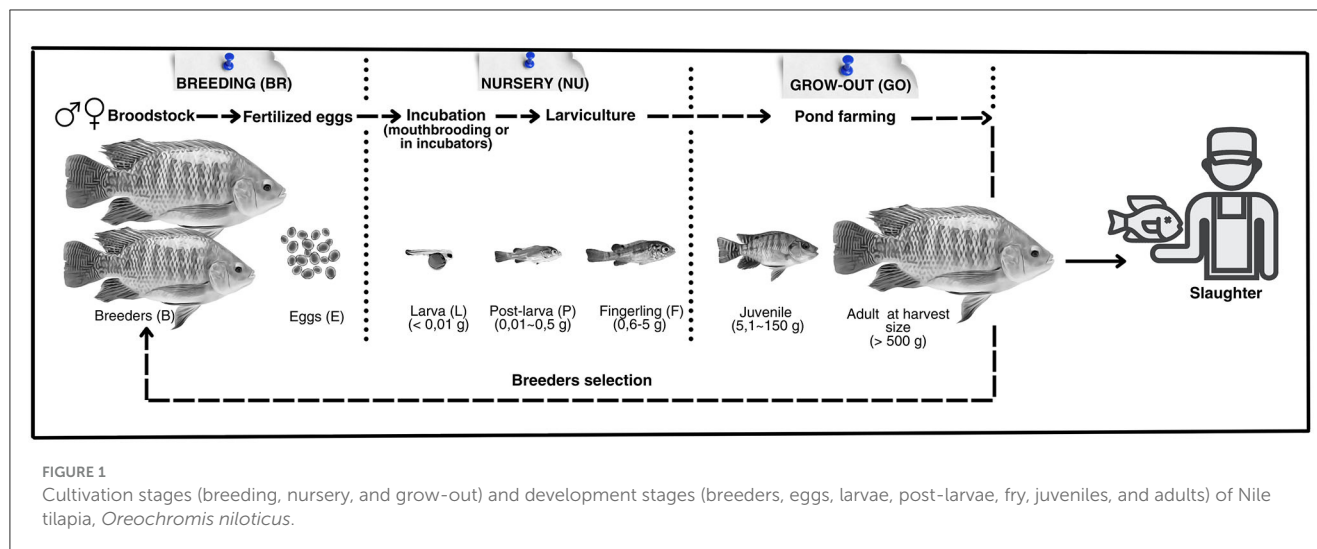
### 2.1. Organization of the welfare protocol for tilapia into categories

The operational welfare indicators for tilapia were organized according to four of the five freedoms for animals established by the Farm Animal Council (24): environmental, health, nutritional, and behavioral. The fifth freedom, psychological, is intrinsically evaluated through the behavioral and physical analysis of the animals. The methodology employed followed the same principles already used in the creation of the protocols previously developed by our group for *O. niloticus* (23) and grass carp, *Ctenopharyngodon idella* (25), during the grow-out phase of both and also for white-leg shrimp, *Penaeus vannamei* throughout its entire production cycle (26).

### 2.2. Systematic review for the definition of indicators and their respective reference values

An integrative systematic review (27) was conducted using the Google Scholar platform as the research base. The aspects related to the environment, health, nutrition, and behavior associated with the species, as well as the specific welfare indicators and their respective reference values for each cultivation stage (Figure 1), were studied and defined. For the grow-out phase of tilapia, emphasis was given to cultivations carried out in earth ponds, the primary fish farming system used globally (18). The research included books, technical and scientific articles, case studies, manuals, and technical reports developed by international institutions, as well as theses and dissertations. Materials containing the following terms were selected: “*Oreochromis niloticus*” AND welfare indicator AND production stage, in the titles, abstracts, or keywords. The welfare indicators defined for each stage of the cultivation process are listed in Table 1. The search period extended from 1985 to 2023.

The reference values for each indicator were classified through a system involving three possible scores (1, 2, or 3). A score of 1 indicates the limits of variation of a particular indicator within the ranges considered ideal for the target species. A score of 2 pertains to variations the animals tolerate, which can cause deleterious effects, provided they are non-lethal. A score of 3 indicates significant levels of variation in a specific indicator that significantly compromises the health and even the survival of the animals, which is deemed unacceptable from an animal welfare perspective. The maximum tolerated mortality rates, the primary indicator of the degree of welfare of the fish, were set at levels much lower than those found in nature, taking into account each stage of the production cycle. The reference values for each indicator and score were established based on the literature available for the species.



### 2.3. Preliminary assessment of the indicators for the early life stages of tilapia

An experimental trial was conducted to establish health protocols for the early life stages of tilapia and to test the pre-selected indicators, allowing for the assessment of their applicability under operational conditions and the evaluation techniques for each indicator. The experiment was conducted at the Laboratory of Research with Aquatic Organisms (LAPOA) of the Integrated Group of Aquaculture and Environmental Studies (GIA) at the Federal University of Paraná, Curitiba, Brazil. All husbandry and experimental procedures were approved by the Animal Use Ethics Committee of the Agricultural Sciences Campus (CEUA) of the Federal University of Paraná (protocol number 021/2023).

A total of 480 newly hatched tilapia larvae were used, subdivided into 12 aquariums of 30 liters in volume, each linked to a chemical-biological filtration system, under controlled conditions of temperature ( $27.01 \pm 1.0^\circ\text{C}$ ), pH ( $7.83 \pm 0.2$ ), and dissolved oxygen ( $5.40 \pm 0.4 \text{ mg/L}$ ). Over 15 days, a tilapia larva from each aquarium was randomly selected using a catch net and carefully and individually transferred to cell culture dishes, duly labeled, containing 10 mL of water on a daily basis. This procedure ensured that the samplings were representative, giving greater precision to the results. The collected fish were anesthetized with clove oil at a concentration of 500 mg/L and kept until they reached stage V anesthesia ( $\sim 5 \text{ min}$ ). At that point, the fish have a medullary collapse and permanent unconsciousness (28). The cell culture dishes containing the collected animals were then transferred to the imaging laboratory, where the animals underwent physical evaluation procedures and photographic recording to determine their health status and evaluate the welfare indicators. The following organs were assessed under a stereo microscope (Figure 2): eyes, mouth and jaws, skin, fins, gill covers, spinal column, and yolk sac. Next, an individual and bilateral photographic record of each individual was made. The indicators that proved unfeasible to measure under operational conditions

and on a commercial scale were excluded from the final version of the protocol. After this final verification, the protocols were reorganized in the format and content presented in Tables 1–10.

### 2.4. Application of welfare indices

Calculation of tilapia welfare indices utilized the same mathematical equations proposed by Pedrazzani et al. (25) for evaluating the welfare degree of *C. idella*. The weights assigned to each indicator in the respective indices were established based on the number of valid bibliographic references found for *O. niloticus* in each production phase, using Google Scholar as the research platform. The variable “Y” was calculated as the integer part of the natural logarithm ( $\ln$ ) of the number of articles identified through specific keywords, as shown in Equation 1. Consequently, partial welfare indices (PWIs) were proposed for each category, along with the general welfare index (GWI), calculated from the PWIs. The PWIs were computed using Equation 2, which considers the weights assigned to each indicator and their respective scores.

$$Y = \text{INT}(\ln(n)) \quad (1)$$

$$PWI_x = \left( \frac{\sum Y}{\sum (S \times Y)} \times (1.4925 - 0.4925) \right) \quad (2)$$

where:

PWI: Partial welfare index standardized to vary continuously between 0 (critical risk of harm to farmed fish welfare) and 1 (maximum welfare or, otherwise, minimum risk of injury to animal welfare), regardless of the number of indicators used in each freedom.

X: Freedom (En, environmental; Be, behavioral; Nu, nutritional, or He, health).

Y: Weight assigned to the specific indicator.

S: Score assigned to the indicators in the analyzed fish farm.

**TABLE 1 Welfare indicators organized according to animal freedoms and production stages of Nile tilapia (BR, breeding; NU, nursery; GO, grow-out).**

Freedom	Indicator	Breeding (BR)	Nursery (NU)	Grow-out (GO)
Environmental	Alkalinity	✓	✓	✓
	Aquatic predators and other interspecific inhabitants	✓	✓	✓
	Dissolved oxygen	✓	✓	✓
	Hapas net cleaning	✓		
	Nitrite	✓	✓	✓
	Non-ionized ammonia	✓	✓	✓
	pH	✓	✓	✓
	Photoperiod	✓	✓	
	Sex ratio (male: female)	✓		
	Temperature	✓	✓	✓
	Terrestrial predators	✓		✓
	Transparency	✓		✓
Health	Breeding control	✓		
	Conditioning/ breeding interval (days)	✓		
	Eggs-macroscopical aspects	✓		
	Emaciation state	✓	✓	
	Eyes	✓	✓	✓
	Fins		✓	✓
	Gills	✓		✓
	Hatching rate		✓	
	Invasive procedures	✓		✓
	Jaws/lips/head	✓	✓	✓
	Mortality (%)	✓	✓	✓
	Operculum	✓	✓	✓
	Sexual maturation	✓		
	Skin	✓	✓	✓
	Spine	✓	✓	✓
	Tail		✓	
	Yolk sac		✓	

(Continued)

**TABLE 1 (Continued)**

Freedom	Indicator	Breeding (BR)	Nursery (NU)	Grow-out (GO)
Nutritional	Amount of feed	✓	✓	✓
	Feed conversion ratio (FCR)			✓
	Feed crude protein	✓	✓	✓
	Feeding frequency	✓	✓	✓
	Food distribution	✓	✓	✓
Behavioral	Anesthesia	✓		✓
	Feed intake	✓	✓	✓
	Swimming behavior	✓	✓	✓
	Stunning during slaughter-reflexes			✓

The general welfare index (GWI) was calculated as the arithmetic mean of the PWIs, multiplied by an elimination factor (*kl*, Equation 3). The *kl* is defined based on the observed mortality rate. Thus, mortality becomes mathematically the most critical indicator for measuring the welfare degree of farmed fish in captivity. By definition, if the mortality rate exceeds 30%, the value of the elimination factor will be equal to zero ( $kl = 0$ ), automatically indicating a “critical” classification for the GWI of the evaluated fish. If the mortality rate is below 30%, the elimination factor will be equal to 1 ( $kl=1$ ), and the welfare of the fish will be determined based on the respective indicators analyzed, their scores, and weights.

$$WGI = \frac{((PWI_{En} + PWI_{Be} + PWI_{Nu} + PWI_{He}) \times kl)}{4} \quad (3)$$

where:

*WGI*: General welfare index, que varia de 0 (critical risk of harm to farmed fish welfare) and 1 (maximum welfare or, otherwise, minimum risk of injury to animal welfare).

*kl*: Knockout level (risk of total impairment of the degree of welfare).

The partial confidence levels (*CLx*) proposed for each PWi are determined based on the number of indicators effectively analyzed in the field. The more indicators evaluated compared to the proposed indicators, the higher the confidence level of the results. The general confidence level (*GCL*) is calculated as the arithmetic mean of the *CLx* (Equation 5). Finally, the *PWix*, *GWI*, *CLx*, and *GCL* are classified and interpreted according to the obtained values.

$$CL_x = \left( \frac{\sum W_{An}}{\sum W_{max}} \right) \quad (4)$$

*CLx*: *PWix* confidence level.

$\sum W_{An}$ : Sum of the weights of the indicators analyzed for the freedom *x*.

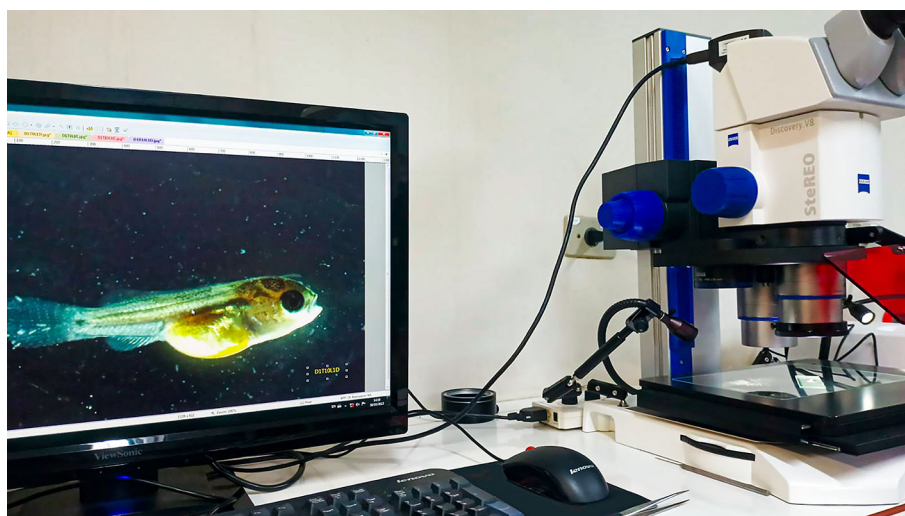


FIGURE 2

Image analysis system used to evaluate and record any health alterations in larvae and post-larvae of *Oreochromis niloticus*.

$\sum W_{max}$ : Sum of the weights of all the defined indicators for the freedom x.

$$GCL = \frac{(IR_{En} + IR_{Be} + IR_{Nu} + IR_{He})}{4} \quad (5)$$

### 3. Results

The general protocol is divided into four categories/freedoms analyzed (environmental, health, nutritional, and behavioral), each with indicators applicable to their respective cultivation phases (Table 1).

#### 3.1. Environmental welfare indicators for Nile tilapia

In the scope of environmental freedom, a set of 12 indicators has been established for different cultivation phases (Table 2). The physicochemical indicators of water have been standardized across all stages. This allows for the prevention of shocks during the transfer of animals between phases whilst maintaining strict adherence to the adopted environmental indicator reference values. Predators and aquatic and terrestrial cohabitants have also been included as environmental welfare indicators, with scores assigned based on their control or presence/absence. A score of 1 should be considered when there is no evidence of other terrestrial or aquatic species in the fishpond. In the case of terrestrial predators, a score of 2 should be applied when the fish farmer adopts control measures, such as filters or screens, but indirect contact still occurs (e.g., a visual connection between tilapia and their predators). For aquatic interspecific predators or cohabitants, a score of 2 should be applied to polyculture systems. The evaluator should assign a score of 3 when there is no control over predators or the presence of cohabitants from other species.

The photoperiod should be considered a relevant indicator during the reproduction and larviculture phases. In the reproduction phase, additional indicators such as hapa cleanliness and the proportion of males and females used in tanks and fishponds were considered (Table 2).

#### 3.2. Health welfare indicators for Nile tilapia

##### 3.2.1. Breeding (BR) and grow-out (GO) phases

Health indicators were established considering the morphological abnormalities in tilapia during their ontogenetic development, observing aspects corresponding to different life and production phases (Tables 3, 4). For breeders (BRs) and animals in the grow-out phase (GO), welfare should be assessed based on physical features such as eye appearance, jaw and lip condition, gill covers, fins, skin, gills, spine, as well as mortality rates and the conduct of invasive procedures with the fish (e.g., vaccination, microchipping, mouth cutting, and removal of dorsal spines). These procedures should be scored as 2 or 3, depending on whether anesthesia is used during execution. For breeders, other important factors for assessing welfare include the stage of sexual maturation, the interval between conditioning periods for mating, and the use of techniques to prevent inbreeding during stock formation and reproductive management. For eggs, only the macroscopic aspect was identified as a practical health indicator, with eggs that are translucent and uniform in appearance considered healthy.

##### 3.2.2. Nursery phase

The health indicators during the nursery phase were subdivided according to the ontogenetic developmental stage of the fish: larvae (L), post-larvae (PL), and fingerlings (F). The laboratory experiment proved essential for evaluating the feasibility of applying the pre-selected indicators in each phase (Table 4). The health status of the eyes, jaw/lips, and head as a whole, as well as the skin, spinal column, and fish mortality, could be easily observed

TABLE 2 Environmental welfare reference values for different tilapia production phases (BR, breeding; NU, nursery; GO, grow-out).

Production stages			Indicators	Scores	Reference values	References
BR	NU	GO				
✓	✓	✓	Temperature (°C)	1	24.0–31.0	(29–35)
				2	21.0–23.9 or 31.1–34.9	
				3	≤ 20.9 or ≥ 35.0	
✓	✓	✓	pH	1	6.0–8.5	(29, 30, 34, 36)
				2	5.5–5.9 or 8.6–9.5	
				3	≤ 5.4 or ≥ 9.6	
✓	✓	✓	Oxygen saturation (%)	1	≥ 6	(29, 37, 38)
				2	40–59	
				3	≤ 39	
✓	✓	✓	Non-ionized ammonia (mg/L of NH <sub>3</sub> )	1	0.00–0.05	(30, 36, 39–42)
				2	0.06–0.09	
				3	≥ 0.10	
✓	✓	✓	Nitrite (mg/L of NO <sub>2</sub> <sup>-</sup> )	1	0.00–0.30	(43–47)
				2	0.31–0.49	
				3	≥ 0.50	
✓	✓	✓	Alkalinity (mg/L of CaCO <sub>3</sub> )	1	30–100	(48–51)
				2	≥ 101	
				3	≤ 29	
✓	✓	✗	Photoperiod (Light: Dark)	1	Natural or 12L: 12D 16L:8D	(32, 39, 52–55)
				2	17L:7D–18L:6D	
				3	19L:5D or lighter; 11L:13D or darker	
✓	✗	✓	Transparency (cm)	1	30–4	(29, 48, 56)
				2	21–29 or 46–60	
				3	≤ 20 or ≥ 61	
✓	✗	✓	Terrestrial predators	1	Absence	(30)
				2	Controlled presence	
				3	Uncontrolled presence	
✓	✓	✓	Aquatic predators and other interspecific inhabitants	1	Absence	(30)
				2	Controlled presence	
				3	Uncontrolled presence	
✓	✗	✗	Hapas net cleaning	1	7–15	(34, 57–59)
				2	≤ 6	
				3	≥ 16	
✓	✗	✗	Sex ratio (male: female)	1	1:2–1:3	(30, 34, 39)
				2	1:1 or 1:4	
				3	Any other configuration	

Indicators included (✓) or not included (✗) in each production phase.

with the aid of a stereoscopic loupe (10x magnification) until the animals reached ~2 cm in standard length. After that, it is possible

to evaluate the external organs of the fish using only a handheld loupe with a 3x magnification capability.














TABLE 3 Health welfare reference values for tilapia breeding (BR, breeding; NU, nursery; GO, grow-out) phases.

Production phase			Indicators	Scores	Description or reference values	References
BR		GO				
✓	✗	✓	Eyes	1	Normal and healthy appearance	(39, 60, 61)
				2	Unilateral hemorrhage, exophthalmos, or traumatic injury	
				3	Bilateral bleeding, exophthalmos, or traumatic injury; chronic condition, impaired vision	
✓	✗	✓	Jaw/lips	1	Normal and healthy appearance	(60, 62, 63)
				2	Mild injury or deformity (without affecting eating)	
				3	Bleeding, redness, severe injury or deformity (affecting eating)	
✓	✗	✓	Operculum	1	Normal and healthy appearance	(63–65)
				2	Absence of tissue (<25%)	
				3	Bleeding, redness, absence of tissue (≥ 25%)	
✓	✗	✓	Skin	1	Normal, healthy appearance, scar tissue	(39, 66, 67)
				2	Punctual loss of scales, ulcers, or superficial lesions <1 cm <sup>2</sup>	
				3	Generalized bristling or loss of scales, ulcers, or superficial lesions >1 cm <sup>2</sup> , redness, necrosis, darkening or lightening, bleeding, swelling, presence of parasites	
✓	✗	✓	Fins	1	Normal and healthy appearance	(67–69)
				2	Scar tissue, mild necrosis, or splitting	
				3	Severe necrosis, splitting or bleeding, redness, exposure to rays, adhered foreign body, ectoparasite	
✓	✗	✓	Gills	1	Normal and healthy appearance	(40, 60, 61, 66, 68, 70, 71)
				2	Light injury, mild necrosis, splitting or thickening	
				3	Bleeding, redness, pallor, severe necrosis, splitting or thickening, excess of mucus, spots, swelling, deformation, adhered foreign body, ectoparasite	
✓	✗	✓	Spine	1	Normal and healthy appearance	(61, 63, 65)
				2	Light deformity (kyphosis, lordosis or scoliosis, normal body weight)	
				3	Severe deformity (kyphosis, lordosis or scoliosis, emaciation)	
✓	✗	✗	Conditioning/breeding interval (days)	1	≥10	(34, 39, 72, 73)
				2	5–9	
				3	≤ 4	
✓	✗	✓	Invasive procedures	1	No invasive procedure	(23, 34, 74)
				2	Microchipping with anesthesia, mouth egg collection	
				3	Microchipping without anesthesia; mouth clipping; up-rooting of dorsal spines	
✓	✗	✗	Sexual maturation	1	Mature animals. Male: Reddish colouration under the jaw; release milt when slight pressure is applied to the abdomen. Female: Ready to spawn grayish colouration under the jaw, pink to red and protruding genital papilla, opened genital pore, distended abdomen	(29)
				2	Male: do not release milt when the abdomen is pressed. Female: Pink to yellow, slightly opened genital pore, slightly distended abdomen	
				3	Male: do not release milt when the abdomen is pressed. Female: Spawned: Red genital papilla, compressed abdomen aspect or; Immature: White to clear and flat genital papilla, regular abdomen aspect	

(Continued)

TABLE 3 (Continued)

Production phase			Indicators	Scores	Description or reference values	References
BR		GO				
			Breeding control	1	Microchipping and family physical restrain	(74, 75)
				2	Family physicals restrain without individual identification	
				3	No breeding control	
			Mortality (%)	1	≤ 10	(23, 25)
				2	11–24	
				3	≥ 25	
			Macroscopical aspect	1	≥ 90 spherical and translucent, with yellowish colouration; the remaining with an opaque aspect	(76–79)
				2	70–89 spherical and translucent, with yellowish colouration; the remaining with an opaque aspect	
				3	≤ 69% spherical and translucent eggs and or detection of some reddish or clustered eggs; presenting white or yellow spots	

Indicators included () or not included () in each production phase.

The indicators exclusively adopted for the larvae were hatching rate, caudal fin formation, and yolk sac. On the other hand, gill covers and fins could only be observed in the post-larvae and fingerlings, as these structures were not fully developed in the larvae. Similarly, fish emaciation was observed only after yolk consumption i.e., in the post-larval stage. Therefore, the “emaciation” indicator was included in the protocols for assessing the welfare level of post-larvae and fingerlings. Due to the difficulty of handling fish during the early life stages and the fragility of organs during physical examination, it was impractical to evaluate the gill condition during the nursery phase. Simply manipulating the larvae during physical examination in the early days of life can cause damage to the skin, eyes, and internal organs, thereby biasing the assessment results. Thus, the convenience and feasibility of applying the protocol for tilapia larvae should be evaluated on a case-by-case basis.

### 3.3. Nutritional welfare indicators for Nile tilapia

Four relevant nutritional indicators applicable to all cultivation phases have been defined: crude protein content in the feed provided to the fish, feed quantity with the biomass of the batch, feeding frequency, and feeding distribution range in the respective cultivation system. The reference values for these indicators were determined based on the weight of tilapia (Table 5). Considering the physiological and energy demands and the management practices adopted during the reproduction phase, two sub-stages were established for calculating the amount of feed provided to the broodstock: “maintenance” and “breeding”. In the grow-out phase, the feed conversion rate (FCR) was incorporated in addition to the mentioned indicators.

### 3.4. Behavioral welfare indicators for Nile tilapia

#### 3.4.1. Breeding (BR) and grow-out (GO)

Behavioral welfare indicators have been established under the management practices commonly adopted during the breeding (BR) and grow-out (GO) phases (Table 6). In both phases, the selected indicators include the effectiveness of anesthesia during invasive procedures and feeding behavior. Regarding feeding, monitoring the time required for fish to capture and entirely consume the provided food is essential. Additionally, during the grow-out phase, the swimming behavior of tilapia during harvesting was considered, along with the total time until the loss of consciousness during slaughter. This latter indicator encompasses the period from the start of the procedure until the point where the animal demonstrates a complete absence of clinical reflexes.

#### 3.4.2. Nursery (NU) phase

Swimming behavior was selected as the sole practical and viable indicator for the visual analysis of larvae and post-larvae, establishing swimming characteristics across different phases (Table 7). The feeding behavior was also included as a welfare indicator for the post-larval stage.

### 3.5. Weights for calculating the partial (PWIs) and general welfare index (GWI)

The weights assigned to determine the importance of each adopted welfare indicator were established based on the number of publications identified in the Google Scholar platform. These values were obtained through a combination of general search

TABLE 4 Health welfare reference values for tilapia nursery (NU) phase, more specifically during larvae (L), post-larvae (P), and fingerlings (F) stages.

Stages			Indicators	Score	Reference values	References
L	P	F				
✓	✗	✗	Hatching rate (% of eggs)	1	≥ 9	(77)
				2	75–89	
				3	≤ 74	
✓	✓	✓	Eyes	1	Normal and healthy appearance	(40, 80–82)
				2	Unilateral: malformation or absence; exophthalmos, redness, darkening, corneal opacity, impaired vision	
				3	Bilateral: malformation or absence; exophthalmos, redness, darkening, corneal opacity, impaired vision	
✓	✓	✓	Jaws/lips/head	1	Normal and healthy appearance	(63, 83, 84)
				2	Malformation without possible feeding restriction	
				3	Malformation with possible feeding restriction, injury, ulcers, necrosis	
✓	✗	✗	Skin	1	Fully pigmented (melanophores throughout the dorsal, ventral, and mediolateral region of the body)	(60, 63, 82, 83, 85)
				2	Partially pigmented (melanophores for some regions of the body)	
				3	Completely translucent or grayish-pale body; redness, paleness, darkening, ectoparasites, white or black spots, bleeding, swelling, ectoparasites, or increase in mucus secretion	
✗	✓	✓	Skin	1	Normal and healthy appearance	(39, 66, 67)
				2	Scar tissue, ulcers, or superficial lesions	
				3	Severe ulcers or lesions, redness, necrosis, white or black spots, cysts, darkening or lightening, bleeding, swelling, ectoparasites, or increase in mucus secretion	
✓	✗	✗	Tail	1	Normal and healthy appearance	(82, 83)
				2	Malformation without movement restriction	
				3	Malformation with movement restriction, darkening, redness	
✓	✓	✓	Spine	1	Functional and healthy appearance	(83)
				2	Malformation without movement restriction	
				3	Malformation with movement restriction	
✓	✗	✗	Yolk sac	1	Functional and healthy appearance	(81, 86)
				2	Malformation without size reduction	
				3	Malformation with size reduction (atrophy), hemorrhage, red spots, sub-epithelial oedema	
✗	✓	✓	Operculum	1	Normal and healthy appearance	(63, 84)
				2	Malformation or lesion causing the absence of tissue (<25% of gills covering)	
				3	Bleeding, swelling, redness, absence of tissue (≥ 25% of gills covering)	
✗	✓	✓	Fins	1	Functional and healthy appearance	(71, 82–85)
				2	Malformation (partial absence), ray deviations	
				3	Necrosis, redness, darkness, white spots, total absence	

(Continued)

TABLE 4 (Continued)

Stages			Indicators	Score	Reference values	References
L	P	F				
✕	✓	✓	Emaciation state	1	No signs of emaciation	(83, 87)
				2	Discrete emaciation	
				3	Advanced emaciation	
✓	✓	✓	Mortality of the batch (%)	1	≤ 10	(81, 88, 89)
				2	11–15	
				3	≥ 16	

Indicators included (✓) or not included (✕) in each production phase.

terms (“*Oreochromis niloticus*” AND “aquaculture” AND the respective life phase) and specific search terms presented in Tables 8–10.

Figure 3 illustrates the application of the welfare protocol for *O. niloticus* during the grow-out phase. The data were simulated using a Microsoft Excel spreadsheet and are derived from a hypothetical but commonly observed scenario in commercial tilapia farming. Based on the example, partial welfare indices related to the environment (PWI<sub>En</sub>), health (PWI<sub>He</sub>), nutrition (PWI<sub>Nu</sub>), and behavior (PWI<sub>Be</sub>) of the fish are calculated. In the analyzed, the first three indices indicate a moderate level of welfare for the cultivated fish. However, the behavioral index suggests a low level of welfare due to improper management practices during harvesting and slaughter. The simulated scenario calculated the general welfare index (GWI) at 0.59, considered moderate. The confidence level in this result was the highest, as all indicators were analyzed. Examples of these index calculations during the breeding and nursery phases are included in Supplementary Figures 1, 2.

## 4. Discussion

The current paradigm regarding the welfare of farmed fish suggests that the evaluation parameters used should be specific to each species, considering the animals' developmental stage and the system in which they are raised. These parameters should also encompass indicators that address the fish's physical, nutritional, environmental, and behavioral aspects (13, 23, 25, 26, 87, 119, 122). However, the understanding and investigation of the psychological dimensions, although constituting one of the animal freedoms, represent a field of scientific knowledge still in its early and nascent stages, especially when compared to the progress achieved regarding animals involved in terrestrial agriculture. Notably, the impacts of domestication on the welfare of farmed fish are more complex to analyse than those faced in welfare studies of land animals that serve as human food sources (123). This is because fish have significantly different genetic, physiological, and behavioral characteristics compared to land animals, as well as experiencing a completely different sensory universe (124, 125). Thus, developing empathy for fish and understanding their needs entails a series of challenges to be evaluated in the field,

which makes it impractical to include them in operational welfare protocols for tilapia.

The welfare of any organism is a dynamic state (126, 127), and the systems used to monitor it should be flexible enough to adapt to changes in its welfare state (17). There is also an understanding of the need for protocols to consider the interaction between different indicators (119, 122, 128), providing relevant information about the overall quality of life of the animals (129, 130). In this context, despite the recent trend of an increasing number of physiological or molecular parameters being tested and recommended and despite the effectiveness of these indicators in laboratory conditions (131–135), the proposed welfare indicators should apply to the practical requirements and routines of commercial fish breeding, larviculture, and grow-out operations (13, 17, 136).

Thus, good operational welfare indicators for farmed fish can be defined as those that address biologically relevant aspects, are easy to use, preferably non-invasive and low-cost (130, 137); reliable, comparable, suitable for aquaculture practices, and appropriate for specific systems or routines (16). They should identify welfare problems and risks to animal welfare and serve as a basis for technical decision-making by producers (138), enabling timely corrections. In contrast, it is highly unlikely that expensive, complex, unreliable, or time-consuming tools or techniques will be adopted and incorporated by the aquaculture industry (139).

### 4.1. Changes in the welfare protocol developed for the grow-out phase of tilapia

When we originally proposed the method and indices that our group has already applied to assess the degree of welfare in grass carp and white-leg shrimp (25, 26), we emphasized that the indicators, their reference values, scores, and weights would need to be periodically reviewed and updated as scientific knowledge advances on the subject. In this article, we put this concept into practice by applying our metrics to assess the welfare of tilapia not only during the grow-out phase but throughout all life stages, whilst also revising and advancing the knowledge generated previously. The indicators now applied to tilapia in the grow-out phase had already been tested and validated by our group under field

TABLE 5 Nutritional welfare reference values for tilapia breeding (BR), nursery (NU), and grow-out (GO) phases and different weights\*.

Indicators	Phases							References
	Scores	BR	NU		GO			
			(≤ 0.5 g)	(0.6–5.0 g)	(5.1–30 g)	(31–150 g)	(151–1,000 g)	
Feed crude protein (%)	1	30–45	40–50	32–40	35–40	28–36	28–	(32, 39, 90–99)
	2	25–29	28–39	28–31	28–34	20–27	20–27	
	3	≤ 24–≥ 46	≤ 27–≥ 51	≤ 27–≥ 41	≤ 27–≥ 41	≤ 19–≥ 37	≤ 19–≥ 37	
Amount of feed (% biomass)**	1	✗	15–30	4–15	4–8	3–6	≥ 2	(30, 45, 88, 99–104)
	2	✗	10–14	3–14	3	2	1	
	3	✗	≤ 10–≥ 31	≤ 2	≤ 2	≤ 1	< 1	
Amount of feed during maintenance (% biomass)*	1	≥ 2	✗	✗	✗	✗	✗	(94, 105–107)
	2	1						
	3	< 1						
Amount of feed during mating (% biomass)*	1	3–5						(94, 105–107)
	2	2	✗	✗	✗	✗	✗	
	3	≤ 1						
Feeding frequency (times/day)	1	✗	5–8	≥ 3	≥ 3	≥ 2	≥ 2	(108–111)
	2	✗	2–4	2	2	1	1	
	3	✗	≤ 1	≤ 1	< 1	< 1	< 1	
Feeding frequency during maintenance (times/day)	1	≥ 2						(108–110)
	2	1	✗	✗	✗	✗	✗	
	3	< 1						
Feeding frequency during mating (times/day)	1	> 1						(112, 113)
	2	1	✗	✗	✗	✗	✗	
	3	< 1						
Food distribution (% of water surface area reach)	1				≥ 75 of surface area			(23, 111)
	2				50–74 of surface area			
	3				≤ 49 of surface area			

(Continued)



TABLE 5 (Continued)

Indicators	Phases							References
	Scores	BR	NU		GO			
			(≤ 0.5 g)	(0.6–5.0 g)	(5.1–30 g)	(31–150 g)	(151–1,000 g)	
FCR	1	✗	✗	✗	≤ 1.0	≤ 1.3	≤ 1.6	(110, 114, 115)
	2				1.1–1.6	1.4–1.7	1.7–2.0	
	3				≥ 1.7	≥ 1.8	≥ 2.1	

Not included (✗) in each production phase.

\*Fish weight adapted from Borges (116). \*\*Always, when rounding a number from one decimal place to none, if the first number after the decimal point is 5 or greater, add 1 to the number before the decimal point; if it is <5, keep the number before the decimal point unchanged.

TABLE 6 Behavioral welfare reference values for tilapia breeding (BR) and grow-out (GO) phases.

Stages		Management	Indicators	Scores	Reference values	Reference
BR	GO					
✓	✗	Invasive procedures (chipping, tagging, clipping)	Anesthesia–surgical stage (lack of balance and swimming; reduction of the opercular rate)	1	Induction in 1–3 min; recovery in ≤ 5 min	(117, 118)
				2	Induction and or recovery in > 5 min	
				3	No induction or no recovery; death	
✓	✓	Feeding	Feed intake (minutes)	1	180–300	(23)
				2	120–179 or 301–419	
				3	≤ 119 or ≥ 420	
✗	✓	Invasive procedure (Vaccination)	Anesthesia–surgical stage (lack of balance and swimming; reduction of the opercular rate)	1	Induction in 1–3 min; recovery in ≤ 5 min	(117, 118)
				2	Induction and or recovery in > 5 min	
				3	No induction or no recovery; death	
✗	✓	Harvest (partial or total)	Swimming behavior	1	Most fish with regular swimming and or few body parts on the surface	(23, 119)
				2	Most of the fish show restless swimming behavior, swimming in different directions and or jumping	
				3	Most fish with decreasing activity; fish trapped against the net or swimming sideways; exposure of the body to air; exhaustion	
✗	✓	Stunning during slaughter	Reflexes*	1	Instantaneous loss of EQ, TGR, VER, OR	(119–121)
				2	Instantaneous loss of EQ and TGR, progressive loss of VER and OR in ≤ 30s	
				3	Progressive loss of E.Q., T.G.R., VER and OR in ≥ 31s	

\*EQ, equilibrium; TGR, tail grab reflex; OR, opercular beating rate; VER, vestibule-ocular reflex.

conditions (23). In this article, in addition to revising and updating the indicators and reference values for this cultivation phase, we simplified the evaluation structure by reducing the number of scores for each indicator from four to three. This approach improves field evaluation, making it more objective and dynamic than protocols with higher scores (87, 140). This reduction in scores

regarding changes affecting eyes, gills, and fins, for example, which often indicate pain and significant diseases, reduces subjectivity in interpreting moderate lesions, thus enhancing evaluation accuracy. However, the blood glucose indicator was removed from the current protocol due to the invasiveness of the method. Additionally, it should be considered that the blood sampling

procedure itself can alter the parameters, causing acute stress to the animals.

Regarding environmental indicators such as temperature, pH, and dissolved oxygen, it is necessary to consider that these parameters naturally vary throughout the day in cultivation ponds. However, the animals can adapt (141) as long as the changes occur within tolerable limits for the species. To avoid conflicts between different values considered acceptable in the literature, we adjusted the new scores to reflect these possible changes in water quality in tilapia cultivation ponds.

The stocking density indicator was removed from the current protocol. Although it is evident that stocking density directly influences the degree of welfare and fish health (142–144), establishing fixed values for this parameter is highly subjective. Defining the ideal density regarding welfare depends significantly on the characteristics of the fish, environmental and nutritional resources provided to them, management practices, fish size, genetic characteristics, and other factors (145).

We also found several operational and conceptual difficulties regarding the shading indicator proposed in the previous protocol. Recent studies present conflicting data that do not accurately reflect the reality of pond cultures, mainly due to inadequate consideration of light intensity and its impact on water quality (148, 149). Initially, we proposed a uniform percentage of shading over the surface of the net pen or pond, typically achieved by using protective and shading screens. However, ponds often have localized shade caused by trees or topographic features in their surrounding areas. This led to misinterpretations during the application of the protocol since localized shading is detrimental to the welfare of tilapia (150). Therefore, we chose to exclude this indicator from the current protocol.

In the original protocol for tilapia grow-out, we established crude protein content (CP), feed conversion ratio (FCR), condition factor (K), and feeding behavior as nutritional indicators. However, after reviewing the data from applying the protocol in various countries (Pedrazzani, unpublished data), we found significant genetic variability amongst cultivated *O. niloticus* strains, which led to morphological variations in the fish. Sometimes the strains were naturally broader than long, and *vice versa*, which affected the value of the condition factor (K) without any relation to fish welfare. Therefore, adjusting the formula for each population or strain of the same species proved impractical for standardization.

The nutritional freedom indicator has now been adjusted to calculate welfare indices by category. Thus, to facilitate tracking the feeding history on farms, we standardized four indicators for all cultivation phases: feed quantity, protein content, feeding frequency, and feed distribution within the pond. Due to the difficulty of capturing and weighing fish in early cultivation, FCR monitoring was suggested only for the grow-out phase.

The behavioral indicators during feeding, harvesting, and slaughter were kept the same as in the original protocol, but their scores were readjusted, aiming to reduce subjectivity during the evaluation, as previously discussed regarding animal suffering associated with health freedom.

**TABLE 7 Behavioral welfare reference values for tilapia during the nursery (NU) phase.**

Indicators	Score	Reference values	References
Larvae swimming behavior	1	Most of the sampled animals presented active swimming against the current	(70, 71, 146, 147)
	2	Most of the sampled animals presented reduced swimming activity against the current	
	3	> 10% of the sample gathered and remained immobile in the center of the container; swimming rapidly; loose equilibrium; present sideways swimming; rubbing against hard surfaces; gasping at the surface	
Post-larvae and fingerlings' swimming behavior	1	Most of the sampled animals presented active swimming against the current and swimming vertically or horizontally in the water column and with short periods at the tank bottom	(40, 70, 71, 146)
	2	Most of the sampled animals presented reduced swimming activity in the water column, or < 10% of the sample present at the tank bottom	
	3	> 10% of the sampled animals presented spiral swimming, efforts to swallow air or float on the water surface; rubbing against hard surfaces; gasping at the surface or being immobile at the tank bottom	
Post-larvae and fingerlings' feed intake (min)	1	180–300	
	2	120–179 or 301–419	(23)
	3	≤ 119 or ≥ 420	

An essential conceptual aspect concerns fish slaughter, which we associate with the grow-out phase. Although the industrial-scale slaughter of tilapia is already a reality (151, 152), globally, this practice remains an exception rather than the rule. Transporting water and live tilapia to the processing plants is costly and requires complex logistics and appropriate slaughter facilities. In most cases, fish are sold alive in markets or slaughtered on-site at the farms where they are cultivated (29, 39, 57, 152–154). Under these circumstances, it is common for animals to be slaughtered without prior stunning, often asphyxiated in the air or ice (155–157). Thus, we believe that a more practical way to improve fish welfare is to consider slaughter as part of the grow-out process. However, considering the expansion of international tilapia trade, the evolution of industrial aquaculture, and the increasing interest of consumers in the quality of the product and the welfare of farmed fish, it is plausible to project that slaughter will soon become a genuinely autonomous stage in the tilapia production cycle.

**TABLE 8** Number of documents and the respective weights of the indicators, established from the general search terms (“*Oreochromis niloticus*” AND “aquaculture” AND “breeding”-larva -nursery AND the specific search terms used in Google Scholar in July 2023).

Freedom	Indicator	Specific search terms	Number of documents ( <i>n</i> )	Weight [ln( <i>n</i> )]
Environmental	Alkalinity	“alkalinity”	1.510	7
	Aquatic predators and other interspecific inhabitants	“aquatic predators” OR “interspecific inhabitants”	43	4
	Dissolved oxygen	“dissolved oxygen”	6.780	9
	Hapas net cleaning	“hapa” AND “clean”	76	4
	Nitrite	“nitrite”	2.560	8
	Non-ionized ammonia	“ammonia”	5.250	9
	pH	“pH”	13.500	10
	Photoperiod	“photoperiod”	3.460	8
	Sex ratio (male: female)	“sex ratio”	2.550	8
	Temperature	“temperature”	15.500	10
	Terrestrial predators	“terrestrial” AND “predator”	627	6
	Transparency	“transparency”	976	7
Health	Breeding control	“breeding control”	10	2
	Conditioning/breeding interval (days)	“conditioning” OR “breeding interval”	729	7
	Eggs–macroscopical aspects	“eggs” AND “macroscopic”	240	5
	Eyes	“eyes”	674	7
	Fins	“fins”	2.480	8
	Gills	“gills”	4.000	8
	Invasive procedures	“chipping” OR “tagging” OR “clipping.”	847	7
	Jaws/lips	“jaw” OR “lips”	857	7
	Mortality (%)	“mortality”	8.020	9
	Operculum	“operculum”	573	6
	Sexual maturation	“maturation”	5.510	9
	Skin	“skin”	5.090	9
	Spine	“spine”	657	6
Nutritional	Amount of feed	“amount of feed”	910	7
	Feed Crude Protein	“crude protein”	4.700	8
	Feeding frequency	“feeding frequency”	834	7
	Food distribution	“food distribution”	91	5
Behavioral	Anesthesia–surgical stage	“anesthesia” OR “anesthesia”	486	6
	Feed intake	“feed intake”	3.290	8
	Swimming behavior	“swimming behavior” OR “swimming behavior”	432	6

## 4.2. First protocol of tilapia welfare for the reproduction and nursery stages

The sanitary indices employed during the grow-out phase were maintained with those developed for adult fish and further expanded. The additions were intended to integrate essential management practices commonly used in commercial tilapia

reproduction and larviculture facilities. Indicators such as the stage of sexual maturation, the time interval adopted for the recovery/conditioning of breeders between reproductive cycles, and the assessment of the adoption or not of methods or mechanisms of control to avoid inbreeding amongst breeder batches were included. In addition to these, other less intuitive indicators were proposed, such as, for instance, the maximum

**TABLE 9** Number of documents and the respective weights of the indicators, established from the general search terms ("*Oreochromis niloticus*" AND "aquaculture" AND nursery" OR "larviculture" – breeding AND the specific search terms used in Google Scholar in July 2023 + specific search terms).

Freedom	Indicator	Specific search terms	Number of documents ( <i>n</i> )	Weight [ln( <i>n</i> )]
Environmental	Alkalinity	"alkalinity"	932	7
	Aquatic predators and other interspecific inhabitants	"aquatic predators" OR "interspecific inhabitants"	9	2
	Dissolved oxygen	"dissolved oxygen"	2.810	8
	Nitrite	"nitrite"	1.550	7
	Non-ionized ammonia	"ammonia"	2.620	8
	pH	"ph"	4.850	8
	Photoperiod	"photoperiod"	979	7
	Temperature	"temperature"	4.660	8
Health	Emaciation state	"emaciation"	35	4
	Eyes	"eyes"	273	6
	Fins	"fins"	569	6
	Hatching rate	"hatching"	1.010	7
	Jaws/lips/head	"jaw" OR "lips" OR "head"	1.350	7
	Mortality (%)	"mortality"	2.780	8
	Operculum	"operculum"	134	5
	Skin	"skin"	1.230	7
	Spine	"spine"	134	5
	Tail	"tail"	523	6
	Yolk sac	"yolk sac"	364	6
Nutritional	Amount of feed	"amount of feed"	550	6
	Feed crude protein	"crude protein"	2.090	8
	Feeding frequency	"feeding frequency"	62	4
	Food distribution	"food distribution"	37	4
Behavioral	Feed intake	"feed intake"	13,50	7
	Swimming behavior	"swimming behavior"	918	7

intervals adopted for cleaning the hapas where the breeders are kept during reproduction—an essential factor to prevent the obstruction of the screens since this hinders the renewal of water and compromises the health of the breeders and can cause not only a reduction of zootechnical indices but also a decrease in immunity and the emergence of diseases (30, 57, 158). The proportion between males and females used during reproduction was another indicator included in the protocol since it interferes with population dynamics and the degree of aggressiveness of the males during the mating phase (30).

For the behavioral assessment of the breeders, we emphasize the recommendation for using anesthesia during any invasive procedures. We kept the "feeding behavior" indicator, although it is relevant to note that female tilapias reduce their food intake during the mating and spawning periods by incubating the eggs in the mouth, whilst males can increase their feed intake in the same period (159). This protocol included the photoperiod due to its central role in the natural

induction of sexual maturation and in defining reproductive rates (160, 161).

Concerning the welfare of eggs, larvae, and post-larvae, considering that this is still a controversial topic, that it is in its initial stages of unveiling scientific knowledge, and that there is a tremendous natural vulnerability of larvae and post-larvae to handling, which requires even greater caution and rigor; we advocate for the implementation of the protocol proposed in this phase of tilapia cultivation. However, at the same time, we suggest that the suitability and necessity of its application be determined on a case-by-case basis, considering the overall conditions of the batch and the local structural and operational capacity to assess the proposed indicators.

The welfare of eggs and larvae involves the parents' nutritional, social, and environmental experiences during their development (162). The environment likely influences the epigenetic pattern of gametes, embryos, or adult organisms (163). During gametogenesis, the DNA is reprogrammed, and

**TABLE 10** Number of documents and the respective weights of the indicators established from the general search terms ("*Oreochromis niloticus*" AND "aquaculture" AND "farming" AND "pond" AND the specific search terms used in Google Scholar in July 2023).

Freedom	Indicator	Specific search terms	Number of documents (n)	Weight [ln(n)]
Environmental	Alkalinity	"alkalinity"	2.850	8
	Aquatic predators and other interspecific inhabitants	"aquatic predators" OR "interspecific inhabitants"	44	4
	Dissolved oxygen	"dissolved oxygen"	8.500	9
	Nitrite	"nitrite"	4.310	8
	Non-ionized ammonia	"ammonia"	7.380	9
	pH	"pH"	12.500	9
	Temperature	"temperature"	12.900	9
	Terrestrial predators	"terrestrial" AND "predator"	533	6
	Transparency	"transparency"	202	5
Health	Eyes	"eyes"	196	5
	Fins	"fins"	255	6
	Gills	"gills"	370	6
	Invasive procedures	"chipping" OR "tagging"	505	6
	Jaws/lips	"jaw" OR "lips"	467	6
	Mortality (%)	"mortality"	7.740	9
	Operculum	"operculum"	449	6
	Skin	"skin"	3.830	8
	Spine	"spine"	355	6
Nutritional	Amount of feed	"amount of feed"	1.710	7
	Feed conversion ratio (FCR.)	"F.C.R."	4.550	8
	Feed crude protein	"crude protein"	107	5
	Feeding frequency	"feeding frequency"	1.260	7
	Food distribution	"food distribution"	199	5
Behavioral	Feed intake	"feed intake"	87	4
	Harvest (partial or total)—swimming behavior	"swimming behavior" OR "swimming behavior"	330	6
	Invasive procedure (vaccination)—Anesthesia—surgical stage	"vaccination"	1.060	7
	Stunning during slaughter—reflexes	"stunning"	129	5

this information will be transmitted to the offspring, resulting in transgenerational effects that directly impact the quantity, viability, social status, neurogenesis, and adaptation of future generations (162). Sneddon et al. (1) highlight that fish larvae have various brain structures that process emotions and learning, although they are not identical to the human brain. Lopez et al. (164) demonstrated that zebrafish larvae at 5 days post-fertilization (5 daf) respond to harmful and potentially painful stimulation caused by environmental acidification, exhibiting similar behaviors to adult fish and reducing their activities. This response was alleviated by analgesic drugs such as lidocaine and morphine. Furthermore, different larval rearing protocols can have a significant impact on larval size and mass, survival rates, and the sex ratio of larvae (165). Therefore, it is necessary to consider that welfare should be understood as continuous and intergenerational, as there is a direct

link between offspring adaptation and the resources provided by parents (162, 166). On the other hand, the quality of the eggs and larvae will also significantly impact the welfare and health of tilapias throughout their lives (167).

Another argument to be considered is that, despite the legislation and regulatory frameworks of the vast majority of countries still not protecting fish larvae, we must consider the scientific arguments linked to the presence of sentience in larvae and fry (1, 162, 164). In this sense, we advocate the application of the proposed protocol based on the precautionary principle, which establishes that when evidence of sentience is inconclusive, we should "give the benefit of the doubt" to the animal or "err on the side of caution" (168). Thus, we understand that the theme "welfare in the early stages of fish life" has relevance to be applied in current aquaculture but, mainly, that it will have a significant



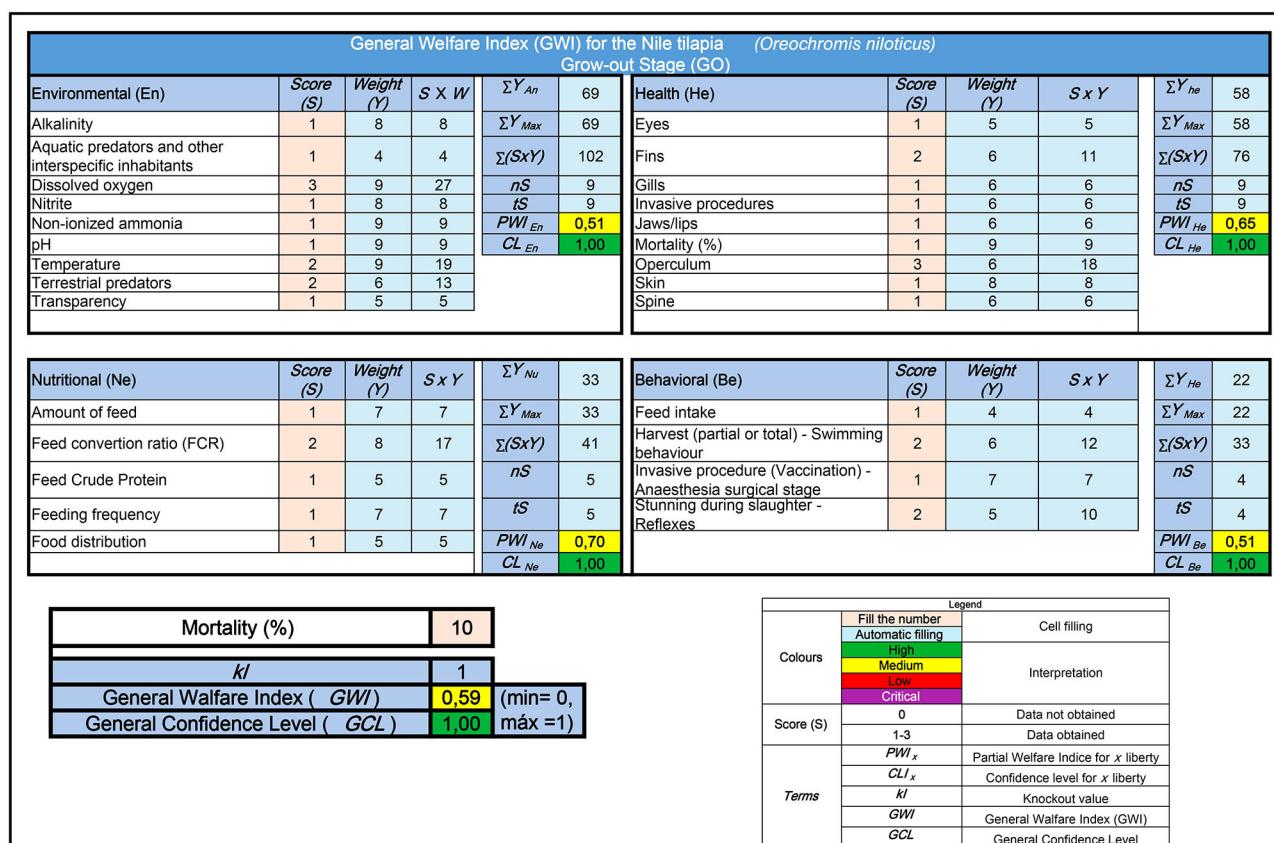


FIGURE 3

Example of calculating partial welfare indices for tilapia during the grow-out phase in land-based ponds and the overall welfare index using a calculation model developed in the Microsoft Excel application. In this hypothetical case, the model indicates a moderate level of welfare and a high level of confidence concerning this result.

impact on the aquaculture that will be practiced in the coming years, possibly under a scenario of regulatory restrictions and rigorous governance practices (169). Therefore, be it for biological, ethical, moral, or commercial reasons, and even recognizing the fragility of the current stage of knowledge on the subject and the need for subsequent discussions on the effectiveness of the application of welfare protocols for the larval and post-larval stages of *O. niloticus*, we understand it to be recommendable and, at the same time, almost “inevitable” that the early stages of life be included in animal welfare assessment protocols in fish farming.

We tested and validated their operational feasibility in the laboratory to assess the indicators proposed here for the early stages of tilapia life. The experiments carried out made it possible to identify the most suitable indicators and, at the same time, exclude those that did not meet the established prerequisites.

Some health indicators were not incorporated into the larval protocols, as structures such as fins and gills are still in development during the ontogenetic processes that occur during the nursery phase, making their visual assessment difficult. For post-larvae and fry, it is relevant to consider the “degree of emaciation” as an indirect indicator of feeding effectiveness. The “yolk sac” indicator was included in the assessment of larvae, as organisms at this stage of life still have endogenous energy reserves and, therefore, do not

show apparent signs of emaciation (76). In all life stages, we kept the mortality rate as an indicator of welfare, as most fish deaths in captivity are likely preceded or accompanied by suffering. Thus, long-term mortality rates may serve as indicators of the degree of retrospective welfare and signal possible future impacts on the success rates of the enterprise (170).

In the analysis of tilapia larvae and post-larvae, we identified a significant number of articles focussed solely on swimming behavior and the changes commonly related to water contamination or the occurrence of diseases. Therefore, we included only this behavioral indicator for the larval stage once they have endogenous feeding. We also excluded some indicators from this stage, as larvae are produced in laboratories, rendering indicators such as the presence of “terrestrial predators” or “water transparency” irrelevant, for instance.

#### 4.3. Partial and general welfare index (PWIs) and general welfare index (GWI)

Animal welfare should not be linked to cultural differences or subjective criteria but to the species' biology (171). Therefore, quantitative animal welfare assessment is essential for promoting humane and responsible management practices in the animal production industry (172, 173). Moreover, using quantitative and

standardized approaches in welfare measurement allows producers to tangibly demonstrate their commitment to animal welfare, which helps build consumer trust and generate new market opportunities (174, 175).

The metrics proposed in this article aim, pioneeringly, to provide a holistic and quantifiable assessment of the welfare of Nile tilapia throughout all stages of its captive life cycle. These metrics were established based on indicators that were simultaneously simple, understandable, and already part of the routine production of the species on a commercial scale. To achieve this, we used indicators representing the nutritional, behavioral, health, and critical environmental conditions to which tilapia are exposed throughout their production process.

Furthermore, this study's partial (PWI) and overall (GWI) welfare indices offer an objective animal welfare assessment. They are based on data and scientifically supported metrics rather than opinions or subjective factors, providing excellent reliability and accuracy. The proposed indices can provide producers with a valuable tool for retrospective and prospective analyses within the same production cycle, enabling informed strategic decisions for the welfare of farmed fish and the profitability and efficiency of their businesses.

There is already recognition within the scientific community that a single score simplifies data interpretation and constitutes a valuable tool for researchers, producers, certifying bodies, and regulatory agencies (87, 122). This characteristic allows the proposed indices to establish a solid foundation for developing animal welfare regulations and guidelines, enabling authorities to define clear and measurable standards to ensure ethical and humane treatment in tilapia farming operations, one of the most critical species in global aquaculture (18).

In this study, PWIs and GWIs follow the same conceptual and mathematical logic applied and extensively discussed concerning grass carp (25) and white-leg shrimp (26). However, the indicators and their respective reference values, scores, and weights are specific to *O. niloticus*, covering aspects of breeding, larviculture, fingerling rearing, and the grow-out phase—in this case, in earthen ponds. Like in previous studies, the weights assigned to each indicator were identified using Google Scholar. These weights ranged from 2 to 10 and were determined based on the number of scientific documents related to each indicator. It should be noted, however, that despite being practical, this method has limitations. For example, the number of publications on a specific topic may not reflect its relevance in a practical context; it may underestimate the importance of less-researched welfare indicators; it can be influenced by the availability of funding for research in specific areas, which may not reflect the importance of those areas for animal welfare; there may also be variations due to the language used in search terms and differences in the consulted databases. Despite these limitations, the approach is robust, standardisable, and encompasses several advantages, including the comprehensiveness and objectivity of evaluating welfare in tilapia farming, recognizing differences that each indicator presents in fish welfare, and the ability to update and refine as new research is published. Subsequent analysis can explore alternative methods for

assigning weights to welfare indicators and further examine how the relative importance of these indicators may vary across different life stages.

## 5. Conclusion

In this study, we have proposed a comprehensive and quantitative approach to assess the welfare of Nile tilapia throughout their life cycle (eggs, larvae, post-larvae, juveniles, and adults) and all cultivation phases (breeding, nursery, and grow-out) in captivity. This approach has generated a valuable and standardized tool for aquaculturists to monitor and improve their production systems, with the potential to enhance the welfare of *O. niloticus* in aquaculture significantly.

The proposed methods will allow for a comprehensive, precise, and tailored analysis of welfare throughout the entire life cycle and all stages of tilapia farming. The developed quantitative indices will enable a standardized comparison of animal welfare amongst different enterprises, locations, and periods, serving as relevant tools to evaluate the effectiveness of other management practices and identify areas that need improvement. This approach can potentially enhance farming practices and promote the welfare of tilapia whilst providing a valuable tool for advancing more sustainable and ethical aquaculture practices.

## Data availability statement

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

## Ethics statement

The animal study was approved by Animal Use Ethics Committee of the Agricultural Sciences Campus (CEUA) of the University of Paraná under protocol 021/2023. The study was conducted in accordance with the local legislation and institutional requirements.

## Author contributions

AP: Conceptualization, Data curation, Formal analysis, Investigation, Validation, Writing—original draft, Writing—review and editing. NC: Investigation, Methodology, Resources, Validation, Writing—original draft. MQ: Project administration, Supervision, Writing—review and editing. CT: Investigation, Validation. VB: Validation. AO: Conceptualization, Methodology, Investigation, Formal analysis, Writing—original draft, Writing—review and editing, Supervision, Project administration.

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## Conflict of interest

AP was employed by Wai Ora Aquaculture and Environmental Technology Ltd. MQ was employed by FAI Farms from Brazil.

## References

1. Sneddon LY, Lopez-Luna J, Wolfenden DCC, Leach MC, Valentim AM, Steenbergen PJ, et al. Fish sentience denial: muddying the waters. *Anim Sentience*. (2018) 115:1–12. doi: 10.51291/2377-7478.1317
2. Chandroo KP, Duncan IJH, Moccia RD. Can fish suffer? Perspectives on sentience, pain, fear and stress. *Appl Anim Behav Sci*. (2004) 86:225–50. doi: 10.1016/j.applanim.2004.02.004
3. Ashley PJ. Fish welfare: Current issues in aquaculture. *Appl Anim Behav Sci*. (2007) 104:199–235. doi: 10.1016/j.applanim.2006.09.001
4. Braithwaite V. *Do Fish Feel Pain?* Oxford: Oxford University Press (2010) 208 p.
5. UK Public General Acts. *Animal Welfare (Sentience) Act 2022 Chapter 22*. (2022). Available online at: <https://www.legislation.gov.uk/ukpga/2022/22/contents> (accessed June 10, 2023).
6. Gismervik K, Tørd B, Kristiansen TS, Osmundsen T, Størkersen KV, Medaas C, et al. Comparison of Norwegian health and welfare regulatory frameworks in salmon and chicken production. *Rev Aquac*. (2020) 12:2396–410. doi: 10.1111/raq.12440
7. Giménez-Candela MM, Saraiva JL, Bauer H. The legal protection of farmed fish in Europe – analysing the range of EU legislation and the impact of international animal welfare standards for the fishes in European aquaculture. *Derecho Animal Forum Animal Law Stud*. (2020) 11:65–119. doi: 10.5565/rev/da.460
8. Seibel H, Weirup L, Schulz C. Fish welfare – between regulations, scientific facts and human perception. *Food Ethics*. (2020) 5:1–11. doi: 10.1007/s41055-019-00063-3
9. Poli BM. Farmed fish welfare-suffering assessment and impact on product quality. *Ital J Anim Sci*. (2009) 8:139–60. doi: 10.4081/ijas.2009.s1.139
10. Newton R, Zhang W, Xian Z, McAdam B, Little DC. Intensification, regulation and diversification: the changing face of inland aquaculture in China. *Ambio*. (2021) 50:1739–56. doi: 10.1007/s13280-021-01503-3
11. Maesano G, Di Vita G, Chinnici G, Pappalardo G, D'Amico M. The role of credence attributes in consumer choices of sustainable fish products: a review. *Sustainability*. (2020) 12:10008. doi: 10.3390/su122310008
12. Kaimakoudi E. Policy initiatives towards enhancing consumer knowledge and tackling consumer confusion in aquaculture sector. *Aquac Int*. (2023) 2023:1–9. doi: 10.1007/s10499-023-01143-2
13. Barreto MO, Rey Planellas S, Yang Y, Phillips C, Descovich K. Emerging indicators of fish welfare in aquaculture. *Rev Aquac*. (2022) 14:343–61. doi: 10.1111/raq.12601
14. Toni M, Manciooco A, Angiulli E, Alleve E, Cioni C, Malavasi S. Review: assessing fish welfare in research and aquaculture, with a focus on European directives. *Animal*. (2019) 13:161–70. doi: 10.1017/S1751731118000940
15. Størkersen KV, Osmundsen TC, Stien LH, Medaas C, Lien ME, Tørd B, et al. Fish protection during fish production. Organizational conditions for fish welfare. *Mar Policy*. (2021) 129:104530. doi: 10.1016/j.marpol.2021.104530
16. Nilsson J, Stien LH, Iversen MH, Kristiansen TS, Torgersen T, Oppedal F, et al. Part A. Knowledge and theoretical background. In: Noble C, Gismervik K, Iversen MH, Kolarevic J, Nilsson J, Stien LH, et al., editors. *Welfare Indicators for Farmed Atlantic Salmon: Tools for Assessing Fish Welfare*. Tromsø: Nofima (2018). p. 351.
17. Segner H, Reiser S, Ruane N, Rösch R, Steinhagen D, Vehanen T. *Welfare of Fishes in Aquaculture*. Budapest: FAO (2019). 18 p.
18. FAO. *The State of World Fisheries and Aquaculture 2022. Towards Blue Transformation*. Rome, Italy: FAO (2022). 266 p.
19. El-Sayed A-FM, Fitzsimmons K. From Africa to the world—the journey of Nile tilapia. *Rev Aquac*. (2023) 15:6–21. doi: 10.1111/raq.12738
20. Rodriguez-Barreto D, Rey O, Uren-Webster TM, Castaldo G, Consuegra S, Garcia de Leaniz C. Transcriptomic response to aquaculture intensification in Nile tilapia. *Evol Appl*. (2019) 12:1757–71. doi: 10.1111/eva.12830
21. Huntingford F, Kadri S. Defining, assessing and promoting the welfare of farmed fish. *Rev Sci Tech*. (2014) 33:233–44. doi: 10.20506/rst.33.1.2286
22. Daskalova A. Farmed fish welfare: stress, post-mortem muscle metabolism, and stress-related meat quality changes. *Int Aquat Res*. (2019) 11:113–24. doi: 10.1007/s40071-019-0230-0
23. Pedrazzani AS, Quintiliano MH, Bolfe F, Sans ECdO, Molento CFM. Tilapia on-farm welfare assessment protocol for semi-intensive production systems. *Front Vet Sci*. (2020) 7:1–16. doi: 10.3389/fvets.2020.606388
24. FAWC. *Prevention of Cruelty to Animals Act*. (1979). Available online at: <https://www.gov.uk/government/groups/farm-animal-welfare-committee-fawc> (accessed March 15, 2023).
25. Pedrazzani AS, Tavares CSP, Quintiliano M, Cozer N, Ostrensky A. New indices for the diagnosis of fish welfare and their application to the grass carp (*Ctenopharyngodon idella*) reared in earthen ponds. *Aquac Res*. (2022) 53:5825–45. doi: 10.1111/are.16105
26. Pedrazzani AS, Cozer N, Quintiliano MH, Tavares CPdS, Silva UAT, Ostrensky A. Non-invasive methods for assessing the welfare of farmed white-leg shrimp (*Penaeus vannamei*). *Animals*. (2023) 13:807. doi: 10.3390/ani13050807
27. Whittemore R, Knapp K. The integrative review: updated methodology. *J Adv Nurs*. (2005) 52:546–53. doi: 10.1111/j.1365-2648.2005.03621.x
28. Ross LG, Ross B. *Anaesthetic and sedative techniques for aquatic animals*. Ross LG, Ross B, editors. *Nova Jersey*. Hoboken, NJ: John Wiley & Sons (2009). 222 p.
29. Nandlal S, Pickering T. *Tilapia Fish Farming in Pacific Island Countries*. Noumea, New Caledonia: Secretariat of the Pacific Community (2004). 190–203 p.
30. El-Sayed AFM. *Tilapia Culture: Second Edition*. Massachusetts: Elsevier (2019). 348 p.
31. López-Olmeda JF, Noble C, Sánchez-Vázquez FJ. Does feeding time affect fish welfare? *Fish Physiol Biochem*. (2012) 38:143–52. doi: 10.1007/s10695-011-9523-y
32. El-Naggar G, El Nady M, Kamar M, Al-Kobabay A. Effect of photoperiod, dietary protein and temperature on reproduction in Nile tilapia (*Oreochromis niloticus*). *Tilapia Culture in the 21st Century*. In: *Proceedings from the Fifth International Symposium on Tilapia Aquaculture*. Rio de Janeiro, Brazil: American Tilapia (2000). p. 352–8.

The remaining authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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## Supplementary material

The Supplementary Material for this article can be found online at: <https://www.frontiersin.org/articles/10.3389/fvets.2023.1268396/full#supplementary-material>

33. FAO. *The State of World Fisheries and Aquaculture*. (2009). Available online at: <https://www.fao.org/fishery/en> (accessed May 24, 2023).
34. Bhujel RC, A. review of strategies for the management of Nile tilapia (*Oreochromis niloticus*) broodfish in seed production systems, especially hapa-based systems. *Aquac*. (2000) 181:37–59. doi: 10.1016/S0044-8486(99)00217-3
35. Alam SMA, Sarker MSI, Miah MMA, Rashid H. Management strategies for Nile tilapia (*Oreochromis niloticus*) hatchery in the face of climate change induced rising temperature. *Aquac Stud*. (2021) 21:55–62. doi: 10.4194/2618-6381-V21\_2\_02
36. Kubitz F. Tilápias: Qualidade da água, sistemas de cultivo, planejamento da produção, manejo nutricional e alimentar e sanidade. Parte II. In: Filho JC, editor. *Tilapias*. Rio de Janeiro, Brazil: Revista Panorama da Aquicultura (2015). p. 7.
37. SEBRAE. *Guia técnico para empreender na criação de tilápias em tanques-rede*. 1 ed. Brasília, DF: Serviço Brasileiro de Apoio às Micro e Pequenas Empresas–SEBRAE (2016). 84 p.
38. SENAR. *Piscicultura: manejo da produção de peixes em viveiros*. 1 ed. Brasília: Serviço Nacional de Aprendizagem Rural (2017). 124 p.
39. El-Sayed AFM. *Tilapia Culture*. Massachusetts, USA: Cabi Publishing (2006). 277 p.
40. Benli AÇK, Koksall G. The acute toxicity of ammonia on Tilapia (*Oreochromis niloticus* L) larvae and fingerlings. *Turk J Vet Anim Sci*. (2005) 29:339–44.
41. Costa FFB. *Efeito agudo e subcrônico da amônia sobre a Tilápia do Nilo*. (dissertation/master's thesis). Belo Horizonte, Brazil: Universidade Federal de Minas Gerais (2018).
42. El-Shafai SA, El-Gohary FA, Nasr FA, Van Der Steen NP, Gijzen HJ. Chronic ammonia toxicity to duckweed-fed tilapia (*Oreochromis niloticus*). *Aquac*. (2004) 232:117–27. doi: 10.1016/S0044-8486(03)00516-7
43. Atwood HL, Fontenot QC, Tomasso JR, Isely JJ. Toxicity of nitrite to Nile tilapia: effect of fish size and environmental chloride. *N Am J Aquac*. (2001) 63:49–51. doi: 10.1577/1548-8454(2001)063<0049:tontnt>2.0.co;2
44. Ferreira GdS, Marcondes Maciel L, Dalmass MV, Tiepo MG. *Tilápia-do- Nilo: Criação e cultivo em viveiros no estado do Paraná*. Curitiba, PR: GIA (2015). 290 p.
45. Silva BC, Massago H, Marchiori NdC. *Monocultivo de tilápia em viveiros escavados em Santa Catarina*. Florianópolis, SC: Empresa de Pesquisa Agropecuária e Extensão Rural de Santa Catarina (Epagri) (2019). 126 p.
46. Silva MdS. *Efeito agudo da amônia e do nitrito em tilápias Oreochromis niloticus mantidas em baixa salinidade* [dissertation/master's thesis]. Belo Horizonte, Brazil: Universidade Federal de Minas Gerais (2013).
47. Wang Y, Zhang W, Li W, Xu Z. Acute toxicity of nitrite on tilapia (*Oreochromis niloticus*) at different external chloride concentrations. *Fish Physiol Biochem*. (2006) 32:49–54. doi: 10.1007/S10695-005-5744-2
48. Boyd CE, Tucker AC. *Pond Aquaculture Water Quality Management*. Massachusetts, USA: Kluwer Academic Publishers (1998). 700 p.
49. Kubitz F, Kubitz LMM. Tilápias: qualidade da água, sistemas de cultivo, planejamento da produção, manejo nutricional e alimentar e sanidade—parte I. In: Filho JC, editor. *Tilapia: Um bom planejamento gera alta rentabilidade*. Rio de Janeiro, Brazil: Revista Panorama da Aquicultura (2000). p. 45–7.
50. Rojas N, Mainardes-Pinto C, Rocha O, Silva A. Larviculture of *Oreochromis niloticus* Linnaeus, 1758 (Perciformes, Cichlidae) in ponds with different levels of water alkalinity. *Acta Limnol Bras*. (2004) 16:341–9.
51. Cavalcante DdH. *Relação dureza/alcalinidade da água e seus efeitos sobre a qualidade da água, do solo e desempenho zootécnico de juvenis de tilápia do Nilo, Oreochromis niloticus, mantidos em condições laboratoriais* (dissertation/master's thesis). Ceará, Brazil: Universidade Federal do Ceará (2012).
52. Baroiller J-F, Desprez D, Carteret Y, Tacon P, Borel F, Hoareau M-C, et al. Influence of environmental and social factors on the reproductive efficiency in three tilapia species, *Oreochromis niloticus*, *O. aureus*, and the red tilapia (Red Florida strain). In: Baroiller J-F, editor. *Proceedings of the Fourth International Symposium on Tilapia in Aquaculture. Actes du 4ème congrès international sur tilapia en aquaculture*. Orlando, US: Northeast Regional Agricultural Engineering Service (1997). p. 808.
53. Biswas AK, Morita T, Yoshizaki G, Maita M, Takeuchi T. Control of reproduction in Nile tilapia *Oreochromis niloticus* (L) by photoperiod manipulation. *Aquac*. (2005) 243:229–39. doi: 10.1016/J.AQUACULTURE.2004.10.008
54. El-Sayed AFM, Kawanna M. Effects of photoperiod on growth and spawning efficiency of Nile tilapia (*Oreochromis niloticus* L) broodstock in a recycling system. *Aquac Res*. (2007) 38:1242–7. doi: 10.1111/J.1365-2109.2007.01690.X
55. Ridha MT, Cruz EM. Effect of light intensity and photoperiod on Nile tilapia *Oreochromis niloticus* L. seed production. *Aquac Res*. (2000) 31:609–17. doi: 10.1046/J.1365-2109.2000.00481.X
56. Abdel-Satar AM, Al-Khabbas MH, Alahmad WR, Yousef WM, Alsomadi RH, Iqbal T. Quality assessment of groundwater and agricultural soil in Hail region, Saudi Arabia. *Egypt J Aquat Res*. (2017) 43:55–64. doi: 10.1016/j.ejar.2016.12.004
57. Bhujel R. A manual for Tilapia business management. Bhujel RC, editor. *Wallingford*. Leesburg: Cabi (2014). 216 p.
58. Akian DD, Quenum CL, Koua DNZ, Houra JAK, Clota F, Bégout ML, et al. Comparative study of larvae production by the Nile tilapia (*Oreochromis niloticus*, Linné, 1758) Bouaké strain between earthen ponds and hapas. *Aquac Res*. (2022) 53:6049–55. doi: 10.1111/are.16075
59. Huisman EA, Zonneveld N, Bouwmans AH. *Aquacultural Research in Asia: Management Techniques and Nutrition: Proceedings of the Asian Seminar on Aquaculture*. Wageningen, Netherlands: Pudoc (1989). 271 p.
60. Noga EJ. *Fish disease: Diagnosis and Treatment*. Second ed. North Carolina, US: Wiley-Blackwell (2010). 544 p.
61. Smith SA. *Fish Diseases and Medicine*. First ed. Florida, US: CRC Press (2019). 412 p.
62. Eiras JC, Segner H, Wahli T, Kapoor BG. *Fish Diseases (2 Vols.)*. First ed. Einfield, UK: Taylor and Francis Inc (2008). 1338 p.
63. Tave D, Jo JY, Kim DS. Gross abnormalities in tilapia. *Fish Aquat Sci*. (2011) 14:148–60. doi: 10.5657/FAS.2011.0148
64. Handwerker TS, Tave D. Semioperculum: a nonheritable deformity in mozambique tilapia. *J Aquat Anim Health*. (1994) 6:85–8. doi: 10.1577/1548-8667(1994)006<0085:SANDIM>2.3.CO;2
65. Evans JJ, Klesius PH, Pasnik DJ, Evans JJ, Klesius PH. Development of skeletal deformities in a *Streptococcus agalactiae*-challenged male Nile tilapia (*Oreochromis niloticus*) broodfish and in its offspring. *Bull Eur Ass Fish Pathol*. (2007) 27:151–69.
66. Cuesta A, Esteban MA. Pathologies in farmed tilapia and the use of immunostimulants and vaccines to prevent or treat diseases: An overview. In: López-Olmeda JF, Sánchez-Vázquez FJ, Rodrigo Fortes-Silva R, editors. *Biology and Aquaculture of Tilapia*. 1. First ed. Florida, US: CRC Press (2021). p. 119–36.
67. Woo PTK, Bruno DW. *Diseases and Disorders of Finfish in Cage Culture*. 2 ed. Wallingford, UK: Cab International (2014). 355 p.
68. Chitmanat C, Lebel P, Whangchai N, Promya J, Lebel L. *Tilapia Diseases and Management in River-Based Cage Aquaculture in Northern Thailand*. Myanmar: Programme inland component – MYSAP (2016) 9–16 p.
69. Ellis T, Oidtmann B, St-Hilaire S, Turnbull J, North B, Mac-Intyre C, et al. Fin erosion in farmed fish. In: Branson EJ, editor. *Fish welfare*. Oxford, UK: Blackwell Publishing Ltd (2008). p. 121–49.
70. Ibrahim T. Diseases of Nile tilapia with special emphasis on water pollution. *J Environ Sci Technol*. (2019) 13:29–56. doi: 10.3923/jest.2020.29.56
71. Nguyen VV, Dong HT, Senapin S, Kayasamruaj P, Pirarat N, Rung-ruangkijrai T, et al. Synergistic infection of *Ichthyophthirius multifiliis* and *Francisella noatunensis* subsp. *orientalis* in hybrid red tilapia (*Oreochromis* sp). *Microb Pathog*. (2020) 147:1–18. doi: 10.1016/j.micpath.2020.104369
72. Ali MS, Griffiths D, Turner WA. *Practical Training Manual: Tilapia Breeding and All-Male Fry Production*. Myanmar: MYSAP Inland (2020). 28 p.
73. Mashaii N, Rajabipour F. A review on the breeding of Nile tilapia, *Oreochromis niloticus* in brackish water hatchery, Iran. *Int J Environ Sci Educ*. (2022) 2:29–45. doi: 10.52547/injoar.2.2.29
74. Puttaraksar N. *GIFT Technology Manual: An Aid to Tilapia Selective Breeding*. Penang, Malaysia: WorldFish Center (2004). 56 p.
75. Fessehay Y, Bovenhuis H, Rezk MA, Crooijmans R, van Arendonk JAM, Komen H. Effects of relatedness and inbreeding on reproductive success of Nile tilapia (*Oreochromis niloticus*). *Aquac*. (2009) 294:180–6. doi: 10.1016/j.aquaculture.2009.06.001
76. Fujimura K, Okada N. Development of the embryo, larva and early juvenile of Nile tilapia *Oreochromis niloticus* (Pisces: Cichlidae). Developmental staging system. *Dev Growth Differ*. (2007) 49:301–24. doi: 10.1111/j.1440-169X.2007.00926.x
77. Bobe J, Labbé C. Egg and sperm quality in fish. *Gen Comp Endocrinol*. (2010) 165:535–48. doi: 10.1016/j.ygcen.2009.02.011
78. Guan B, Cai Y, Zhou Y, Zhao Z, Wang Y, Zhang D, et al. Pathogen identification, risk factor and preventive measure of a columnaris disease outbreak in Tilapia (*Oreochromis niloticus*) eggs and larvae from a tilapia hatchery. *Aquac*. (2022) 561:738718. doi: 10.1016/j.aquaculture.2022.738718
79. Senapin S, Dong HT, Meemetta W, Siriphongphaew A, Charoensapri W, Santimanawong W, et al. Hahella chejuensis is the etiological agent of a novel red egg disease in tilapia (*Oreochromis* spp) hatcheries in Thailand. *Aquac*. (2016) 454:1–7. doi: 10.1016/j.aquaculture.2015.12.013
80. Aich N, Paul A, Choudhury TG, Saha H. Tilapia Lake Virus (TiLV) disease: current status of understanding. *Aquac Fish*. (2022) 7:7–17. doi: 10.1016/j.aaf.2021.04.007
81. El-Greisy ZAE-B, Ahmed NAM. Effect of prolonged ammonia toxicity on fertilized eggs, hatchability and size of newly hatched larvae of Nile tilapia, *Oreochromis niloticus*. *Egypt J Aquat Res*. (2016) 42:215–22. doi: 10.1016/j.ejar.2016.04.001
82. Machimbiri VI, Jansen MD, Senapin S, Khunrae P, Rattanarojpong T, Dong HT. Viral infections in tilapines: more than just tilapia lake virus. *Aquac*. (2019) 503:508–18. doi: 10.1016/j.aquaculture.2019.01.036



83. Fariedadh F, Widodo MS, Nuswantoro S, Rizal MK. Hatching of Nile tilapia (*Oreochromis niloticus*) egg in the hatching medium use salinity media and bromelain enzyme. In: *IOP Conference Series: Earth and Environmental Science*, Vol. 718. Surabaya (2020). doi: 10.1088/1755-1315/718/1/012027
84. Passos Neto OP, Santos ABD, Mota S. Synthetic and natural hormones impact the zootechnical and morphological characteristics of Nile tilapia (*Oreochromis niloticus*). *Eng Sanit Ambient.* (2022) 27:325–33. doi: 10.1590/s1413-415220210098
85. Wahbi OM, Sangak Y. Enhancement of reproductive performance of Nile tilapia *Oreochromis niloticus* using phytochemical *Spirulina platensis*. *J Biol Sci.* (2017) 17:305–11. doi: 10.3923/jbs.2017.305.311
86. Fridman S, Bron JE, Rana KJ. Ontogenic changes in the osmoregulatory capacity of the Nile tilapia *Oreochromis niloticus* and implications for aquaculture. *Aquac.* (2012) 356–7:243–9. doi: 10.1016/j.aquaculture.2012.05.010
87. Stien LH, Bracke MBM, Folkedal O, Nilsson J, Oppedal F, Torgersen T, et al. Salmon Welfare Index Model (SWIM 10): a semantic model for overall welfare assessment of caged Atlantic salmon: Review of the selected welfare indicators and model presentation. *Rev Aquac.* (2013) 5:33–57. doi: 10.1111/j.1753-5131.2012.01083.x
88. El-Sayed A-FM, Kawanna M. Effects of photoperiod on the performance of farmed Nile tilapia *Oreochromis niloticus*: I. growth, feed utilization efficiency and survival of fry and fingerlings. *Aquac.* (2004) 231:393–402. doi: 10.1016/j.aquaculture.2003.11.012
89. Pandit NP, Wagle R, Ranjan R. Alternative artificial incubation system for intensive fry production of Nile tilapia (*Oreochromis niloticus*). *Int J Fish Aquat Stud.* (2017) 5:425–9.
90. Al Hafedh YS, Siddiqui AQ, Al-Saiady MY. Effects of dietary protein levels on gonad maturation, size and age at first maturity, fecundity and growth of Nile tilapia. *Aquaculture International.* (1999) 7:319–32. doi: 10.1023/A:1009276911360
91. Mabroke RS, Tahoun AM, Suloma A, El-Haroun ER. Evaluation of meat and bone meal and mono-sodium phosphate as supplemental dietary phosphorus sources for broodstock Nile Tilapia (*Oreochromis niloticus*) under the conditions of hapa-in-pond system. *Turk J Fish Aquat Sci.* (2013) 13:02. doi: 10.4194/1303-2712-V13\_1\_02
92. Suloma A, Tahoun A-A, Mabroke RS. Development of brood-stock diets for Nile tilapia under hapa-in-pond hatchery system; optimal dietary vitamin C level for the optimum reproductive performance and fry survival. *J Aquac Res Dev.* (2017) S2:010. doi: 10.4172/2155-9546.S2-010
93. Abdel-Tawwab M, Shukry M, Farrag FA, El-Shafai NM, Dawood MAO, Abdel-Latif HMR. Dietary sodium butyrate nanoparticles enhanced growth, digestive enzyme activities, intestinal histomorphometry, and transcription of growth-related genes in Nile tilapia juveniles. *Aquac.* (2021) 536:736–467. doi: 10.1016/j.aquaculture.2021.736467
94. Ng WK, Romano N, A. review of the nutrition and feeding management of farmed tilapia throughout the culture cycle. *Rev Aquac.* (2013) 5:220–54. doi: 10.1111/RAQ.12014
95. Sweilum MA, Abdella MM, Salah El-Din SA. Effect of dietary protein-energy levels and fish initial sizes on growth rate, development and production of Nile tilapia, *Oreochromis niloticus* L. *Aquac Res.* (2005) 36:1414–21. doi: 10.1111/j.1365-2109.2005.01362.x
96. NRC. *Nutrient Requirements of Fish and Shrimp*. Washington, DC: National Academies Press (2011). 392 p.
97. Gomes VDS, Silva JHV, Cavalcanti CR, Filho JJ, Almeida JLS, Amâncio ALL, et al. Avanços do uso de enzimas na nutrição de tilápias. *Visão Acadêmica.* (2018) 19:57380. doi: 10.5380/acd.v19i1.57380
98. Li C, Zhao Y, Wang Y, Li L, Yang X, Chen S, et al. Microbial community changes induced by *Pediococcus pentosaceus* improve the physicochemical properties and safety in fermented tilapia sausage. *Food Res Int.* (2021) 147:110476. doi: 10.1016/j.foodres.2021.110476
99. Kubitz F. *Nutrição e alimentação dos peixes cultivados*. São Paulo, Brazil: Acqua Supre (1999). 123 p.
100. Xie S, Jokumsen A. Replacement of fish meal by potato protein concentrate in diets for rainbow trout, *Oncorhynchus mykiss* (Walbaum): growth, feed utilization and body composition. *Aquac Nutr.* (1997) 3:65–9. doi: 10.1046/j.1365-2095.1997.00074.x
101. Clark AE, Watanabe WO, Olla BL, Wicklund RI. Growth, feed conversion and protein utilization of Florida red tilapia fed isocaloric diets with different protein levels in seawater pools. *Aquac.* (1990) 88:75–85. doi: 10.1016/0044-8486(90)90321-D
102. Sanches LEF, Hayashi C. Densidade de estocagem no desempenho de larvas de tilápia-do-Nilo (*Oreochromis niloticus* L.), durante a reversão sexual. *Acta Sci.* (1999) 21:619–25. doi: 10.4025/actascianimsci.v21i0.4299
103. Bombardelli RA, Hayashi C, Natali MRM, Sanches EA, Piana PA. Níveis de energia digestível sobre o desempenho reprodutivo e zootécnico e a deposição de lipídios nos hepatócitos de machos de tilápia-do-nilo. *Rev Bras Zootec.* (2010) 39:941–9. doi: 10.1590/S1516-35982010000500001
104. Chowdhury DK. *Optimal Feeding Rate for Nile Tilapia (*Oreochromis niloticus*)* (PhD thesis). Norwegian: Norwegian University of Life Sciences (2011).
105. Bhujel RC, Little DC, Hossain A. Reproductive performance and the growth of pre-stunted and normal Nile tilapia (*Oreochromis niloticus*) broodfish at varying feeding rates. *Aquac.* (2007) 273:71–9. doi: 10.1016/j.aquaculture.2007.09.022
106. Venturini FP, Vargas Baldi SC, Parisi G, Costa TD, Rucinke DS, Pires Melo M, et al. Effects of different stunning methods on blood markers and enzymatic activity of stress responses of tilapia (*Oreochromis niloticus*). *Ital J Anim Sci.* (2018) 17:1094–8. doi: 10.1080/1828051X.2018.1426396
107. Santiago CB. Nutrition and feeds of Nile tilapia broodstock and fry. In: Fortes RD, Darvin LC, de Guzman DL, editors. *Fish and Crustacean Feeds and Nutrition?: Proceedings of the Seminar-Workshop On Fish And Crustacean Feeds and Nutrition Held on 25-26 February 1985 at UPV, Iloilo City*. Laguna: Philippine Council for Aquatic and Marine Research and Development (1989). p. 40–9.
108. Malik Daudpota A, Abbas G, Bux Kalhor I, Sajjad Shah SA, Kalhor H, Hafeez-ur-Rehman M, et al. Effect of feeding frequency on growth performance, feed utilization and body composition of juvenile Nile Tilapia, *Oreochromis niloticus* (L.) reared in low salinity water. *Pakistan J Zool.* (2016) 48:171–7.
109. Ferdous Z, Nahar N, Hossen MS, Sumi KR, Ali MM. Performance of different feeding frequency on growth indices and survival of monosex tilapia, *Oreochromis niloticus* (Teleostei: Cichlidae) fry. *Int J Fish Aquat Stud.* (2014) 1:80–3.
110. Garcia JA, Villarroel M. Effect of feed type and feeding frequency on macrophage functions in tilapia (*Oreochromis niloticus* L.). *Fish Shellfish Immunol.* (2009) 27:325–9. doi: 10.1016/j.fsi.2009.05.018
111. Riche M, Haley D, Oetker M, Garbrecht S, Garling D. Effect of feeding frequency on gastric evacuation and the return of appetite in tilapia *Oreochromis niloticus* (L.). *Aquac.* (2004) 234:657–73. doi: 10.1016/j.aquaculture.2003.12.012
112. Morgan IJ, Metcalfe NB. Deferred costs of compensatory growth after autumnal food shortage in juvenile salmon. *Proc R Soc B: Biol Sci.* (2001) 268:295–301. doi: 10.1098/rspb.2000.1365
113. Jobling M, Johansen S, Foshaug H, Burkow I, Jørgensen E. Lipid dynamics in anadromous Arctic charr, *Salvelinus alpinus* (L.): seasonal variations in lipid storage depots and lipid class composition. *Fish Physiol Biochem.* (1998) 18:225–40. doi: 10.1023/A:1007747201521
114. Lovell T. *Nutrition and Feeding of Fish*. Alabama, US: Springer New York, NY (1989). 262 p.
115. Tacon AGJ, Metian M. Feed Matters: Satisfying the Feed Demand of Aquaculture. *Rev Fish Sci Aquac.* (2015) 23:1–10. doi: 10.1080/23308249.2014.987209
116. Borges AM. *Criação de tilápias*. Sete NMDc, editor. Brasília, DF: Emater (2009). 44 p.
117. Marking LL, Meyer FP. Are better anesthetics needed in fisheries? *Fisheries.* (1985) 10:2–5. doi: 10.1577/1548-8446(1985)010<0002:abanif>2.0.co;2
118. Ross LG, Ross B. *Anaesthetic and Sedative Techniques for*. 3 ed. Oxford, UK: Blackwell (2008). 240 p.
119. Noble C, Gismervik K, Iversen MH, Kolarevic J, Nilsson J, Stien LH, et al. *Welfare Indicators for Farmed Atlantic Salmon: Tools for Assessing Fish Welfare*. Norway: Nofima (2018). 352 p.
120. Davis MW. Fish stress and mortality can be predicted using reflex impairment. *Fish Fish.* (2010) 11:1–11. doi: 10.1111/j.1467-2979.2009.00331.x
121. Robb D, Kestin SC. Methods used to kill fish: field observations and literature reviewed. *Anim Welf.* (2002) 11:269–82. doi: 10.1017/S0967278600024854
122. Müller-Graf C, Berthe F, Grudnik T, Peeler E, Afonso A. Risk assessment in fish welfare, applications and limitations. *Fish Physiol Biochem.* (2012) 38:231–41. doi: 10.1007/s10695-011-9520-1
123. IGN. *Fish Welfare in Aquaculture-Problems and Approaches*. Munich, Germany: International Society of Livestock Husbandry (2020). 96 p.
124. Foster C. What a fish knows: the inner lives of our underwater cousins by Jonathan Balcombe. *Common Knowle.* (2018) 24:315–6.
125. Saraiva JL, Arechavala-Lopez P, Castanheira MF, Volstorff J, Heinzpeter Studer B. A global assessment of welfare in farmed fishes: the FishEthoBase. *Fishes.* (2019) 4:30. doi: 10.3390/fishes4020030
126. Flores-García L, Camargo-Castellanos JC, Pascual-Jiménez C, Almazán-Rueda P, Monroy-López JF, Albertos-Alpuche PJ, et al. Welfare indicators in tilapia: an epidemiological approach. *Front Vet Sci.* (2022) 9:882567. doi: 10.3389/fvets.2022.882567
127. Norris IM. *Identification of Potential Welfare Indicators for Commercially Farmed King Salmon (*Hämana*, *Oncorhynchus ishawytscha*): a Scoping Review to Inform the Development of a National Code of Welfare* (dissertation/master's thesis). Manawatu, New Zealand: Massey University (2022).
128. Conte FS. Stress and the welfare of cultured fish. *Appl Anim Behav Sci.* (2004) 86:205–23. doi: 10.1016/j.applanim.2004.02.003
129. Broom DM. Animal welfare: concepts and measurement. *J Anim Sci.* (1991) 69:4167–75. doi: 10.2527/1991.69104167x



130. Whittaker AL, Golder-Dewar B, Triggs JL, Sherwen SL, McLelland DJ. Identification of animal-based welfare indicators in captive reptiles: a delphi consultation survey. *Animals*. (2021) 11:2010. doi: 10.3390/ani11072010
131. Raposo d Magalhães C, Schrama D, Farinha AP, Revets D, Kuehn A, Planchon S, et al. Protein changes as robust signatures of fish chronic stress: a proteomics approach to fish welfare research. *BMC genom*. (2020) 21:1–16. doi: 10.1186/s12864-020-6728-4
132. Afonso LO. Identifying and managing maladaptive physiological responses to aquaculture stressors. In: Tillmann J, Benfey APF, Colin J. Brauner, editor. *Fish Physiology*. Amsterdam: Elsevier (2020). p. 163–91.
133. Gesto M. Characterization of the neuroendocrine stress status as part of the multiparametric assessment of welfare in fish. In: Monzón IF, Fernandes JMO, editors. *Cellular and Molecular Approaches in Fish Biology*. Amsterdam: Elsevier (2022). p. 285–308.
134. Dara M. *Fish Welfare in Aquaculture: From Physiology to Molecular Activities and New Tools For Study Innovative Diets, Social and Spatial Stress* (PhD thesis). Palermo, Italy: Università degli Studi di Palermo (2022).
135. Swirplies F, Wuertz S, Baßmann B, Orban A, Schäfer N, Brunner RM, et al. Identification of molecular stress indicators in pikeperch *Sander lucioperca* correlating with rising water temperatures. *Aquac*. (2019) 501:260–71. doi: 10.1016/j.aquaculture.2018.11.043
136. Sejian V, Lakritz J, Ezeji T, Lal R. Assessment methods and indicators of animal welfare. *Asian J Anim Vet Adv*. (2011) 6:301–15. doi: 10.3923/ajava.2011.301.315
137. Wemelsfelder F, Mullan S. Applying ethological and health indicators to practical animal welfare assessment. *Rev Sci Tech*. (2014) 33:111–20. doi: 10.20506/rst.33.1.2259
138. Rousing T, Bonde M, Sørensen JT. Aggregating welfare indicators into an operational welfare assessment system: a bottom-up approach. *Acta Agric Scand A Anim Sci*. (2001) 51:53–7. doi: 10.1080/090647001300004790
139. Saberion M, Gholizadeh A, Cisar P, Pautsina A, Urban J. Application of machine vision systems in aquaculture with emphasis on fish: state-of-the-art and key issues. *Rev Aquac*. (2017) 9:369–87. doi: 10.1111/raq.12143
140. Pettersen JM, Bracke MBM, Midtlyng PJ, Folkedal O, Stien LH, Steffenak H, et al. Salmon welfare index model 20: An extended model for overall welfare assessment of caged Atlantic salmon, based on a review of selected welfare indicators and intended for fish health professionals. *Rev Aquac*. (2014) 6:162–79. doi: 10.1111/raq.12039
141. Arechavala-Lopez P, Cabrera-Álvarez MJ, Maia CM, Saraiva JL. Environmental enrichment in fish aquaculture: a review of fundamental and practical aspects. *Rev Aquac*. (2022) 14:704–28. doi: 10.1111/raq.12620
142. Bui S, Oppedal F, Sievers M, Dempster T. Behaviour in the toolbox to outsmart parasites and improve fish welfare in aquaculture. *Rev Aquac*. (2019) 11:168–86. doi: 10.1111/raq.12232
143. Carbonara P, Alfonso S, Gai F, Gasco L, Palmegiano G, Spedicato MT, et al. Moderate stocking density does not influence the behavioural and physiological responses of rainbow trout (*Oncorhynchus mykiss*) in organic aquaculture. *Aquac Res*. (2020) 51:3007–16. doi: 10.1111/are.14640
144. Martos-Sitcha JA, Prunet P, Mancera JM, Magnoni LJ. Welfare and stressors in fish: challenges facing aquaculture. *Front Physiol*. (2020) 11:199. doi: 10.3389/fphys.2020.00162
145. Saraiva JL, Rachinas-Lopes P, Arechavala-Lopez P. Finding the “golden stocking density”: a balance between fish welfare and farmers’ perspectives. *Front Vet Sci*. (2022) 9:930221. doi: 10.3389/fvets.2022.930221
146. MacKinnon B, Debnath PP, Bondad-Reantaso MG, Fridman S, Bin H, Nekouei O. Improving tilapia biosecurity through a value chain approach. *Rev Aquac*. (2023) 15:57–91. doi: 10.1111/raq.12776
147. Shlapobersky M, Sinyakov MS, Katzenellenbogen M, Sarid R, Don J, Avtalion RR. Viral encephalitis of tilapia larvae: primary characterization of a novel herpes-like virus. *Virology*. (2010) 399:239–47. doi: 10.1016/j.virol.2010.01.001
148. Bhujel R, Perera A. Shading of breeding hapas enhances reproductive performance of Nile tilapia (*Oreochromis niloticus*) and seed output. *J Aqua Trop*. (2018) 32:187–97.
149. Dayrit GB, Vera Cruz EM, Rodkhum C, Mabrok M, Ponza P, Santos MD. Potential influence of shading in freshwater ponds on the water quality parameters and the hematological and biochemical profiles of Nile tilapia (*Oreochromis niloticus* Linnaeus, 1758). *Fishes*. (2023) 8:322. doi: 10.3390/fishes806322
150. Saraiva JL, Nogueirinha M, Teodósio R, Aragão C, Engrola S, Arechavala-Lopez P. The effect of tank cover on welfare of farmed Nile tilapia. *Appl Anim Behav Sci*. (2021) 241:105396. doi: 10.1016/j.applanim.2021.105396
151. Napoli MAS. (2015) *Consumo de água na industrialização da Tilápia: estudo de caso do oeste do Paraná* (dissertation/master's thesis). Paraná, Brazil: Universidade Estadual do Oeste do Paraná.
152. Coelho M, Pedrazzani A, Quintiliano M, Bolfe F, Molento C. Fish slaughter practices in Brazilian aquaculture and their consequences for animal welfare. *Anim Welf*. (2022) 31:187–92. doi: 10.7120/09627286.31.2.003
153. Hossain MD, Jónsson Á. (2013) *Effect of Combined Blast and Contact (cbc) Cooling and Gutting on the Quality of Tilapia (Oreochromis niloticus) during Chilled Storage* (dissertation/master's thesis). Reykjavik, Iceland: United Nations University Fisheries Training Programme.
154. Eltholth M, Fornace K, Grace D, Rushton J, Häslar B. Characterisation of production, marketing and consumption patterns of farmed tilapia in the Nile Delta of Egypt. *Food Policy*. (2015) 51:131–43. doi: 10.1016/j.foodpol.2015.01.002
155. Lines JA, Spence J. Humane harvesting and slaughter of farmed fish. *Rev Sci Tech*. (2014) 33:255–64. doi: 10.20506/rst.33.1.2284
156. NFDB. Guidelines for responsible farming of tilapia in India. In: *Department of Animal Husbandry DF, Ministry of Agricultura and Farmers Welfare, Govt. of India*. New Delhi, India: National Fisheries Development Board (2015). p. 16.
157. Nasr-Allah A, Dickson M, Al-Kenawy D, Ahmed Ibrahim N, Ali S, Karisa H. *Better Management Practices For Tilapia Culture in Egypt*. Penang, Malaysia: CGIAR Research Program on Fish Agri-Food Systems (2021). 30 p.
158. Gonçalves-De-Freitas E, Bolognesi MC, Gauy ACDS, Brandão ML, Giaquinto PC, Fernandes-Castilho M. Social behavior and welfare in Nile Tilapia. *Fishes*. (2019) 4:1–14. doi: 10.3390/fishes4020023
159. Toguyeni A, Fauconneau B, Boujard T, Fostier A, Kuhn ER, Mol KA, et al. Feeding behaviour and food utilisation in tilapia, *Oreochromis niloticus*: effect of sex ratio and relationship with the endocrine status. *Physiol Behav*. (1997) 62:273–9. doi: 10.1016/S0031-9384(97)00114-5
160. El-Sayed A-FM, Kawanna M. Effects of photoperiod on growth and spawning efficiency of Nile tilapia (*Oreochromis niloticus* L) broodstock in a recycling system. *Aquac Res*. (2007) 38:1242–7. doi: 10.1111/j.1365-2109.2007.01690.x
161. Campos-Mendoza A, McAndrew BJ, Coward K, Bromage N. Reproductive response of Nile tilapia (*Oreochromis niloticus*) to photoperiodic manipulation; effects on spawning periodicity, fecundity and egg size. *Aquac*. (2004) 231:299–314. doi: 10.1016/j.aquaculture.2003.10.023
162. Ramos J, Balasch JC, Tort L. About welfare and stress in the early stages of fish. *Front Vet Sci*. (2021) 8:634434. doi: 10.3389/fvets.2021.634434
163. Labbé C, Robles V, Herraez MP. Epigenetics in fish gametes and early embryo. *Aquac*. (2017) 472:93–106. doi: 10.1016/j.aquaculture.2016.07.026
164. Lopez-Luna J, Al-Jubouri Q, Al-Nuaimy W, Sneddon LU. Reduction in activity by noxious chemical stimulation is ameliorated by immersion in analgesic drugs in zebrafish. *J Exp Biol*. (2017) 220:1451–8. doi: 10.1242/jeb.146969
165. Wilson C. Aspects of larval rearing. *ILAR Journal*. (2012) 53:169–78. doi: 10.1093/ilar.53.2.169
166. Bordeleau X, Pardo S, Chaput G, April J, Dempson B, Robertson M, et al. Spatio-temporal trends in the importance of iteroparity across Atlantic salmon populations of the northwest Atlantic. *ICES J Mar Sci*. (2020) 77:326–44. doi: 10.1093/icesjms/fsz188
167. Cuevas-Rodríguez BL, García-Ulloa M, Hernández-Llamas A, Racotta I, Valdez-González FJ, Polanco-Torres A, et al. Evaluating quality of Nile tilapia (*Oreochromis niloticus*) eggs and juveniles from different commercial hatcheries. *Lat Am J Aquat Res*. (2017) 45:213–7. doi: 10.3856/vol45-issue1-fulltext-23
168. Birch J. Animal sentience and the precautionary principle. *Anim Sentience*. (2017) 16:1200. doi: 10.51291/2377-7478.1200
169. Engle CR, van Senten J, Clark C, Boldt N. Has the regulatory compliance burden reduced competitiveness of the US tilapia industry? *Fishes*. (2023) 8:151. doi: 10.3390/fishes8030151
170. Ellis T, Berrill I, Lines J, Turnbull JF, Knowles TG. Mortality and fish welfare. *Fish Physiol Biochem*. (2012) 38:189–99. doi: 10.1007/s10695-011-9547-3
171. Tallentire CW, Edwards SA, Van Limbergen T, Kyriazakis I. The challenge of incorporating animal welfare in a social life cycle assessment model of European chicken production. *Int J Life Cycle Assess*. (2019) 24:1093–104. doi: 10.1007/s11367-018-1565-2
172. Folkedal O, Pettersen J, Bracke M, Stien L, Nilsson J, Martins C, et al. On-farm evaluation of the Salmon Welfare Index Model (SWIM 10): theoretical and practical considerations. *Anim Welf*. (2016) 25:135–49. doi: 10.7120/09627286.25.1.135
173. Bartlett H, Balmford A, Holmes MA, Wood JL. Advancing the quantitative characterization of farm animal welfare. *Proc R Soc B: Biol Sci*. (2023) 290:20230120. doi: 10.1098/rspb.2023.0120
174. Clark B, Stewart GB, Panzone LA, Kyriazakis I, Frewer LJ. A systematic review of public attitudes, perceptions and behaviours towards production diseases associated with farm animal welfare. *J Agric Environ Ethics*. (2016) 29:455–78. doi: 10.1007/s10806-016-9615-x
175. Peng W, Robinson BE, Zheng H, Li C, Wang F, Li R. The limits of livelihood diversification and sustainable household well-being, evidence from China. *Environ Dev*. (2022) 43:100736. doi: 10.1016/j.envdev.2022.100736



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# Qualitative Behavioural Assessment as a welfare indicator for farmed Atlantic salmon (*Salmo salar*) in response to a stressful challenge

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Animal welfare assessments have struggled to investigate the emotional states of animals while focusing solely on available empirical evidence. Qualitative Behavioural Assessment (QBA) may provide insights into an animal's subjective experiences without compromising scientific rigor. Rather than assessing explicit, physical behaviours (i.e., what animals are doing, such as swimming or feeding), QBA describes and quantifies the overall expressive manner in which animals execute those behaviours (i.e., how relaxed or agitated they appear). While QBA has been successfully applied to scientific welfare assessments in a variety of species, its application within aquaculture remains largely unexplored. This study aimed to assess QBA's effectiveness in capturing changes in the emotional behaviour of Atlantic salmon following exposure to a stressful challenge. Nine tanks of juvenile Atlantic salmon were video-recorded every morning for 15 min over a 7-day period, in the middle of which a stressful challenge (intrusive sampling) was conducted on the salmon. The resultant 1-min, 63 video clips were then semi-randomised to avoid predictability and treatment bias for QBA scorers. Twelve salmon-industry professionals generated a list of 16 qualitative descriptors (e.g., relaxed, agitated, stressed) after viewing unrelated video-recordings depicting varying expressive characteristics of salmon in different contexts. A different group of 5 observers, with varied experience of salmon farming, subsequently scored the 16 descriptors for each clip using a Visual Analogue Scale (VAS). Principal Components Analysis (correlation matrix, no rotation) was used to identify perceived patterns of expressive characteristics across the video-clips, which revealed 4 dimensions explaining 74.5% of the variation between clips. PC1, ranging from 'relaxed/content/positive active' to 'unsettled/stressed/spooked/skittish' explained the highest percentage of variation (37%). QBA scores for video-clips on PC1, PC2, and PC4 achieved good inter- and intra-observer reliability. Linear Mixed Effects Models, controlled for observer variation in PC1 scores, showed a significant difference between PC1 scores before and after sampling ( $p = 0.03$ ), with salmon being perceived as more stressed afterwards. PC1 scores also correlated positively with darting behaviours ( $r = 0.42$ ,  $p < 0.001$ ). These results are the first to report QBA's sensitivity to changes in expressive characteristics of salmon following a putatively stressful challenge, demonstrating QBA's potential as a welfare indicator within aquaculture.

## KEYWORDS

emotional state, aquaculture, positive welfare, behavioural analysis, qualitative behaviour assessment

## 1. Introduction

Animal welfare science has faced the challenge of addressing all aspects of welfare without compromising objectivity and the need for empirical evidence. Physical health has long been recognised as an essential component of animal welfare (1–3). However, a widely held perspective now is that animal welfare is ultimately a state that is perceived by the animal itself, and we should therefore also include concerns for the animal's mental well-being (4–6). There is thus a growing demand that welfare assessments, including those for fish, adopt a more holistic approach that places additional focus on monitoring the animal's positive experiences (2, 6–11). Welfare appraisals that adopt this integrated approach, however, inevitably enter the murky waters that are an animal's subjective experiences (12, 13). Despite decades of research trying to resolve this issue, the only progress thus far has been reaching a consensus that there is no single “measure” that can adequately cover what welfare entails (13–16). This dilemma has resulted in the mental well-being of fish often being overlooked in welfare assessments (17).

In 2018, Atlantic salmon accounted for 4.5% of global aquaculture production by tonnage (18). In 2021, production of Scottish Atlantic salmon reached an all-time high of 205,393 tonnes, with more than 50 million smolts transferred to sea in the same year (19). Total tonnage of Scottish farmed salmon, relative to the number of employees on-site, has increased 11-fold within seawater and 6-fold within freshwater between 1985 and 2016 (20). This increase in the numbers of fish relative to farm staff, unavoidably reduces the time available for monitoring the salmon. There is also mounting scientific evidence supporting the sentience of fish (21–24). A UK National survey, involving 1963 members of the public, found that 77% agreed or strongly agreed that fish can feel pain, and 80% agreed that this should therefore be of concern (25). Considering the scale of this industry, there is a clear ethical and economic incentive to develop welfare indicators that are not only practical, but attempt to include aspects of mental well-being (both positive and negative) in their assessment.

To achieve such an assessment, a framework was proposed in which welfare assessments are viewed in the context of a simple question: “Is the animal healthy, and does it have what it wants?” (13). Answering the second, difficult part of this question (i.e., delving into an animal's subjective experiences) may require accepting two arguments. Firstly, that consciousness still presents an impasse for scientific study (3). Secondly, given that animals cannot express their desires/needs in human language, behavioural analysis may provide some of the best insights into what they “want” (12). Behaviours exhibited by an animal are, in essence, the final product of all its own decision-making processes (1, 26). They are the “final common path,” as described by Sherrington (27), or, in Charles Darwin's words, the “ultimate phenotype” and “expression of the emotions” (27). Behavioural analysis provides a number of additional advantages over physiological/morphological measures in welfare assessments. Such analyses are frequently non-intrusive (the animal is unaware it is being

assessed), and are often quick to observe (1, 28, 29). Behaviour is also gaining recognition as a general, pre-clinical ‘early warning system’ for issues that may be emerging within the stock (1, 28, 30–32).

Observant farmers are capable of detecting changes in the demeanour and behaviour of their animals (33). Such knowledge, typically gained through years of experience, enables farmers to detect subtle shifts in how animals express themselves when issues arise, even though the exact nature of the problem may remain unclear (34). Qualitative Behavioural Assessment (QBA) is a behavioural assessment tool that benefits from this approach, with its reproducibility and validity demonstrated in previous research (35). QBA is an integrative assessment of the “whole-animal,” where observations are made on the animal's dynamic body language (including their appearance, behaviour, and interaction with others and the surrounding environment) as an indicator of its welfare state (36–39). Different aspects of this body language are summarised through a number of ‘descriptors’ (or terms) such as: relaxed, inquisitive, agitated, or stressed (17, 36, 40). Such terms focus not so much on what an animal does (e.g., feeding), as on how it does this; the expressive characteristics shown in the way it moves (40). Descriptions of such characteristics typically have an emotional connotation, indicating how the animal is experiencing the situation it is in, and QBA is therefore hypothesized to be able to contribute valuable information on an animal's emotional state to studies of animal welfare. Interest in the study of fish emotion has grown in recent years. A variety of reviews conclude that fishes are neuro-physiologically and behaviourally similar enough to mammals to warrant assuming a comparable emotional range in fishes, both in terms of negative and positive emotions (6, 41, 42). Indeed, fish are found to be so intelligent that they are frequently used as models for the study of cognition in mammals (43). There is thus a need for developing and testing methods able to address emotional behaviour in fish, and QBA has been recognised as one such potentially promising method (11).

Previous studies for various livestock species have validated the use of QBA against other welfare indicators, and demonstrated high degrees of inter-observer reliability between observers (33, 44). Additionally, QBA allows for simple, time-efficient, and non-intrusive assessments of an animal's well-being (35, 39). QBA is also the only measure currently included in the EU Welfare Quality® welfare assessment protocols to assess positive emotional states in cattle, pigs, and poultry (45, 46). To date, however, the only QBA study to be applied to fish examined solely the inter/intra-observer reliability and QBA's association with ethograms of salmon behaviour, without the inclusion of any treatments (17). No studies have yet examined fish exposed to stressors, or compared QBA scores in this context to other welfare indicators. Comparing QBA scores against other welfare indicators for salmon may help to further explore what potential role QBA may have as a welfare assessment tool. Darting represents a behavioural response previously recorded in fear-conditioning studies of fish, and is commonly associated with predator avoidance (47–50).

It is considered a stress response which, when increasing in frequency/intensity, may indicate impaired welfare (47, 48, 51). Feed intake is also generally considered a reliable indicator within health and welfare assessments of farmed fish (52). A loss in appetite is potentially a sign of impaired welfare (28, 53). The main aim of this study was therefore to examine QBA's ability to detect differences in the expressive characteristics of Atlantic salmon after exposure to a stressful challenge (i.e., an intrusive sampling event). In addition, this study also aimed to compare these QBA scores against other welfare indicators for salmon; their daily feed intake (as a proxy for appetite) and darting behaviours (i.e., sudden, rapid movements of the salmon).

## 2. Materials and methods

### 2.1. Ethical review

Ethical approval for the recording of salmon and QBA work was obtained from the University of Stirling's Animal Welfare & Ethical Review Body (Approval reference no. 2022-6783-5196).

### 2.2. Experimental set-up

#### 2.2.1. Animals

The juvenile Atlantic salmon used in this study were transferred on November 16<sup>th</sup>, 2021, from the Niall Bromage Freshwater Research Unit (NBFRU), Denny, to the Marine Environmental Research Laboratory (MERL) in Campbeltown, Argyll and Bute, Scotland. The salmon were around 14 months of age, and weighed on average 285–360 grams. There were ~80 smolts in each tank at the start of the recording, with an average stocking density of ~34 kg/m<sup>3</sup>.

#### 2.2.2. Husbandry

The salmon were housed in a total of 9 identical flow-through tanks (1.4 m diameter, 750 L volume). Seawater was filtered through a Lacron sand filter (4×100 micron bag filters) before flowing into the tanks to minimise turbidity. Feed was formulated to satisfy the nutrient requirements for Atlantic salmon (54) and contained 46% protein and 24% fat. Automatic belt feeders provided formulated diet salmon pelleted dry feed to all tanks every 20 min between 05:00–09:00 and 16:30–23:30. Dirty water and uneaten feed were flushed out of the tanks through standpipes daily, between 09:00 and 09:15. Any mortalities found during this period were immediately removed. Lights were turned on at exactly 10:30 am each morning.

#### 2.2.3. Treatments (including stressful challenge)

Video clips for this study were recorded around a stressful challenge, conducted on February 18<sup>th</sup>, 2022. This stressful challenge involved a sampling event which was carried out for another study on these salmon. This required capturing, anaesthetising, and handling each of the salmon out of water for measuring their weight, length, and condition factor. While feed withdrawal was also required 24 h before sampling could be carried out, the recording schedule was designed on the assumption that the main disturbances (i.e., stressful challenge) to the salmon would occur largely as a result of this sampling event. For the purposes of the study that involved the sampling event, a subset of the salmon that were sampled were then

euthanised. Fish were euthanised with anaesthetic overdose of Tricaine Methanesulfonate (MS-222) and pithing (Schedule 1) in order to obtain their hepatosomatic index. Following the sampling event, there were approximately 50 salmon left in each tank, with an average stocking density of ~21 kg/m<sup>3</sup>.

#### 2.2.4. Camera and tanks set-up

Cameras were installed in the tanks to record video clips for the QBA and behavioural assessments. To do this, every morning at 9 am, GoPro Hero9 Black® cameras were installed at 1 m depth using a fixed metal pole, which was positioned flush against the inside of each tank to ensure the same angle and field of view (FOV) for recordings. This was carried out 90 min before lights went on to allow time for salmon to habituate to the cameras. These cameras were also installed each morning for 2 days before recording commenced to allow the salmon to further habituate to these novel objects. To minimise any additional disturbances, cameras were turned on before being submerged with recording controlled remotely through the GoPro Quik® mobile application. Connectivity from mobile phone to each underwater camera was achieved through the use of coaxial cables taped to each device. Coaxial cables conduct electrical signals (including Wi-Fi) through an insulated shield, extending network connections to a submerged device (e.g., camera). Recordings for each tank were taken on a strict daily schedule, after lights went on, to ensure consistency. A minimum of 15 min were recorded for each tank once lights went on. All personnel on-site strictly avoided carrying out any procedures around the tanks during filming.

#### 2.2.5. Recording schedule

A 7-day period of video recording was scheduled to gather footage for all behavioural analysis (i.e., QBA and darting behaviours), with the stressful challenge (i.e., sampling) conducted during the middle of this period. Sampling was carried out on all 9 tanks of salmon on February 18<sup>th</sup>, 2022. To obtain a 'baseline' and account for any potential day to day variation in behaviour, 3 consecutive days were recorded before the stressful challenge occurred. A further 3 consecutive 'post-sampling' days were required for recording the salmon's recovery from this stressful challenge. Figure 1 provides a summary of the recording schedule. The ability for QBA to reflect any impacts on salmon behavioural expressions, as a result of these disturbances, could then be assessed from these recordings (Section 2.3.1.1 outlines how the video clips were prepared for QBA).

### 2.3. Qualitative Behavioural Assessment

The QBA process consisted of two main stages. Stage 1 involved 12 observers in the generation of the QBA terms for describing the salmon's expressive characteristics and stage 2 involved 5 different observers scoring the QBA terms for each of the video clips.

#### 2.3.1. Stage 1 – term generation

Twelve professionals employed in the Scottish salmon farming industry were recruited for the term generation stage, which involved two separate meetings. All participants had at least 1 year of experience working directly with farmed salmon, with a number of participants in senior/management roles. During term generation, various video clips were used which were taken from



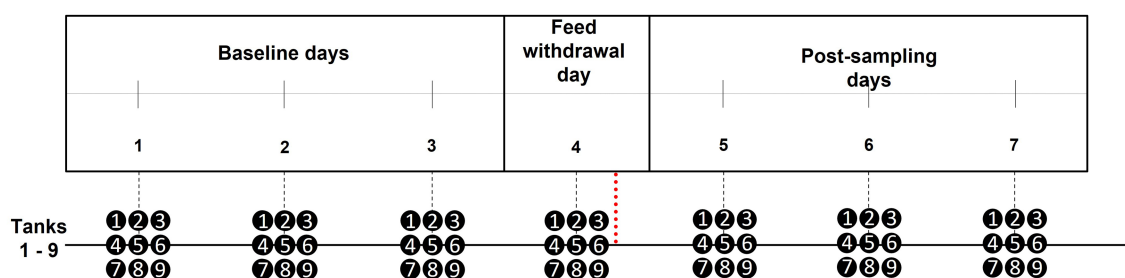


FIGURE 1

Recording schedule and timeline for experiment. Black dots represent each time a tank was recorded for the day, and the dashed red line (after day 4) illustrates when the stressful challenge (sampling event) occurred.

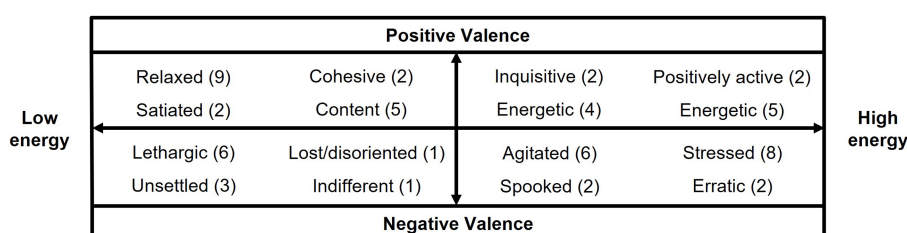


FIGURE 2

Final list of QBA terms generated from stage 1. Valence (positive/negative) and energy (high/low) were used to help describe and discuss terms across the 4 quadrants. Numbers in brackets indicate the total number of participants who brought each term to the initial meeting.

different farm sites under different contexts (e.g., during the middle of the day or during feeding, after treatments/transportation etc.). In this study, we define ‘expressive characteristics’ as the extent to which qualitative characteristics of salmon behaviour (e.g., relaxed, purposeful, lethargic, agitated) are expressed. The video clips were selected to represent all 4 aspects, or ‘quadrants’, of behavioural expression (high to low energy, positive to negative valence) as outlined by Mendl et al. (26) (see Figure 2).

Before terms were generated by participants, the theory and practice of QBA was explained to them and they were provided with guidance on how to generate appropriate terms. To avoid bias, examples of terms from terrestrial farming systems were used. After the first meeting, participants were asked to individually watch the video clips in advance of the second meeting and generate their own personal list of terms. During the second meeting the participants discussed these terms, including how they should be divided between the 4 quadrants of behavioural expression (high to low energy, positive to negative valence). Participants were then asked to select a maximum of 20 terms which were balanced across the 4 quadrants, and best described the range of salmon behavioural expression. By the end of the meeting, the group had agreed on 16 terms. These included the terms “diving deep” and “flighty,” which were excluded by the experimenters from the final list used in the second stage. QBA requires terms that convey some aspect of emotional behaviour and the term “deep diving” did not. Other terms (e.g., spooked, erratic, unsettled, agitated) already covered aspects of the term “flighty.” The final QBA term list therefore had 16 terms, with 4 in each quadrant of behavioural expression (Figure 2). These terms were then used in the QBA scoring stage.

### 2.3.1.1. Video preparation before stage 2

For use in the QBA scoring stage, shorter video clips were extracted from each of the 63, 15-min videos. These clips were the first full minute that the salmon remained clearly in view, starting from 30 s after the lights were turned on. This excluded the initial “noise” from the salmon’s startle responses to the lights. Video clips were first randomised with respect to their chronological order and their occurrence before or after the sampling treatment. To facilitate observer concentration and motivation, they were then arranged so that clips showing contrasting expressive characteristics (e.g., primarily high energy, negative valence vs. low energy, positive valence) were distributed evenly throughout the scoring sessions. Unknown to observers, 4 of the original 63 video clips were duplicated to allow for an assessment of intra-observer reliability (the degree to which participants showed agreement within their own scoring sessions). This resulted in a total of 67 video clips being scored by each observer.

### 2.3.2. Stage 2 – QBA training and scoring session

Scoring sessions for the QBA were carried out with a new group of 5 observers. These 5 observers consisted of 2 Post-doctoral fish welfare researchers from the University of Stirling, and 3 industry professionals all with higher degree education and between 3 and 20 years of aquaculture industry experience. These observers consequently had a varied level of experience in working with and observing salmon. All but 1 observer had hands-on experience in salmon husbandry in a commercial setting.

Observers were given online training in QBA. A brief introduction was given on the principles of QBA and the general purpose of this study (i.e., exploring the use of QBA within fish). Observers were kept

blind to treatment (i.e., the stressful sampling challenge), and were instead only informed about the general context behind the video clips (location of filming, number of tanks and days involved in the recording). It was explained to observers that qualitative descriptors overlap and complement each other in characterising expressive patterns and are not mutually exclusive in the way ethogram categories are. To capture subtle differences when scoring it is important to consider the meaning of each individual descriptor in its own right. Associations between different terms are complex; fish could for example appear stressed and spooked but not too agitated, or a bit unsettled and agitated but not too stressed. To support such use of the descriptors it is important that everyone's understanding of the terms is aligned as much as possible. To this end, an open discussion of the meaning of terms was conducted, aided by a sheet with brief characterisations of each term created by the experimenters (Table 1). The meaning of each term was discussed and adjusted where required, and observers were invited to raise any questions about terms which required clarification. General instructions were given on how to assess whole animal expressivity and how to use the Visual Analogue Scales (VAS) to score the prevalence of each term within a video clip. A VAS is a measurement instrument that allows for the scoring of characteristics (such as those of behavioural expressions) that are believed to range across a continuum of values (55). Observers were reminded that terms must be scored independently from each other, so that in situations where there were contrasting expressive characteristics among different salmon (e.g., some appearing agitated and others relaxed), those contrasting terms could both receive high scores for the same clip.

All QBA scoring was carried out on scoring sheets developed on SurveyMonkey®. For each term, a horizontal line with a 100-step scale was presented as a VAS, along which a single mark could be made. The distance from the left end of the scale would correspond to the participant's assessment of the intensity for each term observed. The left end of the scale represented complete absence of an expressive characteristic described by a term, whereas the right end represented the maximum expression for the term (e.g., the salmon could not be more erratic). While scoring, observers would not be aware of any quantitative values associated with the VASs. They were encouraged to use the entire scale when judging the intensity of each expressive characteristic. Video clips were labelled according to their order in the scoring sheets and transferred electronically to the group. Due to the large number of clips, observers were instructed to avoid scoring them all in a single session, but also to carry out their scoring sessions with minimal delay between each other (i.e., within the same week) to minimise potential variation introduced by scoring on different days.

## 2.4. Additional welfare measures – feed intake and darting events

### 2.4.1. Feed intake

Feed input and feed waste were recorded for each tank daily alongside the 7 days of QBA recordings. The experimental feeds were given by auto-feeders (Arvo-tec TD2000) twice a day from 05:00–09:00 and 16:30–23:30 with uneaten feed collected to measure daily feed intake and calculate apparent feed intake through standpipes

TABLE 1 QBA term list with term characterisations.

Term	Term characterisation
Relaxed	Salmon are moving at ease, free from tension or agitation. Does not just apply to resting - animals can be relaxed also while active.
Agitated	Salmon are restless, excessively moving around, over-responding to unexpected stimuli.
Inquisitive	Salmon show an interest/curiosity towards their surroundings - explorative, investigating features (novel or familiar) of their environment.
Unsettled	Salmon are uncertain, ill at ease, twitchy, vigilant.
Cohesive	Salmon are moving together in synchrony/unison; the shoal appears to behave as one organism.
Spooked/skittish	Salmon are easily scared en-masse (even by small disturbances), abruptly changing behaviour/direction of travel, and avoiding rather than investigating.
Positive active	Salmon are absorbed in activity in a relaxed way, interacting in a positive manner with their environment.
Indifferent	Salmon are unfocused, moving around without much engagement, lacklustre. Not dull or lethargic.
Purposeful	Salmon are self-motivated, focused, determined. Carrying out their actions without hesitation.
Erratic	Salmon movements are un-coordinated, randomly (over)reacting, disorganised. Erratic is more a more vigorous expression than being unsettled.
Energetic	Salmon move in a vigorous, lively way; appearing bright & animated.
Lost/disoriented	Salmon do not know what to do, appearing aimless, confused, worried, and searching on their own.
Satiated	Salmon appear to be satisfied in their physical needs.
Lethargic	Salmon are dull, morose, unresponsive, slow-moving and without any vigour. Looking unwell.
Stressed	Salmon are troubled, tense, not behaving normally. Something is not right.
Content	Salmon are healthy, calm, satisfied, looking well. At ease but could lack positive engagement/purpose.

Terms are listed in the order in which they were presented for scoring to observers.

daily, between 09:00 and 09:15. Recovered feeds were placed in a 300 µm mesh sieve and rinsed in fresh water in order to remove salt and faeces. The leaching of the feeds was calculated through a nutrient dissolution factor. Briefly, 10 g of each feed was incubated in system water in duplicates for 6 h before drying (24 h, 110°C) and weighing. Feed waste daily collected was weighed wet and converted to dry weight using the nutrient dissolution factor.

### 2.4.2. Darting behaviour

For the purpose of this study, darting behaviours were defined as a “rapid, burst of movement clearly distinct from the salmon's regular swimming behaviours; this includes sudden changes in direction, acceleration, and/or positioning of the salmon in the tank.” A scan sampling method (56) was created to record ‘darting events’ in the same 63, 1-min video clips used for the QBA. Since any of these darting events would have also been



observable during the QBA, and thus potentially affected the scoring of certain QBA terms, another second set of video clips were also investigated. This second, separate set involved an additional 63, 1-min video clips that were taken immediately after the QBA clips.

To allow multiple darting events to be recorded in one clip, any darting behaviour must have stopped before the next event could be recorded. The number of salmon involved in each darting event was first recorded and categorised by the proportion to the total number of salmon in the tank (Table 2). Weighted scores were then assigned to each of these categories, relative to their proportions (Table 2). A final score was then calculated for each clip, based on the sum of weighted scores from all darting events recorded. Video playback speed was altered to ensure the number of salmon involved were counted correctly. Where the number of salmon darting was too high to allow for counting, the event was then categorised as involving more than 15% of the fish in the tank.

## 2.5. Statistical analyses

### 2.5.1. Data handling of QBA scores

For each QBA score, the distance of each observers' marks from the zero point of the scales was automatically measured and recorded by SurveyMonkey. The complete dataset of these raw QBA scores were then imported from SurveyMonkey into Microsoft Excel (Version 2,301). Data was organised into a matrix, with QBA terms listed horizontally in the first row and video clip numbers and labels in the first few columns. Unless otherwise stated, all statistical analyses were run in R Studio (version 4.2.2). The threshold of significance for any statistical test was  $p < 0.05$ . For the Linear Mixed Effects Model (LMEM) analyses conducted later in the study, the package "nlme" was applied.

### 2.5.2. Principal Component Analysis

A Principal Component Analysis (PCA) was carried out, using a correlation matrix on the entire dataset of QBA scores to reduce the dimensionality of the QBA terms. PCA allows for the 16 terms scored within each video clip to be summarised by a numerical value for each Principal Component (i.e., the PC "score"). No post-processing step of 'rotation' was carried out, as the only goal of the PCA was to reduce the dimensionality of terms.

The highest positively and negatively loaded terms for each component were identified which, together, represented the larger pattern of expressive characteristics illustrated within each PC. To determine whether PCs were eligible for further analysis, a

combination of criteria was used. Following the "Kaiser criterion," which states that the number of factors to retain should correspond to the number of eigenvalues greater than one, only PCs with eigenvalues  $> 1$  were considered (57). Within each component, there also had to be good inter-observer reliability in the PC scores (Section 2.5.3). There also needed to be a coherent biological interpretation of the terms that had the highest positive and negative loadings within each component. For example, a higher score for PC1 suggested that salmon were more unsettled/stressed, whereas a lower score suggested that salmon were more relaxed/content.

For the complete set of PC scores obtained, Q-Q plots, histogram symmetry, skewness and kurtosis values, sphericity, and Leven's test were inspected to ensure all assumptions required for carrying out further parametric tests were met (including normality of data). The scree plot and proportion of variance for each PC were also used as additional guidance for determining the inclusion of PCs in further analysis.

### 2.5.3. Inter/intra-observer reliability

Kendall's coefficient of concordance (W) was used to calculate the level of agreement between the 5 participants' PC scores in the combined data set, for each of the PCs. Any value of W less than 0.4 was considered to reflect unacceptable inter-observer variability. This analysis was carried out using IBM SPSS Statistics 28 (58). The degree to which observers showed agreement between their scores of the duplicated video clips was, given normal distribution of the scores, determined using Pearson's correlation, performed on each of the relevant PC scores.

### 2.5.4. Comparing pre vs. post disturbances

QBA scores of the salmon before and after the stressful challenge were analysed by applying separate Linear Mixed Effects Models (LMEM) to each of the relevant PCs (PC1, PC2, PC3 and PC4). For each LMEM, the PC score was the dependent variable, 'Pre vs. post disturbance' and 'Observer' were fixed factors, and tank number was a random factor. Before the LMEMs were applied, ANCOVAs were first carried out (with day number as a covariate) to ensure that there were no significant time trends within each subset of days 1–3 and 5–7. Since no additional time trends were present within these subset of days, day number was also included in the LMEMs as a random factor.

Although Kendall's coefficient determines whether there is good agreement between observers for PC1, PC2, and PC4, the actual "treatment" effect of observers still needed to be accounted for, hence the inclusion of 'Observer' as a fixed factor.

### 2.5.5. Comparing feed intake and darting events with QBA scores

For each tank every day, feed intake and two separate sets of darting scores were recorded (2 x separate sets of 63 video clips). Similar LMEMs were applied, with tank and day number as random factors, to first determine whether 'Pre vs. post disturbance' had a significant impact on each of these additional measures. Spearman correlation tests were then carried out to compare feed intake and the two separate sets of darting scores against the corresponding mean PC scores of the 63 clips used in the QBA. Mean PC scores were derived by averaging the PC scores from the 5 observers.

TABLE 2 Categories of darting events by the proportion of salmon from the tank involved, as well as their corresponding weighted scores.

Proportion of salmon in tank involved in each darting event	Weighted score
Less than 4%	1
Less than 8%	2
Less than 15%	3
More than 15%	4

### 3. Results

#### 3.1. Qualitative Behavioural Analysis

##### 3.1.1. Principal Component Analysis

PC1, PC2, PC3, and PC4 had eigen values >1. PC1 explained the greatest percentage of variation at 37%, with the first four components collectively explaining 74.5% of the variation in the data (Table 3).

As outlined in Table 4, PC1 ranged from relaxed/content/positive active to unsettled/stressed/spooked/skittish/agitated. For PC2, the only positively loading term was relaxed, with the main negatively loading terms being energetic/purposeful/inquisitive. Figure 3 illustrates the relationship that the QBA terms have with both PC1 and PC2. For example, a more negative PC1 score indicates salmon that were more relaxed, content, and positive active.

PC1, PC2, and PC4 demonstrated acceptable inter-observer reliability for their PC scores (PC1:  $W = 0.63$ ,  $X^2 = 207.57$ ,  $p < 0.001$ ; PC2:  $W = 0.46$ ,  $X^2 = 152.19$ ,  $p < 0.001$ ; PC4:  $W = 0.56$ ,  $X^2 = 184.94$ ,  $p < 0.001$ ). All four PCs showed acceptable intra-observer reliability between PC scores of video clips that were duplicated (PC1:  $r = 0.716$ ,  $p < 0.001$ ; PC2:  $r = 0.755$ ,  $p < 0.001$ ; PC3:  $r = 0.552$ ,  $p < 0.05$ ; PC4:  $r = 0.581$ ,  $p < 0.01$ ). PC3 had a  $W$  value below 0.4, which was considered unacceptable and therefore not included in further analysis. PC1, PC2, and PC4 were retained for further analysis.

##### 3.1.2. Effect of the stressful challenge (intrusive sampling) on PC scores

There was a significant difference between PC1 scores when comparing days before and after the stressful challenge ( $p = 0.03$ , Figure 4). PC1 scores (averaged between the 5 observers for each video clip) ranged from  $-4.97$  to  $6.04$ . The mean difference between PC1 scores for pre vs. post-disturbance days was  $+0.82$  (Pre =  $-0.239$ , Post =  $0.584$ ). Overall, all five observers scored PC1 higher for post-disturbance days. 7 out of 9 tanks received a higher average PC1 score for post-disturbance days. Figure 5 illustrates the comparative likelihood of a PC1 score being higher or lower for video clips that were recorded either before or after the sampling event. No significant differences were found for PC2 and PC4 scores ( $p > 0.05$ ). For PC1, PC2, and PC4, there was a significant effect for observers as a fixed effect ( $p < 0.001$ ).

#### 3.2. Feed intake, darting behaviours, and their association with QBA

A significant difference was found in the feed intake of salmon from tanks before and after the stressful challenge ( $p = 0.02$ ). Mean daily feed intakes were  $166$  g for pre-disturbance (SEM =  $4.72$ ) and  $79$  g for post-disturbance (SEM =  $7.5$ ), resulting in an average  $87$  g

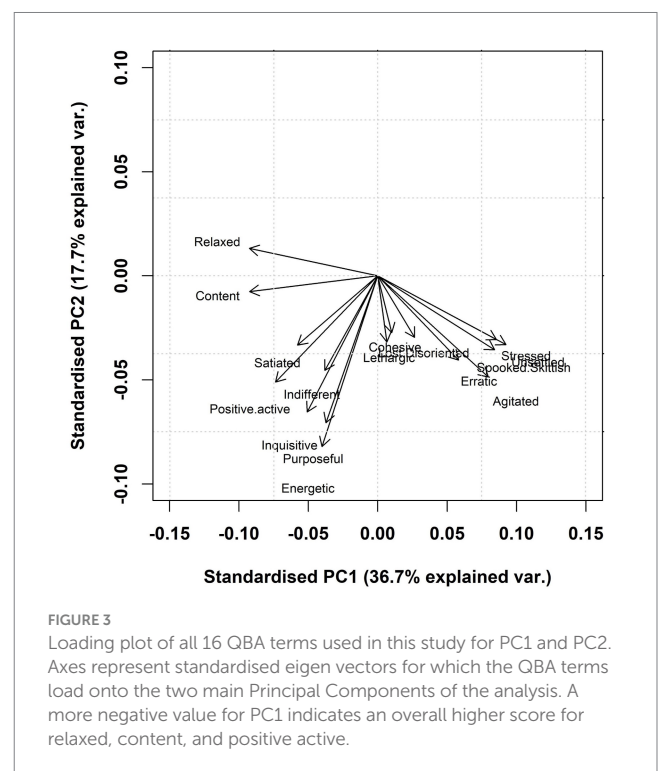
TABLE 3 Eigen analysis of PC1, PC2, PC3, and PC4.

Value	PC1	PC2	PC3	PC4
Eigen value	5.88	2.82	1.95	1.27
% of variation explained	36.7%	17.7%	12.2%	7.9%
Cumulative %	36.7%	54.4%	66.6%	74.5%

TABLE 4 QBA term loading values for each principal component.

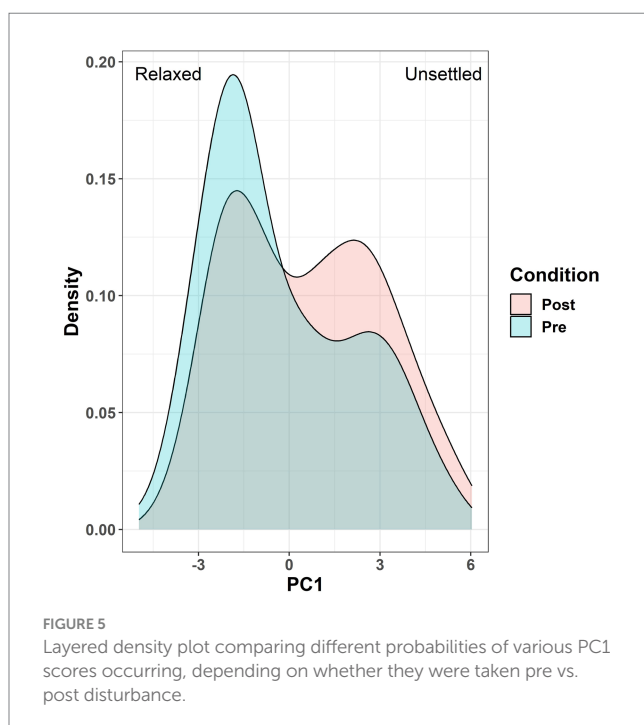
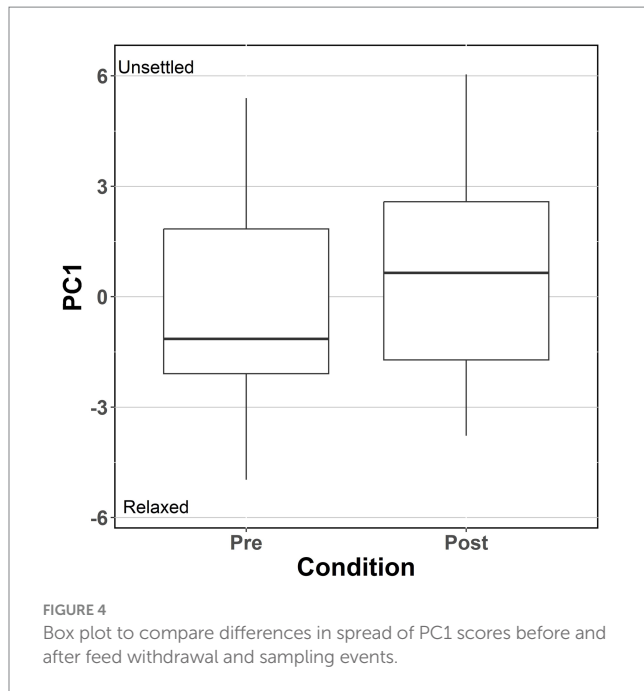
Term	PC1	PC2	PC3	PC4
Relaxed	<b>-0.359</b>	<b>0.074</b>	-0.003	0.073
Agitated	<b>0.309</b>	-0.272	-0.090	-0.109
Inquisitive	-0.197	<b>-0.366</b>	-0.151	0.185
Unsettled	<b>0.358</b>	-0.185	-0.073	-0.049
Cohesive	0.039	-0.153	0.297	<b>-0.491</b>
Spooked/skittish	<b>0.327</b>	-0.199	-0.024	-0.112
Positive active	<b>-0.286</b>	-0.286	<b>-0.168</b>	0.110
Indifferent	-0.148	-0.254	<b>0.477</b>	-0.067
Purposeful	-0.145	<b>-0.395</b>	<b>-0.284</b>	-0.104
Erratic	0.226	-0.226	-0.038	<b>0.431</b>
Energetic	-0.156	<b>-0.459</b>	<b>-0.202</b>	-0.029
Lost/disoriented	0.103	-0.165	<b>0.356</b>	<b>0.585</b>
Satiated	-0.224	-0.186	0.232	<b>-0.290</b>
Lethargic	0.025	-0.178	<b>0.560</b>	0.066
Stressed	<b>0.332</b>	-0.171	-0.056	-0.211
Content	<b>-0.358</b>	-0.042	0.009	-0.050

The highest negatively and positively loaded terms for each PC are in bold.



reduction in daily feed intake post-disturbance. However, there was no significant association found between mean PC1 scores and feed intake ( $r = -0.19$ ,  $p > 0.05$ ).

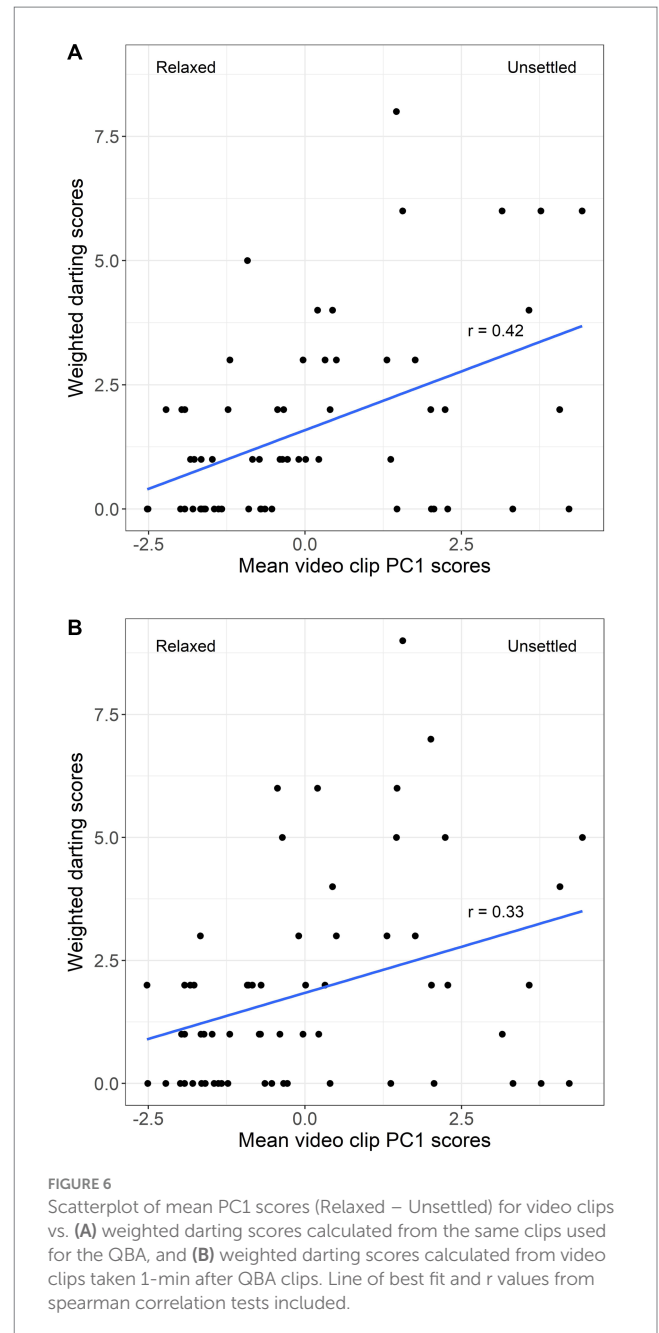
No significant difference was found between darting scores before and after the stressful challenge, in either set of 63 video clips used (same clips as QBA:  $p > 0.05$ ; 1-min post QBA clips:  $p > 0.05$ ). However, PC1 scores showed a moderate positive correlation with the darting scores taken from either set of video clips (same clips as QBA:



$r = 0.42$ ,  $p < 0.001$ ; 1-min post QBA clips:  $r = 0.33$ ,  $p < 0.01$ , see Figure 6).

## 4. Discussion

Integrating indicators of the emotional state of animals within welfare assessments has previously proven to be problematic for many reasons. This study's aim was to determine QBA's ability to detect the effects of a stressful event on Atlantic salmon. We applied QBA to



quantify and evaluate the expressive characteristics of Atlantic salmon before and after exposure to a putatively stressful challenge. While feed withdrawal was required before sampling could be carried out, the sampling event was the focal point as the experimental treatment of this study. The process of capturing, anaesthetising, and handling salmon out of water for sampling has been described as intrusive, stressful, and detrimental for welfare (59–61). Previous studies that have assessed how salmon recover from stressful events (e.g., handling/anaesthesia/invasive sampling) often monitored the recovery over a 24–72 h period (59, 62, 63). Thus, a 3-day period for both the baseline and 'recovery' stage was considered to be sufficient for the purpose of this study.

There was acceptable agreement between the five observers in this study, who were blind to the treatment and had varied

experience in monitoring fish behaviour/welfare. There was one main dimension of QBA that proved effective in capturing changes in the emotional state of the salmon within this study; relaxed/content/positive active – unsettled/stressed/spooked/skittish/agitated (PC1). PC1 explained the largest proportion of variation in expressive characteristics of the salmon (36.7%). There were significant differences between PC1 scores before and after the stressful challenge (sampling), with salmon being scored as more unsettled/stressed/spooked/skittish/agitated after sampling. This reflected a shift from low energy, positive valence to high energy, negative valence after sampling, a contrast that was consistently recorded by all observers and in the majority of tanks. In addition, the single recording that was perceived the most “positively” (i.e., the most relaxed/content/positive active) was taken before any potential impacts from sampling had occurred. These results are in agreement with numerous papers that have previously used QBA to assess the emotional state of terrestrial farmed animals (e.g., for cattle, horses, pigs, and hens), with PC1 typically being characterised by terms such as relaxed and content vs. agitated (64–68). Furthermore, these past studies have used similar descriptors to describe the other main terms used in PC1 for this study; unsettled (uneasy), stressed (nervous), spooked/skittish (scared/fearful/nervous), and stressed (tense).

With lights being switched on at precisely 10:30am every morning, this was considered a routine event that could be methodically recorded and expected to help stimulate activity in the fish. This would potentially maximise what expressive characteristics could be captured without causing additional stress to the salmon. The initial 30 s were cut out to exclude the salmon's startle responses to the lights, which may have otherwise drowned out any potential differences reflected by the QBA scores.

The LMEM determined that there was significant variation between observers in the mean scores they attributed to the 63 video clips on each PC. This suggests that observers may have been interpreting and using the ranges within the VASs differently, while still agreeing on the direction in which the scores should change from one video to another. Such an occurrence is not uncommon when multiple individuals use the same continuous scales (69). In most QBA studies, the directionality of scores, as indicated by Kendall's W, is taken as the most important indicator for inter-observer agreement (17, 44, 70). However, crucial to the aims of this study, the observer effect was accounted for by the LMEM when analysing the treatment effect, and thus a significant difference between PC1 scores was found before and after the stressful challenge.

Previous studies have suggested that significant associations between QBA and other welfare measures help support the validity of QBA as a welfare assessment tool (17, 36, 44, 68). However, as noted by (37), the purpose of QBA is to examine subtle expressive aspects of an animal's demeanour in ways that would be otherwise difficult to quantify for other measures of behaviour. It is important to be reminded of the multi-faceted nature of welfare (16, 71, 72), and that QBA should be regarded as a complementary addition to an integrated approach involving various welfare indicators. QBA is thus used with the intention of gaining unique insights into an animal's emotional state in a way that is complementary to other indicators, allowing for a more comprehensive evaluation of animal welfare (17, 37). Welfare assessments should also aim to minimise redundancies

and include measures that are, at least to some degree, independent from each other (73). Feed intake, on average, more than halved following the stressful challenge. Similar reductions in feed intake have been reported in a number of studies exposing fish to stressful challenges (28, 53, 74). As there were significant differences in both PC1 scores and feed intake before and after the stressful challenge, and yet they were not correlated with each other, these results should further support the notion of QBA being a unique welfare assessment tool. In somewhat of a contrast to this, darting scores showed a moderately positive correlation to PC1 scores. Put simply, as the salmon were observed to be more unsettled, stressed, spooked/skittish, and agitated, there was a corresponding increase in the frequency and/or intensity of darting events. However, the darting scores alone showed no treatment effect from the stressful challenge. While these two measures were not entirely independent from one another, QBA was capable of capturing a significant treatment effect when the darting scores could not. This finding highlights the sensitivity of QBA, indicating that the PC1 scores were more capable of capturing the effects of the stressful challenge on the salmon's welfare than the darting scores.

PC2 and PC4 showed acceptable inter-observer reliability, explaining proportions of variation that were comparable to other studies applying QBA to terrestrial animals (36, 44, 75, 76). For PC2, the only positively loading QBA term was relaxed, with the main negatively loading terms being energetic, purposeful, and inquisitive. This meant that PC2 mainly reflected the salmon's degree of relaxation against ‘high energy’; lower PC2 scores reflected more lively, energetic salmon. PC4 was characterised by terms that reflected a shift in how “harmonious” or “consistent” the behaviour of the salmon was as a collective (i.e., cohesive vs. lost/disoriented). PC3 explained one third of the proportion of variation explained by PC1, with poor inter-observer reliability. The terms most heavily loaded for this dimension (indifferent and purposeful) may help partially explain this inconsistency between observers. Such terms could have been more difficult to perceive and assess in salmon, in comparison to the terms used within PC1.

There was no statistically significant difference between the pre- and post- sampling event stages in PC2 or PC4. Sampling was specifically chosen as a presumably intrusive, stressful event, with the intentions of then assessing QBA's ability to detect the putative impacts of such an event on the salmon's emotional state. Considering the terms used to characterise PC2 and PC4, these dimensions may not be so relevant to addressing the effects of stress, but may be very relevant to assess fish welfare in other contexts and treatments. For example, the potential benefits of environmental enrichment or the impacts of transportation/transfer to new enclosures. Considering that the most relevant dimension in the context of this study (PC1) reflects a combined shift in both valence (positive – negative) and energy (low – high), this dimension could be of significant use for on-farm welfare assessments of Atlantic salmon. Additional research is needed to further explore and validate the relevance of other dimensions found in this study (i.e., PC2 and PC4), under different experimental treatments, to expand the potential applications of QBA for salmon welfare assessments.

Integrating QBA into future welfare assessments (for research or farming) will first require appropriate training in the observing, scoring, and understanding of terms involved (70, 77). While this may



require a significant initial investment towards developing the observers' assessment capabilities, doing so will help ensure acceptable inter-observer reliability and, over the long term, help with the integration of a unique and efficient welfare assessment tool (17). Welfare assessments that include QBA have the advantage of evaluating emotional states of the animals, and the consequent monitoring of positively valenced terms (e.g., content, relaxed, inquisitive, cohesive, purposeful, energetic etc.) also allows for the consideration of positive aspects of fish welfare.

The various ways in which sampling can cause stress and impair fish welfare demonstrates another advantage with implementing QBA; as a non-intrusive method of welfare assessment. QBA avoids any negative impacts from its measurement, an issue that is inherent in many animal-based measures. A large proportion of animal-based measures of welfare are also retrospective, only identifying problems long after they have occurred (72). Analyses of behavioural expression could help minimise this delay, perhaps even to the point of providing early warning signs for pre-clinical health issues (13). QBA has for example been used successfully to detect early clinical signs of mastitis in dairy cows (78). Through virtue of being able to assess behavioural expressions through video monitoring, QBA is also capable of being carried out remotely. Considering the remote locations in which these salmon are often kept (79), as well as issues surrounding monitoring when site access is limited, this feature provides a significant advantage. The need for such welfare monitoring tools was highlighted to the Scottish salmon farming sector when farm staff were restricted from accessing their sites during the 2020 COVID-19 pandemic, and in-person audits for welfare certification schemes had to be replaced with virtual assessments for 2 months (80, 81). During a recent industry survey carried out within the salmon farming sector, various professionals employed in the production process ranked the development of remote, non-intrusive welfare indicators as one of the highest research priorities for farmed salmon welfare (32). The effective implementation of QBA on-site would help meet this demand.

## 5. Conclusion

This is the first study to demonstrate QBA's ability to capture changes in the expressive characteristics of Atlantic salmon following exposure to putatively stressful events. Five observers from various professional backgrounds achieved acceptable inter- and intra-observer reliability in 3 dimensions of QBA scores. PC1 showed a significant treatment effect, with salmon becoming more unsettled, stressed, spooked/skittish, and agitated after the stressful challenge. Both PC1 scores and feed intake recorded a significant difference before and after the stressful challenge, but were not correlated to each other. PC1 scores showed a moderate positive correlation with darting scores, however the darting scores did not show a significant treatment effect, indicating the QBA scores to be more sensitive to the stressful challenge. These results support QBA's ability to provide unique insights that are relevant to the evaluation of farmed salmon welfare. Future experiments should explore the other dimensions found within QBA (e.g., PC2 and PC4) under different treatment conditions, and across other species of fish, to further investigate QBA's applicability within aquaculture. The results from this study demonstrate that QBA is a promising welfare indicator that, with

further research, could act as a time-efficient and complimentary tool for on-farm welfare assessments.

## Data availability statement

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

## Author contributions

TW: Conceptualization, Data curation, Formal analysis, Investigation, Methodology, Visualization, Writing – original draft, Writing – review & editing. SR: Conceptualization, Supervision, Validation, Visualization, Writing – review & editing. MB: Methodology, Validation, Writing – review & editing, Data curation, Resources. MH: Conceptualization, Funding acquisition, Methodology, Supervision, Validation, Visualization, Writing – review & editing. SJ: Validation, Writing – review & editing, Conceptualization, Funding acquisition, Methodology, Supervision, Visualization. AD: Validation, Writing – review & editing. FW: Conceptualization, Validation, Writing – review & editing, Methodology. JT: Conceptualization, Funding acquisition, Supervision, Validation, Visualization, Writing – review & editing.

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## Conflict of interest

AD was employed by Aquascot Ltd.

The remaining authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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## References

- Dawkins MS. Using behaviour to assess animal welfare. *Anim Welf.* (2004) 13:S3–7. doi: 10.1017/S0962728600014317
- Franks B, Ewell C, Jacquet J. Animal welfare risks of global aquaculture. *Sci Adv.* (2021) 7:1–8. doi: 10.1126/sciadv.abg0677
- Fraser D, Weary DM, Pajor EA, Milligan BN. A scientific conception of animal welfare that reflects ethical concerns. *Anim Welf.* (1997) 6:187–205. doi: 10.1017/S0962728600019795
- Hemsworth PH, Mellor DJ, Cronin GM, Tilbrook AJ. Scientific assessment of animal welfare. *N Z Vet J.* (2015) 63:24–30. doi: 10.1080/00480169.2014.966167
- Green TC, Mellor DJ. Extending ideas about animal welfare assessment to include “quality of life” and related concepts. *N Z Vet J.* (2011) 59:263–71. doi: 10.1080/00480169.2011.610283
- Fife-Cook I, Franks B. Positive welfare for fishes: rationale and areas for future study. *Aust Fish.* (2019) 4:31. doi: 10.3390/fishes4020031
- Boissy A, Manteuffel G, Jensen MB, Moe RO, Spruijt B, Keeling LJ, et al. Assessment of positive emotions in animals to improve their welfare. *Physiol Behav.* (2007) 92:375–97. doi: 10.1016/j.physbeh.2007.02.003
- Rault J, Waiblinger S, Boivin X, Hemsworth P. The power of a positive human – animal relationship for animal welfare. *Front Vet Sci.* (2020) 7:1–13. doi: 10.3389/fvets.2020.590867
- Mellor DJ. Updating animal welfare thinking: moving beyond the “five freedoms” towards “a life worth living”. *Animals.* (2016) 6:1–20. doi: 10.3390/ani6030021
- Veit W, Browning H. Perspectival pluralism for animal welfare. *Eur J Philos Sci.* (2021) 11:1–14. doi: 10.1007/s13194-020-00322-9
- Browning H. Improving welfare assessment in aquaculture. *Front Vet Sci.* (2023) 10:1–10. doi: 10.3389/fvets.2023.1060720
- Dawkins MS. Animal welfare and the paradox of animal consciousness. *Adv Study Behav.* (2015) 47:5–38. doi: 10.1016/b.s.asb.2014.11.001
- Dawkins MS. Behaviour as a tool in the assessment of animal welfare. *Zoology.* (2003) 106:383–7. doi: 10.1078/0944-2006-00122
- Broom DM. Welfare, stress, and the evolution of feelings. *Adv Study Behav.* (1998) 27:317–403. doi: 10.1016/S0065-3454(08)60369-1
- Mason G, Mendl M. Why is there no simple way of measuring animal welfare? *Anim Welf.* (1993) 2:301–19. doi: 10.1017/S0962728600016092
- Stien LH, Bracke MBM, Folkedal O, Nilsson J, Oppedal F, Torgersen T, et al. Salmon welfare index model (SWIM 1.0): a semantic model for overall welfare assessment of caged Atlantic salmon: review of the selected welfare indicators and model presentation. *Rev Aquac.* (2013) 5:33–57. doi: 10.1111/j.1753-5131.2012.01083.x
- Jarvis S, Ellis MA, Turnbull JF, Rey Planellas S, Wemelsfelder F. Qualitative behavioral assessment in juvenile farmed Atlantic Salmon (*Salmo salar*): potential for on-farm welfare assessment. *Front Vet Sci.* (2021) 8:1–11. doi: 10.3389/fvets.2021.702783
- FAO. *The state of world fisheries and aquaculture - sustainability in action.* Rome: FAO (2020).
- Munro LA. *Scottish fish farm production survey 2021.* (2022). Available at: <https://www.gov.scot/publications/scottish-fish-farm-production-survey-2021/pages/5/>
- Ellis T, Turnbull JF, Knowles TG, Lines JA, Auchterlonie NA. Trends during development of Scottish salmon farming: an example of sustainable intensification? *Aquaculture.* (2016) 458:82–99. doi: 10.1016/j.aquaculture.2016.02.012
- Kristiansen TS, Fernö A, Pavlidis MA, Van de Vis H. The welfare of fish In: *Animal Welfare*, vol. 20. Cham: Springer (2020).
- Chervova LS, Lapshin DN. Behavioral control of the efficiency of pharmacological anesthesia in fish. *J Ichthyol.* (2011) 51:1126–32. doi: 10.1134/S0032945211110026
- Sneddon LU, Braithwaite VA, Gentle MJ. Do fishes have nociceptors? Evidence for the evolution of a vertebrate sensory system. *Proc R Soc B Biol Sci.* (2003) 270:1115–21. doi: 10.1098/rspb.2003.2349
- Brown C. Fish intelligence, sentience and ethics. *Anim Cogn.* (2015) 18:1–17. doi: 10.1007/s10071-014-0761-0
- Rethink Priorities. Rethink priorities: submission of evidence to animal welfare (sentience) Bill (2019). Available at: <https://committees.parliament.uk/writtenevidence/37632/pdf/>
- Mendl M, Burman OHF, Paul ES. An integrative and functional framework for the study of animal emotion and mood. *Proc R Soc B Biol Sci.* (2010) 277:2895–904. doi: 10.1098/rspb.2010.0303
- Darwin C. *The expression of the emotions in man and animals.* London UK: University of Chicago Press (1872).
- Huntingford FA, Adams C, Braithwaite VA, Kadri S, Pottinger TG, Sandoe P, et al. Current issues in fish welfare. *J Fish Biol.* (2006) 68:332–72. doi: 10.1111/j.0022-1112.2006.001046.x
- Martins CIM, Galhardo L, Noble C, Damsgård B, Spedicato MT, Zupa W, et al. Behavioural indicators of welfare in farmed fish. *Fish Physiol Biochem.* (2012) 38:17–41. doi: 10.1007/s10695-011-9518-8
- Duthie CA, Bowen JM, Bell DJ, Miller GA, Mason C, Haskell MJ. Feeding behaviour and activity as early indicators of disease in pre-weaned dairy calves. *Animal.* (2020) 15:100150–7. doi: 10.1016/j.animal.2020.100150
- Oppedal F, Dempster T, Stien LH. Environmental drivers of Atlantic salmon behaviour in sea-cages: a review. *Aquaculture.* (2011) 311:1–18. doi: 10.1016/j.aquaculture.2010.11.020
- Wiese TR, Haskell M, Jarvis S, Rey Planellas S, Turnbull J. Concerns and research priorities for Scottish farmed salmon welfare – an industry perspective. *Aquaculture.* (2023) 566:739235. doi: 10.1016/j.aquaculture.2023.739235
- Fleming PA, Clarke T, Wickham SL, Stockman CA, Barnes AL, Collins T, et al. The contribution of qualitative behavioural assessment to appraisal of livestock welfare. *Anim Prod Sci.* (2016) 56:1569–78. doi: 10.1071/AN15101
- Hoffmann V, Probst K, Christinck A. Farmers and researchers: how can collaborative advantages be created in participatory research and technology development? *Agric Human Values.* (2007) 24:355–68. doi: 10.1007/s10460-007-9072-2
- Browning H. Assessing measures of animal welfare. *Biol Philos.* (2022) 37:1–24. doi: 10.1007/s10539-022-09862-1
- Vasdal G, Muri K, Stubbsøen SM, Moe RO, Kittelsen K. Qualitative behaviour assessment as part of a welfare assessment in flocks of laying hens. *Appl Anim Behav Sci.* (2022) 246:105535. doi: 10.1016/j.applanim.2021.105535
- Wemelsfelder F. How animals communicate quality of life: the qualitative assessment of behaviour. *Anim Welf.* (2007) 16:25–31. doi: 10.1017/S0962728600031699
- Cooper R, Wemelsfelder F. Qualitative behaviour assessment as an indicator of animal emotional welfare in farm assurance. *Livestock.* (2020) 25:180–3. doi: 10.12968/live.2020.25.4.180
- Ellingsen K, Coleman GJ, Lund V, Mejdell CM. Using qualitative behaviour assessment to explore the link between stockperson behaviour and dairy calf behaviour. *Appl Anim Behav Sci.* (2014) 153:10–7. doi: 10.1016/j.applanim.2014.01.011
- Wemelsfelder F, Hunter TEA, Mendl MT, Lawrence AB. Assessing the ‘whole animal’: a free choice profiling approach. *Anim Behav.* (2001) 62:209–20. doi: 10.1006/anbe.2001.1741
- Braithwaite VA, Huntingford FA, Van den Bos R. Variation in emotion and cognition among fishes. *J Agric Environ Ethics.* (2013) 26:7–23. doi: 10.1007/s10806-011-9355-x
- Ferno A, Folkedal O, Nilsson J, Kristiansen TS. Inside the fish brain: cognition, learning, and consciousness In: *The welfare of fish.* eds. C. Phillips and M. C. Gartner Cham: Springer (2020).
- Brown C, Dorey C. Pain and emotion in fishes – fish welfare implications for fisheries and aquaculture. *Anim Stud J.* (2019) 8:175–201. doi: 10.14453/asj.v8i2.12
- Minero M, Dalla E, Dai F, Canali E, Barbieri S, Zanella A, et al. Using qualitative behaviour assessment (QBA) to explore the emotional state of horses and its association with human-animal relationship. *Appl Anim Behav Sci.* (2018) 204:53–9. doi: 10.1016/j.applanim.2018.04.008
- Keeling L, Evans A, Forkman B, Kjaernes U. Welfare quality® principles and criteria In: H Blokhuis, M Miele, I Veissier and B Jones, editors. *Improving farm animal welfare.* Wageningen: Wageningen Academic Publishers (2013)
- Welfare Quality®. *Welfare quality® assessment protocol for poultry (broilers, laying hens).* Lelystad, Netherlands: Welfare Quality® consortium (2009).
- Magurran AE, Pitcher TJ. Foraging, timidity and shoal size in minnows and goldfish. *Behav Ecol Sociobiol [Internet].* (1983) 12:147. Available from:–52. doi: 10.1007/BF00343206
- Ashley PJ, Sneddon LU. Pain and fear in fish In: EJ Branson, editor. *Fish welfare.* Oxford, UK: Blackwell (2008)
- Cantalupo C, Bisazza A, Vallortigara G. Lateralization of predator-evation response in a teleost fish (*Girardinus falcatus*). *Neuropsychologia.* (1995) 33:1637–46. doi: 10.1016/0028-3932(95)00043-7
- Domenici P, Blake RW. The kinematics and performance of fish fast-start swimming. *J Exp Biol.* (1997) 200:1165–78. doi: 10.1242/jeb.200.8.1165
- Nomura M, Sloman KA, von Keyserlingk MAG, Farrell AP. Physiology and behaviour of Atlantic salmon (*Salmo salar*) smolts during commercial land and sea transport. *Physiol Behav.* (2009) 96:233–43. Available from:–. doi: 10.1016/j.physbeh.2008.10.006
- Jobling M, Coves D, Damsgård B, Kristiansen HR, Koskela J, Petursdottir TE, et al. Techniques for measuring feed intake In: D Houlihan, T Boujard and M Jobling, editors. *Food intake in fish.* Oxford: Blackwell (2001).
- Schreck CB, Olla BL, Davis MW. Behavioural responses to stress In: GK Iwama, AD Pickering, JP Sumpter and CB Schreck, editors. *Fish stress and health in aquaculture.* Cambridge: Cambridge University Press (1997). 155.
- National Research Council. *Nutrient requirements of fish and shrimp. Nutrient requirements of fish and shrimp.* Washington, DC: The National Academies Press (2011).
- Gould D. Visual analog scale (VAS) - in depth. *J Clin Nurs.* (2001) 10:697–706. doi: 10.1046/j.1365-2702.2001.00525.x



56. Martin P, Bateson P. *Measuring behaviour - an introductory guide*. (2007). 1–101 Cambridge University Press, Cambridge, United Kingdom.
57. Kaiser HF. The application of electronic computers to factor analysis. *Educ Psychol Meas*. (1960) 20:141–51. doi: 10.1177/001316446002000116
58. IBM Corp. *IBM SPSS statistics for windows*. Armonk, NY: IBM Corp (2021).
59. Djordjevic B, Kristensen T, Øverli O, Rosseland BO, Kiessling A. Effect of nutritional status and sampling intensity on recovery after dorsal aorta cannulation in free-swimming Atlantic salmon (*Salmo salar* L.). *Fish Physiol Biochem*. (2012) 38:259–72. doi: 10.1007/s10695-009-9362-2
60. Santurtun E, Broom DM, Phillips CJC. A review of factors affecting the welfare of Atlantic salmon (*Salmo salar*). *Anim Welf*. (2018) 27:193–204. doi: 10.1017/S09627286.27.3.193
61. Zahl IH, Samuelsen O, Kiessling A. Anaesthesia of farmed fish: implications for welfare. *Fish Physiol Biochem*. (2012) 38:201–18. doi: 10.1007/s10695-011-9565-1
62. Sandodden R, Finstad B, Iversen M. Transport stress in Atlantic salmon (*Salmo salar* L.): anaesthesia and recovery. *Aquac Res*. (2001) 32:87–90. doi: 10.1046/j.1365-2109.2001.00533.x
63. Iversen M, Finstad B, Nilssen KJ. Recovery from loading and transport stress in Atlantic salmon (*Salmo salar* L.) smolts. *Aquaculture*. (1998) 168:387–94. doi: 10.1016/S0044-8486(98)00364-0
64. Sant'Anna AC, Paranhos da Costa MJR. Validity and feasibility of qualitative behavior assessment for the evaluation of Nellore cattle temperament. *Livest Sci*. (2013) 157:254–62. Available from: doi: 10.1016/j.livsci.2013.08.004
65. Fleming PA, Paisley CL, Barnes AL, Wemelsfelder F. Application of qualitative behavioural assessment to horses during an endurance ride. *Appl Anim Behav Sci*. (2013) 144:80–8. doi: 10.1016/j.applanim.2012.12.001
66. Napolitano F, Rosa GD, Grasso F. Qualitative behaviour assessment of dairy buffaloes (*Bubalus bubalis*). *Appl Anim Behav Sci*. (2012) 141:91–100. doi: 10.1016/j.applanim.2012.08.002
67. Rutherford KMD, Donald RD, Lawrence AB, Wemelsfelder F. Qualitative Behavioural assessment of emotionality in pigs. *Appl Anim Behav Sci*. (2012) 139:218–24. doi: 10.1016/j.applanim.2012.04.004
68. Muri K, Stubbsøen SM, Vasdal G, Moe RO, Granquist EG. Associations between qualitative behaviour assessments and measures of leg health, fear and mortality in Norwegian broiler chicken flocks. *Appl Anim Behav Sci*. (2019) 211:47–53. Available from: doi: 10.1016/j.applanim.2018.12.010
69. Bryce D, Bratzke D. Are introspective reaction times affected by the method of time estimation? A comparison of visual analogue scales and reproduction. *Atten Percept Psychophys*. (2015) 77:978–84. doi: 10.3758/s13414-014-0804-2
70. Clarke T, Pluske JR, Fleming PA. Are observer ratings influenced by prescription? A comparison of free choice profiling and fixed list methods of qualitative behavioural assessment. *Appl Anim Behav Sci*. (2016) 177:77–83. Available from: doi: 10.1016/j.applanim.2016.01.022
71. Weary DM, Robbins JA. Understanding the multiple conceptions of animal welfare. *Anim Welf*. (2019) 28:33–40. doi: 10.1017/S09627286.28.1.033
72. Noble C, Gismervik K, Iversen MH, Kolarevic J, Nilsson J, Stien LH, et al. *Welfare indicators for farmed Atlantic Salmon: tools for assessing fish welfare*. Tromsø: Nofima (2018).
73. Botreau R, Veissier I, Butterworth A, Bracke MBM, Keeling LJ. Definition of criteria for overall assessment of animal welfare. *Anim Welf*. (2007) 16:225–8. doi: 10.1017/S0962728600031390
74. Pankhurst NW, Ludke SL, King HR, Peter RE. The relationship between acute stress, food intake, endocrine status and life history stage in juvenile farmed Atlantic salmon, *Salmo salar*. *Aquaculture*. (2008) 275:311–8. doi: 10.1016/j.aquaculture.2008.01.001
75. Fleming PA, Wickham SL, Stockman CA, Verbeek E, Matthews L, Wemelsfelder F. The sensitivity of QBA assessments of sheep behavioural expression to variations in visual or verbal information provided to observers. *Animal*. (2015) 9:878–87. Available from: doi: 10.1017/S1751731114003164
76. Temple D, Manteca X, Velarde A, Dalmau A. Assessment of animal welfare through behavioural parameters in Iberian pigs in intensive and extensive conditions. *Appl Anim Behav Sci*. (2011) 131:29–39. doi: 10.1016/j.applanim.2011.01.013
77. Grosso L, Battini M, Wemelsfelder F, Barbieri S, Minero M, Dalla Costa E, et al. On-farm qualitative behaviour assessment of dairy goats in different housing conditions. *Appl Anim Behav Sci*. (2016) 180:51–7. doi: 10.1016/j.applanim.2016.04.013
78. de Boyer des Roches A, Lussert A, Faure M, Herry V, Rainard P, Durand D, et al. Dairy cows under experimentally-induced *Escherichia coli* mastitis show negative emotional states assessed through qualitative behaviour assessment. *Appl Anim Behav Sci*. (2018) 206:1–11. doi: 10.1016/j.applanim.2018.06.004
79. Natural Scotland. Scotland's aquaculture. Joint Nature Conservation Committee (2016). p. 1. Available at: <http://aquaculture.scotland.gov.uk/map/map.aspx>
80. Murray AG, Ives SC, Smith RJ, Moriarty M. A preliminary assessment of indirect impacts on aquaculture species health and welfare in Scotland during COVID-19 lockdown. *Vet Anim Sci*. (2021) 11:1–5. doi: 10.1016/j.vas.2021.100167
81. FishFarmingExpert. RSPCA assured targets return of farm visits [internet]. *FishFarmingExpert News*. (2020) Available at: <https://www.fishfarmingexpert.com/article/rspca-assured-takes-first-step-towards-normal-assessments/>
82. Sherrington CS. *The integrative action of the nervous system*. New York, USA: Scribner (1906).



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# Farmed fish welfare during slaughter in Italy: survey on stunning and killing methods and indicators of unconsciousness

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Information on slaughter procedures for farmed fish in aquaculture is limited, both in Europe and in Italy, due to a general lack of field data. The aim of this study was to gather information on the procedures used to slaughter fish in Italy and to discuss them considering the WOA and EFSA recommendations on fish welfare. Using a questionnaire survey, data were collected by official veterinarians in 64 slaughtering facilities where 20 different species of fish were slaughtered. The main species slaughtered were rainbow trout (*Oncorhynchus mykiss*; 29/64), followed by European sea bass (*Dicentrarchus labrax*; 21/64), sea bream (*Sparus aurata*; 21/64), Arctic char (*Salvelinus alpinus*; 14/64), European eel (*Anguilla anguilla*; 11/64), sturgeon (*Acipenser* spp; 11/64), common carp (*Cyprinus carpio*; 6/64), and brown trout (*Salmo trutta fario* L.; 5/64). The most applied stunning/killing methods were "asphyxia in ice/thermal shock" and "electric in water bath," followed by "percussion," "asphyxia in air," and "electric dry system." After the application of the method, the assessment of the fish level of unconsciousness was practiced in 72% of the facilities using more than one indicator, with "breathing" and "coordinated movements" the most practiced. The collected data showed a discrepancy between the available recommendations about the welfare of fish at slaughter and what is practiced in many production sites, but for many species precise recommendations are still not available.

## KEYWORDS

aquaculture, fish unconsciousness, thermal shock, electrical stunning, trout, seabream, seabass, sturgeon

## 1. Introduction

Aquaculture production has experienced significant growth over the past two decades (1–3), reaching 87.5 million tons in 2020 at a global level (2). This rapid expansion has led to considerable interest in welfare issues by scientists, consumers and policy makers. The number of scientific studies claiming that fish are sentient animals, able to perceive emotions and thus

of experiencing fear, psychological stress, and pain is increasing, as recently reviewed (4). Consumers demand information about the origins of food products and the conditions under which farmed animals are kept (5). Animalist groups claim more than half of the consumers are aware that fish are capable of experiencing pain and that the current protection of fish welfare is insufficient if compared to the other farmed species (6). In fact, despite that the number of farmed and slaughtered fish for human consumption is largely higher than that of farmed mammals (approximately 2.5-fold) (3, 7), the welfare of farmed fish is inadequately protected by the European legislation (3, 6). Specifically for the protection of animals at the time of killing (EC Reg. 1099/2009) (8) fish are included only within the general framework of basic principles. Only the Article 3 (1) of the general provision is applicable for fish, according to which animals for slaughter must be protected from avoidable pain, distress or suffering. Unlike for other species, the methods for fish stunning and killing are not defined. Actually, in the absence of scientific updates and EU rules, Article 27 (1) allows Member States to maintain or adopt national rules regarding the protection of fish at the time of killing (8). However, only a few Member States have implemented national laws which refer only to a few species (e.g., salmon) (9). In order to cope with these legislative gaps, the Farm to Fork strategy of the European Green Deal commits to review and update European animal welfare legislation (10) and for the slaughter of fish, in particular, it is intended to add more guidelines for the main farmed species and to revise welfare assessment criteria that are often inapplicable in the field (10).

In Europe, approximately 50 species of fish are reared for human consumption purpose. The most farmed species is rainbow trout (*Oncorhynchus mykiss*), followed by European seabass (*Dicentrarchus labrax*), gilthead seabream (*Sparus aurata*), common carp (*Cyprinus carpio*), and atlantic salmon (*Salmo salar*) (2, 11). Currently, farmed fish can be subjected to a range of different stunning/killing methods at slaughter. Many of these methods present relevant welfare problems, since they expose fish to prolonged suffering and pain before death (12–19).

Within such a heterogeneous framework of farmed species and stunning and killing methods, both the World Organization for Animal Health (WOAH) and the European Food Safety Authority (EFSA) have provided recommendations and opinions, in order to safeguard the welfare of fish at the time of killing (12–20). Considering the main farmed species, WOAHS suggested the following stunning/killing methods as humane: percussion stunning for carp and salmonids; spiking or coring for tuna; and electrical stunning for carp, eel, and salmonids (20). According to EFSA, new methods of stunning/killing are needed, since most of those commonly practiced in Europe do not allow many fish species to be humanely slaughtered (12–19).

In order to be considered humane, a method must involve killing the animal in a state of unconsciousness and insensibility. For this reason, major improvements are necessary regarding the evaluation of unconsciousness or death after the application of the method (12). This control is very difficult under field conditions, as demonstrations of stress or pain in fish are not obvious and complicated to detect. In fact, the best method for recognizing fish unconsciousness is the electroencephalogram (EEG) in a laboratory condition, with the observation of visual evoked responses (VERs) (12, 21–24), which is difficult to be used in the field. According to WOAHS and EFSA, spontaneous behavior, responses to stimuli and reflexes, such as the loss of body and respiratory movement

and the loss of the vestibulo-ocular reflex should be assessed. However, this assessment presents several technical and practical problems in the field due to the large number of slaughtered species and the lack of standard procedures to follow. Fish are slaughtered in groups while most of the indicators refer to an individual fish (25). The indicators have only been validated in the laboratory for a few species (12) and due to physiological and morphological differences among species, they are not always applicable in the same way. For the assessment of unconsciousness, a specific and adequate training of the operators, likewise Standard Operating Procedures (SOPs) to clarify how to perform it (e.g., number of fish checked/cage) are required.

Italy plays an important role on the European panorama, representing the third largest country in terms of aquaculture fish production (11) with around 55 thousand tonnes in the last 10 years (11, 26). Italy is the largest European producer of sturgeon (*Acipenser* spp.), the second of rainbow trout and catfish (*Ictalurus punctatus* and *Ameiurus melas*) and the third of European seabass, gilthead seabream, and European eel (*Anguilla anguilla*) (11, 26, 27). The farms currently operating to rear fish for human consumption are 558, mainly located in the north: Veneto (114 farms), Piedmont (70), Friuli-Venezia Giulia (69), Trentino-South Tyrol (59), and Lombardy (56), (28).

Currently, there is very little information on the practices used in Italy to protect the welfare of fish at the time of killing as there is no a specific national legislation for the protection of the fish at farm and at slaughter (9). The only available data is that collected in 2018 by the European Commission about the methods used to kill three farmed species (9). It is also not known if the effectiveness of the methods applied are evaluated, and what indicators are used. There are not national guidelines on fish welfare at slaughter but guidelines for fish welfare at slaughtering were recently issued by associations of fish farmers. However, they are generic, not mandatory and do not provide any SOPs on stunning or on the assessment of fish unconsciousness.

In this context, where the information available on what happens in the field is limited, it is of utmost importance to have data on the different realities present in Europe in order to produce applicable legislation without an excessive economic impact on the producers.

For these reasons, the aim of this study was to provide a detailed presentation of the current situation regarding fish welfare at the time of slaughtering in Europe's third largest aquaculture producer, Italy. Using a questionnaire, we collected information on the fish-slaughter facilities currently operating on the Italian territory and on the practices used at the key moments of slaughter: the methods used to stun and kill fish and the indicators used to assess the unconsciousness. The assessment of the conformity of the procedures reported by the facilities was carried out in accordance with WOAHS guidelines and EFSA recommendations on the protection of fish at slaughter (12–20).

## 2. Materials and methods

### 2.1. Data collection

The data were collected with the support of the Italian National Health System. The Italian Ministry of Health entrusted the Regional Competent Authorities to select an Official Veterinarian from the Regional and Local Health Units (LHU—ASL) responsible for safety and hygiene of food of animal origin for each Region. The selection was

made based on their background and their expertise on slaughter as well as on their knowledge of the reality of aquaculture in their territory. The data collection was made through a questionnaire to fill with the information of the fish-slaughter facilities operating within their district. The completed questionnaires have been gathered in different moments as it has not been easy to register or verify the necessary data in all facilities simultaneously. The first questionnaire was collected in August 2022, while the last one was delivered by March 2023.

Official veterinarians were instructed on how to collect the data by means of an explanatory document prepared and disseminated by the Italian Ministry of Health. Support to the official veterinarians was given by the authors belonging to the CReNBA (Italian Reference Center for Animal Welfare—Istituto Zooprofilattico della Lombardia e dell'Emilia Romagna, Brescia) throughout the entire data collection period in order to clarify any possible doubt regarding how to fill in the questionnaire. The completed questionnaires were sent via email by each Italian region to the Italian Ministry of Health, which subsequently sent the questionnaire to the CReNBA for the analysis.

## 2.2. Development of the questionnaire

The questionnaire was developed by the CReNBA taking into account the structure and technique used in a veterinary questionnaire survey on welfare issues (29); the questions asked in a recent similar survey carried out in Brazil (30); and the generic recommendation of the European Regulation 1099/2009 on the protection of animals at time of killing. The targets of the questionnaire were the Italian Regional Competent Authorities, which sent it to their selected Official Veterinarians.

The questionnaire consists of an Excel document containing six open and closed questions, with the aim of collecting information about fish-slaughter facilities (Table 1). The first two questions related to the identification of the facilities, like their physical location (question 1) and their registration details (question 2). The third question was an open question on the annual processed fish volumes (question 3). This question was fundamental to be able to assess the volumes processed by the surveyed facilities in order to understand their relevance to the overall Italian aquaculture fish production. The following questionnaire question focused on the slaughtered species (question 4), with the possibility to select from seven fish species representing the main species farmed in Italy considering the most recent censuses of FEAP and FAO (26, 43). The last two questions regarded the stunning and/or killing methods (question 4) and indicators of unconsciousness applied (question 5), which had to be reported for each slaughtered species. The list of possible answers was built considering the main methods and indicators reported in literature as indicated in Table 1. The last three questions also allowing the veterinarians to add other possible options (open answer to fill in if necessary) in order to have the most accurate overview of the applied procedures.

## 2.3. Data analysis

The results of the questionnaire survey were subjected to descriptive statistical analysis in order to provide the geographical distribution of the slaughtering facilities, the stunning/killing practices, and the indicators used for the unconsciousness assessment

TABLE 1 Areas and type of information requested in the questionnaire used for data collection in fish-slaughter facilities in Italy.

Questionnaire area	Type of information
Geographic area	Italian region
	Local Competent Authority (veterinary public health system)
Identification of the slaughter facility	Company name
	Address
	Official registration number (approval number)
What is the annual processing volume (tonnes)?	Open Answer
Which species are slaughtered?	Rainbow Trout ( <i>Oncorhynchus mykiss</i> )
	European seabass ( <i>Dicentrarchus labrax</i> )
	Gilthead seabream ( <i>Sparus aurata</i> )
	Arctic char ( <i>Salvelinus alpinus</i> )
	Common carp ( <i>Cyprinus carpio</i> )
	European eel ( <i>Anguilla anguilla</i> )
	Sturgeon ( <i>Acipenser</i> spp.)
	Other species, please specify
What is the stunning/killing method used? (13–21, 31–37)	Percussion
	Electric in water bath
	Electric dry system
	Carbon monoxide (CO)
	Carbon dioxide (CO <sub>2</sub> )
	Asphyxia in ice/Thermal shock
	Asphyxia in air
	Other method, please specify
Which indicators are used to assess fish unconsciousness? (13–20, 24, 32, 38–42)	Breathing
	Eye movements
	Coordinated movement
	Response to stimuli
	Righting ability
	Other indicator, please specify

during slaughter. For a clearer presentation of the results, slaughtered fish species were divided into two groups: main species (i.e., fish species slaughtered in more than five facilities) and minor species (i.e., species slaughtered in less than five facilities).

## 3. Results

Fourteen Italian regions were confirmed to practice fish slaughtering in a total of 67 facilities: Veneto, Piedmont, Apulia, Lombardy, Friuli Venezia Giulia, Tuscany, Sardinia, Marche, Trentino-South Tyrol, Sicily, Liguria, Emilia Romagna, Lazio, and Campania. The Competent Authorities of the remaining six regions communicated the absence of fish-slaughter facilities on their territory.



Out of the 67 facilities, two slaughter facilities of rainbow trout (one in Veneto and one in Lombardy) and one of European seabass and gilthead seabream (in Sicily) were excluded from the analysis because they did not slaughter fish in the period used for collection of data. Therefore, the number of the facilities included for the data analysis on the fish stunning/killing methods in Italy was 64, where a total volume of 22,229 tonnes of fish was processed. The regions of Veneto, Piedmont, Apulia, Lombardy, and Friuli Venezia Giulia were the only ones which reported more than five fish-slaughter facilities in their territory (Figure 1). Marche, Veneto, Tuscany, Lombardy, and Lazio were the regions processing more than 1,500 tonnes of fish each (Table 2).

Fish slaughtering in Italian facilities involved 20 different species, and the 64% of the facilities processed more than one species. The main slaughtered fish species in Italy were rainbow trout (29/64), followed by European seabass and gilthead seabream (21/64), Arctic char (*Salvelinus alpinus*; 14/64), European eel (12/64), sturgeon (11/64), common carp (6/64), and brown trout (*Salmo trutta fario* L.; 5/64). The minor slaughtered fish species: meager (*Argyrosomus regius*), hybrid striped bass common called “persico-spigola” (*Morone chrysops* × *Morone saxatilis*), trout perch (*Micropterus salmoides*), gray mullet (*Mugil cephalus*), catfish (channel catfish—*Ictalurus punctatus* and black bullhead—*Ameiurus melas*), lavaret (*Coregonus lavaretus*), Danube salmon (*Hucho hucho*), pike-perch (*Sander lucioperca*), royal perch (*Perca fluviatilis*), turbot (*Psetta maxima*), sargo (*Diplodus sargus*), tench (*Tinca tinca*), and greater amberjack (*Seriola dumerili*; Table 2).

The majority of facilities (86%) practiced only one method, while a few used to practice more than one (14%). The most common methods applied were “asphyxia in ice/thermal shock” and “electric in water bath,” which were used in 30 and 24 of the facilities respectively,

followed by “percussion” (13), “asphyxia in air” (5), and “electric dry system” (3) as shown in Figure 2. Two facilities did not perform any stunning or killing and sold live carps and eels directly to the consumer.

The different stunning/killing methods and the frequently used indicators to assess the effectiveness of the reported methods for the slaughtered species are shown in Tables 3, 4, respectively.

Looking in detail at the main species, data showed that thermal shock method was used in 95% of the cases for slaughtering of European seabass and gilthead seabream; the electric in water bath method was utilized in more than half of the facilities for rainbow trout, Arctic char, European eel, common carp, and brown trout; percussive method was practiced for sturgeon in more than half of the facilities. Also for minor species, the most used method is thermal shock while the electric in water bath, percussion and asphyxia were less practiced. Overall, the assessment of unconsciousness or death was carried out routinely in 45 of the surveyed facilities (72%). In details, for the main species, more than 90% of the facilities assessed unconsciousness or death, with the exception of European seabass and gilthead seabream which were assessed in only 28% of facilities. For minor species, the assessment was performed in 66% of facilities. The 50% of all facilities assessed the unconsciousness using more than one indicator (Figure 3). “Breathing” and “coordinated movements” were the indicators most frequently used (34 facilities), followed by “response to stimuli” (18), “righting ability” (15), and “eye movements” (14).

“Breathing” and “coordinated movements” were the most frequently assessed indicators for rainbow trout (24/29 and 19/29 respectively), Arctic char (13/14 and 10/14), European seabass and gilthead seabream (4/21 and 6/21), and brown trout (5/5 and 3/5). The indicator “response to external stimuli” is also assessed in about half of the facilities that slaughtered European eel (6/12) and common carp (4/6). For sturgeon “righting ability” is also assessed (6/11). Relating the stunning/killing methods to the assessment of unconsciousness, the electric and percussion methods are those for which the assessment of unconsciousness is practiced in more than 90% of cases in contrast to the ice and air asphyxia methods, for which the assessment is practiced around the 55% of the times (Figure 2). The distribution of the considered indicators used to assess fish unconsciousness differentiated by the methods used, is shown in Figure 4.

## 4. Discussion

### 4.1. General consideration on slaughter procedures

#### 4.1.1. Stunning/killing methods

The 14 Italian regions that participated in the survey and confirmed to practice fish slaughtering correspond to the regions with the highest presence of fish farms for human consumption in Italy (95% of all farms). In particular, the top five regions in terms of number of fish farms are Veneto, Piedmont, Friuli-Venezia Giulia, Trentino-South Tyrol, and Lombardy (28). Therefore, we can consider the collected data a significant representation of the current Italian situation of fish farmed and slaughtered for human

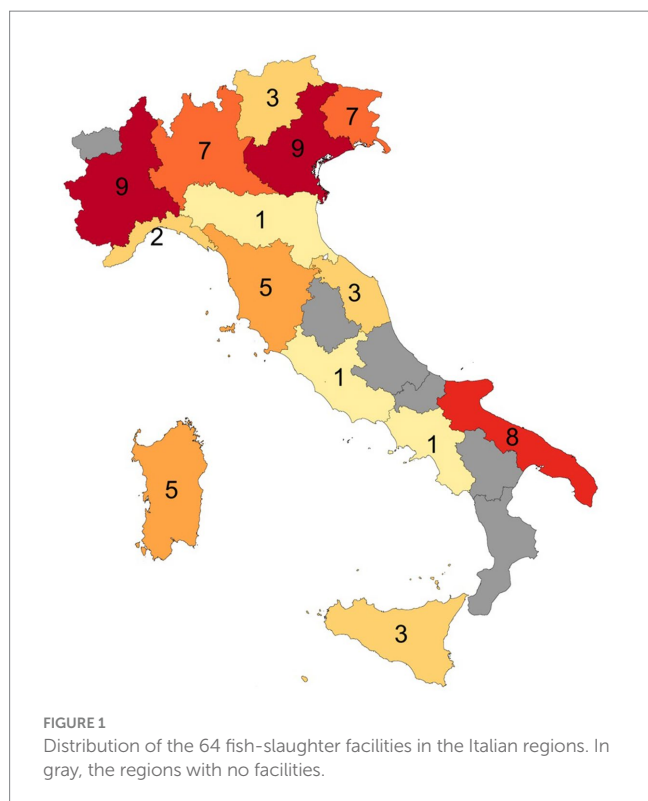


TABLE 2 Distribution of the different fish species slaughtered in the 64 Italian slaughtering facilities surveyed for data collection (n=number of fish-slaughter facilities in the region).

	Italian region														Total (n = 64)
	Veneto (n = 9)	Piedmont (n = 9)	Apulia (n = 8)	Lombardy (n = 7)	Friuli V.G. (n = 7)	Tuscany (n = 5)	Sardinia (n = 5)	Marche (n = 3)	Trentino ST (n = 3)	Sicily (n = 3)	Liguria (n = 2)	E. Romagna (n = 1)	Lazio (n = 1)	Campania (n = 1)	
	5,559 t.	62 t.	1,266 t.	2,114 t.	1,121 t.	2,663 t.	752 t.	6,249 t.	61 t.	55 t.	438 t.	38 t.	1849 t.	2 t.	
Rainbow trout	6	5	-	4	4	1	1	3	3	2	-	-	-	-	29
Seabass-seabream	-	-	7	-	3	3	4	-	-	-	2	-	1	1	21
Arctic char	2	4	-	1	2	-	-	2	3	-	-	-	-	-	14
European Eel	4	3	1	1	-	1	1	-	-	-	-	1	-	-	12
Sturgeon	3	3	-	4	-	-	-	-	-	-	-	1	-	-	11
Common carp	4	1	-	1	-	-	-	-	-	-	-	-	-	-	6
Brown trout	-	-	-	1	1	-	-	2	1	-	-	-	-	-	5
Meager	-	-	2	-	-	-	-	-	-	-	-	-	1	-	4
Trout perch	-	1	-	-	-	-	-	-	-	1	-	1	-	-	3
Hybrid strip. Bass	-	1	-	-	-	-	-	-	-	1	-	1	-	-	3
Royal perch	1	-	-	-	-	-	-	-	-	-	-	1	-	-	2
Gray mullet	-	-	-	-	-	-	2	-	-	-	-	-	-	-	2
Catfish	-	1	-	-	-	-	-	-	-	-	-	1	-	-	2
Lavaret	-	-	-	-	-	-	-	-	1	-	-	-	-	-	1
Danube salmon	-	-	-	-	-	-	-	-	1	-	-	-	-	-	1
Pikeperch	1	-	-	-	-	-	-	-	-	-	-	-	-	-	1
Tench	-	1	-	-	-	-	-	-	-	-	-	-	-	-	1
Greater amb.	-	-	-	-	-	1	-	-	-	-	-	-	-	-	1
Sargo	-	-	-	-	-	-	1	-	-	-	-	-	-	-	1
Turbot	-	-	-	-	-	-	-	-	-	-	-	-	-	1	1

The annual volume of fish processed for human consumption is reported by each region.



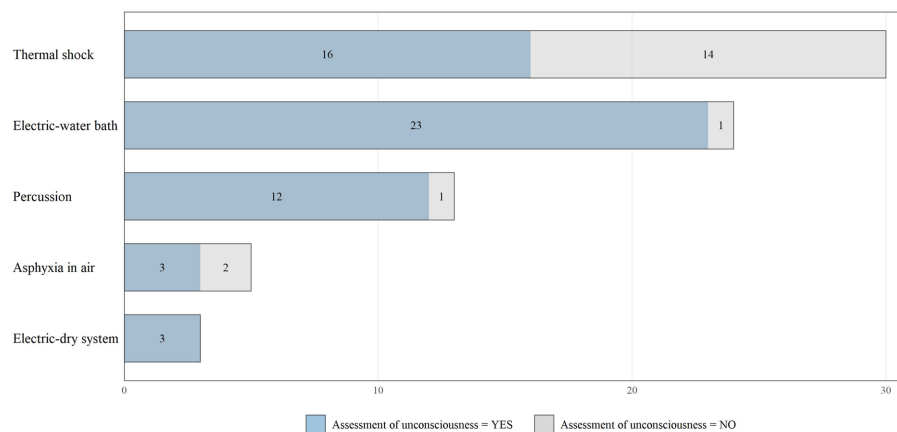


FIGURE 2

Distribution of different stunning/killing methods reported by the 64 fish-slaughter facilities surveyed, in association with the practice of assessing fish unconsciousness.

consumption. As a matter of fact, considering the volumes processed by the 64 facilities, the results show that the data in this survey refer to the 41% of the total volume processed in Italy in 2022 (53,900 t.) (44). Consistent with production data (43), the results of the survey show that rainbow trout is the species most commonly slaughtered, followed by European seabass, gilthead seabream.

The survey also showed that the several facilities practiced slaughtering of more than one species, which could be a critical point for fish welfare, since WOA (20) recommends that stun and slaughter facilities should be designed and constructed for one specific species or group of species in order not to compromise their welfare and not to cause injuries or stress.

A wide variety of methods were used in the surveyed facilities for the slaughtering of rainbow trout, Arctic char and sturgeon, whereas mainly one method was used for European seabass and gilthead seabream. According to the European legislation on the slaughter of farmed species (EC Reg. 1099/2009) and the recommendations of the WOA and EFSA, animals must be subjected to a stunning process before being slaughtered in order to become unconscious and insensible. However, based on the results of the survey, in Italy, more than half of the fish-slaughter facilities (35/64) practiced the asphyxia in air or the thermal shock method. These procedures are considered non-humane methods of slaughtering fish by WOA and EFSA. In fact, the asphyxia in air is considered only as a 'killing' method; the thermal shock does not stun the fish but produces only sedation, leading the conscious animal to death by asphyxia (12, 20). The remaining methods used in the Italian fish-slaughter facilities are generally recommended by the WOA and EFSA as acceptable methods. They are namely the electrical methods and the percussion method, followed by gill-cutting, which, if carried out correctly, can induce a state of instantaneous unconsciousness before slaughtering (12). Specifically, the electrical method induces unconsciousness, but its efficacy can vary considerably according to the parameters set (mainly V, A, Hz and time of application) (12, 20). The details for these parameters were not collected in our study and need further investigations.

#### 4.1.2. Assessment of unconsciousness or death

One of the key aspects of evaluating the welfare of fish slaughtered in the field is to assess their state of unconsciousness or death after the application of the stunning/killing method. In our survey, there were no facilities that reported the assessment of unconsciousness or death by EEC, due to evident practical problems, whereas the majority of the surveyed facilities (72%) used the evaluation of the reflexes, spontaneous behavior or response to external stimuli, as recommended by EFSA and WOA (12, 20). Indeed, there was a big difference in the frequency of this assessment, depending on the fish species slaughtered and method used (Figure 4). The low number of facilities (28%) assessing unconsciousness or death on European seabass and gilthead seabream was actually striking. This data negatively influences the overall result, since unconsciousness was assessed in almost all facilities for all the other main species. Probably, this result is related to the problem of the assessment of the individual fish unconsciousness, which is particularly difficult when the production volumes are high. For this reason, it is recommended to have group indicators available, as suggested by Noble et al. in their handbooks (45, 46).

Most Italian facilities carried out this assessment using the indicators "breathing" and "coordinated movements" which are probably more easily to check compared to eye movement assessment on an individual fish. However, assessment by breathing and body movements can be ineffective or unreliable indicators (47). In fact, as previously reported, asphyxia in ice and incorrectly performed electrical methods are recognized to paralyze fish, thus making it unable to behaviorally express pain (12). There are currently no behavioral indicators that can fully differentiate paralysis from unconsciousness or death (12). In the absence of alternatives, the assessment of a combination of different indicators can improve the evaluation (48, 49), as it occurred in half of the surveyed facilities (Figure 3).

#### 4.2. Species-specific consideration on slaughter procedures

Since one of the objectives of the survey was to relate the collected slaughtered practices with the available species-specific

TABLE 3 Stunning/killing methods and indicators of consciousness used for the main slaughtered fish species in the surveyed fish-slaughter facilities.

	Stunning/killing methods						Indicators of unconsciousness				
	<i>Thermal shock</i>	<i>Electric water bath</i>	<i>Percussion</i>	<i>Asphyxia in air</i>	<i>Electric dry system</i>	<i>Sold alive*</i>	<i>Breathing</i>	<i>Eye movements</i>	<i>Coord. movements</i>	<i>Response to stimuli</i>	<i>Righting ability</i>
Rainbow trout ( <i>n</i> = 29)	7	17	7	4	1	-	24	10	19	10	10
Seabass-seabream ( <i>n</i> = 21)	20	-	-	1	-	-	4	1	6	2	1
Arctic char ( <i>n</i> = 14)	2	10	3	1	-	-	13	6	10	6	7
European eel ( <i>n</i> = 12)	3	8	-	1	-	2	5	1	9	6	2
Sturgeon ( <i>n</i> = 11)	3	3	6	-	2		7	4	6	5	6
Common carp ( <i>n</i> = 6)	2	4	-	1	-	1	5	1	2	4	2
Brown trout ( <i>n</i> = 5)	-	4	1	-	-	-	5	2	3	2	1

\*The practice “sold alive” is also reported.

recommendations (Table 5) among the species slaughtered in Italy, only those covered by these recommendations were taken into account for this purpose, i.e., rainbow trout, common carp, European eel, turbot, European seabass, gilthead seabream and, in some aspects, also the other species belonging to the *Salmonidae* family (Arctic char, brown trout, Danube salmon and lavaret). For the other eleven slaughtered species, the general recommendations on welfare at slaughter of the WOA and EFSA (12, 20) were used.

#### 4.2.1. Rainbow trout

Based on the results of the present survey, trout was the species slaughtered with the highest variety of slaughtering methods in Italian facilities, which is consistent with available data about the other major European producers of trout (France and Denmark) and more generally whole of Europe (9, 16). The most commonly used stunning/killing method was the electric one, like in Denmark (9), which, if carried out correctly, satisfies the recommendations of both WOA and EFSA (16, 20). This result confirms the use of the electric practice on trout in Italy, as previously reported (9). The second most practiced method was percussion, which is the most used method in France (9). If correctly done within 10 s from the moment the fish is pulled out of the water, percussion is also considered a humane method for trout (16). As in other European countries, asphyxia in ice and air were also practiced, although they should be avoided, since they do not induce effective loss of consciousness (9, 12, 20, 50).

The assessment of trout unconsciousness was mainly carried out using “breathing” and “coordinated movements” as recommended by WOA (20). According to EFSA (16), however, these indicators are considered acceptable as indicative of unconsciousness but are not really robust or validated in laboratory conditions. As a matter of fact, although some authors have recently stated that loss of respiratory movement could be related to unconsciousness (24), others found no clear relationship between loss of ventilation and brain failure in rainbow trout under laboratory conditions (47). Thus, we can conclude that the assessment of unconsciousness/death for trout in Italy can be considered to be correctly carried out in the majority of the fish slaughter facilities.

#### 4.2.2. European eel and common carp

Different slaughter methods have been reported for eel in our survey in Italy, as in other European countries (15), with the electric method the most widely used. This method is among those

recommended by WOA (20) for this species and has recently been introduced by the first two major European producers, i.e., the Netherlands and Germany (31). Non-humane procedures such as live evisceration or baths in ammonium salts (15), reported to be carried out in other European countries, were not practiced in Italian slaughtering facilities. Nevertheless, both eel and carp can be sold alive to the consumer in Italy (practice previously reported by EFSA in Europe only for carp) (19) which poses possible risks to the welfare of the fish at the time of killing (19). However, this practice was not widespread in Italy compared to northern European countries (19) and was confined only to two regions (Puglia and Veneto). The electric method was the main method practiced in Italy for carp, as in the top two European carp producers, Germany and Poland (9). This method satisfies WOA recommendations, whereas EFSA highlighted some critical points on the methods and required more investigations in order to be able to express itself more accurately (19). However, recent studies by Daskalova et al. (51) have pointed out that the electrical method can be considered an acceptable method for carp, with a low impact on the welfare of the slaughtered carps. In particular, the use of electrical stunning alone could not make the carp unconscious for a long time, as demonstrated by quick VER recovery after stunning (VER were recorded already at 30 s post stunning) (32). Interestingly, in the field study conducted by Retter et al. (32) in Germany, the majority of farms used a combination of electrical stunning immediately followed by manual percussive stunning (59%). Under this condition, 92.6% of stunned carps displayed no behavioral indicators of consciousness and significantly fewer injuries related to mishits compared to sole percussive stunning. Thus, using a combination of electrical stunning and percussion could be a better option for this species, as the use of the singular methods could not be exhaustive in inducing unconsciousness under field conditions. In our survey, for both eel and carp, killing by ice or air asphyxia were used. However, this can pose risks for fish welfare since carp can survive up to 5 h in apnea (52) and eel, due to its peculiarity of being able to partially breathe with its skin, can survive even for days (12).

The assessment of unconsciousness for carps and eels was carried out mainly using “breathing,” “coordinated movements,” and “response to stimuli,” that are among those recommended by WOA (20). However, this assessment suggests caution in its application because of the physiological peculiarities of these species: e.g. resistance to breathing out of the water (described above) and as for the eel, the ability to move the body even when

TABLE 4 Stunning/killing methods and indicators of consciousness used for the minor slaughtered fish species in the surveyed fish-slaughter facilities.

	Stunning/killing methods					Indicators of unconsciousness				
	Thermal shock	Electric water bath	Percussion	Asphyxia in air	Electric dry system	Breathing	Eye movements	Coord. movements	Response to stimuli	Righting ability
Meager ( <i>n</i> = 4)	3	-	-	-	-	1	1	1	-	-
Trout perch ( <i>n</i> = 3)	3	-	-	-	-	3	3	2	3	3
Hybrid strip. Bass ( <i>n</i> = 3)	3	-	-	-	-	3	3	2	3	3
Royal perch ( <i>n</i> = 2)	2	1	-	-	-	2	1	1	1	1
Gray mullet ( <i>n</i> = 2)	1	-	-	1	-	-	-	1	-	-
Catfish ( <i>n</i> = 2)	2	-	-	-	-	2	2	1	2	2
Lavaret ( <i>n</i> = 1)	-	-	1	-	-	1	1	1	-	1
Danube salmon ( <i>n</i> = 1)	-	-	1	-	-	1	1	1	-	1
Pikeperch ( <i>n</i> = 1)	1	1	-	-	-	1	-	-	-	-
Tench ( <i>n</i> = 1)	-	1	-	-	-	-	-	1	1	-
Greater amb ( <i>n</i> = 1)	1	-	-	-	-	-	-	-	-	-
Sargo ( <i>n</i> = 1)	1	-	-	-	-	-	-	1	-	-
Turbot ( <i>n</i> = 1)	1	-	-	-	-	1	-	1	1	-

the brain is dead. In fact, for eel, in the surveyed facilities, the assessment was carried out taking into account the fish movements and breathing, which are considered not reliable indicators by EFSA (15).

#### 4.2.3. European seabass, gilthead seabream, and turbot

For seabass, seabream, and turbot, asphyxia ice in was confirmed as the most widely used practice in Italy for slaughtering. This method is not accepted by WOA (20), although it represents the most common method used also by the other main European producers, Spain and Greece (9). According to EFSA, an alternative slaughter method is requested for these species (14, 17). In fact, thermal shock has been shown to cause immobility and paralysis of fish: seabream remains conscious for up to 5 min after immersion in water and ice (53) and turbot even up to 90 min (54). The use of anesthetics (i.e., clove oil) (55, 56) or the diffusion of gases (i.e., carbon dioxide or nitrogen) (35, 57) in the stunning tank allow fish to reach faster unconsciousness and death than ice slurry application alone. For the same reason, research testing of electrical stunning reached promising results in seabream and seabass. This method, followed by a thermal shock, has recently been introduced in Europe in some seabass and seabream farms on an experimental basis (9). Recent studies have shown that the electrical method can be a valid alternative for the slaughter of turbot also (58). The introduction of this method also in Italian facilities should be considered if experimental data confirm an improvement of the current slaughter conditions for these species.

In the majority of the surveyed facilities, the slaughter of these fish took place without the assessment of unconsciousness or death. When assessed, the main indicators used namely “breathing” and “coordinated movements,” have been found in some studies to be not reliable and not robust indicators, even more so in regard to the main method applied, which causes paralysis (12, 17).

#### 4.2.4. Salmonids (different from rainbow trout)

For Arctic char, brown trout, Danube salmon and lavaret, the main methods currently used in the surveyed Italian facilities are among those reported by WOA specific recommendations on stunning/killing methods for species of the *Salmonidae* family (20), i.e., percussion and electric.

For these species, fish unconsciousness was monitored in almost all facilities, mainly by assessing the respiration and coordinated movements, among the indicators recommended by WOA (20). To the knowledge of the authors, no scientific study or species-specific opinions/guidelines are available about the reliability of these indicators for these species.

Certain recommendations contained in EFSA's opinions on Atlantic salmon and trout can generally be extended to Arctic char and brown trout and other salmonids. However, caution should be exercised when using these recommendations as there are different species within the family that react differently to stunning and killing methods (e.g., Arctic char has been shown to be strongly resistant to electricity).

#### 4.2.5. Other species

In the surveyed facilities, most of the species in this group were slaughtered using methods considered non-humane, which should be replaced (12, 20). Only asphyxia (in ice and air) was used for meager, trout perch, hybrid striped bass, gray mullet, catfish, greater amberjack, and sargo. For royal perch and pikeperch, electrical stunning was also utilized and should be preferred from a welfare point of view (12, 20).

Sturgeon and tench were the only species of this group to be slaughtered mainly in accordance with the general recommendations of WOA and EFSA (12, 20) for appropriate slaughter, i.e., with percussion and electric methods. However, it should be highlighted that these are very different species both in terms of size and behavior, thus there is an evident criticality when following the general guidelines for the analysis. For sturgeon, the

mechanical or electrical methods perfectly fall within the generically recommended methods by WOA and EFSA (12, 20). However, stunning procedures can considerably differ from facility to facility. Based on data collected by the survey, some facilities used rubber hammers, others used steel hammers, sometimes inside, some other outside the aquatic environment. In one case, the use of the sheep-specific stunner was also reported. It is evident how these methods have numerous practical critical issues, including dependence on the subjectivity of the operator and on his/her ability, training or state of fatigue. Last, but not least, while slaughtering methods should be species-specific, the use of the same methods and the same working conditions for different species, especially when minor species are involved, is a critical point. For the sturgeon in particular, that is a species so different from the other farmed and slaughtered species, there is a general lack of specific recommendation. Recently,

Williot et al. (59) reviewing the existing Siberian Sturgeon farming procedures in relation to their welfare, highlighted that there is no published study focused on slaughter of Sturgeon. Finally, while most of surveyed facilities monitored unconsciousness in fish belonging to this group by the indicators recommended by WOA (20), species-specific and validated indicators are lacking for these species.

## 5. Conclusion

This is the first study providing data about the systems used for stunning and killing fish and the methods used to evaluate the unconsciousness in Italian fish-slaughter facilities. In view of the revision of the European legislation about welfare of farmed animals and in the absence of field data about fish, extensive information about the systems currently used for slaughtering is pivotal for all the stakeholders in order to have an overview of fish welfare at slaughter.

Based on the collected information, methods considered non-humane are still widely used for the slaughtering of fish in Italy, especially for sea bass, sea bream, and the less slaughtered species, whereas for the other species, slaughtering mainly follow the WOA recommendations. The lack of scientific data and validated indicators make it difficult to obtain a clear picture of the welfare condition at slaughter for many species. On the other hand, as for rainbow trout, i.e., the major farmed fish in Italy, both the used stunning/killing methods and the indicators of unconsciousness are consistent with species-specific recommendations about welfare at slaughter.

In conclusion, more studies are necessary to clearly identify the best species-specific methods to protect fish at slaughter and the best indicators to be used to assess unconsciousness and death. Similarly, it is pivotal to gather more data on the feasibility and efficacy of these methods and indicators in field conditions as well as their economic impact.

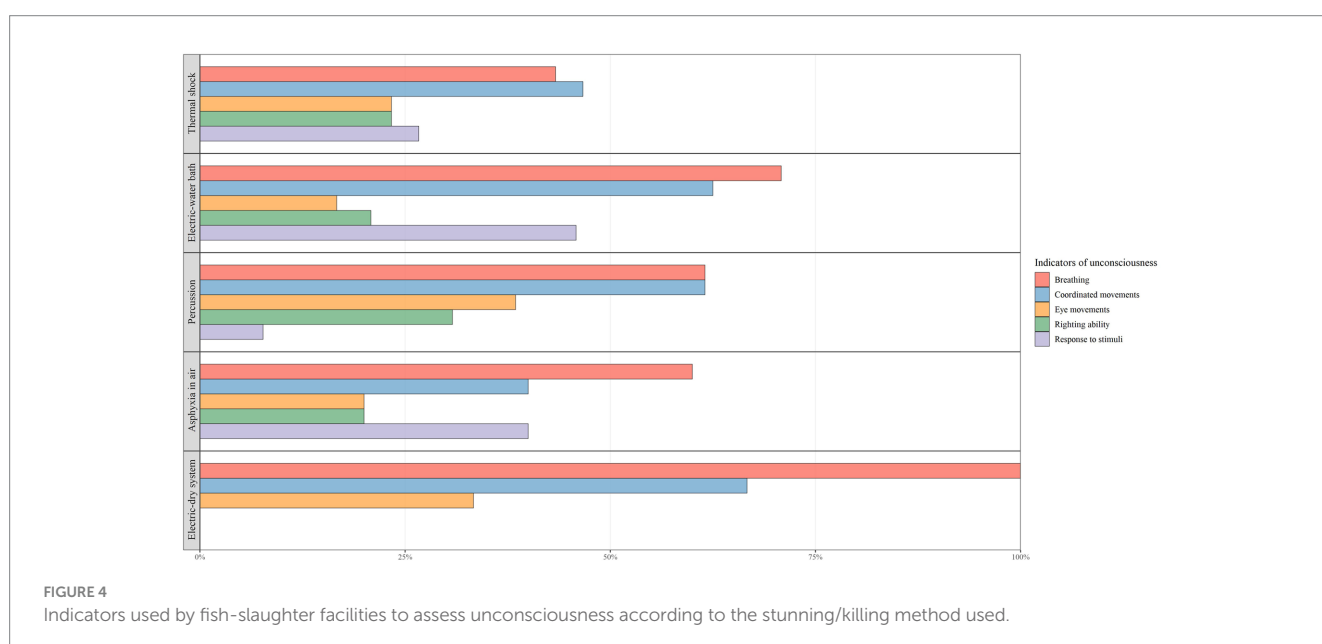
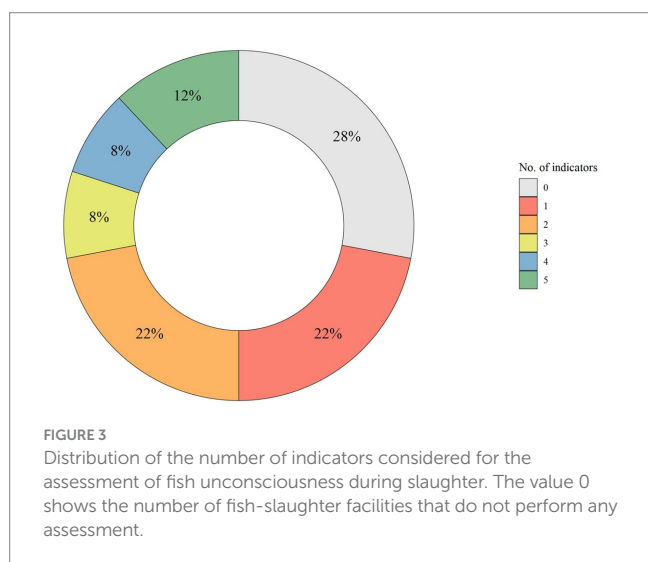


TABLE 5 General and specific WOA (20) and EFSA recommendations (12–19) on stunning/killing methods of farmed fish for human consumption.

	Percussion		Electric		Spiking or coring		Asphyxia in air or ice	
	WOAH	EFSA	WOAH	EFSA	Free bullet		CO2	
					WOAH	EFSA	WOAH	EFSA
General considerations	✓	✓	✓	✓	✓	✓	X	X
Rainbow trout	✓	✓	✓	✓	X	X	X	X
Eel	X	*	✓	✓	X	X	X	X
Carp	✓	*	✓	*	X	X	X	X
Seabream—Seabass	n/a	n/a	n/a	*	n/a	X	n/a	X
Turbot	n/a	*	n/a	*	n/a	X	n/a	X
Atlantic Salmon	✓	✓	✓	✓	X	X	X	X
Other salmonids	✓	✓	✓	✓	X	X	X	X
Tuna	X	X	X	X	✓	✓	X	X

✓, recommended/acceptable method; X, not recommended method; \*, recommended/acceptable method which requires further studies on this species; n/a., species-specific recommendations not available.

## Data availability statement

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

## Author contributions

SR and GC: conceptualization and funding acquisition. SR, GC, AT, TD, DB, FB, and VL: methodology. GC, SR, CT, and AB: formal analysis and investigation. GC, SR, and AT: writing—original draft preparation. GC, SR, AT, AM, CT, and VL: writing—review and editing. SR and AT: supervision. All authors contributed to the article and approved the submitted version.

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## References

1. FAO The State of World Fisheries and Aquaculture 2020. (2020).
2. FAO. *Towards Blue Transformation*. Rome: FAO (2022). 236 p.
3. Mood A, Lara E, Boyland NK, Brooke P. Estimating global numbers of farmed fishes killed for food annually from 1990 to 2019. *Anim Welf*. (2023) 32:e12. doi: 10.1017/awf.2023.4
4. Lambert H, Cornish A, Elwin A, D'Cruze N. A kettle of fish: a review of the scientific literature for evidence of fish sentience. *Animals*. (2022) 12:1182. doi: 10.3390/ani12091182
5. European Commission Directorate-general for health, food safety. Attitudes of Europeans towards animal welfare: Report. European Commission (2016).
6. Eurogroup for Animals Looking beneath the surface: fish welfare in European aquaculture. (2018) Available at: <https://www.slu.se/globalassets/ew/org/centrb/scaw-nationellt-centrum-for-djurvalfard/kontaktpunkter/kontaktpunkt-slakt/efsa-fisk/fish-welfare-in-european-aquaculture.pdf> (Accessed June 18, 2023).
7. European Court of Auditors Special report, animal welfare in the EU: closing the gap between ambitious goals and practical implementation. (2018). Available at: [https://www.eca.europa.eu/Lists/ECADocuments/SR18\\_31/SR\\_ANIMAL\\_WELFARE\\_EN.pdf](https://www.eca.europa.eu/Lists/ECADocuments/SR18_31/SR_ANIMAL_WELFARE_EN.pdf) (Accessed June 18, 2023)
8. Council of the European Union. Council regulation 1099/2009/EC of 24 September 2009 concerning the protection of animals at time of killing. *J Eur Commun*. (2009) L303:1–29.
9. European Commission Report from the commission to the European Parliament and the council on the possibility of introducing certain requirements regarding the protection of fish at the time of killing. (2018). Available at: <https://eur-lex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:52018DC0087&rid=3> (Accessed June 18, 2023).
10. European Commission A farm to fork strategy—for a fair, healthy and environmentally-friendly food system. Revision of the animal welfare legislation. (2022) Available at: <https://food.ec.europa.eu/animals/animal-welfare/evaluations->

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## Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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and-impact-assessment/revision-animal-welfare-legislation\_en (Accessed June 18, 2023).

11. EU Directorate General for Maritime Affairs and fisheries, EUMOFA. *The EU Fish Market*. 2022nd ed. LU: Publications Office (2022).

12. European Food Safety Authority (EFSA). Opinion of the scientific panel on animal health and welfare (AHAW) on a request from the commission related to welfare aspects of the main systems of stunning and killing the main commercial species of animals. *EFSA J.* (2004) 2:45. doi: 10.2903/j.efsa.2004.45

13. European Food Safety Authority (EFSA). Species-specific welfare aspects of the main systems of stunning and killing of farmed Atlantic Salmon. *EFSA J.* (2009) 7:1011. doi: 10.2903/j.efsa.2009.1011

14. European Food Safety Authority (EFSA). Species-specific welfare aspects of the main systems of stunning and killing of farmed turbot. *EFSA J.* (2009) 7:1073. doi: 10.2903/j.efsa.2009.1073

15. European Food Safety Authority (EFSA). Species-specific welfare aspects of the main systems of stunning and killing of farmed eels (*Anguilla Anguilla*). *EFSA J.* (2009) 7:1014. doi: 10.2903/j.efsa.2009.1014

16. European Food Safety Authority (EFSA). Species-specific welfare aspects of the main systems of stunning and killing of farmed fish: rainbow trout. *EFSA J.* (2009) 7:1012. doi: 10.2903/j.efsa.2009.1012

17. European Food Safety Authority (EFSA). Species-specific welfare aspects of the main systems of stunning and killing of farmed seabass and seabream. *EFSA J.* (2009) 7:1010. doi: 10.2903/j.efsa.2009.1010

18. European Food Safety Authority (EFSA). Species-specific welfare aspects of the main systems of stunning and killing of farmed tuna. *EFSA J.* (2009) 7:1072. doi: 10.2903/j.efsa.2009.1072

19. European Food Safety Authority (EFSA). Species-specific welfare aspects of the main systems of stunning and killing of farmed carp. *EFSA J.* (2009) 7:1013. doi: 10.2903/j.efsa.2009.1013

20. WOA, "Welfare aspects of stunning and killing of farmed fish for human consumption", Aquatic Animal Health Code (2022). p. 1–4. Available at: [https://www.woah.org/fileadmin/Home/eng/Health\\_standards/aahc/current/chapitre\\_welfare\\_stunning\\_killing.pdf](https://www.woah.org/fileadmin/Home/eng/Health_standards/aahc/current/chapitre_welfare_stunning_killing.pdf)

21. Robb DHF, Kestin SC. Methods used to kill fish: field observations and literature reviewed. *Anim Welf.* (2002) 11:269–82. doi: 10.1017/S0962728600024854

22. Lambooi E, Grimsbø E, de Vis JW, Van RHGM, Nortvedt R, Roth B. Percussion and electrical stunning of Atlantic salmon (*Salmo salar*) after dewatering and subsequent effect on brain and heart activities. *Aquaculture.* (2010) 300:107–12. doi: 10.1016/j.aquaculture.2009.12.022

23. Lambooi E, Van De Vis JW, Kuhlmann H, Münkner W, Oehlenschläger J, Kloosterboer RJ, et al. A feasible method for humane slaughter of eel (*Anguilla anguilla* L.): electrical stunning in fresh water prior to gutting. *Aquac Res.* (2002) 33:643–52. doi: 10.1046/j.1365-2109.2002.00677.x

24. Jung-Schroers V, Hildebrandt U, Retter K, Esser K-H, Hellmann J, Kleingeld DW, et al. Is humane slaughtering of rainbow trout achieved in conventional production chains in Germany? Results of a pilot field and laboratory study. *BMC Vet Res.* (2020) 16:197. doi: 10.1186/s12917-020-02412-5

25. European Commission Inception Impact assessment document. (2021) Available at: [https://food.ec.europa.eu/system/files/2023-01/aw\\_eval\\_revision\\_iaa\\_food-labelling.pdf](https://food.ec.europa.eu/system/files/2023-01/aw_eval_revision_iaa_food-labelling.pdf) (Accessed June 18, 2023).

26. Federation of European Aquaculture producers (FEAP) European Aquaculture Production Report 2014–2020. (2021). Available at: <https://feap.info/wp-content/uploads/2022/03/production-report-v1.1.pdf> (Accessed June 18, 2023).

27. EU Directorate General for Maritime Affairs and Fisheries, EUMOFA Sturgeon meat and other by-products of caviar: Production, trade, and consumption in and outside the EU. LU: Publications Office (2023). Available at: <https://data.europa.eu/doi/10.2771/18974> (Accessed June 29, 2023).

28. Italian livestock registry Number of aquaculture farms. (2022). Available at: [https://www.vetinfo.it/j6\\_statistiche/#/report-pbi/106](https://www.vetinfo.it/j6_statistiche/#/report-pbi/106)

29. Herskin MS, Hels A, Anneberg I, Thomsen PT. Livestock drivers' knowledge about dairy cow fitness for transport – a Danish questionnaire survey. *Res Vet Sci.* (2017) 113:62–6. doi: 10.1016/j.rvsc.2017.09.008

30. Coelho MEG, Pedrazzani AS, Quintiliano MH, Bolfe F, Molento CFM. Fish slaughter practices in Brazilian aquaculture and their consequences for animal welfare. *Anim Welf.* (2022) 31:187–92. doi: 10.17120/09627286.31.2.003

31. Lines JA, Spence J. Humane harvesting and slaughter of farmed fish. *Rev Sci Tech OIE.* (2014) 33:255–64. doi: 10.20506/rst.33.1.2284

32. Retter K, Esser K-H, Lüpke M, Hellmann J, Steinhagen D, Jung-Schroers V. Stunning of common carp: results from a field and a laboratory study. *BMC Vet Res.* (2018) 14:205. doi: 10.1186/s12917-018-1530-0

33. Zampacavallo G, Parisi G, Mecatti M, Lupi P, Giorgi G, Poli BM. Evaluation of different methods of stunning/killing sea bass (*Dicentrarchus labrax*) by tissue stress/quality indicators. *J Food Sci Technol.* (2015) 52:2585–97. doi: 10.1007/s13197-014-1324-8

34. Gräns A, Niklasson L, Sandblom E, Sundell K, Algers B, Berg C, et al. Stunning fish with CO<sub>2</sub> or electricity: contradictory results on behavioural and physiological stress responses. *Animal.* (2016) 10:294–301. doi: 10.1017/S1751731115000750

35. van de Vis H, Abbink W, Lambooi B, Bracke M. Stunning and killing of farmed fish: how to put it into practice? In: M Dikeman and C Devine, editors. *Encyclopedia of Meat Sciences*. 2nd ed. Oxford: Academic Press (2014). 421–6.

36. Robb DHF, O'Callaghan M, Lines JA, Kestin SC. Electrical stunning of rainbow trout (*Oncorhynchus mykiss*): factors that affect stun duration. *Aquaculture.* (2002) 205:359–71. doi: 10.1016/S0044-8486(01)00677-9

37. Marx H, Brunner B, Weinzierl W, Hoffmann R, Stolle A. Methods of stunning freshwater fish: impact on meat quality and aspects of animal welfare. *Z Lebensm Unters Forsch.* (1997) 204:282–6. doi: 10.1007/s002170050078

38. Anders N, Roth B, Grimsbø E, Breen M. Assessing the effectiveness of an electrical stunning and chilling protocol for the slaughter of Atlantic mackerel (*Scomber scombrus*). *PLoS One.* (2019) 14:e0222122. doi: 10.1371/journal.pone.0222122

39. Robb DHF, Roth B. Brain activity of Atlantic salmon (*Salmo salar*) following electrical stunning using various field strengths and pulse durations. *Aquaculture.* (2003) 216:363–9. doi: 10.1016/S0044-8486(02)00494-5

40. Kestin SC, Robb DH, van de Vis JW. Protocol for assessing brain function in fish and the effectiveness of methods used to stun and kill them. *Vet Rec.* (2002) 150:302–7. doi: 10.1136/vr.150.10.302

41. Vis JW, De V, Oehlenschläger J, Kuhlmann H, Muenkner W, Robb DHF, et al. Effect of the commercial and experimental slaughter of eels *Anguilla anguilla* L on quality and welfare In: Kestin, Steve C, Warriss, Paul D, editors. *Farmed Fish Quality*. Oxford: Fishing News Books (2001). 234–48.

42. Kestin SC, Wotton SB, Gregory NG. Effect of slaughter by removal from water on visual evoked activity in the brain and reflex movement of rainbow trout (*Oncorhynchus mykiss*). *Vet Rec.* (1991) 128:443–6. doi: 10.1136/vr.128.19.443

43. FAO. *FAO Yearbook. Fishery and Aquaculture Statistics 2019*. Rome: FAO (2021).

44. Italian Fish Farmers Association Production data 2022. Italian Fish Farmers Association. (2023). Available at: <https://www.acquacoltura.org/dati-produttivi-2022/> (Accessed July 28, 2023).

45. Noble C, Gismervik K, Iversen MH, Kolarevic J, Nilsson J, Stien LH, AS N Institute NV et al. Welfare indicators for farmed rainbow trout: Tools for assessing fish welfare. *Nofima* (2020). 310 p. Available at: <http://dspace.stir.ac.uk/handle/1893/31242> (Accessed August 2, 2023)

46. Noble C, Gismervik K, Iversen MH, Kolarevic J, Nilsson J, Stien LH, et al. Welfare indicators for farmed Atlantic salmon: Tools for assessing fish welfare. *Nofima* (2018). 351 p. Available at: <http://dspace.stir.ac.uk/handle/1893/28435> (Accessed August 2, 2023).

47. Hjelmstedt P, Sundell E, Brijis J, Berg C, Sandblom E, Lines J, et al. Assessing the effectiveness of percussive and electrical stunning in rainbow trout: does an epileptic-like seizure imply brain failure? *Aquaculture.* (2022) 552:738012. doi: 10.1016/j.aquaculture.2022.738012

48. Terlouw C, Bourguet C, Deiss V. Consciousness, unconsciousness and death in the context of slaughter. Part I. neurobiological mechanisms underlying stunning and killing. *Meat Sci.* (2016) 118:133–46. doi: 10.1016/j.meatsci.2016.03.011

49. Bowman J, van Nuland N, Hjelmstedt P, Berg C, Gräns A. Evaluation of the reliability of indicators of consciousness during CO<sub>2</sub> stunning of rainbow trout and the effects of temperature. *Aquac Res.* (2020) 51:5194–202. doi: 10.1111/are.14857

50. Bermejo-Poza R, Fernández-Muela M, De la Fuente J, Pérez C, González de Chavarri E, Díaz MT, et al. Effect of ice stunning versus electronarcosis on stress response and flesh quality of rainbow trout. *Aquaculture.* (2021) 538:736586. doi: 10.1016/j.aquaculture.2021.736586

51. Daskalova A, Pavlov A, Kyuchukova R, Daskalov H. Humane slaughter of carp – a comparison between three stunning procedures. *Turkish J Fish Aquat Sci.* (2016) 16:753–8. doi: 10.4194/1303-2712-v16\_4\_01

52. Rahmanifarrah K, Shabanpour B, Sattari A. Effects of clove oil on behavior and flesh quality of common carp *Cyprinus carpio* L in comparison with pre-slaughter CO<sub>2</sub> stunning chilling and asphyxia. *Turk J Fish Aquat Sci.* (2011) 11:139–47. doi: 10.4194/trjfas.2011.0118

53. van de Vis H, Kestin S, Robb D, Oehlenschläger J, Lambooi B, Münkner W, et al. Is humane slaughter of fish possible for industry? *Aquac Res.* (2003) 34:211–20. doi: 10.1046/j.1365-2109.2003.00804.x

54. Roth B, Imsland AK, Foss A. Live chilling of turbot and subsequent effect on behaviour, muscle stiffness, muscle quality, blood gases and chemistry. *Anim Welf.* (2009) 18:33–41. doi: 10.1017/S096272860000004X

55. Simitzis PE, Tsopelakos A, Charismiadou MA, Batzina A, Deligeorgis SG, Miliou H. Comparison of the effects of six stunning/killing procedures on flesh quality of sea bass (*Dicentrarchus labrax*, Linnaeus 1758) and evaluation of clove oil anaesthesia followed by chilling on ice/water slurry for potential implementation in aquaculture. *Aquac Res.* (2014) 45:1759–70. doi: 10.1111/are.12120

56. López-Cánovas AE, Cabas I, Ros-Chumillas M, Navarro-Segura L, López-Gómez A, García-Ayala A. Nanoencapsulated clove essential oil applied in low dose decreases stress in farmed gilthead seabream (*Sparus aurata* L.) during slaughter by hypothermia in ice slurry. *Aquaculture*. (2019) 504:437–45. doi: 10.1016/j.aquaculture.2019.02.003
57. Roque A, Gras N, Rey-Planellas S, Fatsini E, Palliser J, Duncan N, et al. The feasibility of using gas mixture to stun seabream (*Sparus aurata*) before slaughtering in aquaculture production. *Aquaculture*. (2021) 545:737168. doi: 10.1016/j.aquaculture.2021.737168
58. Lambooi B, Digre H, Erikson U, Reimert H, Burggraaf D, van de Vis H. Evaluation of electrical stunning of Atlantic cod (*Gadus morhua*) and turbot (*Psetta maxima*) in seawater. *J Aqua Food Product Technol*. (2013) 22:371–9. doi: 10.1080/10498850.2011.654047
59. Williot P, Chebanov M, Nonnotte G. Welfare in the cultured Siberian sturgeon, *Acipenser baerii* Brandt: state of the art In: P Williot, G Nonnotte and M Chebanov, editors. *The Siberian Sturgeon (Acipenser baerii, Brandt, 1869)2—Farming*. Cham: Springer International Publishing (2018). 403–50.



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# The effects of aerator noise on the swimming, feeding, and growth of *Micropterus salmoides*

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Aquaculture systems, replete with equipment noise originating from aerators, pumps, feeders, and filtration systems, are known to exert substantial influence on fish behavior and growth. In this research, the focus was directed towards comprehending the impacts of aerator noise on the swimming, feeding, and growth progression of largemouth bass. In the course of a 50-day experimental period, the bass population was segmented into two groups: aerator noise (90.3dB re  $\mu\text{Pa RMS}^{-1}$ ) and ambient noise (70.4dB re  $\mu\text{Pa RMS}^{-1}$ ). The findings indicated discernible disparities in the swimming behavior and feeding between the two groups. Specifically, bass in the noise group maintained greater average angular distance and physical separation from their nearest counterparts than the bass in the ambient group, which were  $43.61 \pm 1.89^\circ$  and  $85.47 \pm 1.72\text{mm}$  for the ambient group and  $48.32 \pm 0.49^\circ$  and  $97.01 \pm 0.57\text{mm}$  for the noise group. Furthermore, the feeding kinetic energy was markedly lower in the noise group as compared to the ambient group. For the first time, the Shannon-Wiener diversity index was leveraged to gauge the diversity of fish swimming behavior, with the results signifying the diversity index of the noise group was  $2.69 \pm 0.07$  higher than that of the ambient group, which was  $2.51 \pm 0.02$ . Lastly, the noise group demonstrated compromised growth performance, with a significantly lower average weight as opposed to the ambient group, along with marked variations in the specific growth rate. These findings offer a telling revelation about the profound impacts of aerator noise on the behavioral and growth of largemouth bass, thereby forming a valuable referential base for future research centered on the effects of noise pollution on aquatic organisms.

## KEYWORDS

aerator noise, largemouth bass, swimming behavior, feeding, growth

# 1 Introduction

As the global populace continues to expand, the imperative of food security and sustainable development is gaining paramount importance (Godfray et al., 2010). Aquaculture, an escalating segment of food production, is playing an increasingly significant role in maintaining global food security and bolstering economic growth (Subasinghe et al., 2009). As per a United Nations Food and Agriculture Organization (FAO) report, nearly half of the world's fishery yield is derived from aquaculture, with the prospect of this fraction growing in the foreseeable future (Canton, 2021). Aquaculture not only offers an abundant supply of animal and plant protein but also contributes to mitigating the overexploitation of marine resources and environmental degradation induced by conventional fishing methods (Naylor et al., 2000; Diana, 2009). In light of technological advances and improved aquaculture management, the industry is progressively adopting more ecofriendly and resource-efficient practices.

For instance, through deploying recirculating aquaculture systems (RAS), multitrophic aquaculture models, and precision feeding strategies, we can alleviate environmental pollution associated with aquaculture while augmenting resource utilization efficiency (Buck et al., 2018; Tang and Liu, 2018; Naylor et al., 2021). Amidst rapid advancements in aquaculture technology, the deployment of aquaculture equipment has become increasingly prevalent. While such equipment is instrumental in enhancing aquaculture productivity, promoting environmental quality, and ensuring product integrity, the escalating prevalence of noise pollution within the industry necessitates heightened attention (Bart et al., 2001; Slabbekoorn et al., 2010; Todd et al., 2021).

Noise emissions from various mechanical activities can have detrimental effects on animal health (Craven et al., 2009; New et al., 2014; Kunc et al., 2016; Slabbekoorn et al., 2019), compromising auditory abilities, inducing physiological stress (Popper, 2003; Halvorsen et al., 2012; Mancera et al., 2017), disrupting communication and behavior (Sarà et al., 2007; Gil et al., 2015; Gendron et al., 2020), and increasing stress-response hormones (Wale et al., 2013a; Hawkins and Chapman, 1975; Smith et al., 2004; Duarte et al., 2015), potentially impacting animal health and survival (Popper, 2003; Purser and Radford, 2011), especially in captive settings like farms and reserves (Popper and Hawkins, 2019; Sherwen and Hemsworth, 2019; Duarte et al., 2023). These noise-induced effects may manifest as reduced feeding behavior due to heightened vigilance (Evans et al., 2018; Giordano et al., 2022), leading to diverted attention (Mendl, 1999; Chan and Blumstein, 2011), potentially causing misjudgments in prey identification and feeding responses (Purser and Radford, 2011; Holles et al., 2013). Additionally, noise could obscure feeding-related acoustic signals (Brumm and Slabbekoorn, 2005; Zhang et al., 2021), resulting in missed feeding opportunities (Schaub et al., 2008; Siemers and Schaub, 2011) and altering food detection and classification during foraging (Lupien and McEwen, 1997; De Kloet et al., 1999).

Noise interference can disrupt animals' ability to accurately discern food quantity and quality (Kight and Swaddle, 2011), potentially leading to reduced foraging efficiency (Schaub et al., 2008), contrary to optimal foraging theory (MacArthur and Pianka,

1966). Alternatively, this could also be attributed to noise-induced stress, suppressing appetite and activity, and consequently reducing foraging behavior (Mendl, 1999; Charmandari et al., 2005; Williams et al., 2015; Zhang et al., 2022). For instance, Gendron et al. (2020) discovered that noise-exposed *Pseudopleuronectes americanus* larvae exhibited significantly shorter feeding durations and smaller stomach capacities than controls. Purser and Radford (2011) observed decreased foraging performance in *Gasterosteus aculeatus* due to increased feeding errors. Likewise, Voellmy et al. (2014) reported on the impact of noise on *Phoxinus phoxinus* feeding success. These investigations collectively suggest that noise pollution can significantly affect animal feeding behavior, consequently influencing their growth and health indices.

Noise disturbances can not only alter fish feeding behavior but also significantly impact their swimming behavior and spatial distribution (De Vincenzi et al., 2015; Sabet et al., 2016; Simpson et al., 2016; Mickle et al., 2019; Hang et al., 2021). For instance, Hanache et al. (2020) reported that *Phoxinus phoxinus*, when exposed to noise, allocated less time to foraging in low food density conditions, exhibited changes in spatial distribution, and demonstrated decreased cohesiveness in their swimming. Similarly, noise-affected tuna schools have been observed to alter their coordinated unidirectional swimming structure, increase vertical movements towards the water surface or bottom, and exhibit uncoordinated swimming behavior in more dispersed groups (Sarà et al., 2007).

Largemouth bass (*Micropterus salmoides*), an economically significant freshwater species known for its fast growth rate, high nutritional value, and strong environmental adaptability (Bai and Li, 2018), holds global economic importance in aquaculture. While previous research has explored the effects of noise on fish feeding and swimming behavior, this study takes approach by introducing the Shannon-Wiener Diversity Index (SWDI) analysis method, extending investigation to encompass group behavior. Specifically, this study will assess the impact of oxygenation equipment noise on feeding and swimming behaviors of largemouth bass school, and quantify the differences in noise impacts on fish school behavior. Through a comprehensive analysis of the findings, this study aims to offer scientific insights to the aquaculture industry, mitigate adverse environmental impacts, and bolster the sustainable development of the industry.

## 2 Materials and methods

### 2.1 Fish

In this study, 200 largemouth bass were procured, each with an average body length of  $11.17 \pm 0.36$  cm and a weight of  $14.37 \pm 2.13$  g, from Hangzhou Jianfeng Agricultural Development Co., Ltd. Before initiating the experiment, the fish were temporarily housed in a RAS for environmental acclimatization. The water parameters in the holding pool were sustained at certain levels to ensure optimal conditions for the fish. Specifically, the temperature was kept steady at around 25°C, within a margin of  $\pm 1^\circ\text{C}$ . Similarly, the pH was  $7.2 \pm 0.8$ . Additionally, the concentration of dissolved

oxygen was ensured to be  $6.5 \pm 1 \text{ mg l}^{-1}$ . Monitoring was also carried out on the levels of ammonia nitrogen and nitrite, with both being kept below  $0.3 \text{ mg l}^{-1}$  and  $0.5 \text{ mg l}^{-1}$ , respectively. Lighting for the RAS system was provided by full-spectrum LED lights with a 24-hour photoperiod. The fish were fed with floating feed (Fujian Tianma Science and Technology Group Co., Ltd.) at 8:00 a.m. and 8:00 p.m. daily. These preliminary steps were implemented to ensure optimal environmental and feeding conditions for the smooth progression of the subsequent experiment.

## 2.2 Experimental system setup

The experimental setup incorporated two sets of RAS (Figure 1), each containing three aquaculture pools made of PVC with an internal diameter of 1 meter and a height of 0.8 meters. The biological filter pool, a  $1 \times 0.5 \times 0.8 \text{ m}$  PVC tank, facilitated ammonia removal, UV sterilization, and solid particle filtration, offering a controlled environment for the fish. The soundscape management equipment included a type 8130 hydrophone (Denmark Brüel & Kjær Co. Ltd.), a player, an AVANT MI-2004 power amplifier, and a US-0150 underwater loudspeaker (Hangzhou ECON Technology Co. Ltd., China), collectively used for noise soundscape collection and provision during the experiment. Additionally, the setup comprised a behavior video capture system (Hikvision, Hangzhou) to record the fish group's feeding and swimming patterns.

The experimental system was established in a quiet laboratory with measures like wrapping the stainless steel frame with blackout cloth, padding the culture pool's bottom with a PVC shock-absorbing tray, employing liquid oxygen for oxygen supply, using a fixed biological bed, and wrapping PVC pipes with sound-absorbing materials to mitigate interference from environmental noise and system operational noise.

## 2.3 Experimental design

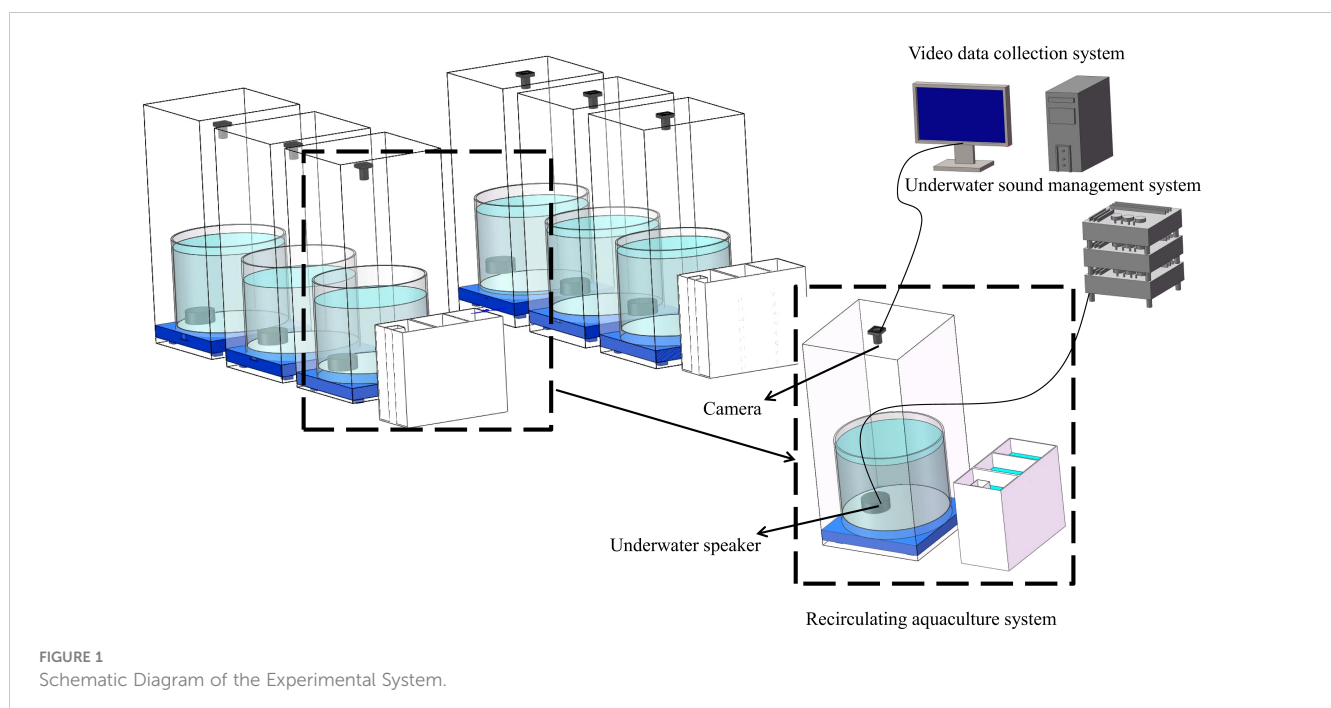
In this study, a noise group was established and subjected to a sound level of  $90.3 \text{ dB re } \mu\text{Pa RMS}^{-1}$ , while an ambient control group with a sound level of  $70.4 \text{ dB re } \mu\text{Pa RMS}^{-1}$  was also established (Figure 2). The noise sample was collected from the large-scale aquaculture facility of Suzhou Jinchengfu Fisheries Co., Ltd. and effectively represents the acoustic characteristics encountered by fish in a closed RAS. The aeration equipment sampled was a GRB-200A three-lobe roots blower (GSD Industrial Co., Ltd.) with an air volume capacity of  $50 \text{ m}^3 \text{ min}^{-1}$  and a pressure of  $78.43 \text{ kPa}$ . To evaluate the acoustic impact of the aeration machine within the fish pond, recordings of the soundscape were obtained when the aeration machine operated exclusively. For the ambient control group, the soundscape was recorded with the complete RAS in the shutdown state.

Each group contained three replicates of the aquaculture pool, each housing 20 fish. The entire experiment spanned 50 days. During the experiment, the recirculating water flow was set at  $300 \text{ L h}^{-1}$ , with a daily water change constituting 15% of the total volume. Fecal waste at the bottom of the fish pond was vacuumed daily. Concurrently, the parameters of the water quality were held in alignment with those established during the acclimation stage, and the fish were fed at 8:00 am and 8:00 pm each day.

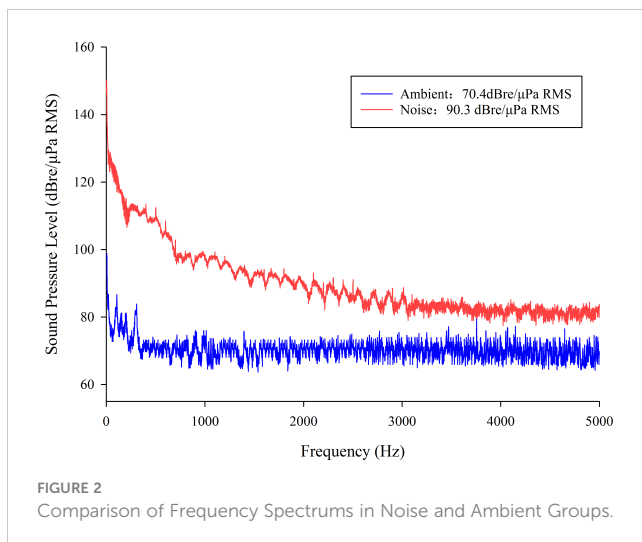
## 2.4 Data collection and analysis

### 2.4.1 Fish swimming and feeding

To capture the feeding and swimming behavior of the largemouth bass, cameras and 24-hour constant LED lights were installed on the stainless steel frame carrying shade cloth, which was positioned directly above the culture tank. Recordings were made at







six designated time points (3:00, 6:00, 9:00, 12:00, 15:00, 18:00) each day, with each session lasting ten minutes, to analyze the group swimming behavior. To analyze feeding behavior, video footage was captured every three days at 8:00 in the morning. The cameras were networked to a switch and recorder, facilitating both video storage and real-time monitoring.

For the analysis of group swimming behavior, the Mask RCNN technique was employed (Figure 3). From each video, 60 frames were extracted, and object instance segmentation was conducted on the experimental fish in the frames. This generated a segmentation mask for each experimental fish, enabling their identification and localization (He et al., 2020). To evaluate the cohesion of swimming, the distance between every experimental fish and its two nearest neighbors within the group was computed. A line connecting a fish's head and tail represented its swimming direction. To evaluate the polarity of swimming, the angle formed by the swimming direction of an experimental fish and its two closest neighbors was utilized.

The SWDI, an index commonly used in biodiversity research proposed by Claude Shannon and Norbert Wiener in the fields of information theory and ecology, measures the richness and evenness of species in an ecosystem. This study proposed to use the SWDI to evaluate the diversity of group swimming behavior.

The average angle and distance during the experiment were divided into nine angle ranges (0-10°, 10-20°, 20-30°..., >90°) and eight distance ranges (<60mm, 60-70mm, 70-80mm..., >120mm), generating 72 unique angle-distance combinations. The frequency of each angle-distance combination was counted, and the relative frequency calculated. The SWDI for group behavior was then determined using the formula:

$$H = -\sum (p_i * \ln(p_i))$$

Where H represents the SWDI, and  $p_i$  is the relative frequency of the  $i_{th}$  angle-distance combination.

In analyzing group feeding behavior, this study implemented a Modified kinetic energy model (MKEM) to quantify the feeding intensity of the group (indicative of the group's appetite). The MKEM model doesn't require tracking or foreground segmentation. It facilitates the extraction and quantification of the spatial behavioral characteristics of the fish group, effectively describing the feeding behavior, independent of light environment and water quality (Wei et al., 2021). Greater intensity of water ripple fluctuations and alterations in the flow field in the reflective area, induced by feeding, result in higher feeding kinetic energy values (Figures 4A, B). Each feeding video clip in this experiment lasted 15 seconds, with a frame rate of 25fps, and the single-time feeding kinetic energy change is shown in Figure 5A. The feeding process was initiated when the feed was introduced. The same quantity of feed was given in each fish tank using a satiety feeding strategy, with leftover feed scooped out, dried, and weighed post-feeding to compute the feed intake quantity.

## 2.4.2 Growth

Prior to the commencement of the experiment, a random sampling of 20 fish was undertaken to determine initial body length and weight. Over the course of the experiment, fish were weighed every 10 days. At the conclusion of the experiment, eight fish were randomly selected from each cultivation pond for growth evaluation, with the concurrent tally of fish in each group also recorded.

The Specific Growth Rate (SGR, % d<sup>-1</sup>) and Survival Rate (SR, %) were calculated as follows:

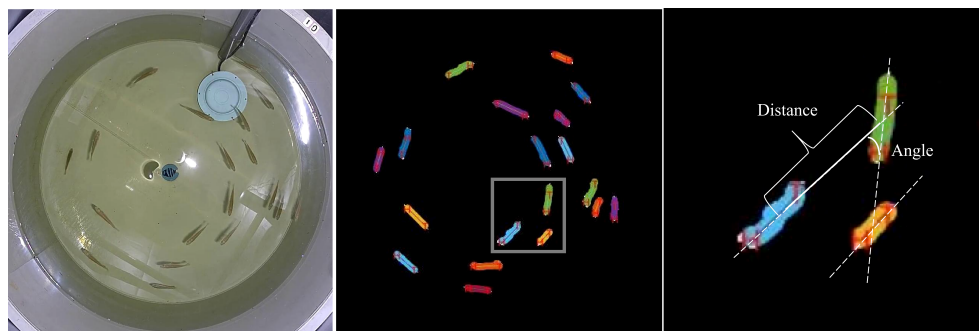


FIGURE 3  
Quantitative Measurement of the Angle and Distance Between a Focal Fish and Its Nearest Neighbor.

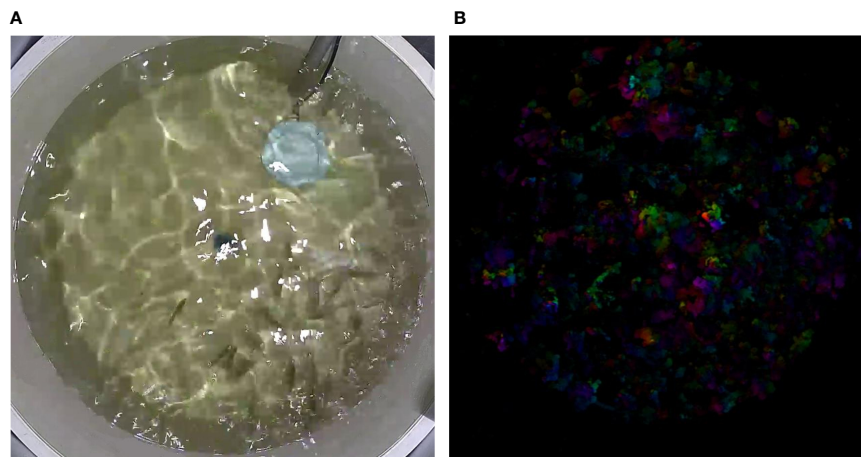


FIGURE 4

Feeding Kinetic Energy Analysis. (A) Snapshot of a Feeding Event: Visual depiction of largemouth bass during the feeding process. (B) Transformation of Feeding Snapshot into an Optical Flow Image: Conversion of the feeding snapshot into an optical flow image for further analysis.

$$\text{SGR} = 100 \times (\ln \text{BW}_2 - \ln \text{BW}_1) \times (T_2 - T_1)^{-1}$$

$$\text{SR} = 100 \times N_f \times N_i^{-1}$$

In these equations,  $\text{BW}_1$  and  $\text{BW}_2$  represent the initial and final body weights of each fish (in grams), while  $T_1$  and  $T_2$  denote the corresponding time intervals.  $N_i$  and  $N_f$  stand for the initial and final numbers of fish, respectively.

### 2.4.3 Data analysis

The data are presented as mean  $\pm$  standard deviation (SD). A comparative analysis of growth, swimming, and feeding across different treatments was conducted using independent t-tests. For SWDI measured at multiple time points, comparisons were analyzed using repeated measures ANOVA. Prior to analysis, normality of the data distribution was assessed through the Kolmogorov-Smirnov test, and homogeneity of variances was examined using Levene's Test. In cases where the assumptions

were met, the threshold for statistical significance was set at  $P < 0.05$ . All statistical analyses were performed using IBM SPSS Statistics version 22.0. Graphical representations were generated using SigmaPlot Version 14.0.

## 3 Results

### 3.1 Swimming behavior

The results exhibited substantial disparities in swimming behavior between the largemouth bass in the ambient group and those exposed to noise. Throughout the experiment, the mean angle and distance between a focal fish and their nearest counterparts in the noise group were significantly greater than those in the ambient group (Figure 6), with average values of  $48.32 \pm 0.49^\circ$  and  $97.01 \pm 0.57\text{mm}$  for the noise group, compared to  $43.61 \pm 1.89^\circ$  and  $85.47 \pm 1.72\text{mm}$  for the ambient group ( $P < 0.05$ ).

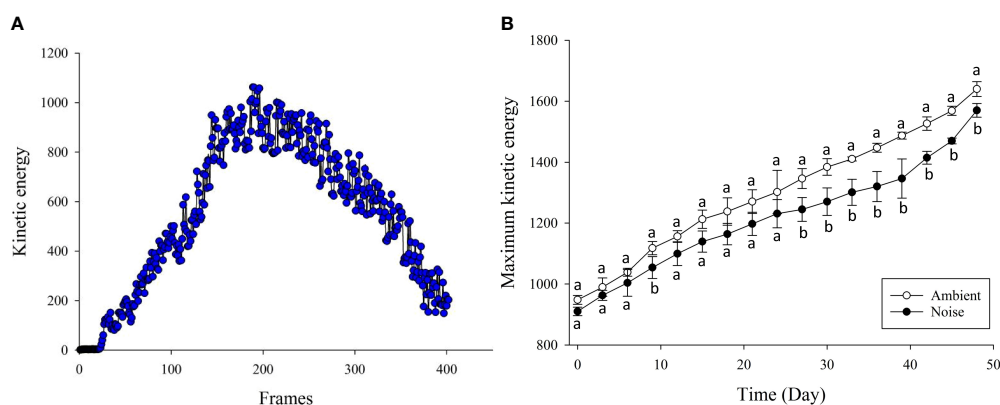


FIGURE 5

Kinetic Energy Changes During Feeding. (A) Change in Kinetic Energy During a Single Feeding: Variation in kinetic energy throughout a singular feeding event. (B) Peak Feeding Kinetic Energy Changes in Noise and Ambient Groups: Comparison of peak feeding kinetic energy changes in both groups from Day 1 to Day 50 of the experiment.

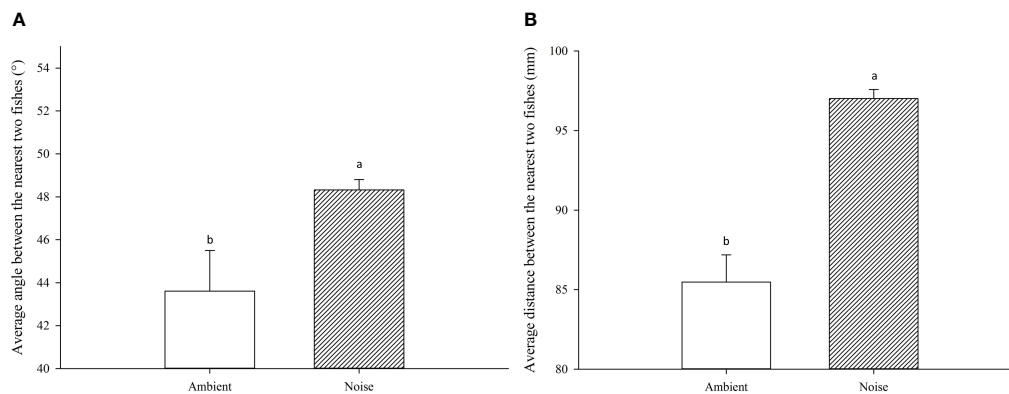


FIGURE 6

Comparative Analysis of Swimming Behavior in the Noise and Ambient Groups. (A) The average angle between the nearest two fish in each group. (B) The average distance between the nearest two fish in each group.

When considering the diversity of swimming behavior, notable differences were also observed between the ambient and noise groups. An examination of various angle-distance combination categories (Figure 7) reveals that in the ambient group, the combination of an average angle of 30-40° and an average distance of 80-90mm was observed in 12.73% of fish, followed by the combination of 30-40° and 90-100mm, which accounted for 11.23% of observations. Conversely, the fish in the noise group demonstrated more dispersed swimming behavior. The highest proportion of observations in the noise group was in the combination of 30-40° and 100-110mm (9.09%), followed by 30-40° and 90-100mm (8.72%). Swimming behavior in the noise group was characterized by larger average distances.

The mean SWDI also differed significantly between the two groups throughout the experiment (Figure 8A), with the noise group ( $2.69 \pm 0.07$ ) being significantly higher than the ambient group ( $2.51 \pm 0.02$ ) ( $P < 0.05$ ). As depicted in Figure 8B, the SWDI of the fish population in the noise group exhibited an upward trend

over time. The two groups started to show significant difference one month after noise intervention, and this disparity persisted until the conclusion of the experiment. The present study further analyzed the average SWDI at various time points (3/6/9/12/15/18 o'clock), revealing diurnal patterns in fish activity (Figure 8C). At 6:00 and 18:00, the diversity of fish swimming behavior was lower ( $P < 0.05$ ), suggesting more stable formations. While similar trends were observed in both the noise and ambient groups, the average SWDI for fish swimming behavior in the ambient group was significantly lower at 3/6/18 o'clock compared to the noise group ( $P < 0.05$ ).

### 3.2 Feeding

With respect to feeding behavior, the largemouth bass in the ambient group exhibited a more robust feeding inclination, while those in the noise group displayed a diminished appetite (Figure 5B).

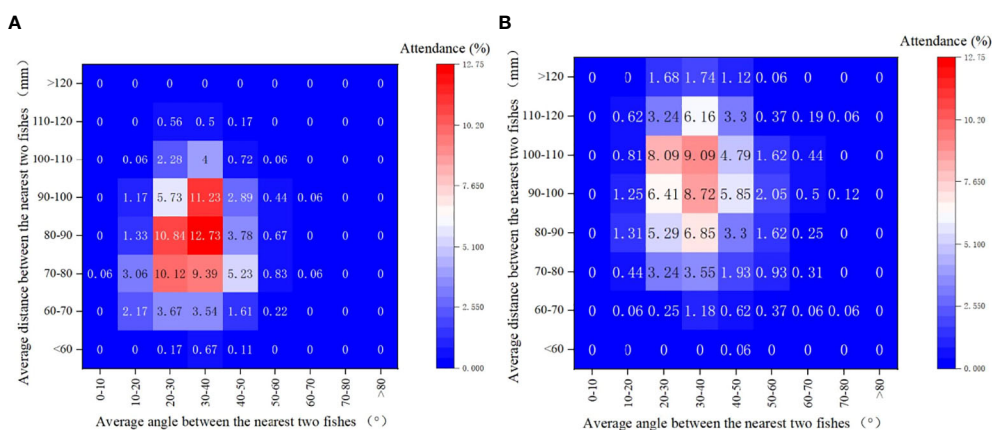


FIGURE 7

Heat Map Depicting the Distribution of Largemouth Bass Swimming Behavior in Both Experimental Tanks: The heat map visualizes the variations in the swimming behavior of largemouth bass in response to the experimental conditions. Both the average angle (horizontal axis) and average distance (vertical axis) between fish are represented. (A) Ambient Group: Depicts the swimming behavior distribution under normal conditions. (B) Noise Group: Illustrates the altered swimming behavior under the influence of equipment noise.

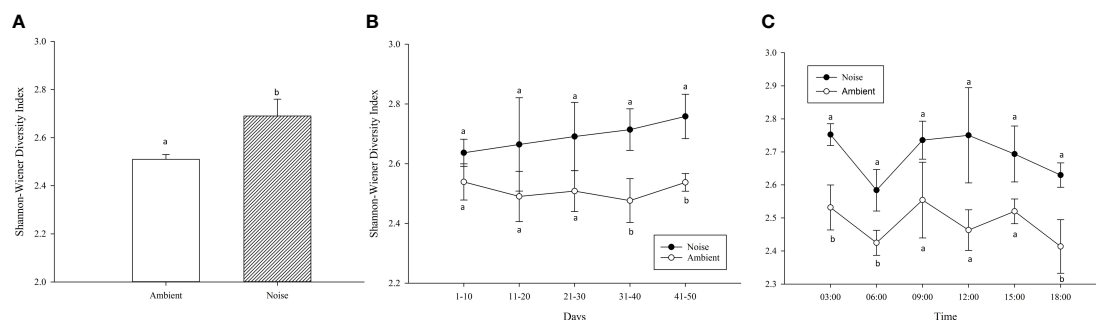


FIGURE 8

Variation in Shannon-Wiener Diversity Index (SWDI) Reflecting Swimming Behavior Diversity in Largemouth Bass Populations Across Noise and Ambient Groups: This figure illustrates the diverse swimming behaviors of largemouth bass in response to both experimental conditions using the SWDI. (A) Mean SWDI Throughout the Experiment. (B) Mean SWDI for Each 10-Day Period. (C) Mean SWDI at Different Time Points of the Day.

Specifically, the feeding kinetic energy in both groups increased as the experiment unfolded, albeit at different rates. During the early stage of the experiment, the feeding kinetic energy of the noise group was lower than the ambient group, though the difference was not statistically significant. However, from the 27th day onwards, the feeding performance of the ambient group consistently outstripped that of the noise group, a trend that persisted until the experiment's conclusion. Measurements taken on the 48th day indicated that the feeding kinetic energy of the ambient group stood at  $1639.99 \pm 24.30$ , compared to  $1569.97 \pm 22.53$  for the noise group.

Moreover, the daily average feed intake of the largemouth bass in the ambient group exceeded that of the noise group, with the discrepancy widening over time. For most of the experimental period, the divergence in daily average feed intake between the two groups did not reach statistical significance, except on days 33, 47, 49, and 50 (Figure 9). Nevertheless, a marked contrast in total feed consumption was observed between the two groups (Figure 10), with the noise group's intake ( $24.95 \pm 1.58\text{g}$ ) significantly lower than that of the ambient group ( $29.25 \pm 0.86\text{g}$ ) ( $P < 0.05$ ).

### 3.3 Growth

The present study revealed that the SR remained 100% across all groups throughout the experiment, and that the noise generated by aeration equipment significantly impacted the growth performance of largemouth bass. At the experiment's inception, the initial weight of fish across all groups was identical, averaging  $11.06 \pm 0.68\text{g}$ . After 50 days of experimental treatment, the mean weight of largemouth bass in the ambient group ( $30.61 \pm 1.51\text{g}$ ) was significantly higher than that in the noise group ( $26.36 \pm 1.45\text{g}$ ) ( $P < 0.05$ ), indicating that aeration noise impeded growth. Figure 11A illustrates the weights of the fish measured at ten-day intervals throughout the experiment. The data show that during the experiment's first month, there was no significant disparity in weight between the ambient and noise groups. However, from day 40 onwards, the weight of fish in the aerator noise group was persistently and markedly below the level in the ambient group, a trend that endured until the experiment's end. Over the course of the experiment, the weight difference between the ambient and noise groups progressively increased (Figure 11A). The SGR of

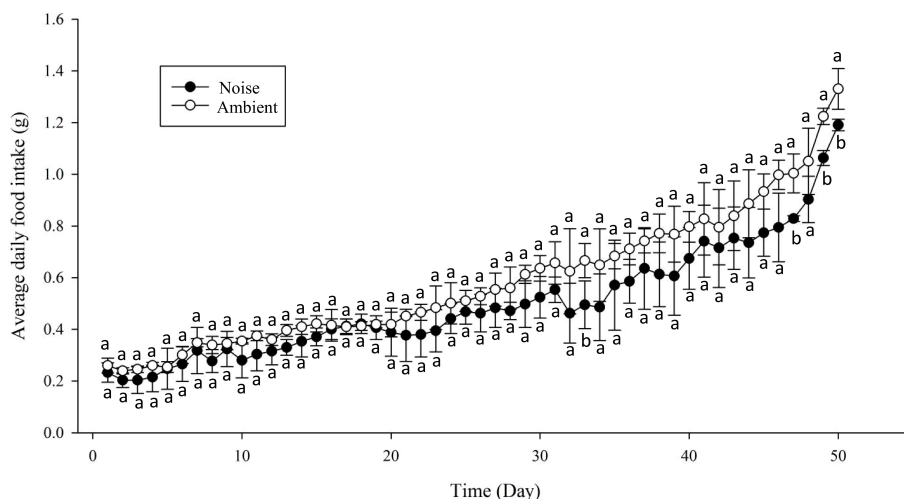


FIGURE 9

Comparison of Daily Feed Intake: This graph depicts the change in the average daily feed intake for both the Noise and Ambient groups over the course of the 50-day experimental period. Differences marked with different lowercase letters are considered significant.

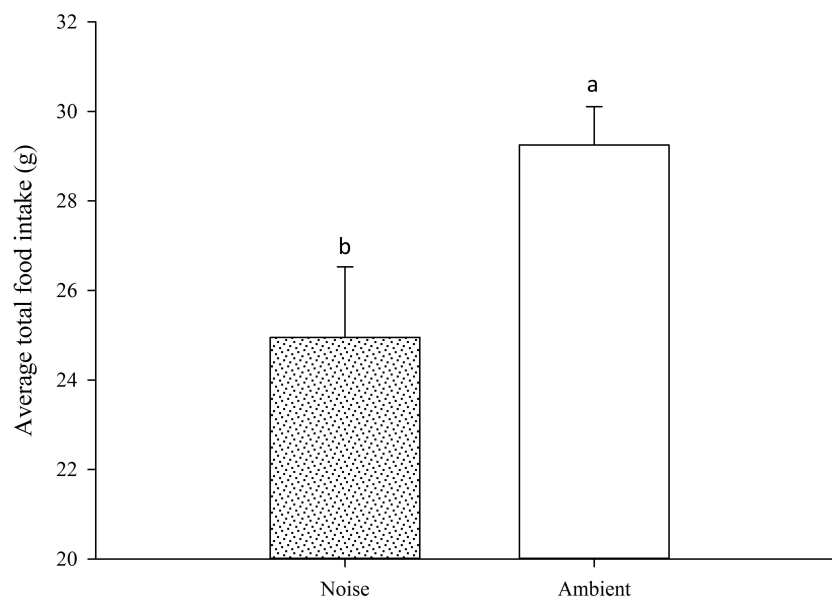


FIGURE 10

Total Food Intake Comparison: This figure illustrates the overall contrast in food consumption between the Noise and Ambient groups throughout the entire experimental duration. Differences marked with different lowercase letters are considered significant.

largemouth bass in the noise group and the ambient group also displayed significant divergence (Figure 11B), with the noise group's SGR ( $3.38 \pm 0.09$ ) markedly lower than that of the ambient group ( $3.68 \pm 0.08$ ) ( $P < 0.05$ ).

## 4 Discussion

In this investigation, the profound influence of aeration equipment noise on the swimming patterns, feeding habits, and growth of largemouth bass was distinctly observed. Evidently, compared to those in the ambient group, the fish in the noise group exhibited substantial differences in swimming behavior, feeding vigor, and quantity of food consumed. The escalated

swimming expenditure and diminished feeding intensity and quantity within a noisy environment adversely affected the growth performance of the largemouth bass in the noise group.

Acoustic signals constitute one of the primary channels through which fish garner information about their environment, with aquatic species relying on auditory data for communication, navigation, orientation, evasion, and foraging (Popper and Hawkins, 2019). Noise acts as a stimulus source that disrupts fish attention and interferes with fish behavior by masking acoustic information (Slabbekoorn et al., 2010). In terms of swimming behavior observed in this study, significant disparities arose between the ambient and aerator noise groups. The average distance and angle between each fish in the noise group and its nearest neighbor were greater, implying that the noise environment

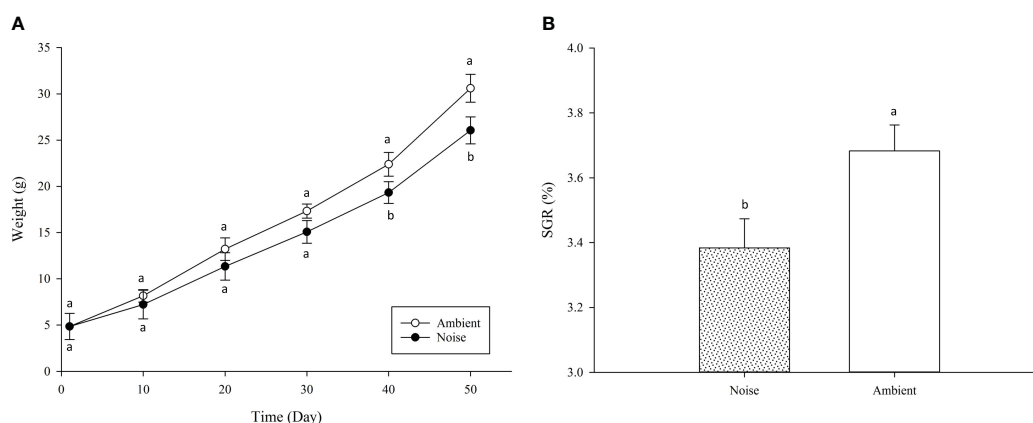


FIGURE 11

Largemouth Bass Growth Comparison in Different Noise Environments. (A) Body Weight Change in Body Weight per 10 Days. (B) Specific Growth Rate (SGR) Comparison.



might have impeded the orientation navigation information of the largemouth bass or induced discomfort, thus compelling them to maintain a more significant distance and angle from other individuals.

The distance and angle between the closest individuals in a fish school can serve as metrics of the school's clustering strategy, which enhances group collaboration and reduces swimming energy expenditure (Chen et al., 2016; Kent et al., 2019). In this study, the noise group fish had larger average distances and angles, indicative of lower fish school cohesion and polarity. This observation aligns with previous research findings. For instance, Sarà et al. (2007) found that noise could alter the swimming direction of tuna, leading to a relatively dispersed cluster structure and more chaotic swimming behavior in the school. Similarly, De Vincenzi et al. (2015) demonstrated that noise impacted the swimming activity and behavior distribution of the lesser spotted dogfish (*Scyliorhinus canicula*) in aquaculture ponds. McLaughlin and Kunc (2015) reported that noise exposure caused the european minnow (*Phoxinus phoxinus*) to modify its spatial distribution, with marked increases in social interaction and swimming distance. These findings resonate with the present observations: under the influence of the noise environment, the distribution of largemouth bass in the culture pond becomes more dispersed.

However, the research further points out that the noise environment not only affects the clustering behavior of largemouth bass schools but also affects the diversity of their swimming behavior. It was discovered that, compared to the ambient group, the diversity of swimming behavior among the largemouth bass in the noise group was higher. This could be due to the largemouth bass being distracted by the disruption of aeration equipment noise, or the occlusion of orientation and navigation acoustic information by noise. Additionally, this might be a response strategy of the largemouth bass to the noise environment, in which behavioral diversity is increased to adapt to the noisy conditions. A higher SWDI may indicate a more diversified distance and angle between fish in a school and more irregular formations during swimming. In some instances, a greater diversity of swimming behavior may correlate with increased swimming energy expenditure (Webb, 1984). For example, irregular formations that lead to larger distances between individuals could cause fish to expend more energy while adjusting their position and speed during swimming. Maintaining this state of elevated energy expenditure over an extended period could potentially affect individual growth performance within the fish school (Jobling et al., 1993). The application of the SWDI to assess the behavior of largemouth bass schools is a novel contribution of this study. Although this index is often used to evaluate the diversity of biological communities (Spellerberg and Fedor, 2003; Thukral, 2017), it is infrequently applied within behavioral ecology. The study's results indicate that this index effectively captures the diversity of swimming behavior in largemouth bass schools, thereby introducing a novel research tool in behavioral ecology. The reliability of the findings was ensured through meticulous experimental design and data analysis methodologies. For example, the swimming behavior of

fish schools was continuously monitored throughout the experiment, encompassing comprehensive classification of angle-distance combinations and rigorous statistical analyses. Additionally, regular inspection and maintenance of experimental equipment and data training models were conducted to ensure the stability of the noise level and the accuracy of the algorithm. As a result, the credibility of the findings is held with a high level of confidence.

While this study primarily aimed to investigate the distinct impact of noise generated by oxygenation equipment on largemouth bass behavior, an additional insight emerged regarding the general behavioral patterns of fish schools. By analyzing the diversity of fish school behavior at different times throughout the day, the present study observed that fish schools maintain a more regular swimming formation prior to feeding times (6:00, 18:00). Post-feeding, fish schools may exhibit a greater diversity of behavior. Once satiated, they might be more inclined to engage in other activities such as reproduction, territorial protection, and social interactions (Ward et al., 2006). At this juncture, the fish school may become more dispersed, reducing the need for foraging, and individuals within the school may have greater freedom of movement. Under conditions of scarcity, the behavior of fish schools may be more uniform. This is due to the need for cooperation in the search for food in the face of reduced availability, causing the fish to form more compact groups. This scenario leads to an increase in fish school cohesion, with the tighter assembly being more effective in food foraging (Keenleyside, 2012). Starving fish schools might rely more heavily on the guidance of a leader (Reebs, 2000), subsequently reducing behavioral diversity.

In this study, it was observed that largemouth bass in the ambient group displayed a stronger feeding inclination, while those in the noise group demonstrated a significantly diminished appetite. This discrepancy might be attributed to noise distractions affecting the largemouth bass's focus, indirectly impacting their feeding desire and resulting in decreased food consumption. These findings align with those of Voellmy et al. (2014), who verified the disruptive influence of sound on fish feeding, showing that significantly more individuals in the control group obtained food, indicating that fish foraging activities were hindered under experimental conditions. Purser and Radford (2011) demonstrated that the stickleback fish, *Gasterosteus aculeatus*, exposed to noise, increased the frequency of attacks on non-food items due to misidentifying them as food, leading to a decline in its foraging performance. Gendron et al. (2020) found that the winter flounder, *Pseudopleuronectes americanus*, preyed less frequently and had a relatively reduced stomach capacity under noise conditions. Wale et al. (2013b) reported that noise affected the feeding behavior and performance of crabs in a tank-based controlled experiment. While the mentioned studies involve different fish species, the present study focus on largemouth bass contributes to understanding of how noise pollution can impact feeding behavior in this particular species.

It's worth noting that this research also observed that the feeding kinetic energy in both groups was initially low, but increased as the experiment progressed with the growth and increased food consumption of the largemouth bass. In the noise

group, the increase in feeding kinetic energy may reflect an adaptive response, suggesting that, after a period of acclimation to the noise environment, largemouth bass could adjust their behavior to cope with the noise conditions (Dooling et al., 2015). Nonetheless, despite these adaptive shifts, the feeding performance of the largemouth bass in the noise-exposed group continued to be significantly inferior compared to the ambient group. This diminished performance might be attributed to the persistent interference and suppression induced by the noisy environment on the largemouth bass's feeding conduct and appetite.

This investigation underscores the significant influence of aeration equipment noise on largemouth bass growth performance. Largemouth bass in the noise-exposed group exhibited a considerably lower average weight and slower growth rate compared to the ambient group. This discrepancy may stem from the noise-induced reduction in feeding desire and intake, leading to decreased energy consumption. Additionally, the less cohesive and polarized swimming patterns observed in the noise group could contribute to heightened energy expenditure during swimming activities. These findings align with prior research. Lagardère (1982) reported elevated noise levels in aquaculture ponds leading to increased metabolic rates in prawns and reduced growth performance. Banner and Hyatt (1973) observed decreased SR and growth performance in carp exposed to noise. Similarly, Wysocki et al. (2007) and Davidson et al. (2009) noted short-term negative effects of noise on rainbow trout growth performance in RAS, with potential adaptability over time. Hang et al. (2021) also explored the negative effects of noise on largemouth bass growth performance in RAS systems, this study uniquely focuses on analyzing the specific impact of aeration equipment noise, which plays a prominent role in the soundscape of intensive aquaculture settings, on the diversity of fish school behavior, providing additional evidence to better assess the impact of noise on fish ecology and welfare. Furthermore, this research expands the understanding by directly linking noise to the total feeding quantity of largemouth bass schools, providing concrete evidence for the reduction in feeding desire and intake caused by noise.

However, the study presents certain limitations. For instance, the sample size employed is relatively modest, potentially lacking comprehensive representation of the entire largemouth bass population. Furthermore, the experiment's duration is comparatively brief, and variations between the experimental conditions and actual production environments may impact the validity of the results. It is also important to acknowledge the influence of the 24-hour photoperiod with full-spectrum LED lights on fish behavior. While this lighting condition was chosen to ensure consistent visibility for our observations, it may have disrupted the natural diurnal rhythms of the tested fish. This is a notable limitation as it could have affected the baseline behavior of largemouth bass, potentially confounding the results. Additionally, the methodologies and instruments implemented for sound environment assessment may inherently possess constraints. While sound pressure was utilized to characterize the noise environment, measurement of the particle motion component remained beyond the scope of this study. Notwithstanding these limitations, the research elucidates the noteworthy impact of noise on the behavior and growth of

largemouth bass within the specific experimental framework. Such insights are crucial for understanding the repercussions of noise pollution on fish ecology, contributing to the enhancement of fish welfare and potentially influencing fish production. These findings also provide a foundation for future investigations into the impacts of noise on largemouth bass and other fish species.

## 5 Conclusion

This investigation has illuminated the influence of aeration equipment noise on the swimming behavior, feeding habits, and growth progression of largemouth bass populations. The results from the experimental trials indicate that largemouth bass residing in noisy environments, when contrasted with the ambient group, display substantial disparities in swimming behavior, characterized by larger average angles and distances from the nearest neighbor fish. Moreover, these bass experience diminished appetite, decreased food consumption, and compromised growth performance. Additionally, the present study has pioneered the use of the SWDI to appraise the diversity of fish swimming behavior, thereby offering a novel evaluation instrument for fish behavior studies. Going forward, it is crucial to delve deeper into the precise mechanisms by which noise impacts largemouth bass and other aquatic organisms, aiming to provide scientific substantiation for the mitigation and regulation of noise pollution. Simultaneously, these research findings assume paramount importance in addressing the challenges posed to fish welfare within aquaculture settings. Efforts to minimize noise pollution within aquacultural settings hold the potential to elevate farming efficiency, foster animal welfare, and establish a robust framework for the sustainable advancement of aquaculture practices.

## Data availability statement

The original contributions presented in the study are included in the article/supplementary material. Further inquiries can be directed to the corresponding author.

## Ethics statement

The animal study was approved by Experimental Animal Welfare Ethics Committee of Zhejiang University. The study was conducted in accordance with the local legislation and institutional requirements.

## Author contributions

YZ conceived the study, carried out field work, analyzed data, and wrote the majority of the manuscript. AS, SH, HZ carried out filed work. WX, JZ, SZ conceived the study, analyzed data and revised the manuscript. All authors contributed to the article and approved the submitted version.

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## References

- Bai, J., and Li, S. (2018). Development of largemouth bass (*Micropterus salmoides*) culture," in *Aquaculture in China*. Eds. J. Gui, Q. Tang, Z. Li, J. Liu and S. S. De Silva. doi: 10.1002/9781119120759.ch4\_5
- Banner, A., and Hyatt, M. (1973). Effects of noise on eggs and larvae of two estuarine fishes. *Trans. Am. Fisheries Soc.* 102 (1), 134–136. doi: 10.1577/1548-8659(1973)102<134:EONOE>2.0.CO;2
- Bart, A. N., Clark, J., Young, J., and Zohar, Y. (2001). Underwater ambient noise measurements in aquaculture systems: a survey. *Aquacultural Eng.* 25 (2), 99–110. doi: 10.1016/S0144-8609(01)00074-7
- Brumm, H., and Slabbekoorn, H. (2005). Acoustic communication in noise. *Adv. Study Behav.* 35, 151–209. doi: 10.1016/S0065-3454(05)35004-2
- Buck, B. H., Troell, M. F., Krause, G., Angel, D. L., Grote, B., and Chopin, T. (2018). State of the art and challenges for offshore integrated multi-trophic aquaculture (IMTA). *Front. Mar. Sci.* 5, 165. doi: 10.3389/fmars.2018.00165
- Canton, H. (2021). "Food and agriculture organization of the United Nations—FAO," in *The Europa directory of international organizations 2021* (Rome, Italy: Routledge), 297–305.
- Chan, A. A. Y. H., and Blumstein, D. T. (2011). Attention, noise, and implications for wildlife conservation and management. *Appl. Anim. Behav. Sci.* 131 (1–2), 1–7. doi: 10.1016/j.applanim.2011.01.007
- Charmandari, E., Tsigos, C., and Chrousos, G. (2005). Endocrinology of the stress response. *Annu. Rev. Physiol.* 67, 259–284. doi: 10.1146/annurev.physiol.67.040403.120816
- Chen, S. Y., Fei, Y. H. J., Chen, Y. C., Chi, K. J., and Yang, J. T. (2016). The swimming patterns and energy-saving mechanism revealed from three fish in a school. *Ocean Eng.* 122, 22–31. doi: 10.1016/j.oceaneng.2016.06.018
- Craven, A., Carton, A. G., McPherson, C. R., and McPherson, G. (2009). Determining and quantifying components of an aquaculture soundscape. *Aquacultural Eng.* 41 (3), 158–165. doi: 10.1016/j.aquaeng.2009.07.003
- Davidson, J., Bebak, J., and Mazik, P. (2009). The effects of aquaculture production noise on the growth, condition factor, feed conversion, and survival of rainbow trout, *Oncorhynchus mykiss*. *Aquaculture* 288 (3–4), 337–343. doi: 10.1016/j.aquaculture.2008.11.037
- De Kloet, E. R., Oitzl, M. S., and Joëls, M. (1999). Stress and cognition: are corticosteroids good or bad guys? *Trends Neurosci.* 22 (10), 422–426. doi: 10.1016/S0166-2236(99)01438-1
- De Vincenzi, G., Filiciotto, F., Maccarrone, V., Mazzola, S., and Buscaino, G. (2015). Behavioural responses of the European spiny lobster, *Palinurus elephas* (Fabricius), to conspecific and synthetic sounds. *Crustaceana* 88 (5), 523–540. doi: 10.1163/15685403-00003430
- Diana, J. S. (2009). Aquaculture production and biodiversity conservation. *Bioscience* 59 (1), 27–38. doi: 10.1525/bio.2009.59.1.7
- Dooling, R. J., Leek, M. R., and Popper, A. N. (2015). Effects of noise on fishes: What we can learn from humans and birds. *Integr. zoology* 10 (1), 29–37. doi: 10.1111/1749-4877.12094
- Duarte, M. H. L., Sousa-Lima, R. S., Young, R. J., Farina, A., Vasconcelos, M., Rodrigues, M., et al. (2015). The impact of noise from open-cast mining on Atlantic forest biophony. *Biol. Conserv.* 191, 623–631. doi: 10.1016/j.biocon.2015.08.006
- Duarte, R. H. L., de Oliveira Passos, M. F., Beirão, M. V., Midamegbe, A., Young, R. J., and de Azevedo, C. S. (2023). Noise interfere on feeding behaviour but not on food preference of saffron finches (*Sicalis flaveola*). *Behav. Processes* 206, 104844. doi: 10.1016/j.beproc.2023.104844
- Evans, J. C., Dall, S. R., and Kight, C. R. (2018). Effects of ambient noise on zebra finch vigilance and foraging efficiency. *PLoS One* 13 (12), e0209471. doi: 10.1371/journal.pone.0209471
- Gendron, G., Tremblay, R., Jolivet, A., Olivier, F., Chauvaud, L., Winkler, G., et al. (2020). Anthropogenic boat noise reduces feeding success in winter flounder larvae (*Pseudopleuronectes americanus*). *Environ. Biol. Fishes* 103, 1079–1090. doi: 10.1007/s10641-020-01005-3
- Gil, D., Honarmand, M., Pascual, J., Pérez-Mena, E., and Macías García, C. (2015). Birds living near airports advance their dawn chorus and reduce overlap with aircraft noise. *Behav. Ecol.* 26 (2), 435–443. doi: 10.1093/beheco/aru207
- Giordano, A., Hunninch, L., and Sheriff, M. J. (2022). Prey responses to predation risk under chronic road noise. *J. Zoology* 317 (2), 147–157. doi: 10.1111/jzo.12968
- Godfray, H. C. J., Beddington, J. R., Crute, I. R., Haddad, L., Lawrence, D., Muir, J. F., et al. (2010). Food security: the challenge of feeding 9 billion people. *science* 327 (5967), 812–818. doi: 10.1126/science.1185383
- Halvorsen, M. B., Casper, B. M., Matthews, F., Carlson, T. J., and Popper, A. N. (2012). Effects of exposure to pile-driving sounds on the lake sturgeon, Nile tilapia and hogchoker. *Proc. R. Soc. B: Biol. Sci.* 279 (1748), 4705–4714. doi: 10.1098/rspb.2012.1544
- Hanache, P., Spataro, T., Firmat, C., Boyer, N., Fonseca, P., and Médoc, V. (2020). Noise-induced reduction in the attack rate of a planktivorous freshwater fish revealed by functional response analysis. *Freshw. Biol.* 65 (1), 75–85. doi: 10.1111/fwb.13271
- Hang, S., Zhao, J., Ji, B., Li, H., Zhang, Y., Peng, Z., et al. (2021). Impact of underwater noise on the growth, physiology and behavior of *Micropterus salmoides* in industrial recirculating aquaculture systems. *Environ. pollut.* 291, 118152. doi: 10.1016/j.envpol.2021.118152
- Hawkins, A. D., and Chapman, C. J. (1975). Masked auditory thresholds in the cod, *Gadus morhua* L. *J. Comp. Physiol.* 103 (2), 209–226. doi: 10.1007/BF00617122
- He, K., Gkioxari, G., Dollár, P., and Girshick, R. (2020). R-CNN mask 2020. *IEEE Trans. Pattern Anal. Mach. Intell.* 42, 386–397. doi: 10.1109/TPAMI.2018.2844175
- Holles, S., Simpson, S. D., Radford, A. N., Berten, L., and Lecchini, D. (2013). Boat noise disrupts orientation behaviour in a coral reef fish. *Mar. Ecol. Prog. Ser.* 485, 295–300. doi: 10.3354/meps10346
- Jobling, M., Baardvik, B. M., Christiansen, J. S., and Jørgensen, E. H. (1993). The effects of prolonged exercise training on growth performance and production parameters in fish. *Aquaculture Int.* 1, 95–111. doi: 10.1007/BF00692614
- Keenleyside, M. H. (2012). *Diversity and adaptation in fish behaviour* Vol. 11 (Berlin, Germany: Springer Science & Business Media).
- Kent, M. I., Lukeman, R., Lizier, J. T., and Ward, A. J. (2019). Speed-mediated properties of schooling. *R. Soc. Open Sci.* 6 (2), 181482. doi: 10.1098/rsos.181482
- Kight, C. R., and Swaddle, J. P. (2011). How and why environmental noise impacts animals: an integrative, mechanistic review. *Ecol. Lett.* 14 (10), 1052–1061. doi: 10.1111/j.1461-0248.2011.01664.x
- Kunc, H. P., McLaughlin, K. E., and Schmidt, R. (2016). Aquatic noise pollution: implications for individuals, populations, and ecosystems. *Proc. R. Soc. B: Biol. Sci.* 283 (1836), 20160839. doi: 10.1098/rspb.2016.0839
- Lagardère, J. P. (1982). Effects of noise on growth and reproduction of Crangon crangon in rearing tanks. *Mar. Biol.* 71, 177–185. doi: 10.1007/BF00394627

- Lupien, S. J., and McEwen, B. S. (1997). The acute effects of corticosteroids on cognition: integration of animal and human model studies. *Brain Res. Rev.* 24 (1), 1–27. doi: 10.1016/S0165-0173(97)00004-0
- MacArthur, R. H., and Pianka, E. R. (1966). On optimal use of a patchy environment. *Am. Nat.* 100 (916), 603–609. doi: 10.1086/282454
- Mancera, K. F., Lisle, A., Allavena, R., and Phillips, C. J. (2017). The effects of mining machinery noise of different frequencies on the behaviour, faecal corticosterone and tissue morphology of wild mice (*Mus musculus*). *Appl. Anim. Behav. Sci.* 197, 81–89. doi: 10.1016/j.applanim.2017.08.008
- McLaughlin, K. E., and Kunc, H. P. (2015). Changes in the acoustic environment alter the foraging and sheltering behaviour of the cichlid *Amititlania nigrofasciata*. *Behav. processes* 116, 75–79. doi: 10.1016/j.beproc.2015.04.012
- Mendl, M. (1999). Performing under pressure: stress and cognitive function. *Appl. Anim. Behav. Sci.* 65 (3), 221–244. doi: 10.1016/S0168-1591(99)00088-X
- Mickle, M. F., Harris, C. M., Love, O. P., and Higgs, D. M. (2019). Behavioural and morphological changes in fish exposed to ecologically relevant boat noises. *Can. J. Fisheries Aquat. Sci.* 76 (10), 1845–1853. doi: 10.1139/cjfas-2018-0258
- Naylor, R. L., Goldburg, R. J., Primavera, J. H., Kautsky, N., Beveridge, M. C., Clay, J., et al. (2000). Effect of aquaculture on world fish supplies. *Nature* 405 (6790), 1017–1024. doi: 10.1038/35016500
- Naylor, R. L., Hardy, R. W., Buschmann, A. H., Bush, S. R., Cao, L., Klinger, D. H., et al. (2021). A 20-year retrospective review of global aquaculture. *Nature* 591 (7851), 551–563. doi: 10.1038/s41586-021-03308-6
- New, L. F., Clark, J. S., Costa, D. P., Fleishman, E., Hindell, M. A., Klanjšček, T., et al. (2014). Using short-term measures of behaviour to estimate long-term fitness of southern elephant seals. *Mar. Ecol. Prog. Ser.* 496, 99–108. doi: 10.3354/meps10547
- Popper, A. N. (2003). Effects of anthropogenic sounds on fishes. *Fisheries* 28 (10), 24–31. doi: 10.1577/1548-8446(2003)28[24:EOASOF]2.0.CO;2
- Popper, A. N., and Hawkins, A. D. (2019). An overview of fish bioacoustics and the impacts of anthropogenic sounds on fishes. *J. fish Biol.* 94 (5), 692–713. doi: 10.1111/jfb.13948
- Purser, J., and Radford, A. N. (2011). Acoustic noise induces attention shifts and reduces foraging performance in three-spined sticklebacks (*Gasterosteus aculeatus*). *PLoS One* 6 (2), e17478. doi: 10.1371/journal.pone.0017478
- Rees, S. G. (2000). Can a minority of informed leaders determine the foraging movements of a fish shoal? *Anim. Behav.* 59 (2), 403–409. doi: 10.1006/anbe.1999.1314
- Sabet, S. S., Wesdorp, K., Campbell, J., Snelderwaard, P., and Slabbekoorn, H. (2016). Behavioural responses to sound exposure in captivity by two fish species with different hearing ability. *Anim. Behav.* 116, 1–11. doi: 10.1016/j.anbehav.2016.03.027
- Sarà, G., Dean, J. M., d'Amato, D., Buscaino, G., Oliveri, A., Genovese, S., et al. (2007). Effect of boat noise on the behaviour of bluefin tuna *Thunnus thynnus* in the Mediterranean Sea. *Mar. Ecol. Prog. Ser.* 331, 243–253. doi: 10.3354/meps331243
- Schaub, A., Ostwald, J., and Siemers, B. M. (2008). Foraging bats avoid noise. *J. Exp. Biol.* 211 (19), 3174–3180. doi: 10.1242/jeb.022863
- Sherwen, S. L., and Hemsworth, P. H. (2019). The visitor effect on zoo animals: Implications and opportunities for zoo animal welfare. *Animals* 9 (6), 366. doi: 10.3390/ani9060366
- Siemers, B. M., and Schaub, A. (2011). Hunting at the highway: traffic noise reduces foraging efficiency in acoustic predators. *Proc. R. Soc. B: Biol. Sci.* 278 (1712), 1646–1652. doi: 10.1098/rspb.2010.2262
- Simpson, S. D., Radford, A. N., Nedelec, S. L., Ferrari, M. C., Chivers, D. P., McCormick, M. I., et al. (2016). Anthropogenic noise increases fish mortality by predation. *Nat. Commun.* 7 (1), 10544. doi: 10.1038/ncomms10544
- Slabbekoorn, H., Bouton, N., van Opzeeland, I., Coers, A., ten Cate, C., and Popper, A. N. (2010). A noisy spring: the impact of globally rising underwater sound levels on fish. *Trends Ecol. Evol.* 25 (7), 419–427. doi: 10.1016/j.tree.2010.04.005
- Slabbekoorn, H., Dalen, J., de Haan, D., Winter, H. V., Radford, C., Ainslie, M. A., et al. (2019). Population-level consequences of seismic surveys on fishes: An interdisciplinary challenge. *Fish Fisheries* 20 (4), 653–685. doi: 10.1111/faf.12367
- Smith, M. E., Kane, A. S., and Popper, A. N. (2004). Noise-induced stress response and hearing loss in goldfish (*Carassius auratus*). *J. Exp. Biol.* 207 (3), 427–435. doi: 10.1242/jeb.00755
- Spellerberg, I. F., and Fedor, P. J. (2003). A tribute to claude shanno–2001 and a plea for more rigorous use of species richness, species diversity and the 'Shannon–Wiener' Index. *Global Ecol. biogeography* 12 (3), 177–179. doi: 10.1046/j.1466-822X.2003.00015.x
- Subasinghe, R., Soto, D., and Jia, J. (2009). Global aquaculture and its role in sustainable development. *Rev. aquaculture* 1 (1), 2–9. doi: 10.1111/j.1753-5131.2008.01002.x
- Tang, Q., and Liu, H. (2018). Development strategies and prospects – driving forces and sustainable development of Chinese aquaculture," in *Aquaculture in China*. Eds. J. Gui, Q. Tang, Z. Li, J. Liu and S. S. De Silva. doi: 10.1002/9781119120759.ch8\_1
- Thukral, A. K. (2017). A review on measurement of Alpha diversity in biology. *Agric. Res. J.* 54 (1), 1–10. doi: 10.5958/2395-146X.2017.00001.1
- Todd, V. L., Williamson, L. D., Jiang, J., Cox, S. E., Todd, I. B., and Ruffert, M. (2021). Prediction of marine mammal auditory-impact risk from Acoustic Deterrent Devices used in Scottish aquaculture. *Mar. pollut. Bull.* 165, 112171. doi: 10.1016/j.marpolbul.2021.112171
- Voellmy, I. K., Purser, J., Flynn, D., Kennedy, P., Simpson, S. D., and Radford, A. N. (2014). Acoustic noise reduces foraging success in two sympatric fish species via different mechanisms. *Anim. Behav.* 89, 191–198. doi: 10.1016/j.anbehav.2013.12.029
- Wale, M. A., Simpson, S. D., and Radford, A. N. (2013a). Size-dependent physiological responses of shore crabs to single and repeated playback of ship noise. *Biol. Lett.* 9 (2), 20121194. doi: 10.1098/rsbl.2012.1194
- Wale, M. A., Simpson, S. D., and Radford, A. N. (2013b). Noise negatively affects foraging and antipredator behaviour in shore crabs. *Anim. Behav.* 86 (1), 111–118. doi: 10.1016/j.anbehav.2013.05.001
- Ward, A. J., Webster, M. M., and Hart, P. J. (2006). Intraspecific food competition in fishes. *Fish Fisheries* 7 (4), 231–261. doi: 10.1111/j.1467-2979.2006.00224.x
- Webb, P. W. (1984). Body form, locomotion and foraging in aquatic vertebrates. *Am. zoologist* 24 (1), 107–120. doi: 10.1093/icb/24.1.107
- Wei, D., Bao, E., Wen, Y., Zhu, S., Ye, Z., and Zhao, J. (2021). Behavioral spatial-temporal characteristics-based appetite assessment for fish school in recirculating aquaculture systems. *Aquaculture* 545, 737215. doi: 10.1016/j.aquaculture.2021.737215
- Williams, R., Wright, A. J., Ashe, E., Blight, L. K., Brintjes, R., Canessa, R., et al. (2015). Impacts of anthropogenic noise on marine life: Publication patterns, new discoveries, and future directions in research and management. *Ocean Coast. Manage.* 115, 17–24. doi: 10.1016/j.ocecoaman.2015.05.021
- Wysocki, L. E., Davidson, J. W. III, Smith, M. E., Frankel, A. S., Ellison, W. T., Mazik, P. M., et al. (2007). Effects of aquaculture production noise on hearing, growth, and disease resistance of rainbow trout *Oncorhynchus mykiss*. *Aquaculture* 272 (1–4), 687–697. doi: 10.1016/j.aquaculture.2007.07.225
- Zhang, X., Guo, H., Chen, J., Song, J., Xu, K., Lin, J., et al. (2021). Potential effects of underwater noise from wind turbines on the marbled rockfish (*Sebastes marmoratus*). *J. Appl. Ichthyology* 37 (4), 514–522. doi: 10.1111/jai.14198
- Zhang, X., Zhou, J., Xu, W., Zhan, W., Zou, H., and Lin, J. (2022). Transcriptomic and behavioral studies of small yellow croaker (*Larimichthys polyactis*) in response to noise exposure. *Animals* 12 (16), 2061. doi: 10.3390/ani12162061





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# Experimental study on the effect of sound stimulation on hearing and behavior of juvenile black rockfish (*Sebastes schlegelii*)

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Assessing the potential impacts of wind farm noise on fish is a crucial aspect of Environmental Impact Assessment (EIA) studies. There is increasing evidence of disturbances and effects on hearing and behavior in animals. The black rockfish (*Sebastes schlegelii*) is a commercially valuable rocky reef fish native to East Asia. However, empirical studies that measure the actual consequences are lacking. In this study, we used auditory evoked potentials (AEP) to assess the effects of dominant frequency noise emitted by offshore wind farms on the auditory sensitivity, hearing threshold, swimming, and feeding behavior of juvenile black rockfish. The experimental findings revealed that the most sensitive sound frequency was 200 Hz, with the lowest hearing threshold recorded at  $86.4 \pm 3.4$  dB re 1  $\mu$ Pa. Following 3 and 7 days of exposure to 200 Hz noise at 110 dB, threshold shifts in black rockfish reached 19.0 dB and 13.3 dB, respectively. During the subsequent recovery phase, these shifts decreased to approximately 9.8 dB after 3 days, respectively. The noise-exposed group exhibited higher swimming duration, moving distance, and caudal fin swing frequency compared to the control group without noise exposure. Furthermore, noise prolonged the feeding rate of black rockfish. Our findings provide the first evidence of noise-induced temporary threshold shift and behavioral disturbances in juvenile black rockfish, implying potential fitness consequences associated with noise pollutant.

## KEYWORDS

hearing sensitivity, auditory evoked potential, temporary threshold shift, behavior, fish welfare

## 1 Introduction

With the increasing energy demand and the pursuit of sustainable development in society, the offshore wind industry has experienced significant growth worldwide in recent decades (Slabbekoorn et al., 2010). Many countries are now focusing on utilizing offshore wind farm (OWF) facilities to operate aquaculture systems, aiming to enhance ocean space utilization and reduce costs. The concept of combining OWFs as fixation points for



aquaculture or co-using OWF sites by installing aquaculture farms between wind turbines has garnered considerable attention in recent years (Lindell, 2003; Wever et al., 2015; Tullio et al., 2018). However, the noise generated by OWFs primarily arises during the construction and operation stages (Nedwell et al., 2003; De Jong et al., 2011). Based on extensive measurements reviewed by Madsen et al. (2006), the underwater noise produced by operating wind turbines is predominantly limited to low frequencies (below 1 kHz) and low intensity. For instance, the frequency peak observed at the Vindeby OWF in Denmark and the Gotland OWF in Sweden was 25 Hz and 160 Hz, respectively (Nedwell, Langworthy and Howell, 2004). Similarly, the Horns Rev OWF in Denmark exhibited frequency peaks at 150 Hz and 300 Hz (Betke et al., 2005). Additionally, measurements conducted at three different types of OWFs in Denmark and Gotland (Middelgrund, Vindeby, and Bockstigen-Valar) revealed a noise frequency range below 500 Hz (Tougaard et al., 2009). Blew et al. (2008) summarized the frequency peaks of four OWFs in Denmark, which were 176 Hz, 150 Hz, 135 Hz, and 134 Hz, respectively, with corresponding sound pressure levels (SPLs) of 114 dB, 117 dB, 110 dB, and 122 dB. In the case of the East China Sea Bridge OWF, the noise frequency range was below 400 Hz, and the SPLs ranged from 81 dB to 99 dB (Zhang et al., 2016). Tougaard et al. (2020) compiled data from 46 measured values obtained from 14 wind farms. Although the distances and wind speeds of the measurements varied, the dominant frequencies were found to be in the range of 10–400 Hz, with equivalent continuous sound levels (LAeq) ranging from 80 dB to 135 dB re 1  $\mu$ Pa.

Previous studies have demonstrated the influence of disturbances on fish behavior, including responses to noise (Stevens and Don, 1979; Russell et al., 2001; Sara et al., 2007; Ladich and Fay, 2013; Andrew et al., 2016; Velasquez et al., 2020; Jones et al., 2020). Andersson et al. (2007) conducted a study on two different fish species, *Rutilus rutilus* L and *Gasterosteus aculeatus*, exposed to single-tone frequencies and sound generated by offshore wind turbines. They observed various swimming patterns in sticklebacks (*Gasterosteus aculeatus*), including forward swimming, twitching, backing, and freezing. Previous studies have indicated that prolonged exposure to noise can lead to damage or loss of hair cells in the fish's inner ear. Severe auditory insults can result in temporary threshold shifts (TTS) or permanent threshold shifts (PTS). In a study by Popper et al. (1976), goldfish were exposed to pure tones, and sound pressure levels (SPLs) of 149 dB re 1  $\mu$ Pa caused threshold shifts of approximately 7–9 dB at 500 Hz and 18–27 dB at 800 Hz. Amoser and Ladich (2003) exposed goldfish to white noise at 158 dB re 1  $\mu$ Pa for 24 hours and found the greatest hearing loss at 800 Hz and 1000 Hz. Smith et al. (2004) reported significant threshold shifts (up to 28 dB) at all frequencies in goldfish due to noise exposure, with larger shifts occurring at frequencies where their hearing sensitivity is highest. The negative impact of OWF (OWF) noise on fish has become a matter of concern. Understanding the auditory threshold shift and recovery of fish is crucial (Amoser and Ladich, 2003). Striking a balance between energy development and the preservation of aquatic life is imperative (Lacroix and Sylvain, 2011; Thompson et al., 2020).

In this study, we focused on the black rockfish (*Sebastes schlegelii*) as our research subject. The black rockfish belongs to

the *Scorpaenidae* family and is an economically important marine ovoviviparous teleost species. It is widely distributed and cultured in Japan, Korea, and the northeast coast of China (Likang et al., 2018). The objectives of our study were as follows: (i) To determine the sensitive sound frequency and hearing threshold of the black rockfish using the auditory evoked potential (AEP) method. (ii) To assess the degree of hearing loss (temporary threshold shift or TTS) in the black rockfish. (iii) To investigate the impact of noise from OWFs at prominent frequency ranges on fish behavior. The findings of this study are expected to provide valuable insights and serve as a reference for the development of OWFs, taking into consideration the effects on black rockfish and other similar species.

## 2 Methods

### 2.1 Study species

Juvenile black rockfish ( $n=80$ ; standard length:  $6.99 \pm 0.51$  cm; wet mass:  $19.47 \pm 2.93$  g) were studied at the Fish Behavior Laboratory at the Dalian Ocean University. Before the experiment, the fish were farmed in a polypropylene water bucket (diameter 52 cm, water depth 46 cm) for 5 days. Water temperature was  $21.4\text{--}22.8^\circ\text{C}$  and dissolved oxygen was  $7.89 \pm 0.03$  mg/L. Fish were fed once daily at noon and 50% of the water was changed after 2 days to clean out excrement. We present the feeding amount based on the experience in the rearing phase. The total quantity fed daily was 1.1% fish body weight (1.7 g/day).

### 2.2 Auditory evoked potential recordings of black rockfish

The auditory evoked potential (AEP) test conducted in this study was based on the method described by Kenyon et al. (1998). The experimental setup is illustrated in Figure 1. To minimize external noise interference, all equipment was placed inside a soundproof room with an internal sound pressure level measuring 35.2 dB. The testing was carried out in a cylindrical Polyvinyl Chloride (PVC) water bucket with an inner diameter of 0.4 m and a height of 0.4 m. The bucket was positioned on a shockproof table, and an aluminum radiation-proof cloth was applied to the outer wall to prevent noise disturbances. Juvenile black rockfish were first anesthetized and then securely positioned at the center of the tank. The water level was maintained at approximately 3 mm below the fish's skull. Sound stimuli were generated using an underwater speaker (UW-30, China). The generated sound was pre-amplified with a Crown amplifier (D-75A, China) and analyzed in real-time using a Tucker-Davis Technologies (TDT) system (RZ6, USA) equipped with a hydrophone probe mounted near the fish's head. The hydrophone probe was connected to an acoustic acquisition system (Aquafeeler IV, Roland, Japan) used for online measurements. The hydrophone probe had a sensitivity of  $-190$  dB re 1  $\mu$ Pa and a bandwidth of 20 Hz to 200 kHz. Ambient noise and stimulus noise were recorded during the experiments. AEP waveforms were generated using the SigGenRZ software.

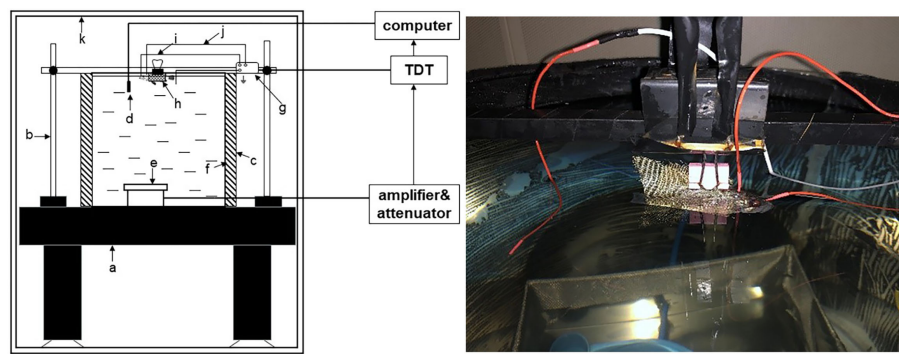


FIGURE 1

Test installation diagram of AEP. a, holographic-table; b, fixed frame; c, acoustic insulation material foam; d, hydrophone; e, underwater speaker; f, water tank; g, preamplifier; h, fish; i, reference electrode; j, recording electrode; k, sound-proof chamber.

Each experimental case in this study involved the use of a single live fish and lasted for a duration of 2 hours. A total of 8 cases were conducted, but unfortunately, one of them resulted in the death of the fish. Prior to any manipulation, the fish were anesthetized by immersing them in water containing 1:14000 tricaine methanesulfonate (MS-222) until their movements ceased, which typically took around 50 seconds. To minimize myogenic noise levels, the fish were temporarily immobilized by injecting gallamine triiodide (Sigma-Hefei Bomeranian Biology) at a dosage of 1.2–1.4  $\mu\text{g/g}$  per body weight. The anesthesia lasted for 4 to 5 hours. For the AEP recordings, two silver needles with a length of 5 mm and a diameter of 0.25 mm were used as electrodes. One needle was inserted through the fish's skull at the medulla region and served as the recording electrode, while a reference electrode was placed subcutaneously between the fish's eyes (Codarin and Wysocki (2009)). Two additional electrodes were inserted subcutaneously at a depth of 2.5 mm. Shielded electrode leads were connected to the preamplifier's input, and the other end of the shielded electrode was positioned posterior to the fish's body. The reference, ground, and recording electrodes were connected to the amplifier. The sound stimuli used in the experiment were short alternating tone bursts with frequencies of 100 Hz, 200 Hz, 300 Hz, and 500 Hz, which were selected based on the dominant frequency of offshore wind turbine noise. The tone bursts were presented at a repetition rate of 10/s and alternated between  $90^\circ$  and  $270^\circ$  phases. The duration of the sound stimuli ranged from 2 cycles at 100 Hz and 200 Hz to 5 cycles at 300 Hz and 500 Hz. The rise and fall times of the stimuli increased from 1 cycle at 100 Hz and 200 Hz to 2 cycles at 300 Hz and 500 Hz. A total of 1000 stimuli were presented for each polarity. The lowest sound pressure level (SPL) at which a repeatable AEP trace could be obtained was considered the threshold. The SPLs were attenuated in 5-dB steps initially. As the threshold was approached, the SPLs were further attenuated in 1–2 dB steps until the threshold level was determined.

### 2.3 Audiometry test under noise exposure

Sixteen juvenile black rockfish were selected for this study and divided into two groups: a noise-exposed group and a controlled and sheltered group, as depicted in Figure 2. Based on previous

research on the dominant noise of single-pile foundation of OWFs, the stimuli used for noise exposure in this study consisted of 200 Hz pure tones at a sound pressure level of 110 dB (Betke et al., 2005; Tougaard et al., 2009; Marmo et al., 2014). The duration of noise exposure for each treatment group was set at 7 days. AEP images were collected separately at two time points: 3 days and 7 days after the initiation of noise exposure for both groups. Additionally, images were recorded 1 and 3 days after the recovery period.

### 2.4 Behavior test under noise exposure

The noise exposure group of black rockfish was subjected to continuous exposure to a pure tone of 200 Hz at an intensity of 110 dB for a period of 7 days. Throughout this period, video recordings capturing the behavior of the fish from a bird's-eye view were conducted at three specific time points each day: 7:00, 13:00, and 21:00. The duration of each video recording was 5 minutes, and the middle 3 minutes were selected for behavior analysis. Specifically, after the swimming video recording at 13:00 each day, additional videos capturing the feeding behavior of both the noise exposure group and the controlled group were recorded. To minimize anthropogenic disturbances, baits were introduced into the tank using a tube rather than directly by hand. To analyze the potential interaction between tailbeat frequencies (TBF) and swimming time (ST), the analysis of locomotor behavior of the fish was conducted from a side view perspective. Additionally, the feeding time and feed rate of the bottom bait were recorded from each feeding behavior video. The quantity of food provided during the test phase was the same as that during the rearing phase. All the criteria used for recording and analysis are summarized in Table 1.

### 2.5 Correlation parameter and data processing

The behavioral data collected were analyzed to determine their normal distribution using the Shapiro-Wilk normality test, which was conducted using SPSS Statistics (Version 17.0). Since the data did not follow a normal distribution, non-parametric tests,

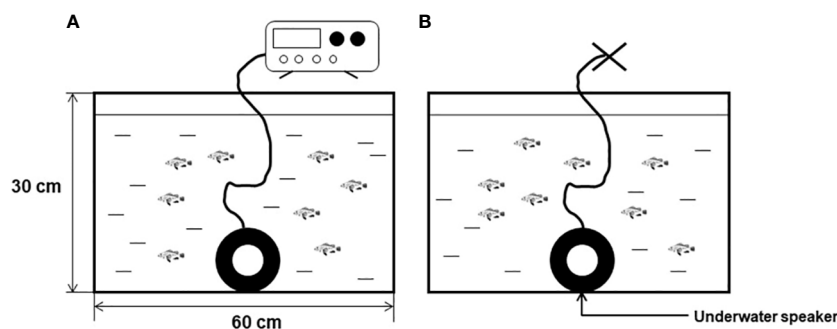


FIGURE 2  
Acoustic exposure behavior test diagram (A) for the acoustic exposure group, (B) for the control group).

specifically the Mann-Whitney test, were performed to compare the differences between the control group and the noise group. The results of the analysis were expressed as the median values with the interquartile range (IQR). To visualize the findings, plots were generated using GraphPad Prism 8 software. The activity behaviors of the black rockfish were tracked and recorded using video footage. The motion trajectory of each individual fish was determined, and the distances traveled were measured using the Digimizer software system. This software facilitated the analysis of the recorded videos to extract relevant behavioral parameters.

## 2.6 Ethical note

All experimental procedures conducted in this study adhered to the guidelines set forth by the Institutional Animal Care and Use Committee (IACUC). The study protocol, including the ethical considerations and animal welfare practices, was reviewed and approved by the Animal Welfare and Ethical Review Committee of Ocean University of China (IACUC—2105003). Throughout the study, all relevant international, national, and institutional guidelines and regulations regarding the care and use of animals were strictly followed to ensure the ethical treatment and well-being of the black rockfish used in the experiment. These measures were implemented to guarantee the humane and responsible conduct of the research.

TABLE 1 Parameter definition used in the collection of behavioral data from black rockfish exposed or not to sound.

Parameter name	Detailed description
Swimming distance	The horizontal distance from the beginning of a fish's displacement to the end of its displacement
Tailbeat frequency	The movement from one extreme lateral position to the opposite extreme lateral position.
Swimming time	The time from the start of swimming until the time when the fish became stationary.
Feeding time	The time from the first fish starts feeding until the left bait is gone.
Feed rate of bottom bait	The number of tails feeding on the bottom bait divided by the total tails

## 3 Results

### 3.1 Sensitive sound frequency and hearing threshold

AEP (auditory evoked potential) waveforms were successfully obtained from 7 black rockfish that were tested in the study. Figure 3 provides an overview of the AEP waveforms specifically generated in response to a 100 Hz sound stimulus. It can be observed that as the sound intensity decreased, the magnitude of the AEP response also declined. Through repeated testing, it was determined that the average number of wave peaks in the AEP waveforms between sound pressure levels (SPLs) of 105 dB and 93 dB was approximately 8. However, when the SPL was further reduced to 92 dB, recognizable and repeatable waveforms could no longer be produced. This suggests that the AEP threshold for the black rockfish at a frequency of 100 Hz is 92 dB, indicating that the fish's auditory response becomes undetectable below this sound pressure level.

Audiograms were recorded using the AEP instrument, and the results are presented in Figure 4. It is important to note that the minimum threshold values obtained from the audiograms were considerably higher than the measured background noise level. This indicates that the audiogram data were not influenced by the background noise. The audiograms reveal a decreasing trend in auditory thresholds within the frequency range of 100–200 Hz, indicating that the black rockfish exhibited relatively lower thresholds for sound stimuli in this frequency range. However, beyond 200 Hz and up to 500 Hz, the thresholds showed a tendency to rise, implying that the fish had higher auditory thresholds in this frequency range. Of particular interest is the lowest threshold observed at 200 Hz, which was measured to be  $86.4 \pm 3.4$  dB.

### 3.2 Effect of 200 Hz noise on the hearing threshold

The degree of hearing loss was quantified by measuring noise-induced threshold shifts during sound stimulation at 110 dB SPL and 200 Hz, as shown in Figure 5. After 3 days of noise exposure, the average auditory thresholds increased by  $19.0 \pm 8.4$  dB. The maximum threshold shift observed was 27.4 dB at 100 Hz. After 7

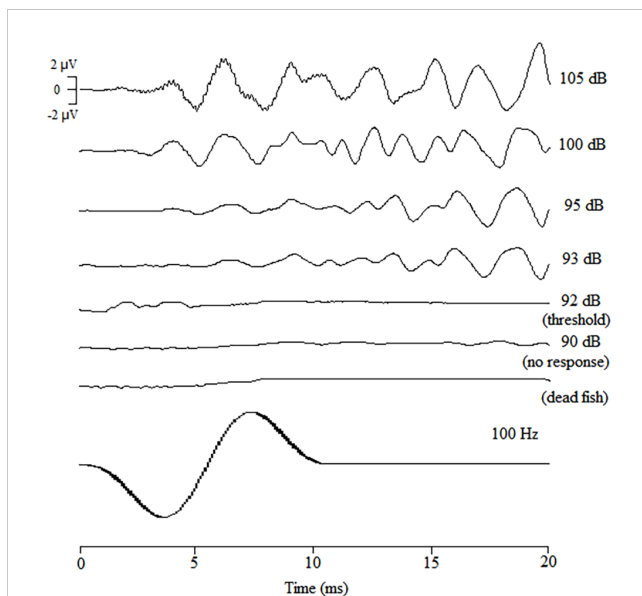


FIGURE 3  
AEP waveforms of black rockfish obtained in response to tone bursts of 200 Hz attenuated from 85 to 100 dB (1000 times repeated).

days of noise exposure, the average threshold shift was  $13.3 \pm 8.3$  dB, with a maximum shift of 21.6 dB at 200 Hz. During the recovery period, the auditory thresholds gradually improved. After the third day of recovery, the decrease was  $9.8 \pm 5.6$  dB. The maximum threshold shifts during recovery was 18.6 dB at 200 Hz, respectively. Overall, the degree of threshold shift decreased with increasing frequency and increased exposure time. However, except for the 3-day exposure group ( $p=0.019$ ), there were no significant differences between the control group and the noise-exposed group ( $p>0.05$ ). The greatest threshold shifts were observed at 100 Hz (27.4 dB) and 200 Hz (26.1 dB) after 3 days of noise exposure. However, hearing thresholds for all frequencies were reduced after 7 days of noise

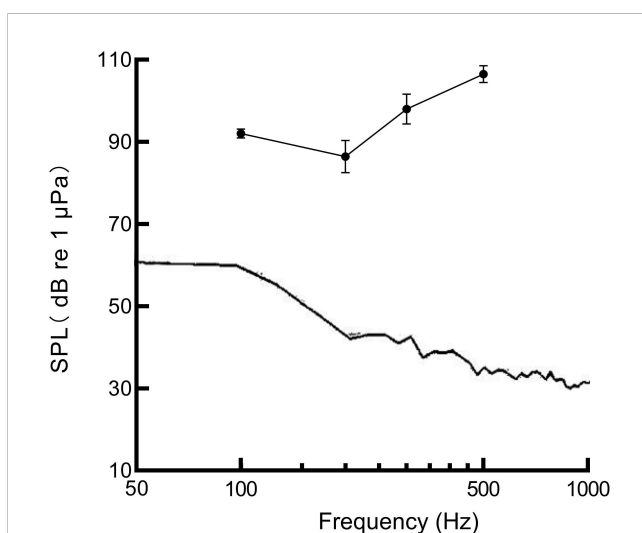


FIGURE 4  
AEP audiogram of black rockfish (Solid line is background noise,  $n=7$ ).

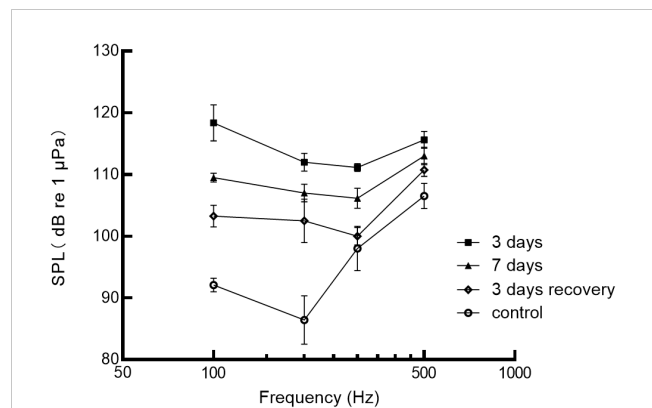


FIGURE 5  
Auditory sensitivity of the black rockfish before (control) and after exposure to 200 Hz noise for 3 or 7 days at 110 dB and 3 days of recovery.

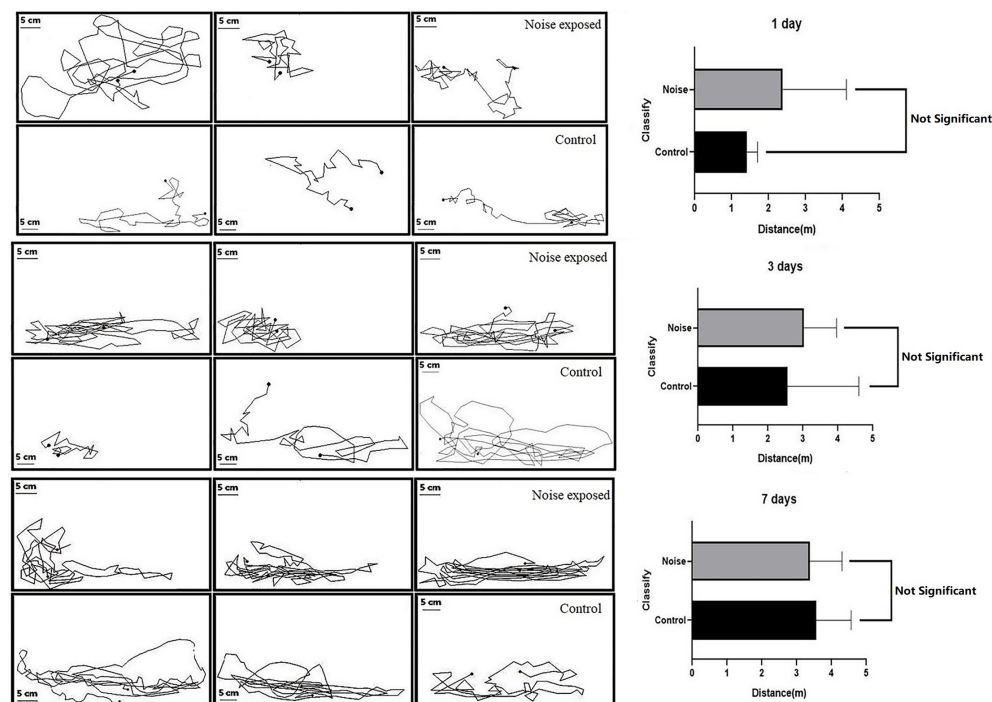
exposure compared to 3 days of exposure. Furthermore, even after 3 days of recovery, the auditory thresholds of the black rockfish did not return to the control levels. These findings suggest that noise exposure can lead to significant threshold shifts in the black rockfish, with the extent of the shifts depending on the frequency and duration of exposure. After 3 days, 7 days of noise exposure, and 3 days of recovery, the best frequency of black rockfish changed from 200 to 300 Hz.

### 3.3 Effects of 200 Hz noise on behavior

#### 3.3.1 Swimming distance

The effect of noise stress on the swimming behavior of black rockfish was examined, and the results are presented in Figure 6. It was observed that the fish in the test group exhibited startle responses and even jumped out of the water during the initial noise stress. The motion tracks of the black rockfish during a 3-minute observation period are depicted. The distribution of the fish was mainly on the sidewalls of the tank, as shown in the figures. The average horizontal movement distance for the test group on day 1 was  $2.38 \pm 1.00$  m, while for the control group it was  $1.42 \pm 0.16$  m. On day 3, the distance for the test group increased to  $3.04 \pm 0.54$  m, and for the control group, it was  $2.56 \pm 1.18$  m. On day 7, the distance for the test group further increased to  $3.38 \pm 0.53$  m, and for the control group, it was  $3.56 \pm 0.58$  m. Initially, the distance moved by the test group was greater than that of the control group, indicating a more active swimming response to the noise stress. However, as the exposure time increased, the gap between the two groups gradually closed. By day 7, the test group's distance reached a constant level similar to that of the control group, and there was no significant difference between the two groups ( $p>0.05$ ). Based on the visualization of the swimming trend, it can be concluded that black rockfish are capable of habituating to noise over a relatively short period, at least in terms of routine swimming behavior. The average distance moved by both the control and experimental groups became similar after 7 days of noise exposure, suggesting a habituation response to the noise stress.





**FIGURE 6**  
Trajectory diagram (left) and movement distance (right) of black rockfish after exposure to 200 Hz noise (110 dB) for 1, 3, and 7 days, 3 min each time ( $p > 0.05$ ).

### 3.3.2 Tailbeat frequencies and swimming time

The average tail beat frequency and swimming duration of the test group and control group are presented in Figure 7 and Table 2. The interquartile range of the tail beat frequency in the test group were 246 times, while in the control group, they were 159 times. The average tail beat frequency in the test group was significantly higher than that in the control group (82 times/min vs 53 times/min,  $p = 0.001$ ). There was a significant difference observed between day 4 ( $p = 0.007$ ) and day 7 ( $p = 0.015$ ) in both the control and test groups. Additionally, the interquartile range of the swimming time in the test group were 60.6 s, while in the control group, they were 45.9 s. The average swimming duration of the test group was higher than that of the control group (33.67% vs 25.50%,  $p < 0.05$ ). This difference was found to be significant on day 1, day 2, and day 5 compared to the control group ( $p < 0.05$ ). These results indicate that the fish in the test group exhibited increased tail beat frequency and longer swimming durations compared to the control group, indicating a heightened level of activity and possibly stress response to the noise exposure.

### 3.3.3 Feeding behavior

Based on video observation, it was observed that the time taken for the first fish to locate the bait after feeding was approximately 1–2 seconds in both the control and test groups. However, except for the feeding duration on day 1 and day 7 of noise exposure, the control group exhibited shorter feeding times compared to the test group (Figure 8). The total feeding time in the test group was longer than that in the control group (62 seconds/day vs 43 seconds/day).

The noise stimulation had a specific impact on the feeding time of the black rockfish, although the difference was not statistically significant ( $p > 0.05$ ). In the test group, as the bait substrates were not completely consumed, it extended the feeding time significantly. During the recovery stage, only one fish in the experimental group was able to locate the bottom bait on the first day, and from the second day onwards, none of the fish in the experimental group were able to see the bait at the bottom of the tank, resulting in the feeding time being unable to be recorded. In contrast, the control group was able to consume all the bait within 3 days, and there was a significant difference between the control and test groups ( $p < 0.01$ ). These findings suggest that the noise exposure affected the feeding behavior of the black rockfish, leading to prolonged feeding time and difficulties in locating the bait during the recovery period.

## 4 Discussion

In this study, the best hearing sensitivity of black rockfish was observed at 200 Hz, with a mean auditory threshold of  $86.4 \pm 3.4$  dB re 1  $\mu$ Pa. Figure 9 compares three different approaches (behavior, electrocardiogram ECG, auditory evoked potentials AEP) for determining auditory thresholds of black rockfish (Ishizaki et al., 1992; Keiichiro, 1997). The general trends of the three methods were found to be similar. At 100 Hz, the auditory threshold was 92.1 dB, which was 0.1 dB higher than the ECG method and 1.5 dB higher than the behavioral method. At 200 Hz, the auditory



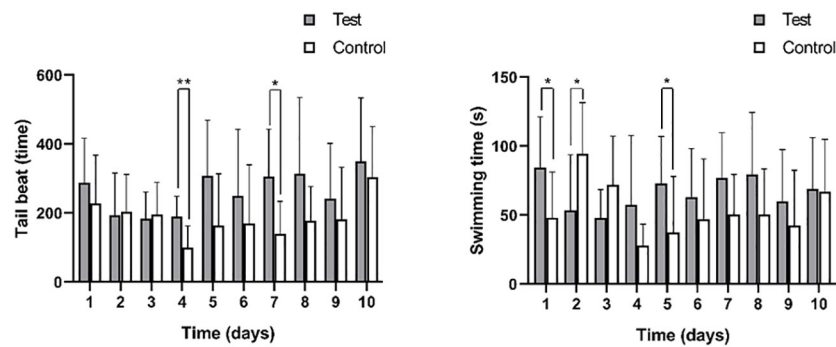


FIGURE 7

The tail beat times (left) and swimming time (right) of black rockfish were measured 3 min each time after exposure to 200 Hz noise (110 dB) for 11 days (7 days acoustic stress+3 days recovery). \* p value < 0.05, \*\* p value < 0.01.

threshold was 86.4 dB, which was 4.1 dB lower than the ECG method and 6 dB lower than the behavioral method. At 300 Hz, the auditory threshold was 98 dB, which was 4 dB lower than the ECG method and 8.4 dB lower than the behavioral method. At 500 Hz, the auditory threshold was 106 dB, which was 14.3 dB lower than the ECG method and 10.6 dB lower than the behavioral method. AEP detects synchronous neural activity in the 8<sup>th</sup> cranial nerve and brainstem auditory nuclei elicited by sound at the surface of the skull. One advantage of the AEP method over behavioral or ECG methods is that it is not affected by the condition or feeding motivation of the fish. The speed and ease with which AEP audiometry can be performed and the fact that there is no need for lengthy and repetitive subject conditioning, make it applicable to studies of hearing sensitivity in fish (Kojima et al., 2005; Ladich and Fay, 2013).

This study provides the first evidence of Noise-Induced Hearing Loss in black rockfish based on changes in auditory sensitivity. After 3 days of noise exposure, the black rockfish exhibited a significant increase in TTS, indicating a temporary shift in their hearing thresholds. However, after 7 days, this shift started to decrease, suggesting that as the duration of noise exposure extends, black rockfish exhibit a potential gradual adaptation to the deleterious effects of noise. After 3 days of noise exposure, the best frequency of black rockfish changed from 200 to 300 Hz. Similar findings have been reported in other studies. For example, Scholik and Yan (2002) examined the effects of white noise and boat noise on *Pimephales promelas* and found a change in the best frequency from 1000 to 300 Hz. Liu et al. (2013) exposed *Myxocyprinus asiaticus* to ship noise and observed a change in the best frequency from 800 to 200 Hz. These differences in findings may be attributed to variations in noise SPL, bandwidth, and species-specific characteristics. In general, fish

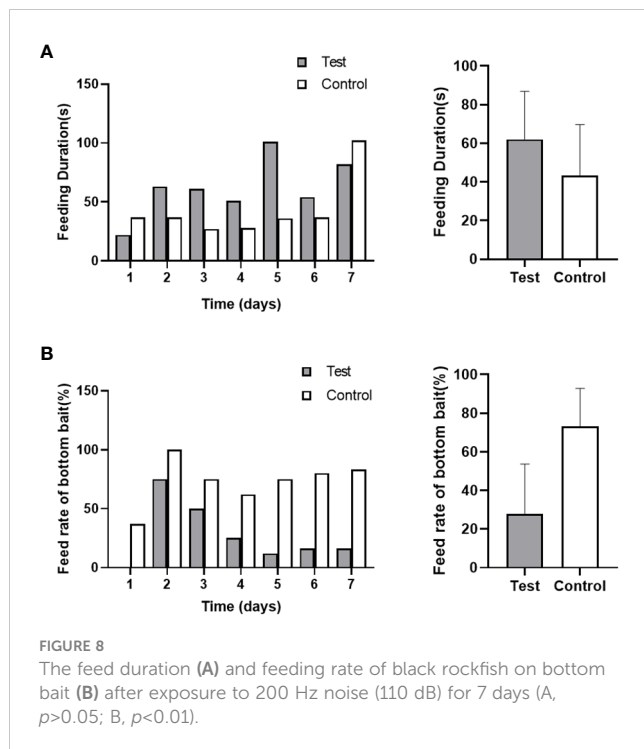
species with limited auditory sensitivity (without Weberian apparatus) are less affected by noise. The auditory threshold tends to decrease as the duration of the recovery period increases. Breitler et al. (2020) measured zebrafish exposed to white noise at different SPLs and found that recovery function occurred within 7 days for fish exposed to 130 dB and 140 dB noise levels, while fish subject to 150 dB only returned to baseline thresholds after 14 days. Further research is needed to determine the time required for the auditory threshold of black rockfish to recover to baseline levels.

Fish behavior research plays a crucial role in noise studies, as fish exhibit various behavioral responses to anthropogenic disturbances in aquatic environments. Changes in swimming activities, tail movements, and foraging behavior have been used as indicators of stress in fish exposed to human-generated disturbances (Mickle and Higgs, 2017; Weilgart, 2018; Faria et al., 2019). Based on the trajectories of black rockfish observed in this study, it was concluded that the fish spent more time irregularly swimming on the sidewalls of the tank, both with and without exposure to 200 Hz noise. Black rockfish are demersal marine fish that typically live in groups on rocky reefs, and their behavior may be influenced by their natural habitat. The onset of underwater sound propagation elicits startled reactions in some juvenile black rockfish, consistent with the findings reported by Spiga et al. (2017). The tail beat frequency (TBF) and swimming time (ST) of the control group were higher than those of the treatment group on days 2 and 3. From the video observations, it appeared that black rockfish tended to remain still and seek cover for extended periods when under constant threat. We hypothesize that fish may undergo a transition from an initial phase to an adaptive phase. During the initial phase, it is plausible that fish would employ strategies to minimize their interactions with noise stimulation, potentially

TABLE 2 Total tail beat times and total swimming time of black rockfish (interquartile range IQR).

Groups	TBF (time)	Z	p	ST (s)	Z	p
Noise exposure	246(141.3~335.5)	3.328	0.001**	60.6(34.92~91.32)	2.046	0.041*
Control	159(80.5~268.8)			45.9(20.08~82.94)		

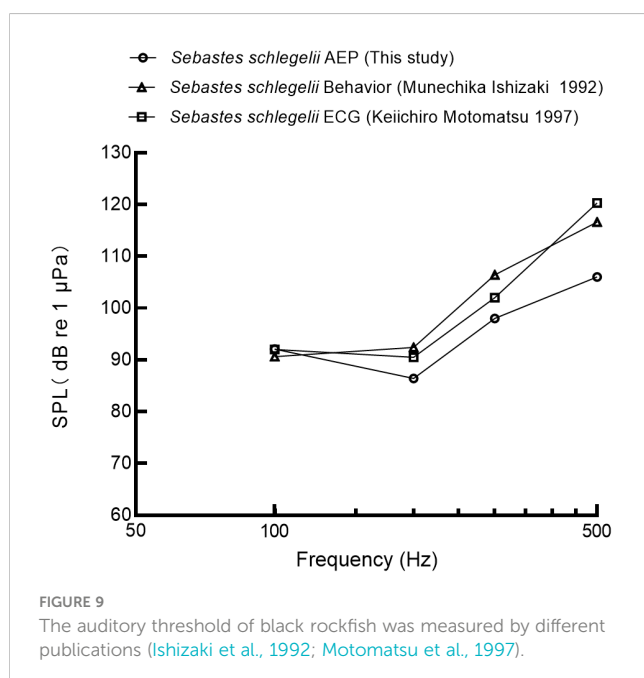
\* p value &lt; 0.05, \*\* p value &lt; 0.01.



resulting in a decrease in TBF and ST. These strategies may include seeking refuge or escaping from the noise source. However, from the fourth day onwards, the TBF and ST of the treatment group surpassed those of the control group. Significant differences were observed between the treatment and control groups ( $*p < 0.05$ ,  $**p < 0.01$ ). Following prolonged exposure to noise stimulation, it is conceivable that fish could enter an over-adaptation phase. Within this phase, fish may gradually readjust their swimming behavior, leading to an increase in TBF and ST. This phenomenon

may arise from the fish attaining a state of equilibrium in their adaptation to the noise or undergoing behavioral modifications to enhance resource utilization or cope with environmental pressures. These behaviors, such as increased tail beat frequency and swimming time, were likely selected as escape strategies. It is worth noting that tail beat consumes energy in fish and can potentially affect their growth rate. Furthermore, in quiet tanks, black rockfish exhibited more frequent and variable gaping behavior compared to fish in loud tanks. During the 7-day experimental period, the number of gaping behavior in the noise-exposed group was recorded as 60, whereas the control group exhibited 74. Both groups displayed a peak occurrence of gaping behavior during the nighttime period, followed by the morning period, and the lowest frequency was observed during the midday interval. Limited observations on gaping behavior in fish in the scientific literature suggest that gaping is associated with low activity levels (Rasa, 1971), which aligns somewhat with the results of this study, indicating less frequent adjustments in animals in quiet tanks (Andersson, 2011). Overall, these findings demonstrate the impact of noise on black rockfish behavior, highlighting the manifestation of stress and anxiety-related responses. The altered swimming distance, increased tail beat frequency, and modified gaping behavior indicate the adaptive strategies employed by the fish in response to the noise stress.

Stressors appear to cause shifts, lapses and narrowing of attention, and can also influence decision speed, which may explain the increase in foraging errors observed in the test group of black rockfish (Mendl, 1999; De Kloet et al., 1999). Interestingly, the black rockfish in the test group displayed rapid reactions to the baits but exhibited specific behaviors such as swimming in the vertical direction and not consuming the baits that fell to the bottom of the tank, especially during the recovery phase. These observations indicate altered feeding behavior and increased avoidance responses in black rockfish under noise conditions. The longer duration of feeding and lower feeding rate observed in the noise-exposed black rockfish align with the findings reported by Mickle and Higgs (2017) and the conclusions drawn by Sabet et al. (2016). Exposure to elevated noise levels can impair foraging behavior through various mechanisms, which are not mutually exclusive. Firstly, noise can trigger stress or fear-related responses in the fish, affecting their feeding behavior. Secondly, noise can act as a distraction, diverting the fish's attention away from the feeding task. Lastly, noise can mask important acoustic information that the fish relies on for successful foraging (Voellmy et al., 2014). The reduced foraging strikes and increased feeding duration observed in black rockfish are consistent with the concept of a defense cascade, where ongoing activities, such as foraging, are interrupted. Such responses are typically associated with stressors and fear-inducing stimuli (Metcalf, Huntingford, & Thorpe, 1987). These behavioral changes indicate that the black rockfish perceive the noise as a threat and exhibit altered feeding strategies as a defensive response. In summary, the impaired foraging behavior, increased feeding duration, and avoidance responses observed in black rockfish exposed to noise are in line with previous research. The effects of noise on feeding behavior can be attributed to stress-related responses, distraction, and the masking of acoustic cues necessary for successful foraging.



Indeed, the research period of this experiment was relatively short, and long-term subsequent studies will provide further insights into the auditory recovery of black rockfish in response to noise. The dominant frequency range of OWF noise falls within the auditory sensitivity range of black rockfish, but it's crucial to recognize that noise levels can vary significantly depending on the turbine power. Therefore, comprehensive field studies are needed to assess the actual impact of OWF noise on fish behavior and its implications for their fitness (Madsen et al., 2006). This research should include realistic noise simulations, considering key factors such as sound pressure levels and particle motion. By conducting such studies and accumulating more data, researchers will be better equipped to provide evidence-based recommendations and guidelines for mitigating the potential impacts of OWF noise on fish.

## 5 Conclusion

The results of this study emphasize the impact of anthropogenic noise on juvenile black rockfish, particularly in relation to their auditory threshold, swimming distance, and foraging behavior. The findings indicate that the minimum auditory threshold of black rockfish is 200 Hz, with a sound intensity of  $86.4 \pm 3.4$  dB re 1  $\mu$ Pa. Following exposure to 200 Hz noise, the most sensitive frequency for black rockfish shifted from 200 Hz to 300 Hz. Furthermore, we observed that continuous noise stimulation increased the swimming distance, swimming time, and tail beat frequency of juvenile black rockfish. Concurrently, the noise also had detrimental effects on their foraging behavior, potentially impacting their ability to effectively search for and capture food. These findings provide mounting evidence that anthropogenic noise can have wide-ranging effects on marine species, including fish.

## Data availability statement

The original contributions presented in the study are included in the article/supplementary material. Further inquiries can be directed to the corresponding author.

## Ethics statement

The animal studies were approved by Animal Ethical Committee of Ocean University of China. The studies were conducted in

accordance with the local legislation and institutional requirements. Written informed consent was obtained from the owners for the participation of their animals in this study.

## Author contributions

YW: Investigation, Methodology, Writing – original draft, Writing – review & editing. LH: Conceptualization, Formal analysis, Project administration, Supervision, Writing – review & editing. BX: Resources, Writing – review & editing, Data curation.

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## Conflict of interest

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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## References

- Amoser, S., and Ladich, F. (2003). Diversity in noise-induced temporary hearing loss in otophysine fishes. *J. Acoustical Soc. America* 113, 2170–2179. doi: 10.1121/1.1557212
- Andersson, M. H. (2011). "OWFs-ecological effects of noise and habitat alteration on fish," (Doctoral dissertation, Department of Zoology, Stockholm University).
- Andersson, M. H., Dockåkerman, E., Ubralhedenberg, R., Ohman, R., Marcus, C., and Sigra, P. (2007). Swimming Behaviour of Roach (*Rutilus rutilus*) and Three-spined Stickleback (*Gasterosteus aculeatus*) in Response to Wind Power Noise and Single-tone Frequencies. *Ambio* 36, 636–638. doi: 10.1579/0044-7447(2007)36[636:SBORRR]2.0.CO
- Andrew, T., Francis, E., Charles, M., Naigaga, I., Jessica, N., Micheal, O., et al. (2016). Mercury concentration in muscle, bellyfat and liver from *Oreochromis niloticus* and *Lates niloticus* consumed in Lake Albert fishing communities in Uganda. *Cogent Food Agric.* 2 (1), 1214996. doi: 10.1080/23311932.2016.1214996

- Betke, K., Glahn, S. V., and Matuschek, R. (2005). Underwater noise emissions from offshore wind turbines. *Pro CFA/DAGA* 18, 25–30. Available at: [https://pub.dega-akustik.de/DAGA\\_1999-2008/data/articles/001773.pdf](https://pub.dega-akustik.de/DAGA_1999-2008/data/articles/001773.pdf) (Accessed March, 2004).
- Blew, J., Hoffmann, M., Nehls, G., and Hennig, V. (2008). BioConsult SH report: Investigations of the bird collision risk and the responses of harbour porpoises in the offshore wind farms Horns Rev, North Sea, and Nysted, Baltic Sea. Available at: [https://tethys.pnnl.gov/sites/default/files/publications/Bird\\_Collision\\_Risk\\_and\\_the\\_Responses\\_of\\_Harbour\\_Porpoi.pdf](https://tethys.pnnl.gov/sites/default/files/publications/Bird_Collision_Risk_and_the_Responses_of_Harbour_Porpoi.pdf) (Accessed 2008).
- Breitzler, L., Ieng, H. L., Fonseca, P. J., and Vasconcelos, R. (2020). Noise-induced hearing loss in zebrafish: investigating structural and functional inner ear damage and recovery. *Hearing Res.* 391, 107972. doi: 10.1016/j.heares.2020.107952
- Codarin, A., Wysocki, L. E., Ladich, F., Picciulin, M., and Lidia, E. (2009). Effects of ambient and boat noise on hearing and communication in three fish species living in a marine protected area (Miramare, Italy). *Mar. Pollut. Bull.* 58, 1880–1887. doi: 10.1016/j.marpolbul.2009.07.011
- De Jong, C. A. F., Ainslie, M. A., and Blacquièrre, G. (2011). *TNO report: Standard for measurement and monitoring of underwater noise, Part II: procedures for measuring underwater noise in connection with offshore wind farm licensing*. Available at: [https://users.ece.utexas.edu/~ling/2A\\_EU1.pdf](https://users.ece.utexas.edu/~ling/2A_EU1.pdf) (Accessed September, 2011).
- De Kloet, E., Melly, R., Oitzl, S., and Joëls, M. (1999). Stress and cognition: are corticosteroids good or bad guys? *Trends Neurosci.* 22, 422–426. doi: 10.1016/S0166-2236(99)01438-1
- Faria, M., Vall, A., Prats, E., Bedrossiantz, J., Orozco, M., Porta, J., et al. (2019). Further characterization of the zebrafish model of acrylamide acute neurotoxicity: gait abnormalities and oxidative stress. *Sci. Rep.* 9(1), 7075. doi: 10.1038/s41598-019-43647-z
- Ishizaki, M., Hiraishi, T., Yamamoto, K., and Nashimoto, K. (1992). Auditory threshold of black rockfish. *Nsugaf* 58, 55–61. doi: 10.2331/suisan.58.55
- Jones, I. T., Stanley, J. A., and Aran Mooney, T. (2020). Impulsive pile driving noise elicits alarm responses in squid (*Doryteuthis pealeii*). *Mar. Pollut. Bull.* 150, 110792. doi: 10.1016/j.marpolbul.2019.110792
- Keichiro, M. (1997). Auditory threshold and masking in black rockfish *Sebastes schlegelii*. *Nippon Suisan Gakkaishi* 63, 110–111. doi: 10.2331/suisan.63.110
- Kenyon, T. N., Ladich, F., and Yan, H. Y. (1998). A comparative study of hearing ability in fishes: the auditory brainstem response approach. *J. Comp. Physiol. A-neuroethology Sensory Neural Behav. Physiol.* 182, 307–318. doi: 10.1007/s003590050181
- Kojima, T., Ito, H., Komada, T., Taniuchi, T., and Akamatsu, T. (2005). Measurements of auditory sensitivity in common carp *Cyprinus carpio* by the auditory brainstem response technique and cardiac conditioning method. *Fisheries Sci.* 71, 95–100. doi: 10.1111/j.1444-2906.2005.00935x
- Lacroix, D., and Sylvain, P. (2011). The multi-use in wind farm projects: more conflicts or a win-win opportunity? *Aquat. Living Resources*. 24, 129–135. doi: 10.1051/alr/2011135
- Ladich, F., and Fay, R. (2013). Auditory evoked potential audiometry in fish. *Springer Open Choice* 23, 317–364. doi: 10.1007/s11160-012-9297-z
- Likang, L., Wen, H., Yun, L., Jifang, L., Simin, Z., Min, S., et al. (2018). Deep transcriptomic analysis of black rockfish (*Sebastes schlegelii*) provides new insights on responses to acute temperature stress. *Sci. Rep.* 8, 9113. doi: 10.1038/s41598-018-27013-z
- Lindell, H. (2003). *Ingemansson report: Utgrunden off-shore wind farm- Measurements of underwater noise*. Available at: [https://mhk.pnnl.gov/sites/default/files/publications/Utgrunden\\_Underwater\\_Noise\\_2003.pdf](https://mhk.pnnl.gov/sites/default/files/publications/Utgrunden_Underwater_Noise_2003.pdf) (Accessed June 17, 2003).
- Liu, M., Wei, Q. W., Du, H., Fu, Z. Y., and Chen, Q. C. (2013). Ship noise-induced temporary hearing threshold shift in the Chinese sucker *Myxocyprinus asiaticus* (Bleeker). *J. Appl. Ichthyology* 29, 1416–1422. doi: 10.1111/jai.12345
- Madsen, P. T., Wahlberg, M., Tougaard, J., Lucke, K., and Tyack, P. L. (2006). Wind turbine underwater noise and marine mammals: Implications of current knowledge and data needs. *Mar. Ecol. Prog. Ser.* 309, 279–295. doi: 10.3354/meps309279
- Marmo, B., Roberts, I., Buckingham, M. P., King, S., and Booth, C. (2013). *Scottish Marine and Freshwater Science report: Modelling of Noise Effects of Operational Offshore Wind Turbines including noise transmission through various foundation types*. Available at: <https://data.marine.gov.scot/dataset/modelling-noise-effects-operational-offshore-wind-turbines-including-noise-transmission-0> (Accessed November 17, 2014).
- Mendl, M. (1999). Performing under pressure: stress and cognitive function. *Appl. Anim. Behav. Sci.* 65, 221–244. doi: 10.1016/S0168-1591(99)00088-X
- Metcalfe, N. B., Huntingford, F. A., and Thorpe, J. E. (1987). The influence of predation risk on the feeding motivation and foraging strategy of juvenile Atlantic salmon. *Anim. Behav.* 35 (3), 901–911. doi: 10.1016/S0003-3472(87)80125-2
- Mickle, M. F., and Higgs, D. M. (2017). Integrating Techniques: a review of the effects of anthropogenic noise on freshwater fish. *Can. J. Fisheries Aquat. Science*. 75, 217–245. doi: 10.1139/cjfas-2017-0245
- Motomatsu, K., Hiraishi, T., Yamamoto, K., and Naishimoto, K. (1997). Auditory Threshold and Masking in Black Rockfish *Sebastes schlegelii*. *Nihon Suisan Gakkaishi* 63 (1), 110–111. doi: 10.2331/suisan.64.792
- Nedwell, J., Langworthy, J., and Howell, D. (2003). *COWRIE report: Assessment of sub-sea acoustic noise and vibration from offshore wind turbines and its impact on marine wildlife; initial measurements of underwater noise during construction of offshore wind farms, and comparison with background noise*. Available at: [https://users.ece.utexas.edu/~ling/2A\\_EU1.pdf](https://users.ece.utexas.edu/~ling/2A_EU1.pdf) (Accessed May, 2003).
- Nedwell, J., Langworthy, J., and Howell, D. (2004). *COWRIE report: Mesurement of underwater noise during construction of offshore wind farm, and comparison with background noise*. Available at: [https://tethys.pnnl.gov/sites/default/files/publications/COWRIE\\_Noise\\_Report\\_Draft.pdf](https://tethys.pnnl.gov/sites/default/files/publications/COWRIE_Noise_Report_Draft.pdf) (Accessed 2004).
- Netherlands Ministry of Infrastructure and the Environment. (2011). *TNO report: Standard for measurement and monitoring of underwater noise, Part II. procedures for measuring underwater noise in connection with OWF licensing*. Available at: <https://tethys.pnnl.gov/sites/default/files/publications/TNO-Report-2011.pdf> (Accessed September 25, 2011).
- Popper, A. N., Nancy, L., and Clarke, (1976). The auditory system of the goldfish (*Carassius auratus*): effects of intense acoustic stimulation. *Comp. Biochem. Physiol. Part A: Physiol.* 53, 11–18. doi: 10.1016/S0300-9629(76)80003-5
- Rasa, (1971). The causal factors and function of 'Yawning' in *Microspathodon chrysurus* (Pisces: Pomacentridae). *Behaviour* 39, 39–57. doi: 10.1163/156853971X00168
- Russell, D. F., Tucker, A., Wetrting, B. A., Neiman, A. B., Wilkens, L., and Moss, F. (2001). Noise effects on the electrosense-mediated feeding behavior of small paddlefish. *Fluctuation Noise Lett.* 1 (02), L71–L86. doi: 10.1142/S021947750100024X
- Sabet, S. S., Wesdorp, K., Campbell, J., Snelderwaard, P., and Slabbekoorn, H. (2016). Behavioural responses to sound exposure in captivity by two fish species with different hearing ability. *Anim. Behav.* 116, 1–11. doi: 10.1016/j.anbehav.2016.03.027
- Sara, G., Dean, J. M., D'Amato, D., Buscaino, G., Oliveri, A., Genovese, S., et al. (2007). Effect of boat noise on the behaviour of bluefin tuna *Thunnus thynnus* in the Mediterranean Sea. *Mar. Ecol. Prog. Ser.* 331, 243–253. doi: 10.3354/meps331243
- Scholik, A. R., and Yan, H. Y. (2002). Effects of boat engine noise on the auditory sensitivity of the fathead minnow, *pimephales promelas*. *Environ. Biol. Fishes* 63, 203–209. doi: 10.1023/A:1014266531390
- Slabbekoorn, H., Bouton, N., Opzeeland, I. V., Coers, A., Cate, C., and Popper, A. N. (2010). A noisy spring: the impact of globally rising underwater sound levels on fish. *Trends Ecol. Evol.* 25, 419–427. doi: 10.1016/j.tree.2010.04.005
- Smith, M. E., Kane, A. S., and Popper, A. N. (2004). Acoustical stress and hearing sensitivity in fishes: does the linear threshold shift hypothesis hold water? *J. Exp. Biol.* 207, 3591–3602. doi: 10.1242/jeb.01188
- Spiga, I., Aldred, N., Gary, S., and Caldwell, (2017). Anthropogenic noise compromises the anti-predator behaviour of the European seabass, *Dicentrarchus labrax* (L.). *Mar. Pollut. Bull.* 122, 297–305. doi: 10.1016/j.marpolbul.2017.06.067
- Stevens, E., and Don, (1979). The effect of temperature on tail beat frequency of fish swimming at constant velocity. *Can. J. Zoology* 57, 1628–1635. doi: 10.1139/z79-214
- Subacoustech Ltd. (2003). *COWRIE report: Assessment of sub-sea acoustic noise and vibration from offshore wind turbines and its impact on marine wildlife*. Available at: <https://www.mendeley.com/catalogue/a7794cec-bdf8-3574-a332-67d20afe4ef3/> (Accessed May 3, 2003).
- Subacoustech Ltd. (2004). *COWRIE report: A review of offshore windfarm related underwater noise sources*. Available at: <https://tethys.pnnl.gov/publications/review-offshore-windfarm-related-underwater-noise-sources> (Accessed October 1, 2004).
- Thompson, P. M., Graham, I. M., Cheney, B., Barton, T. R., and Merchant, N. D. (2020). Balancing risks of injury and disturbance to marine mammals when pile driving at OWFs. *Ecol. Solutions Evidence* 1, e12034. doi: 10.1002/2688-8319.12034
- Tougaard, J., Henriksen, O. D., and Miller, L. A. (2009). Underwater noise from three types of offshore wind turbines: estimation of impact zones for harbor porpoises and harbor seals. *J. Acoustical Soc. America* 125, 3766–3773. doi: 10.1121/1.3117444
- Tougaard, J., Hermannsen, L., and Madsen, P. T. (2020). How loud is the underwater noise from operating offshore wind turbines? *J. Acoustical Soc. America* 148, 2885–2893. doi: 10.1121/10.0002453
- Tullio, G., Mariani, P., Benassai, G., Luccio, D. D., and Grieco, L. (2018). Sustainable use of marine resources through offshore wind and mussel farm co-location. *Ecol. Model.* 367, 34–41. doi: 10.1016/j.ecolmodel.2017.10.012
- Velasquez, J., Laura, E., and Mark, I. (2020). Vessel noise affects routine swimming and escape response of a coral reef fish. *PLoS One* 15, e0235742. doi: 10.1371/journal.pone.0235742
- Voellmy, I. K., Purser, J., Flynn, D., Kennedy, P., Simpson, S. D., and Radford, A. N. (2014). Acoustic noise reduces foraging success in two sympatric fish species via different mechanisms. *Anim. Behav.* 89, 191–198. doi: 10.1016/j.anbehav.2013.12.029
- Weilgart, L. (2018). *OceanCare report: The impact of ocean noise pollution on fish and invertebrates*. 2018. In: *Report for ocean care*. Available at: <https://thegreentimes.co.za/wp-content/uploads/2022/01/impact-of-ocean-noise-pollution-on-fish-and-invertebrates.pdf> (Accessed May 1, 2018).
- Wever, L., Krause, G., and Buck, B. H. (2015). Lessons from stakeholder dialogues on marine aquaculture in OWFs: Perceived potentials, constraints and research gaps. *Mar. Policy* 51, 251–259. doi: 10.1016/j.marpol.2014.08.015
- Zhang, B., Zhang, X. G., Guo, H. Y., Fang, N., and Song, J. K. (2016). Characteristics of underwater noise from Shanghai Donghai Bridge OWF. *J. Shanghai Ocean Univ.* 25, 599–606. doi: 10.1024/sou.20150401425





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# Does sedation with AQUÍ-S® mitigate transport stress and post transport mortality in ballan wrasse (*Labrus bergyltae*)?

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Ballan wrasse (*Labrus bergylta*) are commonly used as cleaner fish in salmon farms as a biological treatment to mitigate sea lice infestation. Improved welfare for cleaner fish both during production of these fish and when in sea-cages with salmon is crucial for the industry's development. A common operational procedure in ballan wrasse production is transporting juveniles from one land-based farm to another for further on-growing. Episodes of increased mortality have been reported after such transportations. In this study, the relationship between transport stress and post-transport mortality at the on-growing facility was examined. It was also investigated if light sedation with AQUÍ-S® can mitigate stress during transport. Stress was quantified by measuring cortisol release rate to the tank water during transport. This was investigated in 10 commercial live carrier truck transports (6 without AQUÍ-S® sedation and 4 with sedation during loading and transport). The total time of transport varied between 12 and 21h. In general, mortality was significantly higher ( $1.0 \pm 0.6\%$  day<sup>-1</sup>) the first five days post-transport compared to 15–20days post transport ( $0.5\%$  day<sup>-1</sup>). There was also a strong relationship between fish weight at transport and post-transport mortality, where higher mean weight at transport reduced mortality. In contrast to what was expected, AQUÍ-S® treatment during transport procedures increased cortisol excretion rate, suggesting a stimulating effect of AQUÍ-S® on the stress axis in ballan wrasse. Considering these results, the value of using AQUÍ-S® to reduce stress during transport of juvenile ballan wrasse might be questioned. However, there was no relationship between cortisol release rate during transport and post-transport mortality. Furthermore, this study emphasizes that water cortisol measurements can be used as a none-invasive tool for monitoring stress and can be integrated into the welfare evaluation during commercial fish transports.

## KEYWORDS

delayed mortality, sedation, transport, Cleanerfish, aquaculture, fish welfare

## 1 Introduction

Sea lice control continues to be the primary challenge for growth and sustainability of the Atlantic salmon farming industry in Norway and elsewhere (1–3). There is currently no effective vaccine against sea-lice. In addition, sea-lice show increased resistance to chemo therapeutics (2). Therefore, non-medicinal solutions for sea-lice control are the most common



mitigation strategies. A biological strategy, using cleaner fish to delouse salmon is now commonly used and both reduces the need for medicinal treatments and is less stressful for the farmed salmon compared to other delousing methods that require extensive handling (4–6).

In the last 10–15 years commercial production of primarily ballan wrasse, *Labrus bergylta*, has developed to meet the increased demand and to address sustainability and biosecurity issues related to using wild-caught wrasse. In recent years, considerable improvements with regards to delousing efficiency, disease management and welfare of cleaner fish have been made (4, 7). However, there are still large knowledge gaps regarding how common operational procedures affect the welfare of ballan wrasse, emphasized by reports of high losses during production and when used in sea cages (8–10). Transport of cleaner fish, by boat and/or vehicle, is an essential procedure of the production process and often occurs over long distances. Most of the previous studies on welfare issues related to transport of cleaner fish have focused on lumpfish when transported from land-based hatcheries to co-stocking with salmon in sea-cages (8, 9). In ballan wrasse production, transport between land-based hatcheries and on-growing facilities is common, and episodes of increased mortality have been reported after such transport.

Generally, the transport of live fish is one of the most critical operations in aquaculture, in which fish are exposed to multiple handling events and large variations in environmental factors within a short period of time. These are events that can initiate a severe stress response in farmed fish. In response to stress cortisol, the main stress hormone in teleost fishes, is released (11, 12) enabling stress coping mechanisms by making the resources needed to cope with the stressful stimuli available. Energy required for stress coping is redistributed from maintenance functions such as growth and immune reactions (12). In this vein, the sum of the high intense stressor of transport together with the post transport low intensity stressor have been suggested to affect post transport survival (13). Accordingly, stress has been discussed as one of the underlying factors for delayed mortality syndrome or hauling mortality, referring to mortality appearing days or even weeks after transport (13). Factors that have been suggested to affect the post-transport outcome include water quality, confinement, density, holding container design, handling procedures, and agonistic behavior reviewed by Harmon (13–15).

There is a strong relationship between plasma cortisol levels and the release of this hormone from the fish to the water, reviewed by Scott and Ellis (16). Accordingly, the release rate of cortisol to the water has been used as a non-invasive welfare indicator in laboratory studies (17–19) and in commercial aquaculture settings (20). Fish transport in closed tanks in hauling vehicles with or without external water exchange and in well-boats is a common practice in aquaculture. In such closed transports, cortisol will accumulate over time. Therefore, measurements of water cortisol concentrations can be used as an integrative tool for assessing the total stress burden associated with transport. However, its use as a monitoring tool for fish welfare during transport is still very limited.

Using anesthetics prior and during transport is a common practice to lower the metabolism of the fish, thus reducing toxic metabolite production, reviewed in Harmon (13). Anesthetics are also used to lessen the stress response due to handling and transport (13, 21–23). This is achieved by sedating the fish enough to reduce the

stress response and the risk of injury, while fish are still able to maintain swimming and an upright position. Therefore the proper dosage is critical and will vary with species and fish size. AQUI-S®, an isoeugenol based sedative, has been used before and during transport to reduce transport stress in several fish species (22, 24, 25). It has proven to be a good sedative for different types of handling procedures, since several fish species go directly into the narcotic (resting) state without a pre excitement phase (26). Although AQUI-S® is a widely used anesthetic the mechanisms of action are not clearly understood. Although AQUI-S has been reported to mitigate stress in several fish species, other studies indicate that the stress mitigating properties of AQUI-S® are species specific (25, 26). If sedation with AQUI-S® affects post transport survival and welfare in ballan wrasse is to our best knowledge unknown.

The objective of this study was to investigate the relationship between transport stress and post transport mortality at a commercial Ballan wrasse on-growing facility. The effect of AQUI-S® sedation on transport stress was quantified by cortisol release rate to the water in the transport tanks, and effects of transport stress on post-transport mortality was investigated by relating cortisol release during transport to mortality rates at the on-growing facility.

## 2 Materials and methods

### 2.1 Transports

The study included transports from 2 commercial hatcheries producing ballan wrasse larvae, all transports were delivered to the same receiving land-based farm for further on-growing. Fish were loaded on to transport trucks by lowering the water level in the tank and then vacuum pumping fish into transport tanks. The loading procedure took between 1 and 3 h. The transports were performed in two trucks with 10 or 12 internal tanks, with a volume of 1.2 and 1.0 m<sup>3</sup>, respectively. Oxygen saturation (%), temperature (°C) and pH were monitored and logged in the transport tanks. Transport time, fish weight and density, temperature, pH and oxygen saturation during the transport are presented in Table 1. All fish from the same transport were offloaded into one tank at the receiving farm, in which they remained throughout the 25-day study. Dead fish from each tank were counted and removed daily.

To investigate if AQUI-S reduced stress during transport procedures, AQUI-S was used both during loading and transport during four transports. 0.5 h before loading AQUI-S (2.5 µL/L) was added directly to the fish tanks in the hatchery. In the tanks aboard the hauling vehicle the same dose of AQUI-S was premixed in the tank water before the fish were loaded. Six of the followed transports fish were not treated with AQUI-S®, Table 1.

### 2.2 Water sampling and analysis

#### 2.2.1 Water sampling

One liter water samples were collected from the water supply used to fill up transport tanks at the hatchery (basal levels). Post transport water samples (1 L) were randomly collected from four of the tanks on the transport vehicles, directly when fish were offloaded. The water samples were frozen and stored at −20°C until analysis.

TABLE 1 Conditions during truck transports of juvenile Ballan wrasse between hatcheries and a land based on-growing facility.

	Treatment									
	AQUI-S®					No AQUI-S®				
Transport time (h)	16.25	19.50	17.25	15.25	16.25	20.75	17.50	18.50	15.25	12.50
Fish weight (g)	2.6	1.7	2	2.2	1.8	1.5	2	1.7	2.2	2.3
Tank volume (m³)	1	1	1	1	1	1	1	1	1,2	1
Fish density (kg m <sup>-3</sup> )	29.3	19.8	12.6	19.8	13.4	12.6	19.6	10.8	28.8	21.1
Temp (min-max; °C)	12.1–13.1	11.8–13.1	12.2–13.3	11.8–13.6	11.8–13.3	11.2–12.5	11.4–12.5	11.4–14.0	11.9–13.5	–
pH (min-max)	6.9–7.6	6.9–7.7	6.7–8.1	7.2–7.4	7.1–7.9	7.3–7.9	7.3–7.5	7.3–7.5	7.2–7.5	–
Oxygen (min-max; %)	107–122	107–132	106–132	103–117	105–138	102–121	106–123	92–128	101–128	85–143

## 2.2.2 Water cortisol analysis

Analysis of accumulated water cortisol during the transport followed a method described by McWhinney et al. (27) with some modifications. Water samples were spiked with 10 ng internal standard (cortisol d4) to correct for matrix effects and for losses in sample extraction, concentration, and analysis. Samples were loaded onto activated Oasis HLB 6 cc (200 mg) solid-phase extraction cartridges (Waters, Milford, MA, United States). After loading 200–300 mL of the samples, the columns were washed with 3 mL milliQ water followed by 3 mL of 20% methanol. Samples were eluted with 5 mL 100% ethyl acetate, dried at 50°C, and reconstituted in 200 µL of 40% methanol with 5 mmol/L ammonium formate and 0.1% formic acid. Separation was achieved on a BEH C8 column (Waters, Milford, MA, United States) using a solvent gradient consisting of 5 mmol/L ammonium formate and 0.1% formic acid in water and methanol. Cortisol content was analyzed with a tandem mass spectrometer (Waters TQ-S, Milford, MA, United States) operated in negative electron spray ionization mode with the following MRM acquisition parameters (precursor and product ions); 407.1 > 331.05, 407.1 > 331.1 for cortisol and 411.1 > 335.05, 411.1 > 335.1 for cortisol d4 (IS).

### 2.2.2.1 Cortisol release rate

In fish, cortisol in unconjugated form is mainly excreted through the gills to the water and it has previously been shown that free cortisol in the water is directly related to the cortisol concentration in the blood of fish (18). On this basis, the excretion rate of cortisol has been used as a stress indicator in commercial fish farms (20). The excretion rate of cortisol to the water during transport from the hatchery to the receiving on-growing facility, was calculated using the formula:

$$\frac{([\text{Cortisol transport}_{\text{end}}] - [\text{Cortisol transport}_{\text{start}}])}{(\text{biomass}_{\text{tank}} * \text{transport time})}$$

[Cortisol transport<sub>start</sub>] is the cortisol concentration in the water before fish are loaded. [Cortisol transport<sub>end</sub>] is the cortisol concentration in the tank at the transport end. Biomass<sub>tank</sub> is the biomass of the fish in the transport tank and transport time is the time between loading and unloading the fish. [cortisol transport<sub>start</sub>] was below the detection limit of the analysis method (0.1 µg L<sup>-1</sup>) and therefore set to 0.

## 2.3 Statistical analyses

All values are presented as mean ± standard error unless otherwise specified. The effects of AQUI-S® treatment on cortisol excretion rate during transport was investigated by a student's *T*-test. Effect of transport on mortality during five-day intervals following transport was investigated by a repeated measure analysis of variance (ANOVA). Dunnett's *post hoc* test was employed to investigate differences in mortality during the first five days interval after transport compared to the following four five-day intervals. Factors affecting accumulated mortality 25 days after transport was investigated with an analysis of covariance (ANCOVA) with transport treatment (AQUI-S® or no AQUI-S® treatment) as categorical factors and cortisol excretion rate during transport and the mean size of the transported fish as continuous predictors. *p* < 0.05 was set as the threshold for statistical significance.

## 3 Results

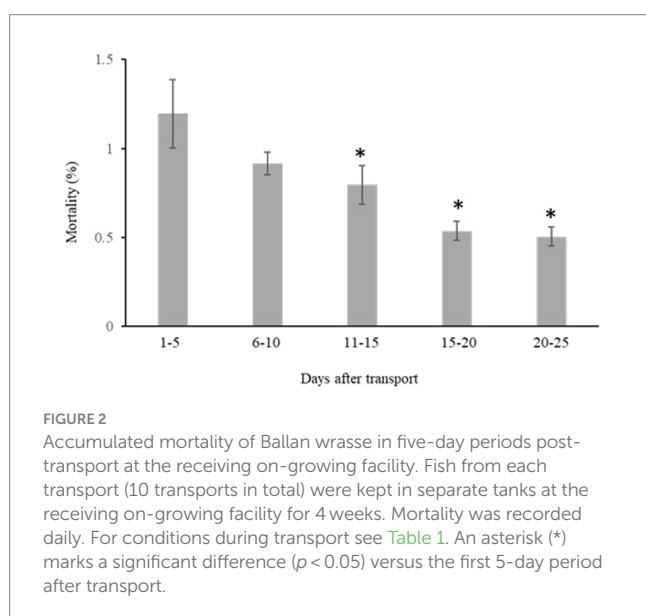
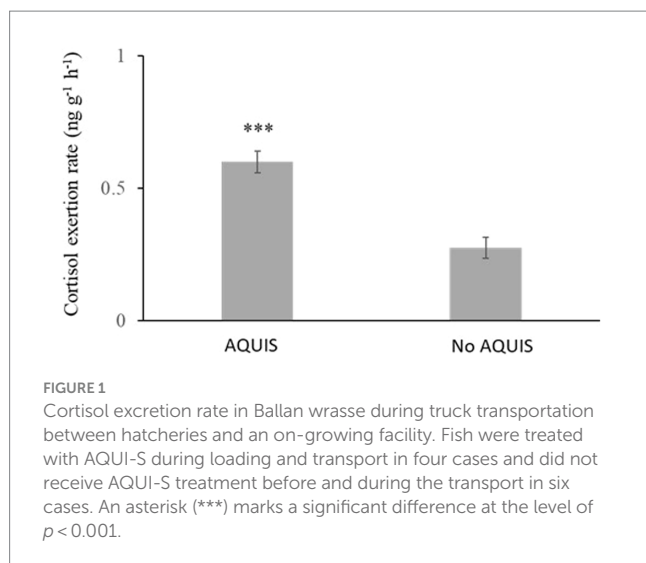
### 3.1 Water cortisol

Ballan wrasse sedated with AQUI-S® during loading and transport had a significantly higher cortisol excretion rate compared to fish transported without AQUI-S® treatment ( $T_{(8)} = 5.5$ , *p* < 0.001) (Figure 1).

### 3.2 Mortality after transport

Mortality changed over time during the first 25 days after transport (Repeated measure ANOVA;  $F_{(4, 36)} = 13$ , *p* < 0.001), leading to significantly lower values day 11–15, 16–20, and 20–25 compared to mortality during the first five-day period after transport (day 1–5, *p* < 0.01, *p* < 0.001 and *p* < 0.001 respectively), (Figure 2). There was no significant difference in mortality between the first (day 1–5) and the second (day 6–10) period (*p* < 0.067).

The results from the ANCOVA showed a significant relationship between fish weight at transport and accumulated mortality 25 days after transport ( $F_{(1, 6)} = 17$ , *p* < 0.01), (Figure 3). However, there were no significant effects of AQUI-S treatment ( $F_{(1, 6)} = 0.0001$ , *p* < 0.98) on accumulated mortality 25 days after transport. Neither, was there a



significant relationship between cortisol excretion rate during transport and accumulated mortality 25 days after transport ( $F_{1,6} = 0.18, p < 0.98$ ).

## 4 Discussion

In general transport is one of the most stressful and critical operations in fish farming, including handling, crowding, pumping and potential sub-optimal water quality, and in addition motion in the transport vessel (13, 23). In this study, the average cortisol excretion rate to the water in the transport tanks ranged between 0.15 and 0.6  $\text{ng g}^{-1} \text{h}^{-1}$ . Laboratory studies with rainbow trout and salmon show that cortisol excretion rate can increase from 0.01 to values up to 0.6–7  $\text{ng g}^{-1} \text{h}^{-1}$  during stressful events, reviewed in Scott and Ellis (18). The initial process of loading to the transport tanks requires several novel stressors within a short time period, i.e., handling, pumping and a new environment, whereas the long transport time

may act as a recovery period if water quality is maintained. The average cortisol extraction rate to the water in the transport tanks can be viewed as the total stress burden during transport. Considering this, and that the cortisol release rate to the transport tanks in this study was just under the values in Scott and Ellis (18), the results might reflect that fish are recovering from an acute-intense stress that is associated with additive stressors during the loading process. However, when comparing cortisol excretion rates, it is important to note that higher baseline levels of plasma cortisol have been reported in ballan wrasse compared to rainbow trout however both species seem to have similar cortisol levels in response to acute intense stress (28). This emphasizes that further studies on stress responsiveness, and the relationship between plasma cortisol and cortisol release rate to the water, are needed to verify the impact of potential stressors on the release rate of cortisol in ballan wrasse.

In this study light sedation with AQUIS-S® during loading and transport increased cortisol excretion rate. Similar results have been reported in Gilthead seabream, showing that AQUIS-S® sedation elevated plasma cortisol levels after 6 h simulated transport (25). In a comparable study with Gilthead seabream clove oil, another isoeugenol based sedative, also increased plasma cortisol levels after transport (29). Based on the elevated post transport plasma cortisol values together with differences in gene expression in the head kidney, the authors suggest that clove oil prolongs the time needed for stress recovery (29). As mentioned above, it is possible that the relatively low cortisol release rate during transport in our study reflects that fish were able to recover from the initial intense stress associated with loading during the long transport. Also in line with (28), our findings suggest that AQUIS-S® sedated fish take longer to recover and habituate to transport tanks thus have a higher overall cortisol release rate. However, these results are in contrast to studies in salmon that show that AQUIS-S® has stress reducing capabilities (22, 30). In Atlantic salmon smolts, sedation just before onloading to the transport truck and again 15 min before offloading resulted in lower post-transport plasma cortisol levels (22). Still, it is important to note that in the studies (22, 30) that demonstrate that AQUIS-S® has stress reducing effects during transport procedures, AQUIS-S® was not used the whole transport and fish were allowed to recover in water without AQUIS-S® before sampled. Thus, it cannot be excluded that the sedation protocol for reducing transport stress in Atlantic salmon smolts (22) might also have stress reducing effects in Ballan wrasse, if only used during handling and onloading operations and not during transport.

In line with the suppressive effects of AQUIS-S® on the cortisol response in Atlantic salmon during transport procedures it has also been demonstrated that channel catfish (*Ictalurus punctatus*) exposed to confinement and hypoxia when AQUIS-S® sedated have a reduced cortisol response (31). There are also other studies that report no effect of AQUIS-S® on the plasma cortisol response such as when used for crowding rainbow trout (*Oncorhynchus mykiss*) (32) or striped bass exposed to low water levels (33). In addition to the mentioned effects of AQUIS-S® on cortisol dynamics during stress recovery, species specific effects of the sedative on the stress response might also contribute to discrepancies between studies (26, 34).

Generally, light sedation is considered to mitigate stress related mortality in transported fish reviewed by Harmon (13). Iversen and Eliassen (22) reported that beside the stress reducing effects of AQUIS-S® it also resulted in higher post transport survival rates in Atlantic salmon smolts. In our study we could not detect any effects of

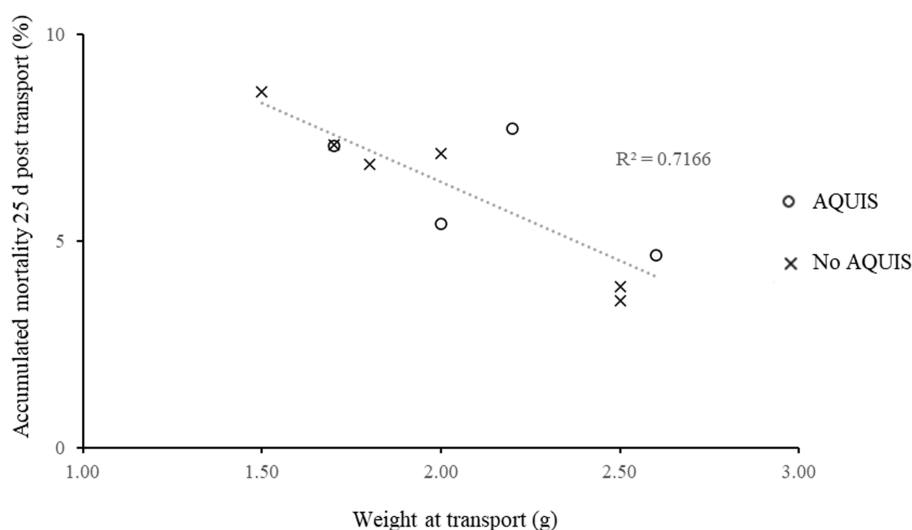


FIGURE 3

Relationship between the weight of Ballan wrasse juveniles at transport and mortality 25 days post-transport, when transported from hatcheries to a land based on-growing facility. The fish from each transport were kept in separate tanks at the on-growing facility. Mortality was recorded daily.

AQUI-S® on post transport mortality. Interestingly, size at transport showed a close negative relationship with accumulated mortality 25 days after transport. Furthermore, in our study, mortality was highest (~1%) the first 5 days after transport to the on-growing facility. After 5 days mortality decreased and 25 days post transport the accumulated mortality was ~4%. In one of the few other studies performed on ballan wrasse transport, Jonassen and Foss (35) showed similar mortality values after transporting larger ballan wrasse from a land-based facility to open sea- cages (by truck and/or well-boat). In that study the cumulative mortality in the sea-cages was 1.1 to 1.3% seven days after transfer and between 4.4 and 5.8% 1 month post-transport. However, in the mentioned study, it is pointed out that the large difference in rearing conditions between the land and sea site and the increased environmental variability at sea has a greater impact on survival than the transport stress. In the present study, fish were reared in similar conditions at the hatchery and at the on-growing facility. Thus, the observed post transport mortality is likely linked to transport conditions and size of the fish. However, it is important to note that data on size related mortality in non-transported fish is needed to verify the actual impact of transport on mortality in farmed ballan wrasse. Still, the strong relationship between fish weight at transport and post-transport mortality in the present study suggests that farmers should include size at transport as an important factor to consider during planning and establishing best practice protocols for juvenile ballan wrasse transports.

In conclusion, this study shows that slight sedation with AQUI-S® neither had a mitigating effect on cortisol release rate during transport nor reduces post-transport mortality in ballan wrasse. Moreover, somewhat in contrast to what was expected, AQUI-S® treatment during loading and transport increased cortisol excretion rate to the water in the transport tanks, suggesting a stimulating effect of AQUI-S® on the stress axis in ballan wrasse. Furthermore, the cortisol release rate in the present study was rather low in comparison with stressed salmonids. This suggests a species difference in cortisol release rate and/or that during the long transport the ballan wrasse have been

recovering from the initial stressful phase of handling and onloading to the transport tanks. The latter is supported by the stimulating effect of AQUI-S® on the cortisol response observed in our study and other studies (25, 29) indicating that using AQUI-S® during transports prolongs the time needed for stress recovery. Overall, this study demonstrates that water cortisol can be used as an integrative tool to assess stress during transport. Still, studies linking cortisol release rate during transport to negative welfare and health effects, and ultimately leading to delayed mortality syndrome or hauling mortality, are required for implying negative transport consequences. Moreover, considering that the negative impact of stress on animal welfare is related to the intensity, duration, and frequency of the stressors (36), further studies linking the dynamics of cortisol release rate to fish health and welfare is needed.

## Data availability statement

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

## Ethics statement

Ethical approval was not required for the study involving animals in accordance with the local legislation and institutional requirements because we followed commercial fish transport with normal routines. We used water born cortisol as a stress indicator of transported fish. Thus, fish was not exposed extra stress due to sampling.

## Author contributions

SC: Investigation, Writing – review & editing. TJ: Methodology, Writing – review & editing. ES: Writing – review & editing. HÅ:



Investigation, Writing – review & editing. AI: Funding acquisition, Project administration, Writing – review & editing. CS: Investigation, Methodology, Writing – original draft, Writing – review & editing. TW: Methodology, Writing – review & editing. EH: Conceptualization, Formal analysis, Investigation, Methodology, Project administration, Writing – review & editing.

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## References

- Igboeli OO, Fast MD, Heumann J, Burka JF. Role of P-glycoprotein in emamectin benzoate (SLICE<sup>®</sup>) resistance in sea lice *Lepeophtheirus salmonis*. *Aquaculture*. (2012) 344–349:40–7. doi: 10.1016/j.aquaculture.2012.03.026
- Igboeli OO, Burka JF, Fast MD. *Lepeophtheirus salmonis*: a persisting challenge for salmon aquaculture. *Anim Front*. (2014) 4:22–32. doi: 10.2527/af.2014-0004
- Torrissen O, Jones S, Asche F, Guttormsen A, Skilbrei OT, Nilsen F, et al. Salmon lice—impact on wild salmonids and salmon aquaculture. *J Fish Dis*. (2013) 36:171–94. doi: 10.1111/jfd.12061
- Skiftesvik AB, Bjelland RM, Durif CM, Johansen IS, Browman HI. Delousing of Atlantic salmon (*Salmo salar*) by cultured vs. wild ballan wrasse (*Labrus bergylta*). *Aquaculture*. (2013) 402–403:113–8. doi: 10.1016/j.aquaculture.2013.03.032
- Imsland AK, Reynolds P, Eliassen G, Hangstad TA, Foss A, Vikingstad E, et al. The use of lumpfish (*Cyclopterus lumpus* L.) to control sea lice (*Lepeophtheirus salmonis* Krøyer) infestations in intensively farmed Atlantic salmon (*Salmo salar* L.). *Aquaculture*. (2014) 424–425:18–23. doi: 10.1016/j.aquaculture.2013.12.033
- Treasurer JW. *Cleaner fish in aquaculture review and looking to the future*. Sheffield, UK: 5M publications (2018).
- Brooker AJ, Papadopoulou A, Gutierrez C, Rey S, Davie A, Migaud H. Sustainable production and use of cleaner fish for the biological control of sea lice: recent advances and current challenges. *Vet Rec*. (2018) 183:383. doi: 10.1136/vr.104966
- Jonassen T, Remen M, Lekva A, Steinarsson A, Árnason T. “Transport of lumpfish and wrasse”, in *Fish Biology and Aquaculture Applications*. ed. J. Treasurer (Sheffield, UK: 5M Publishing Ltd) (2018).
- Remen M, Jonassen T. Utvikling av transport-og mottaksprosedyrer for rognkjeks basert på kartlegging av miljø og stress. *Akvaplan-niva rapport*. (2017) 7707-1:40–48.
- Geitung L, Wright DW, Oppedal F, Stien LH, Vågseth T, Madaro A. Cleaner fish growth, welfare and survival in Atlantic salmon sea cages during an autumn-winter production. *Aquaculture*. (2020) 528:735623. doi: 10.1016/j.aquaculture.2020.735623
- Barton BA, Iwama GK. Physiological changes in fish from stress in aquaculture with emphasis on the response and effects of corticosteroids. *Annu Rev Fish Dis*. (1991) 1:3–26. doi: 10.1016/0959-8030(91)90019-G
- Pottinger TG. “The stress response in fish—mechanisms, effects and measurement”, in *Fish Welfare*. ed. E. J. Branson (Oxford, UK: Blackwell publisher) (2008):32–48.
- Harmon TS. Methods for reducing stressors and maintaining water quality associated with live fish transport in tanks: a review of the basics. *Rev Aquac*. (2009) 1:58–66. doi: 10.1111/j.1753-5131.2008.01003.x
- Davis MW. Fish stress and mortality can be predicted using reflex impairment. *Fish Fish*. (2010) 11:331. doi: 10.1111/j.1467-2979.2009.00331.x
- Budy P, Thiede GP, Bouwes N, Petrosky C, Schaller H. Evidence linking delayed mortality of Snake River salmon to their earlier hydrosystem experience. *N Am J Fish Manag*. (2002) 22:35–51. doi: 10.1577/1548-8675(2002)022<0035:ELDMOS>2.0.CO;2
- Scott AP, Ellis T. Measurement of fish steroids in water—a review. *Gen Comp Endocrinol*. (2007) 153:392–00. doi: 10.1016/j.ygcen.2006.11.006
- Ruane NM, Komen H. Measuring cortisol in the water as an indicator of stress caused by increased loading density in common carp (*Cyprinus carpio*). *Aquaculture*. (2003) 218:685–93. doi: 10.1016/S0044-8486(02)00422-2
- Ellis T, Bagwell N, Pond M, Baynes S, Scott AP. Acute viral and bacterial infections elevate water cortisol concentrations in fish tanks. *Aquaculture*. (2007) 272:707–16. doi: 10.1016/j.aquaculture.2007.07.235
- Fanouraki E, Papandroulakis N, Ellis T, Mylonas C, Scott A, Pavlidis M. Water cortisol is a reliable indicator of stress in European sea bass, *Dicentrarchus labrax*. *Behaviour*. (2008) 145:1267–81. doi: 10.1163/156853908785765818
- Höglund E, Fernandes P, Rojas-Tirado P, Rundberget JT, Hess-Erga O-K. Assessing stress resilience after Smolt transportation by waterborne cortisol and feeding behavior in a commercial Atlantic Salmon (*Salmo salar*) grow-out recirculating aquaculture system. *Front Physiol*. (2022) 12:2419. doi: 10.3389/fphys.2021.771951
- Wedemeyer G. *Physiology of fish in intensive culture systems*. Springer New York, NY: Springer Science & Business Media (1996).
- Iversen M, Eliassen RA. The effect of AQUI-S<sup>®</sup> sedation on primary, secondary, and tertiary stress responses during Salmon Smolt, *Salmo salar* L., transport and transfer to sea. *J World Aquacult Soc*. (2009) 40:216–25. doi: 10.1111/j.1749-7345.2009.00244.x
- Sandodden R, Finstad B, Iversen M. Transport stress in Atlantic salmon (*Salmo salar* L.): anaesthesia and recovery. *Aquac Res*. (2001) 32:87–90. doi: 10.1046/j.1365-2109.2001.00533.x
- Iversen MH, Økland F, Thorstad EB, Finstad B. The efficacy of AQUI-S vet. (iso-eugenol) and metomidate as anaesthetics in European eel (*Anguilla anguilla* L.), and their effects on animal welfare and primary and secondary stress responses. *Aquac Res*. (2013) 44:1307–16. doi: 10.1111/j.1365-2109.2012.03140.x
- Jerez-Cepa I, Fernández-Castro M, Alameda-López M, González-Manzano G, Mancera J, Ruiz-Jarabo I. Transport and recovery of gilthead seabream (*Sparus aurata* L.) sedated with AQUI-S<sup>®</sup> and etomidate: effects on intermediary metabolism and osmoregulation. *Aquaculture*. (2021) 530:735745. doi: 10.1016/j.aquaculture.2020.735745
- Soldatov A. Functional effects of the use of anesthetics on teleostean fishes. *Inland Water Biol*. (2021) 14:67–77. doi: 10.1134/S1995082920060139
- McWhinney BC, Briscoe SE, Ungerer JP, Pretorius CJ. Measurement of cortisol, cortisone, prednisolone, dexamethasone and 11-deoxycortisol with ultra high performance liquid chromatography–tandem mass spectrometry: application for plasma, plasma ultrafiltrate, urine and saliva in a routine laboratory. *J Chromatogr B*. (2010) 878:2863–9. doi: 10.1016/j.jchromb.2010.08.044
- Ellis T, James J, Stewart C, Scott A. A non-invasive stress assay based upon measurement of free cortisol released into the water by rainbow trout. *J Fish Biol*. (2004) 65:1233–52. doi: 10.1111/j.0022-1112.2004.00499.x
- Jerez-Cepa I, Fernández-Castro M, Del Santo O'Neill TJ, Martos-Sitcha JA, Martínez-Rodríguez G, Mancera JM, et al. Transport and recovery of gilthead seabream (*Sparus aurata* L.) sedated with clove oil and MS-222: effects on stress axis regulation and intermediary metabolism. *Front Physiol*. (2019) 10:612. doi: 10.3389/fphys.2019.00612

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## Conflict of interest

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30. Iversen M, Finstad B, McKinley RS, Eliassen RA. The efficacy of metomidate, clove oil, Aquil-S<sup>TM</sup> and Benzoak<sup>®</sup> as anaesthetics in Atlantic salmon (*Salmo salar* L.) smolts, and their potential stress-reducing capacity. *Aquaculture*. (2003) 221:549–66. doi: 10.1016/S0044-8486(03)00111-X
31. Small BC. Effect of isoeugenol sedation on plasma cortisol, glucose, and lactate dynamics in channel catfish *Ictalurus punctatus* exposed to three stressors. *Aquaculture*. (2004) 238:469–81. doi: 10.1016/j.aquaculture.2004.05.021
32. Davidson GW, Davie PS, Young G, Fowler RT. Physiological responses of rainbow trout *Oncorhynchus mykiss* to crowding and anesthesia with AQUIL-S<sup>TM</sup>. *J World Aquacult Soc.* (2000) 31:105–14. doi: 10.1111/j.1749-7345.2000.tb00704.x
33. Davis KB, Griffin BR. Physiological responses of hybrid striped bass under sedation by several anesthetics. *Aquaculture*. (2004) 233:531–48. doi: 10.1016/j.aquaculture.2003.09.018
34. Hill JV, Forster ME. Cardiovascular responses of Chinook salmon (*Oncorhynchus tshawytscha*) during rapid anaesthetic induction and recovery. *Comp Biochem Physiol Toxicol Pharmacol*. (2004) 137:167–77.
35. Jonassen T, Foss A, Remen M, Watts EJ, Hangstad TA. Toleranse for transport og miljøovergang hos berggylt og rognkjeks. (2019). Report No.: 9081-1.
36. Korte SM, Koolhaas JM, Wingfield JC, McEwen BS. The Darwinian concept of stress: benefits of allostasis and costs of allostatic load and the trade-offs in health and disease. *Neurosci Biobehav Rev*. (2005) 29:3–38. doi: 10.1016/j.neubiorev.2004.08.009



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# Humane slaughter in Mediterranean sea bass and bream aquaculture: farm characteristics, stakeholder views, and policy implications

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In many countries, increasing concern for animal welfare is driving retailer commitments and government legislation that aim to improve the lives of farmed fish. One aspect of fish welfare involves stunning fish prior to slaughter. The feasibility of stunning depends on the species of fish and physical farm characteristics. In this article, we provide an overview of stunning before slaughter in European sea bass and sea bream aquaculture, one of the largest finfish farming industries in the developed world that does not yet stun most of its production. Sea bass and sea bream stunning necessitates the use of electrical stunning equipment aboard harvest vessels, often a significant distance from the shoreline; this presents an interesting engineering and policy challenge. Together, Türkiye, Greece, Spain, and Italy produced over 400,000 t of sea bass and sea bream in 2020. In Türkiye and Greece, farms are numerous and located very close to the shoreline. In Spain and Italy, farms are few and located far from the shoreline. The highest average production is found in farms from Türkiye (1,000 t) and Spain (1,300 t), and lower average production is found in Greece (300 t) and Italy (350 t). Producer progress towards the installation of electrical stunning appears comparatively well-developed for Türkiye, Spain, and Greece, though we emphasise that producers and other stakeholders require continued support to realise this opportunity. Producers in Italy appear slower to make progress on this aspect of animal welfare and may require additional support.

## KEYWORDS

*Dicentrarchus labrax*, dry stunning, European Union, *Sparus aurata*, Turkey, wet stunning

## 1 Introduction

The concept of animal welfare is emerging as an increasingly important aspect of sustainable food systems (United Nations Environment Programme, 2019; Coghlan et al., 2021). In many countries, increasing consumer concern for animal welfare is driving retailer commitments and government legislation that aim to improve the lives of animals farmed for food (Alonso et al., 2020; Albalat et al., 2022; Wahlteiz et al., 2022; Wickens, 2022). These policies are increasingly targeted at aquatic animals in particular (Ashley, 2007; Stien et al., 2020; Crump et al., 2022). By scale, fish constitute one of the most numerous groups of farmed animals, exceedingly even the number of farmed chickens and pigs and behind only farmed invertebrates (Waldhorn and Autric, 2022; Klaura et al., 2023; Mood et al., 2023). As such, initiatives that aim to improve the lives of farmed fish can cause large overall benefits in the lives of animals.

Fish welfare interventions typically target one of three periods during the lives of farmed fish. Interventions can target the welfare of fish during breeding (the breeding stock or the juveniles) (Grimsrud et al., 2013; Tørud et al., 2019), the welfare of fish on-farm during the grow-out period (Pettersen et al., 2014; Stien et al., 2020), or the welfare of fish at slaughter (Lines and Spence, 2012; European Commission, 2017; Clemente et al., 2023).

One tractable way to improve the welfare of fish at slaughter is to stun fish before slaughter. Stunning involves rendering fish insensible, thus reducing the amount of suffering experienced by the fish when killed (Lines and Spence, 2012). Fish farming industries in many countries have made progress in implementing stunning before slaughter. However, one of the largest finfish farming industries that has not yet made significant progress in implementing stunning before slaughter is the European sea bass and sea bream farming industry.

In Europe, sea bass and sea bream are farmed along the Mediterranean coast. The countries with the highest production of sea bass and sea bream are Türkiye, Greece, Spain, and Italy (Table 1). Türkiye and Greece each have several hundred farms,

while Spain and Italy have only 24 farms each. The farms in Türkiye and Spain tend to have larger production volumes per farm, while the industries in Greece and Italy tend to be dispersed with farms having smaller production volumes.

Before turning to the details of humane slaughter, it helps to give a brief overview of the industry structure in these four countries. In Türkiye, 257,000 t of sea bass and sea bream are produced by 237 mostly large-scale sea cage farms and 173 mostly small-scale earthen pond farms (Çoban et al., 2020). In Türkiye, there are fish farms along the western and southern coastlines with the Mediterranean and the northern coastline with the Black Sea. In Greece, 100,000 t of sea bass and sea bream are produced by 347 sea cage farms (European Commission, n.d.). The eastern coastline appears to be more important, with many farms clustered within just a few hundred kilometres of Athens. Sea bass and sea bream farms in Greece are farmed using inshore floating sea cages (Pavlidis and Mylonas, 2011). In Spain and Italy, the farms are located along the length of the countries' coastlines with the Mediterranean. In Spain, 35,000 t of sea bass and sea bream are produced by 24 sea cage farms (Fishcount, 2017; Nielsen et al., 2021). In Italy, 14,000 t of sea bass and sea bream are produced by 24 sea cage farms. This equates to an average production per farm of 250 t (Italy – Eurofish.dk, n.d.; Hofherr et al., 2015; WWF, 2021). Spanish and Italian farms also mostly use floating sea cages, though a minority of production takes place in wetlands/brackish water (Nielsen et al., 2021).

When it comes to distance from shore, these four countries can be divided into two pairs (Figure 1A). Greece and Türkiye tend to have fish farms very close to the shoreline. Half of Greece's fish farms are located within 170 metres of the shoreline, while half of Türkiye's fish farms are located within 180 metres of the coastline (Hofherr et al., 2015). In contrast, Italy and Spain tend to have fish farms much further from the shoreline. Half of all fish farms in Italy are located within 990 metres from the coastline, while half of Spain's fish farms are located within 1,600 metres of the shoreline (Hofherr et al., 2015). The distance from shoreline determines the geographical and wave dynamics of farms, which in turn influences

TABLE 1 Description of the sea bass and sea bream aquaculture industry in the four key countries.

	Türkiye	Greece	Spain	Italy
Annual SBSB slaughter (number of fish)	346 million	235 million	86 million	37 million
Annual SBSB slaughter (weight in tonnes)	257,000 t	100,000 t	35,000 t	14,000 t
Description of farms (% of production)	Sea cages (96%) Earthen ponds (4%)	Sea cages	Sea cages (90%) Wetlands (10%)	Sea cages (78%) Land-based systems (16%) Wetlands (6%)
Number of farms in country	237 (sea cage farms) 173 (earthen pond farms)	347	24	24
Average production per farm	1,040 t (sea cages) 59 t (ponds)	305 t	1,270 t	350 t
Exports (percent of production)	75%	92%	40%	40%

Data summarised by the authors from sources provided in-text.

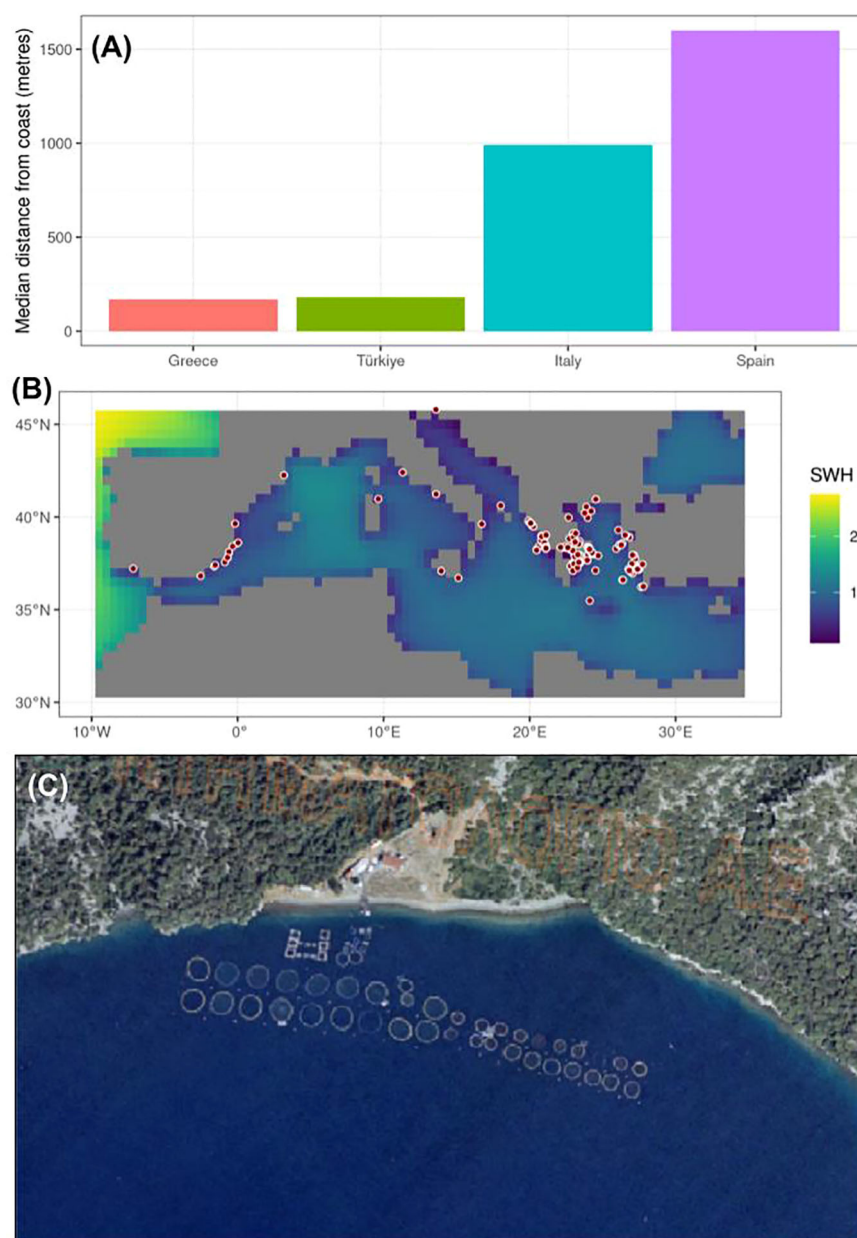


FIGURE 1

Farm physical characteristics. (A) median distance from coast by country; (B) coordinates of some farms (red points) visualised against mean significant wave height (SWH) throughout 2023 in metres; (C) aerial image of a typical fish farm, in this case a near-shore farm in Greece. For (B) data are the mean significant wave height 2001 to 2016 for farms with known coordinates (incomplete for Türkiye and Italy). For (C) the horizontal extent of the aerial photograph is 0.01 decimal degrees (roughly 800 m). Data in (A) from Hofherr et al. (2015); (B) from Hersbach et al. (2023), using Copernicus Climate Change Service information and farm coordinates from European Marine Observation and Data Network and Clawson et al. (2022); (C) from <https://gis.ktimanet.gr/under> CC BY 3.0.

the technology and equipment that farmers can adopt. We discuss this point further below.

In the remainder of this article, we examine the technological options for humane slaughter of sea bass and sea bream, before turning to a discussion of the industry perspectives, economic details, and engineering aspects associated with implementing humane slaughter in the European sea bass and sea bream industry.

## 2 Humane slaughter of sea bass and sea bream

### 2.1 Electrical stunning before slaughter

Today, almost all sea bass and sea bream are slaughtered by live chilling in ice or ice slurry, followed by a gill cut (European

Commission, 2017). During live chilling, fish generally remain conscious for between 5 and 40 minutes and display signs of suffering, including vigorous escape attempts (Van De Vis et al., 2003; Simitzis et al., 2013; Zampacavallo et al., 2015; de la Rosa et al., 2021). Brain activity continues even after fish become immobile (Lines and Spence, 2012).

During slaughter, animal welfare can be improved by stunning. Electrical stunning involves rendering fish unconscious using an electric current. Studies indicate that electrical stunning causes sea bass and sea bream to immediately become unconscious and lose sensibility (Giuffrida et al., 2007; Lambooi et al., 2007; Zampacavallo et al., 2015) (Panel on Animal Health and Welfare, 2009). It is important that stunning parameters (e.g. current and duration) are carefully selected, especially with reference to the method of electrical stunning, the handling method, and the subsequent killing method. Greater efficacy is generally achieved with higher currents and a longer duration (Robb et al., 2002; Lines et al., 2003; Jung-Schroers et al., 2020).

Electrical stunning can involve one of two methods: in-water and in-air. For in-water electrical stunning, fish are transported to a tank in batches using brailing and then exposed to the electrical field. Alternatively, fish can be captured using a pump and passed along a channel through which an electrical current is generated (Lines et al., 2003). In-air stunning is also possible but introduces the requirement to dewater the fish prior to stunning. Currently, producers of electrical stunning equipment for sea bass and sea bream include the UK-based company Ace Aquatec (in-water stunning), the Norway-based company Optimar (dry stunning), and the Türkiye-based company Smilefish.

The other major category of fish stunning is percussive slaughter. Percussive slaughter is not recommended for small fish, which includes sea bass and sea bream. This is because percussive stunning requires accuracy to be consistently maintained over time, which is difficult for small fish in a commercial setting (de la Rosa et al., 2021).

## 2.2 Current uptake of humane slaughter

Many Mediterranean sea bass and sea bream producers have expressed an interest in implementing electrical stunning. Currently, progress appears relatively promising in Spain, Greece, and Türkiye, with Italy appearing less promising.

In Spain and Greece, the major producer Avramar intends to implement electrical stunning this decade. Specifically, Avramar has publicly committed to implementing, by 2027, electrical stunning for 100% of its production in Spain and 50% of its production in Greece. Currently, Avramar has electrical stunning installed on two of its farms in Greece (Avramar, 2023). Likewise, the producer Philosfish has begun installing on its harvest vessels electrical stunnings produced by Ace Aquatec (The Fish Site, 2023).

In Türkiye, a study conducted by the NGO Future for Fish provides insight into the status of electrical stunning at Turkish fish farms (Future for Fish, 2023). In collaboration with academic advisors, Future for Fish contacted and visited numerous aquaculture companies, accounting for 76% of sea bass and sea

bream production in Türkiye. A key finding showed that the majority (90 percent) of Turkish aquaculture companies already have electrical stunning systems. Companies with electrical stunning systems do not always use those systems; only around 40% of companies use electrical stunning for over 95% of the harvesting process. The primary driver of Turkish farmers' adoption of electrical stunning systems is demand from customers (e.g. retailers in the UK and the Netherlands). Turkish fish farmers tend to use electrical stunning systems that are produced and sold in Türkiye, by companies such as Smilefish. Such systems have not been evaluated scientifically for their efficacy in successfully and consistently stunning fish.

In Italy, a survey of 21 sea bass and sea bream farms found that none of the surveyed farms have adopted electrical stunning (Clemente et al., 2023). The slaughter methods used by the surveyed farms were "thermal shock" (20 farms) and "asphyxia in air" (one farm).

## 3 Industry perspectives and implementation of humane slaughter

### 3.1 Industry perspectives

The most detailed information about industry perspectives on humane fish slaughter in Türkiye comes from the recent stakeholder survey conducted by Future for Fish (Future for Fish, 2023). The main challenge with electrical stunning systems in Türkiye is the initial installation aboard harvest vessels. The research also found other valuable insights. Implementing electrical stunning in operations leads to a decrease in labour due to "better organisation and occupational safety on harvest ships", better product quality, and possibly a longer shelf life. Fish producers believe that electrical stunning is more challenging for sea bass in particular, due to the fish getting stuck in the fish pumps leading to haemorrhages on their skin. A similar survey was conducted in Italy, but the negligible uptake of electrical stunning in the Italian sea bass and sea bream industry means that the survey did not produce detailed information about farmers' views on those species (Clemente et al., 2023).

### 3.2 Economic costs and funding

To our knowledge, there are three main manufacturers of electrical stunning equipment for European aquaculture: Ace Aquatec (United Kingdom), Optimar (Norway), and Smilefish (Türkiye). Surveys of manufacturers reveal that installing an in-water electrical stunner aboard a harvest vessel would cost a farmer somewhere in the vicinity of 150,000 €, though stunnings produced by Smilefish may cost a different amount (European Commission, 2017). This cost excludes expenses associated with any necessary modification to ships.

When it comes to making the investment necessary to purchase electrical stunning equipment, company size emerges as a critical factor (European Commission, 2017). Larger enterprises, with more resources, may integrate stunning equipment more seamlessly,



while smaller companies might face greater challenges, indicating a need for tailored strategies. A stakeholder consultation conducted by the European Parliament's Committee on Fisheries found that European fish farmers most often mention investment funds as the main obstacle to adopting new techniques, followed by justification of need (Pavlidis et al., 2023). Direct funding is mentioned as one of the most useful tools for mitigating potential impacts when transitioning to new fish welfare requirements.

Funding through the European Maritime Fisheries and Aquaculture Fund offers support, with specific budgets allocated for animal health and welfare improvements (e.g. 4M € for Italy, 8M € for Spain, and 560,000 € for Greece). A considerable portion of the funds allocated during the 2014–2020 period for aquaculture innovation remained unused. A lack of transparency in grants hitherto makes it difficult to assess how much money was used on animal welfare improvements. There are however known examples of the previous budget being used for a sea bream and sea bass stunning pilot in Spain. In general, transparency in funding allocation is a concern, with unclear information on recipients and purposes.

### 3.3 Engineering aspects

The physical, economic, and geographical details of sea bass and sea bream farms vary by country (Figure 1). Therefore, it is reasonable to expect that some countries will have an easier time than others in implementing electrical stunning. Spain in particular appears to have farms located in areas of the ocean with large waves, which could mean that air entering the fish pumps could be an engineering challenge in that country (Figure 1B). Size differences of fish can also cause the design sizes of pumps to be exceeded, likely leading to fish suffering and damaged product (Future for Fish, 2023).

The difficulty of installing new stunning equipment onboard vessels depends on the space available on the vessel, the type of vessel, and the installation requirements for the stunning equipment. Research and data on the types of vessels being used is scarce. Vessel types include platforms, workboats, and larger wellboats. All vessel types have been observed used for harvesting operations. Platforms are floating barges equipped with outboard engines and a crane. Workboats are larger, better equipped vessels ranging from 12 to 30 metres (Paleo et al., 2000). For larger, workboat-type vessels, installation of stunning equipment and fish pumps should not pose large problems. Equipment can be installed on deck or, for in-water stunners beneath deck. For smaller platform vessels, the limited deck space and the lack of below-deck area pose problems for installing a stunner and pump system. The pre-existing power on smaller vessels might not be enough to power both the crane and stunning system.

The study by Future for Fish found that 60% of interviewed farmers mentioned that stunning works faster than live chilling, meaning that stunning equipment may actually offer advantages for the harvest rate and associated labour requirements (Future for Fish, 2023). For some types of systems adding electrical stunning led to a slower harvest. Other technical obstacles include malfunctions in the equipment, long waiting times for spare parts and increased operating costs.

## 4 Discussion

As consumers develop an appreciation for animal welfare and its role as a component of sustainable food production, the priority placed on animal welfare by producers will only increase. In this report, we have examined the trajectory towards stunning before slaughter in European sea bass and sea bream aquaculture. The current progress towards installing electrical stunners aboard harvest vessels provides optimism, but succeeding in this policy goal will require concerted action by stakeholders throughout the supply chain (McAfee et al., 2019).

Sea bass and sea bream producers have expressed an interest in installing electrical stunners where this is economically attractive, and some producers have begun to do so (Avramar, 2023; Future for Fish, 2023; The Fish Site, 2023). However, success is greatest when there is a clear demand from retailers for stunned fish (Future for Fish, 2023). This mirrors the dynamic in other agricultural sectors like poultry and pork, where demand from retailers drives improvements in on-farm practices (Scrinis et al., 2017; Peacock and Mendez, 2020). As such, retailers have an important responsibility to ensure that their procurement policies account for consumers' increasing demand for animal welfare.

Likewise, producers are likely to respond to clear signals from government, whether at the national or the EU level. The EU is home to some of the world's most progressive pieces of animal welfare legislation, though legislation for the welfare of fish specifically has been lagging behind (McCulloch, 2018; Giménez-Candela et al., 2020). Nevertheless, the EU has expressed interest in legislation that would make stunning before slaughter mandatory for fish farmers (Dullaghan, 2023). The EU is also exploring the possibility of applying EU farm animal welfare legislation to imported meat products, which could have important repercussions for producers in Türkiye who export their product to EU Member States (Dullaghan, 2023). When particular higher-welfare practices become mandatory, this can help ensure that all producers within a country have access to the same markets (Carey et al., 2017; Carey et al., 2020; Department for Environment, Food and Rural Affairs, 2017). It can be useful for governments or other stakeholders to support farmers in the transition to higher-welfare practices by providing funding and other support, a strategy used successfully when installing CCTV in slaughterhouses in Great Britain (Department for Environment, Food and Rural Affairs, 2017; Springlea, 2022). As such, support from stakeholders throughout the supply chain can drive collective action to improve the lives of animals used for food.

## Data availability statement

The original contributions presented in the study are included in the article/supplementary material, further inquiries can be directed to the corresponding author/s.

## Author contributions

KV: Conceptualization, Funding acquisition, Investigation, Writing – original draft, Writing – review & editing. MC:

Investigation, Writing – original draft, Writing – review & editing. RR: Conceptualization, Formal analysis, Funding acquisition, Investigation, Project administration, Visualization, Writing – original draft, Writing – review & editing.

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## References

- Albalat, A., Zacarias, S., Coates, C. J., Neil, D. M., and Planellas, S. R. (2022). Welfare in farmed decapod crustaceans, with particular reference to *Penaeus vannamei*. *Front. Mar. Sci.* 9. doi: 10.3389/fmars.2022.886024
- Alonso, M. E., González-Montaña, J. R., and Lomillos, J. M. (2020). Consumers' Concerns and perceptions of farm animal welfare. *Anim. (Basel)* 10. doi: 10.3390/ani10030385
- Ashley, P. J. (2007). Fish welfare: Current issues in aquaculture. *Appl. Anim. Behav. Sci.* 104, 199–235. doi: 10.1016/j.applanim.2006.09.001
- Avramar (2023) *Responsible Aquaculture* (Avramar). Available online at: <https://avramar.eu/story/sustainability/> (Accessed January 11, 2024).
- Carey, R., Parker, C., and Scrinis, G. (2017). Capturing the meaning of “free range”: The contest between producers, supermarkets and consumers for the higher welfare egg label in Australia. *J. Rural Stud.* 54, 266–275. doi: 10.1016/j.jrurstud.2017.06.014
- Carey, R., Parker, C., and Scrinis, G. (2020). How free is sow stall free? Incremental regulatory reform and industry co-optation of activism. *Law Policy* 42, 284–309. doi: 10.1111/lapo.12154
- Clawson, G., Kuempel, C. D., Frazier, M., Blasco, G., Cottrell, R. S., Froehlich, H. E., et al. (2022). Mapping the spatial distribution of global mariculture production. *Aquaculture* 553, 738066. doi: 10.1016/j.aquaculture.2022.738066
- Clemente, G. A., Tolini, C., Boscarino, A., Lorenzi, V., Dal Lago, T. L., Benedetti, D., et al. (2023). Farmed fish welfare during slaughter in Italy: survey on stunning and killing methods and indicators of unconsciousness. *Front. Vet. Sci.* 10. doi: 10.3389/fvets.2023.1253151
- Çoban, D., Didem Demircan, M., and Tosun, D. D. (2020). *Marine Aquaculture in TURKEY: Advances and Management*. (Istanbul, Türkiye: Turkish Marine Research Foundation).
- Coghlan, S., Coghlan, B. J., Capon, A., and Singer, P. (2021). A bolder One Health: expanding the moral circle to optimize health for all. *One Health Outlook* 3, 21. doi: 10.1186/s42522-021-00053-8
- Crump, A., Browning, H., Schnell, A. K., Burn, C., and Birch, J. (2022). Invertebrate sentience and sustainable seafood. *Nat. Food* 3, 884–886. doi: 10.1038/s43016-022-00632-6
- de la Rosa, I., Castro, P. L., and Ginés, R. (2021). Twenty years of research in Seabass and Seabream welfare during slaughter. *Anim. (Basel)* 11. doi: 10.3390/ani11082164
- Department for Environment, Food and Rural Affairs. (2017). *Improving animal welfare: Closed Circuit Television (CCTV) in Slaughterhouses: Impact Assessment* (Defra, London, United Kingdom: Department for Environment, Food and Rural Affairs). Available at: [https://consult.defra.gov.uk/farm-animal-welfare/cctv-in-slaughterhouses/supporting\\_documents/CCTV%20internal%20impact%20assessment%20%20final.pdf](https://consult.defra.gov.uk/farm-animal-welfare/cctv-in-slaughterhouses/supporting_documents/CCTV%20internal%20impact%20assessment%20%20final.pdf).
- Dullaghan, N. (2023). *EU Farmed Fish Policy Reform Roadmap* (San Francisco, United States: Rethink Priorities). Available at: <https://rethinkpriorities.org/publications/eu-farmed-fish-policy-reform-roadmap-brief>.
- European Commission. (2017). *Welfare of farmed fish: Common practices during transport and at slaughter* (Brussels, Belgium: Directorate-General for Health and Food Safety, European Commission). Available at: [https://publications.europa.eu/resource/cellar/facddd32-cda6-11e7-a5d5-01aa75ed71a1.0001.01/DOC\\_1](https://publications.europa.eu/resource/cellar/facddd32-cda6-11e7-a5d5-01aa75ed71a1.0001.01/DOC_1).
- European Commission. (n.d). *Eurostat* (Luxembourg: Eurostat, European Commission). Available at: <https://ec.europa.eu/eurostat/data/database>.
- Fishcount. (2017). *Estimated numbers of individuals in aquaculture production (FAO) of fish species, (2017)* (United Kingdom: Fishcount). Available at: <http://fishcount.org.uk/studydatascreens2/2017/numbers-of-farmed-fish-B0-2017.php>.
- Future for Fish. (2023). *Electrical Stunning System: Türkiye Review* (Türkiye: Future for Fish). Available at: <https://futureforfish.org/wp-content/uploads/2024/01/ELECTRICAL-STUNNING-SYSTEM-REPORT-TURKIYE.pdf>.
- Giménez-Candela, M., Saraiva, J. L., and Bauer, H. (2020). The legal protection of farmed fish in Europe – analysing the range of EU legislation and the impact of international animal welfare standards for the fishes in European aquaculture. *Derecho Anim. Forum Anim. Law Stud.* 11, 65. doi: 10.5565/rev/da.460
- Giuffrida, A., Pennisi, L., Ziino, G., Fortino, L., Valvo, G., Marino, S., et al. (2007). Influence of slaughtering method on some aspects of quality of gilthead seabream and smoked rainbow trout. *Vet. Res. Commun.* 31, 437–446. doi: 10.1007/s11259-007-3431-8
- Grimsrud, K. M., Nielsen, H. M., Navrud, S., and Olesen, I. (2013). Households' willingness-to-pay for improved fish welfare in breeding programs for farmed Atlantic salmon. *Aquaculture* 372–375, 19–27. doi: 10.1016/j.aquaculture.2012.10.009
- Hersbach, H., Bell, B., Berrisford, P., Biavati, G., Horányi, A., Muñoz Sabater, J., et al. (2023). *ERA5 monthly averaged data on single levels from 1940 to present* (Reading, United Kingdom: European Centre for Medium-Range Weather Forecasts, Reading). doi: 10.24381/cds.f17050d7
- Hofherr, J., Natale, F., and Trujillo, P. (2015). Is lack of space a limiting factor for the development of aquaculture in EU coastal areas? *Ocean Coast. Manage.* 116, 27–36. doi: 10.1016/j.ocecoaman.2015.06.010
- Italy – Eurofish.dk (n.d). Available online at: <https://www.eurofish.dk/Italy> (Accessed November 11, 2021).
- Jung-Schroers, V., Hildebrandt, U., Retter, K., Esser, K.-H., Hellmann, J., Kleingeld, D. W., et al. (2020). Is humane slaughtering of rainbow trout achieved in conventional production chains in Germany? Results of a pilot field and laboratory study. *BMC Vet. Res.* 16, 197. doi: 10.1186/s12917-020-02412-5
- Klaura, J., Breeman, G., and Scherer, L. (2023). Animal lives embodied in food loss and waste. *Sustain. Production Consumption* 43, 308–318. doi: 10.1016/j.spc.2023.11.004
- Lambooi, B., Gerritzen, M. A., Reimert, H., Burggraaf, D., André, G., and Van De Vis, H. (2007). Evaluation of electrical stunning of sea bass (*Dicentrarchus labrax*) in seawater and killing by chilling: welfare aspects, product quality and possibilities for implementation. *Aquacult. Res.* 39, 50–58. doi: 10.1111/j.1365-2109.2007.01860.x
- Lines, J. A., Robb, D. H., Kestin, S. C., Crook, S. C., and Benson, T. (2003). Electric stunning: a humane slaughter method for trout. *Aquacult. Eng.* 28, 141–154. doi: 10.1016/S0144-8609(03)00021-9

- Lines, J. A., and Spence, J. (2012). Safeguarding the welfare of farmed fish at harvest. *Fish Physiol. Biochem.* 38, 153–162. doi: 10.1007/s10695-011-9561-5
- McAfee, D., Doubleday, Z. A., Geiger, N., and Connell, S. D. (2019). Everyone loves a success story: optimism inspires conservation engagement. *Bioscience* 69, 274–281. doi: 10.1093/biosci/biz019
- McCulloch, S. P. (2018). Brexit and animal protection: legal and political context and a framework to assess impacts on animal welfare. *Anim. (Basel)* 8. doi: 10.3390/ani8110213
- Mood, A., Lara, E., Boyland, N. K., and Brooke, P. (2023). Estimating global numbers of farmed fishes killed for food annually from 1990 to 2019. *Anim. Welf.* 32, e12. doi: 10.1017/awf.2023.4
- Nielsen, R., Guillen, J., and Virtanen, J. (2021). *Scientific, Technical and Economic Committee for Fisheries (STECF) - The EU Aquaculture Sector – Economic report 2020 (STECF-20-12)* (Luxembourg: Publications Office of the European Union). Available at: <https://stecf.jrc.ec.europa.eu/documents/43805/2783239/STECF+20-12+-+EU+Aquaculture+economics.pdf/ef242822-3343-43f4-b0a3-dfad889dd52c?version=1.0>.
- Paleo, J. D. B., Muir, J., and Turner, R. (2000). *Offshore mariculture: Workboats*. Available online at: <http://om.ciheam.org/om/pdf/b30/00600659.pdf>.
- Panel on Animal Health and Welfare. (2009). Species-specific welfare aspects of the main systems of stunning and killing of farmed fish: Rainbow Trout. *EFSA J.* 1013, 1–55. doi: 10.2903/j.efsa.2009.1012
- Pavlidis, M. A., and Mylonas, C. C. (2011). *Sea bream: biology and aquaculture of gilthead sea bream and other species* (Ames, Iowa: Wiley-Blackwell).
- Pavlidis, M., Papaharisis, L., Jung-Schroers, M. A. D. S., Kristiansen, T., Theodoridi, A., and Lourido, F. O. (2023). *Animal welfare of farmed fish* (Strasbourg, France: European Parliament). Available at: [https://www.europarl.europa.eu/thinktank/en/document/IPOL\\_STU\(2023\)747257](https://www.europarl.europa.eu/thinktank/en/document/IPOL_STU(2023)747257).
- Peacock, J., and Mendez, S. (2020). *Measuring Better Chicken Commitment-Compliant Chicken Supply*. (Rockville, MD, United States: The Humane Leagues Labs). doi: 10.31219/osf.io/8v2k9
- Pettersen, J. M., Bracke, M. B. M., Midtlyng, P. J., Folkedal, O., Stien, L. H., Steffenak, H., et al. (2014). Salmon welfare index model 2.0: an extended model for overall welfare assessment of caged Atlantic salmon, based on a review of selected welfare indicators and intended for fish health professionals. *Rev. Aquac.* 6, 162–179. doi: 10.1111/raq.12039
- Robb, D. H. F., O'Callaghan, M., Lines, J. A., and Kestin, S. C. (2002). Electrical stunning of rainbow trout (*Oncorhynchus mykiss*): factors that affect stun duration. *Aquaculture* 205, 359–371. doi: 10.1016/S0044-8486(01)00677-9
- Scrinis, G., Parker, C., and Carey, R. (2017). The caged chicken or the free-range egg? The regulatory and market dynamics of layer-hen welfare in the UK, Australia and the USA. *J. Agric. Environ. Ethics* 30, 783–808. doi: 10.1007/s10806-017-9699-y
- Simitzis, P. E., Tsopelakos, A., Charismiadou, M. A., Batzina, A., Deligeorgis, S. G., and Miliou, H. (2013). Comparison of the effects of six stunning/killing procedures on flesh quality of sea bass (*Dicentrarchus labrax*, Linnaeus 1758) and evaluation of clove oil anaesthesia followed by chilling on ice/water slurry for potential implementation in aquaculture. *Aquac. Res.* 45 (11), 1759–1770. doi: 10.1111/are.12120
- Springlea, R. (2022). *CCTV cameras in slaughterhouses: Modest benefits for animal welfare* (London, United Kingdom: Animal Ask). Available at: <https://www.animalask.org/post/cctv-cameras-in-slaughterhouses-modest-benefits-for-animal-welfare>.
- Stien, L. H., Bracke, M., Noble, C., and Kristiansen, T. S. (2020). “Assessing fish welfare in aquaculture,” in *The Welfare of Fish*. Eds. T. S. Kristiansen, A. Fernö, M. A. Pavlidis and H. van de Vis (Springer International Publishing, Cham), 303–321.
- The Fish Site (2023) *Philosophy commits to electrical stunners* (The Fish Site). Available online at: <https://thefishsite.com/articles/philosophy-commits-to-electrical-stunners> (Accessed January 11, 2024).
- Torud, B., Jensen, B. B., Gåsnes, S., Grønbech, S., and Gismervik, K. (2019). *Animal welfare in fish hatcheries - SMÅFISKVEL* (Ås, Norway: Norwegian Veterinary Institute). Available at: <https://dyrevern.no/app/uploads/2019/12/Animal-welfare-in-fish-hatcheries-SMAFISKVEL.pdf>.
- United Nations Environment Programme (2019). *Why is Animal Welfare Important for Sustainable Consumption and Production?* (Nairobi, Kenya: UN Environment). Available at: <https://www.unep.org/resources/perspective-series/issue-no-34-why-animal-welfare-important-sustainable-consumption-and>.
- Van De Vis, H., Kestin, S., Robb, D., Oehlenschläger, J., Lambooi, B., Münkner, W., et al. (2003). Is humane slaughter of fish possible for industry? *Aquac. Res.* 34, 211–220. doi: 10.1046/j.1365-2109.2003.00804.x
- Wahlteiz, S. J., Stacy, N. I., Hadfield, C. A., Harms, C. A., Lewbart, G. A., Newton, A. L., et al. (2022). Perspective: Opportunities for advancing aquatic invertebrate welfare. *Front. Vet. Sci.* 9, 973376. doi: 10.3389/fvets.2022.973376
- Waldhorn, D. R., and Autric, E. (2022). Shrimp: The animals most commonly used and killed for food production. Preprint. doi: 10.31219/osf.io/b8n3t
- Wickens, S. (2022). Review of the evidence of sentience in cephalopod molluscs and decapod crustaceans. *Anim. Welf.* 31, 155–156. doi: 10.1017/S0962728600009866
- WWF (2021). *SEA BASS AND SEA BREAM SUPPLY CHAIN STUDY: FROM TURKEY TO EUROPE*. Available online at: [https://www.fishforward.eu/wp-content/uploads/2021/07/WWF\\_supply\\_chain\\_study\\_2021\\_seabass\\_seabream.pdf](https://www.fishforward.eu/wp-content/uploads/2021/07/WWF_supply_chain_study_2021_seabass_seabream.pdf) (Accessed December 6, 2021).
- Zampacavallo, G., Parisi, G., Mecatti, M., Lupi, P., Giorgi, G., and Poli, B. M. (2015). Evaluation of different methods of stunning/killing sea bass (*Dicentrarchus labrax*) by tissue stress/quality indicators. *J. Food Sci. Technol.* 52, 2585–2597. doi: 10.1007/s13197-014-1324-8



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# Behavior analysis of juvenile steelhead trout under blue and red light color conditions based on multiple object tracking

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**Introduction:** The lighting environment significantly influences fish behavior. This study explores the impact of diverse lighting conditions on the behavior of steelhead trout (*Oncorhynchus mykiss*) to illuminate the mechanisms underlying their behavioral responses.

**Methods:** This experiment was set up with six treatments at a constant light intensity of 150 lx: 12h white light + 12h dark (12 W), 12h blue light + 12h dark (12B), 12h red light + 12h dark (12R), 1.5h blue light + 9h red light + 1.5h blue light + 12h dark (3B9R), 3h blue light + 6h red light + 3h blue light + 12h dark (6B6R), total 12h of blue and red light + 12h dark (T12BR). A multiple object tracking method, YOLOv5 with SORT, was employed to capture the movement trajectory of each fish, quantifying three motion metrics: swimming velocity, swimming angular velocity, and generalized intersection over union.

**Results:** The results revealed that fish exposed to 12R light environment showed significantly higher activity levels than other groups. The mixed light environments (3B9R, 6B6R) formed significant differences in behavioral metrics with 12R earlier than pure light environments (12B, 12W, T12BR), indicating sudden light color changes should be avoided. Fish in the 3B9R environment exhibited the lowest activity level but highest growth performance, with the highest specific growth rate of  $1.91 \pm 0.12$  d<sup>-1</sup>, a value significantly surpassing the lowest recorded rate, supported by a p-value of 0.0054, indicating it is suitable for steelhead trout cultivation.

**Discuss:** Behavioral significant differences were observed as early as week eight, much earlier than physiological differences, which became apparent by week 16. Overall, this paper employs computer vision methods to study the impact of different light colors on fish behavior, found that 3B9R is the optimal lighting condition tested and sudden light color changes should be avoided, offering a new perspective on light conditions and behavior in steelhead trout cultivation.

## KEYWORDS

steelhead trout, fish behavior, behavior quantify, aquaculture environment regulation, light color



# 1 Introduction

In aquaculture, the breeding environment factors, such as light, temperature, and dissolved oxygen, significantly influence fish growth. Light, which consists of light intensity, light color, and photoperiod, serves as a fundamental environmental factor playing a crucial role in the growth process of fish (Villamizar et al., 2011; Zhang et al., 2020). Fish have evolved light-sensing mechanisms throughout their evolution to adapt to environmental light conditions, thereby lighting conditions affect fish behavior (Xu et al., 2019; Noureldin et al., 2021). Existing research has found that light spectrum conditions can influence the growth, immune response, and digestive and metabolic capabilities of steelhead trout (Chen et al., 2022a, 2022b). Light affects fish metabolism, growth, and behavior through visual stimulation, and appropriate lighting conditions are beneficial for improving fish welfare and promoting growth (Ruchin, 2020). Fish behavior reflects their growth and physiological status, and analyzing behavior is conducive to identifying suitable light environments.

Light color (spectrum), as a fundamental factor of light condition (Zhang et al., 2020), has a significant impact on fish physiological traits due to their long-evolutionary adaptations, subsequently affecting fish growth and behavior. Some studies have found that specific light colors can enhance the growth of fish, for instance, yellow light for pearl gourami (*Trichopodus leerii*) (Heydarnejad et al., 2017), green light for barfin flounder (*Verasper moseri*) (Takahashi et al., 2016, 2018; Yamanome et al., 2009), red light for pikeperch (*Sander lucioperca*) (Baekelandt et al., 2019) and yellow perch (*Perca flavescens*) (Head and Malison, 2007), blue light for goldfish (*Carassius auratus*) (Noureldin et al., 2021) and turbot (*Scophthalmus maximus*) (Wu et al., 2021). Additionally, there are previous studies that focus on the influence of light color on fish behavior. For example, juvenile Nile tilapia showed a preference for yellow and red light (Luchiari and Oliveira, 2014). Young grass carp avoid red light and prefer blue (Mu et al., 2019). Juvenile bighead carp (*Aristichthys nobilis*) swim significantly slower under yellow and green light than under red and blue light (Xi et al., 2019). Goldfish (*Carassius auratus*) reared under blue light exhibited a lower ventilation frequency, longer reaction time, and longer swimming duration (Noureldin et al., 2021) in the novel object test, a behavioral assessment where fish are presented with an unfamiliar object to measure their exploratory and stress responses (Norton et al., 2011; Norton and Gutiérrez, 2019). Both behavior and growth are long-term outcomes of fish physiological traits. Analyzing fish behavior is conducive to identify suitable light environments for fish cultivation. However, previous studies mainly focused on the effects of single light color on fish growth and behavior, and there is limited information on the effects of light combination on fish behavior.

With the advancements in deep learning and computer vision in recent years, computer vision based methods are increasingly being employed in the study of fish behavior (Li et al., 2020; Yang et al., 2020; Li, D. et al., 2022). Fish multiple object tracking technology has demonstrated increasingly powerful capabilities,

efficiently and accurately capturing fish trajectories (Li, W. et al., 2022; Mandel et al., 2023; Liu et al., 2024). Compared to manual and acoustic-based methods, computer vision-based fish multiple object tracking approaches offer lower costs and higher efficiency, enabling rapid processing of large volumes of videos to obtain fish behavior information for further study. Currently, fish multiple object tracking methods have been widely applied in fish behavior and aquaculture research. For instance, Wang et al. proposed an end-to-end detection and tracking method to identify abnormal behaviors in porphyry seabream (Wang et al., 2022b). Li et al. introduced a detection and tracking method to study the behavior of *Oplegnathus punctatus* in the ammonia nitrogen environment (Li et al., 2023). Additionally, Xiao et al. developed a tracking method based on attention regions to investigate fish behavior for water quality monitoring (Xiao et al., 2016).

The spectral distribution of light in water is greatly influenced by depth (Liu et al., 2016; Hou et al., 2019). Specially, blue light can penetrate up to 200 meters in clear water, while red light is mostly absorbed within the first 20 meters (Sanchez-Vazquez et al., 2019; Ruchin, 2020). Juvenile steelhead trout reside in shallow freshwater streams (Hartman, 1965; Bugert et al., 1991), and the average swimming depth for adult steelhead trout is around 1.6 meters (Ruggerone et al., 1990), where both blue and red light can be perceived by fish. In this study, considering the natural light environment of steelhead trout, a variety of experimental conditions were established with mixed light colors and varying durations, employing computer vision and artificial intelligence techniques to quantify fish behavior in their daily activities. The growth and immunity performance of steelhead trout under different light color combinations has been studied (Chen et al., 2022b) as well as their digestive and anabolic capabilities (Chen et al., 2022a) based on this experiment. The computer vision-based object detection method YOLOv5 (Jocher et al., 2022) and the multiple object tracking method SORT (Bewley et al., 2016) were introduced to capture fish trajectories. The potential relationship between quantified behavioral traits and physiological characteristics was discussed. Therefore, this study demonstrates the feasibility of using behavioral quantification methods to guide the regulation of fish growth environments. The proposed quantification process in this paper can provide guidance for precise environmental control.

## 2 Materials and methods

### 2.1 Animals

Triploid steelhead trout eyed eggs were purchased from Troutlodge, Inc. (Washington, USA) and hatched at the Wanze Feng Fishery Company (Rizhao, Shandong, China). Before starting the experiment, the juvenile trout were acclimatized to a brackish saltwater environment (salinity:  $14.2 \pm 0.7$ ) for 2 weeks. In this stage, the fish were given twice-daily feedings to apparent satiation at 08:00 and 18:30 with a commercial trout feed from Greatseven Inc. (Qingdao, China). The controlled environment was



maintained with a constant 24-hour oxygen supply and a 12 light (L):12 dark(D) photoperiod.

## 2.2 Trial design

Six light color treatments were established with a photoperiod of 12L:12D (light period: 07:30-19:30) and total light intensity of 150 lx: 12 h white light + 12 h dark (12 W, full spectrum); 12 h blue light + 12 h dark (12B, peak at 454.9 nm); 12 h red light + 12 h dark (12 R, peak at 614.8 nm); 1.5 h blue light + 9 h red light + 1.5 h blue light + 12 h dark (3B9R, peaks at 454.9 nm and 614.8 nm, respectively); 3 h blue light + 6 h red light + 3 h blue light + 12 h dark (6B6R, peaks at 454.9 nm and 614.8 nm, respectively); 12 h blue light and red light + 12 h dark (T12BR, peaks at 454.9 nm + 614.8 nm, respectively), as illustrated in Figure 1. The light intensity of each treatment ensured to be consistent by calculating the average from measurements taken at the surface, middle, and bottom layers of the water. This guaranteed that the spectral composition of light was the only varying factor between treatments. By eliminating the influence of fluctuating light intensities, the differential impacts of light spectra on fish behavior can be analysed more accurately. The light was provided by packaged LED (COB) designed and produced by Qingdao Lanchi Technology Company. High-precision electronic adjustment of light intensity is not supported. Instead, light intensity was ensured by adjusting the distance between the light source and the water surface. Light intensity and spectra were measured using a handheld illuminometer (PLA300; Everfine Inc., Hangzhou, China). The 16-week trial was conducted using a completely randomized block design with four replicates per

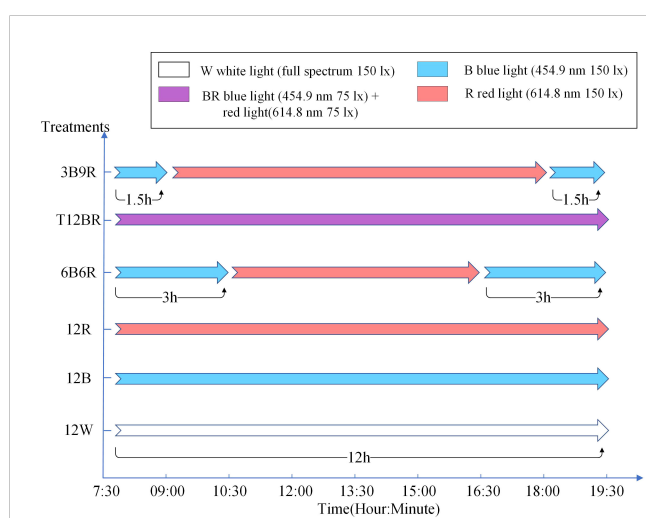
treatment, 20 fish per tank (380 L volume, 0.72 m height × 0.95 m diameter, white background). Each tank was shielded by black light-absorbing fabric, ensuring no light interference occurred between the various treatment groups. Fish were weighed by an electronic balance (Mettler-Toledo International, Inc., Greifensee, Switzerland).

At the the start and conclusion of the trail, there was a 36-hour halt in feeding to guarantee the digestive tracts of fish were empty. Before beginning the experiment, the fish were submerged in a solution containing 30 mg/L of tricaine methanesulfonate (MS-222; Sigma-Aldrich, USA) for anesthesia. And they were delicately dried using tissue before being weighed. Throughout the cultivation period, each trout, with an initial weight of  $34.67 \pm 2.69$  g, received commercial trout feed daily at 08:00 and 18:30 respectively. Residual unconsumed feed was collected after feeding for 30 min. The calculation of daily feed intake involved evaluating the moisture content of the feed and adjusting for any loss. The water was renewed via a single flow system with a flow rate of 1.15 L/min. Parameters such as water temperature, salinity, dissolved oxygen, and pH were measured three times daily using a YSI ProPlus handheld multiparameter meter (YSI Inc., USA). Sampling of the water occurred every three days, with subsequent analysis involving the Cleverchem 380 automatic chemical analyzer (DeChem-Tech Inc., Hamburg, Germany) to determine levels of total ammonia nitrogen (TAN), phosphate, nitrite nitrogen, and nitrate nitrogen.

Several measures were implemented in order to minimize noise interference in the experiment. Firstly, PVC damping trays (1.1m × 1.1m in width × 10 cm) were placed at the bottom of the rearing tanks. Secondly, the exterior of the rearing tanks and the black shielding fabric were covered with sound-absorbing cotton. Thirdly, nano-aeration discs and liquid oxygen were utilized for aeration. Fourthly, adherence to standardized experimental protocols was strictly maintained, minimizing disturbances during operations. Finally, the PVC pipes for water inflow and outflow were covered with shock-absorbing sound insulation materials.

During the experimental period, the following water quality parameters (mean ± standard deviation) were maintained: water temperature at  $16.5 \pm 0.2^\circ\text{C}$ , salinity at  $14.2 \pm 0.7$ , dissolved oxygen at  $8.7 \pm 0.3$  mg/L, pH value at  $7.3 \pm 0.1$ , ammonia nitrogen (TAN) at  $0.03 \pm 0.03$  mg N/L, phosphate at  $0.11 \pm 0.08$  mg P/L, nitrite nitrogen at  $0.09 \pm 0.05$  mg/L, and nitrate nitrogen at  $3.43 \pm 1.9$  mg N/L.

At 4-week intervals, the fish in each tank were anesthetized with 30 mg/L tricaine MS-222, counted, and weighed. The fish behavior was captured using 1080P infrared night vision video webcams (DS-IPC-T12) from Hikvision Digital Technology Co., Ltd. (Hangzhou, China) and saved as videos. The webcams, mounted on a bracket 1.5 m above each tank, recorded videos at a frame rate of 25 frames per second (FPS), capturing the fish swimming behavior from a top-down perspective. Videos were captured daily between 07:30 h and 19:30 h, stored on the hard disk, and each video lasted approximately one hour (resulting in 12 videos for 12 hours per day). The total experimental environment is shown in Figure 2. To prevent the influence of light color on the image-based method, the videos were captured in grayscale. Due to the substantial volume of



**FIGURE 1**  
setting of light color groups in this experiment. In this experiment, the photoperiod was set to 12L:12D, with 12 hours of light (from 7:30 to 19:30) and 12 hours of darkness every day. The naming convention for treatments is derived from the arrangement of light colors within the 12-hour light period, denoted as xCyD: C color light for X hours, D color light for y hours. For instance, '3B9R' denotes blue light for 3 hours and 9 hours of red light, and '12B' denotes blue light for 12 hours.

video data and the limitations in processing and analysis speed, a balance was struck between the representativeness of the sample and analytical efficiency. Segments of the videos that included fish feeding moments (30 minutes each in the morning and evening), were removed to prevent any interference with the swimming behavior analysis. Furthermore, the initial ten minutes of each processed hourly video were selected as a representative segment for that hour, which was then utilized for the subsequent tracking of fish trajectories and behavior analysis. Although the trial lasted 16 weeks, the records from the first 8 weeks were utilized for behavior analysis, as significant behavioral differences became apparent by the eighth week, which was earlier than significant differences in growth.

In this experiment, fish were reared in experimental tanks (380 L volume, 1.0 m height  $\times$  0.95 m diameter, with a white background), with 20 fish per tank. Tanks were covered with a back shielding fabric to prevent interference between different treatment groups. Illumination with various light colors was provided by LED sources (packaged LED (COB), Qingdao Lanchi Technology Company), and fish behavior was recorded from a top-down perspective using a camera (DS-IPC-T12, Hikvision Digital Technology Co., Ltd.), with videos stored on computer hard drives.

## 2.3 Fish behavior analysis

In this experiment, the YOLOv5 with Sort algorithm was employed to perform multiple object tracking on ten-minute video segments of six treatments, with twelve segments recorded each day to correspond with the 12L:12D photoperiod. This process generated frame-by-frame bounding boxes depicting the motion trajectories of individual fish. Subsequently, fish swimming behavior was quantified using three behavioral metrics: Swimming Velocity (SV), Swimming Angular Velocity (SAV), and Generalized Intersection over Union (GIOU). Finally, a statistical analysis was conducted to assess the impact of different light color environments on fish behavior. As for the possible social interactions (hierarchies) among fish, this study primarily focused on investigating the collective behavior of the entire fish population, particularly by measuring average values of metrics such as SV, SAV,

and GIOU under different light conditions. This approach naturally integrates the possible social interactions among fish, as these averages reflect the collective behavioral performance of all individuals under specific light conditions. By concentrating on the behavioral dynamics at the population level, the analysis implicitly including the influence of social structures on overall behavioral patterns without the need for direct quantification of specific social interactions among individual fish.

### 2.3.1 Multiple fish tracking

A subset of fish behavior videos was selected from the experiment to create a fish dataset through annotation. Training and performance comparisons of several popular Multiple Object Tracking (MOT) algorithms, including YOLOv5 with SORT and CenterTrack, were conducted using this dataset, as shown in Table 1 in section 3.1 of the Results. Metrics such as Mean Average Precision (mAP50), Multiple Object Tracking Accuracy (MOTA), and Identity F1 Score (IDF1) were employed for evaluation. YOLOv5 with SORT exhibited superior performance and was chosen in this trial. YOLOv5, an object detection algorithm that has been widely utilized in aquaculture studies (Li, X. et al., 2022; Wang et al., 2022a), can locate fish individuals in each frame with bounding boxes. SORT is an instance association algorithm that uses positional information to associate bounding boxes of the same fish individual across consecutive frames, resulting in fish motion trajectories, as shown in Figure 3. SORT was also widely employed for fish tracking (Wang et al., 2021; Gong et al., 2022; Zhang et al., 2023). In this experiment, YOLOv5 with SORT was employed for multiple object tracking of fish videos recorded, utilizing bounding boxes to represent fish motion trajectories, for subsequent behavioral quantification and analysis. Despite the inevitable presence of some false positives and false negatives, the comprehensive fish dataset, effective behavioral quantification metrics, and scientifically sound statistical methods employed ensure that the results accurately reflect the overall movement trajectories and behavioral states of the fish.

### 2.3.2 Quantification and calculation of behavior metrics

Using the bounding boxes obtained from tracking fish videos as input, three metrics were utilized to quantify fish behavior for subsequent statistical analysis: Swimming Velocity (SV), Swimming Angular Velocity (SAV), and Generalized Intersection over Union (GIOU). These behavioral metrics can be computed over three consecutive frames, as depicted in Figure 3. A sliding window approach was employed to sequentially select  $M$  frames throughout the entire video for calculating metrics, where  $M$  represent the window size with a default value of 3.

The Swimming Velocity reflects the displacement per frame, representing the straight swimming of fish, with higher values denoting a relatively higher activity level. The Swimming Angular Velocity is the change rate in the swimming direction of the fish, representing circular swimming behavior in fish, with higher values denoting a relatively lower activity level. Employing angles and distances to study fish behavior is a commonly used method in fish behavior analysis (Hang et al., 2021). GIOU provides a composite measure of the

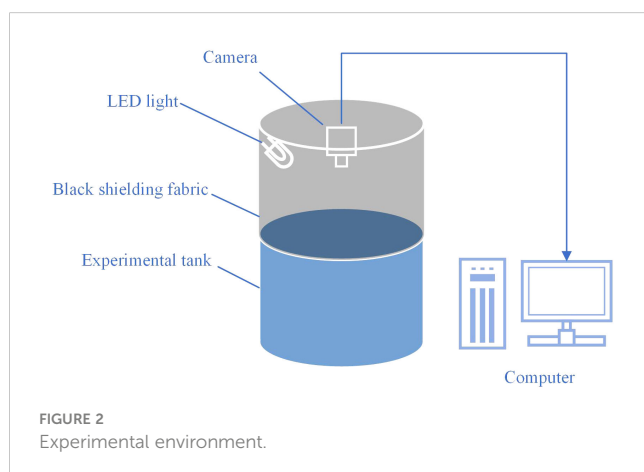


FIGURE 2  
Experimental environment.

TABLE 1 Evaluation Metrics for the performance of MOT(Multi Object Tracking) methods.

Method	mAP50 (%)	MOTA (%)	RcII (%)	IDF1 (%)	IDP (%)	IDR (%)
YOLOv5+SORT	95.7	87.4	92.2	47.7	48.7	46.9
CenterTrack	90.7	64.2	73.4	27.5	30.7	24.9

The mAP50 denotes mean Average Precision under a condition with Intersection over Union (IoU) threshold at 0.5, reflecting the performance of object detection. The MOTA denotes Multiple Object Tracking Accuracy, integrating false positives, misses, and identity switches, and reflects tracking performance. RcII, the recall of the model, evaluates the ability of model to find all positive samples, with higher values suggesting fewer missed objects. IDF1, IDP, and IDR focus on the continuity and identity preservation in tracking. IDP is the proportion of correctly identified matches by the model that maintain the correct identity, while IDR represents the proportion of all true identity matches correctly identified and maintained by the model. IDF1 is the harmonic mean of IDP and IDR, and these metrics collectively reflect the consistency of object identities throughout the tracking process.

overlap, distance, and dimension differences between two bounding boxes. Since the shape and size of the bounding box corresponds to the pose of the fish, and the displacement of the bounding box corresponds to the movement velocity of the fish, GIOU effectively encapsulates the fish behavioral dynamics. It can serve as an indicator of information representation in various fish poses and can be considered an expression of the stress response capability in fish (Huang et al., 2022).

The left side of Figure 3 shows the multiple object tracking results of a single fish across three consecutive frames, represented by moving bounding boxes that indicate the fish trajectory. On the right side, the figure illustrates the underlying principles and calculation methods for the three behavioral metrics: Swimming Velocity (SV), Swimming Angular Velocity (SAV), and Generalized Intersection over Union (GIOU). These metrics correspond to the following Equations 1, 2 and Algorithm 1.

The notation  $f_u$  was used to denote the  $u^{th}$  fish and the center motion of bounding boxes to represent the movement feature of fish swimming.  $(x_{j_{f_u}}, y_{j_{f_u}})$  means the center coordinates at  $j^{th}$  frame and the swimming velocity ( $v^{f_u}$ ), swimming angular velocity ( $\theta^{f_u}$ ) of fish in three consecutive frames were calculated as described in Equation 1 and Algorithm 1.

$$v_j^{f_u} = \frac{\sum_{m=1}^M \sqrt{(x_{j+m-1}^{f_u} - x_{j+m}^{f_u})^2 + (y_{j+m-1}^{f_u} - y_{j+m}^{f_u})^2}}{M} \quad (1)$$

**Input:**  
 $x, y$ : Coordinates of  $u^{th}$  fish positions in each frame  
 $M$ : Window size for computing swimming angular velocity

**Output:**  
 $\theta_j^{f_u}$  Swimming angular velocity of  $u^{th}$  fish in  $j^{th}$  frame

**Algorithm:**  
1: Initialize  $Angles \leftarrow \emptyset$   
2: **for** each  $m$  from 1 to  $M$  **do**  
    /\* Compute changes in angles of fish swimming velocities \*/

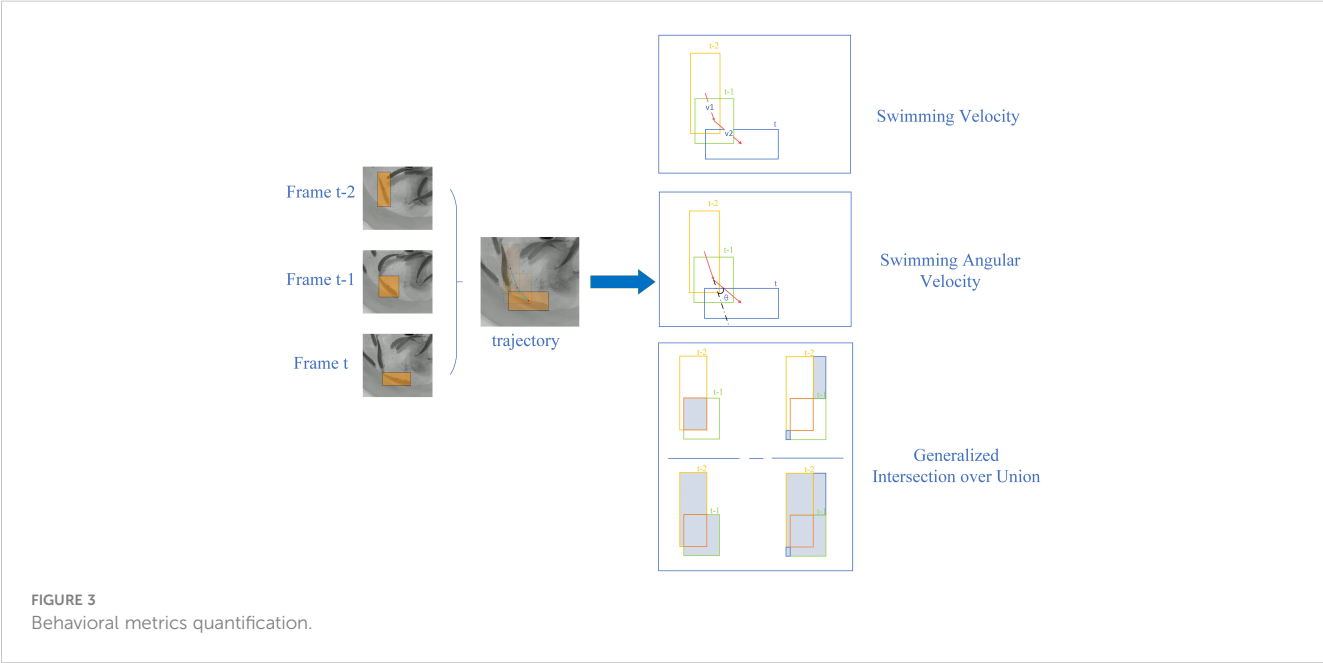


FIGURE 3 Behavioral metrics quantification.

```

3:  $\Delta x_1 = x_{j+m} - x_{j+m-1}$ 

4:  $\Delta y_1 = y_{j+m} - y_{j+m-1}$ 

5:  $\Delta x_2 = x_{j+m-1} - x_{j+m-2}$ 

6:  $\Delta y_2 = y_{j+m-1} - y_{j+m-2}$ 

7:  $\angle 1 = \arctan\left(\frac{\Delta y_1}{\Delta x_1}\right)$ 

8:  $\angle 2 = \arctan\left(\frac{\Delta y_2}{\Delta x_2}\right)$  /* Compute the absolute difference of
angles and convert to degrees */

9:  $\Delta angle_m = |\angle 1 - \angle 2| \cdot \frac{180^\circ}{\pi}$  /* Convert the difference of
angles less than  $180^\circ$  */

10:  $\Delta angle_m = \min(\Delta angle_m, 360^\circ - \Delta angle_m)$  /* Accumulate to
Angles */

11:  $Angles \leftarrow Angles \cup \{\Delta angle_m\}$ 

12: end for /* Compute the average angle change */

13:  $\theta_j^{fu} \leftarrow \frac{1}{|Angles|} \sum_{\alpha \in Angles} \alpha$ 

14: Return  $\theta_j^{fu}$ 

```

**Algorithm 1.** Calculate Swimming Angular Velocity of Fish.

$GIOU_j^{fu}$ , reflecting the degree of overlap of the fish bounding boxes in consecutive frames, which was calculated as Equation 2.  $A_{jk=m}^c$  denotes the area of min box which contains two bounding boxes of the same fish in consecutive two frames, IoU is the area ratio of the intersection and union of the two bounding boxes,  $U_{j+m}$  is the union area of the two bounding boxes.

$$GIOU_j^{fu} = \frac{1}{M} \sum_{m=1}^M \left( IoU_{jk=m} - \frac{A_{jk=m}^c - U_{jk=m}}{A_{jk=m}^c} \right) \quad (2)$$

Based on the principle and calculation of these behavioral metrics, it can be inferred that higher swimming speed indicates a higher fish activity level, while increased swimming angular velocity and GIOU values signify a reduced fish activity level. The integration of these indicators provides a comprehensive perspective to systematically evaluate the impact of different light color conditions on fish behavior by comparing the variations among groups based on these three metrics. This methodology enables us to identify the suitable light color conditions for steelhead trout.

### 2.3.3 Statistical analysis and methods

In the quantification of behavioral metrics, which included twelve values (12 hours per day) for each metric, two parts were conducted. The first part of the study used behavioral metrics from the last five days of the trial, serving as the final behavior effects of the first 8 weeks of the trial, for one-way analysis of variance (ANOVA). The results of the first part are in Sections 3.2 to 3.4 of the Results. The second part

utilized daily data on swimming velocity, each comprising 12 hourly values. The 12W group was used as a baseline to identify when significant differences (represented by p-values) between 12W and other groups stabilized. Given the significant differences of the 12R group with other groups, the 12R group was similarly employed as a baseline for conducting the same analysis. This approach allowed for a comprehensive analysis of the effects of different lighting conditions on fish behavior from various perspectives. The results of the second part are in Sections 3.5 of the Results.

The homoscedasticity ( $P > 0.05$ ) of data was evaluated using the Levene test and the normality ( $P > 0.05$ ) of data was assessed through the Shapiro-Wilk test. Subsequently, the following stratified approach was adopted to assess inter-group differences: For data that conformed to both normality and homogeneity of variance, ANOVA and LSD tests were utilized; for data that adhered to normality but not to homogeneity of variance, Welch ANOVA and Games-Howell tests were employed; for data that did not meet the normality criterion, Kruskal-Wallis H test and Dunn's Test were conducted. Statistical analyses were performed using Python packages including Scipy version 1.7.1, scikit\_posthocs version 0.8.1, pingouin version 0.5.4, and statsmodels version 0.14.1, along with IBM SPSS Statistics version 25.

## 2.4 Animal and ethics approval

All experimental methods of this study were performed under the Regulations of the Administration of Affairs Concerning Experimental Animals of China, as well as the Regulations of the Administration of Affairs Concerning Experimental Animals of Shandong Province.

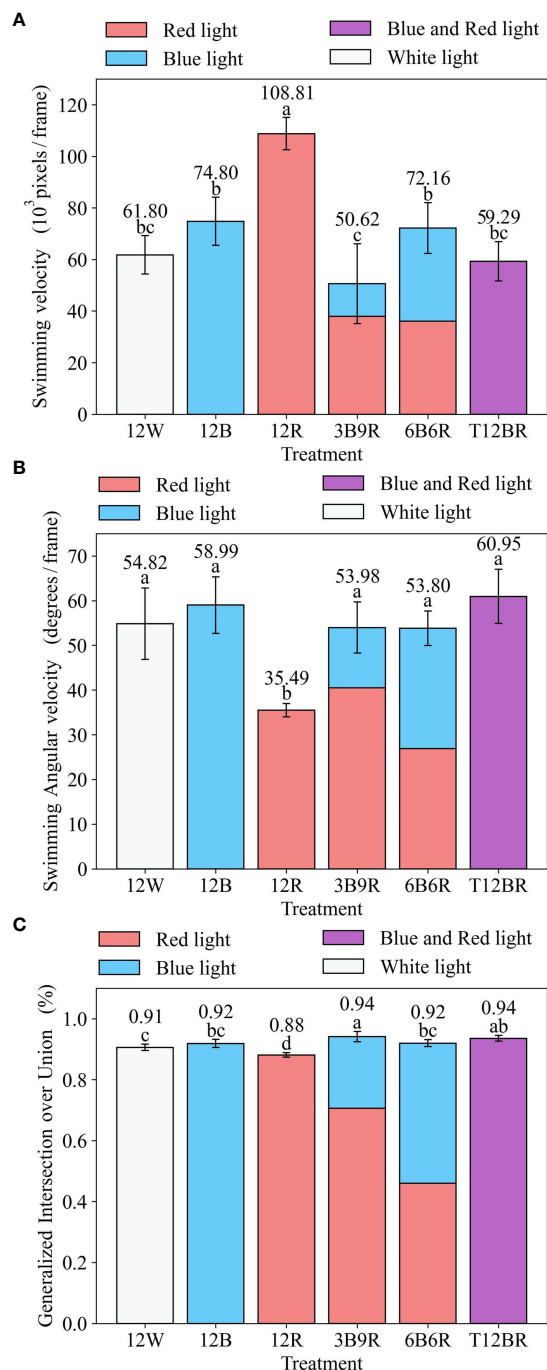
## 3 Results

### 3.1 Comparison of multiple object tracking methods

The comparative results between two Multiple Object Tracking (MOT) methods, YOLOv5+SORT and CenterTrack, are presented in Table 1. Both methods were trained and tested using a dataset collected and annotated as part of the experiment. The results showed that YOLOv5+SORT outperforms CenterTrack in the context of fish multiple object tracking within this study, evidenced by achieving a Mean Average Precision (mAP50) of 95.7%, a Multiple Object Tracking Accuracy (MOTA) of 87.4%, and an Identity F1 Score (IDF1) of 47.7%. These metrics collectively underscore the exemplary proficiency of YOLOv5+SORT method in target detection and identity association.

### 3.2 Swimming velocity

The effects of different light conditions on fish behavior metrics are shown in Figure 4. After the first 8 weeks of the trial, the results revealed that there was a significant difference ( $P < 0.01$ ) in



**FIGURE 4**  
Swimming velocity, Swimming Angular velocity, GIOU for juvenile steelhead trout. (A) Swimming velocity (B) Swimming angular velocity (C) Generalized Intersection over Union.

swimming velocity among the six light color treatments (as shown in Figure 4A). The swimming velocity of the 12R group was the highest ( $P < 0.01$ ), which significantly differed from the other groups. The fish reared under the 3B9R groups had the lowest swimming velocity, which was significantly lower than that of the 12R, 12B, and 6B6R groups.

In this Figure 4, panels (a–c) are statistical analyses of fish behavioral metrics (Swimming Velocity (SV), Swimming Angular Velocity (SAV),

and Generalized Intersection over Union (GIOU), respectively) at the end of the trial, reflecting the behavioral influences of fish in the trial. Different letters denote significant differences ( $P < 0.05$ ) between groups based on a one-way analysis of variance (ANOVA) with the Least Significant Difference (LSD) test. Results are expressed as the mean  $\pm$  standard deviation. The naming convention for treatments is derived from the arrangement of light colors within the 12-hour light period, denoted as xCyD: C color light for X hours, D color light for y hours. For instance, ‘3B9R’ denotes blue light for 3 hours and 9 hours of red light, and ‘12B’ denotes blue light for 12 hours.

### 3.3 Swimming angular velocity

The results indicated a significant variance in swimming angular velocities among the six light color groups ( $P < 0.01$ ), as illustrated in Figure 4B. Specifically, the fish in the 12R group exhibited the slowest swimming angular velocity ( $P < 0.01$ ), significantly differing from the other groups. The swimming angular velocities of the other groups were relatively similar, with T12BR displaying the highest value.

### 3.4 Generalized intersection over union

The results revealed significant differences in GIOU among the six light color groups ( $P < 0.01$ ), as depicted in Figure 4C. The fish in the 12R group exhibited the lowest GIOU, showing a significant distinction from the other groups. Conversely, the 3B9R group had the highest GIOU, significantly differing from the 12R, 12B, 12W, and 6B6R groups.

### 3.5 Dynamics of swimming velocity differences among treatments

Temporal dynamics of significant differences in swimming velocity between treatment groups were demonstrated in Figures 5, 6. Specifically, Figure 5 illustrates the variations in swimming velocity p-values and metrics for the 12W group in contrast with the pure color groups (12B, 12R, and T12BR), while Figure 7 presents these comparisons for the 12W group against the mixed color groups (6B6R and 3B9R). In a parallel fashion, Figure 8 details the changes in swimming velocity p-values and metrics for the 12R group relative to the pure color groups (12B, 12W, and T12BR), and Figure 6 does the same for the 12R group when compared to the mixed color groups (6B6R and 3B9R).

Comparative analysis of the 12W group revealed that stable significant differences were only observed with the 12R group by the conclusion of the trial ( $p < 0.05$ ). During the mid-phase of the trial (weeks 3 to 4), the 3B9R and 6B6R groups exhibited transitory significant differences when compared with the 12W group, which were not maintained in the long term.

Comparative analysis of the 12R group revealed that significant differences eventually manifested between this group and all other



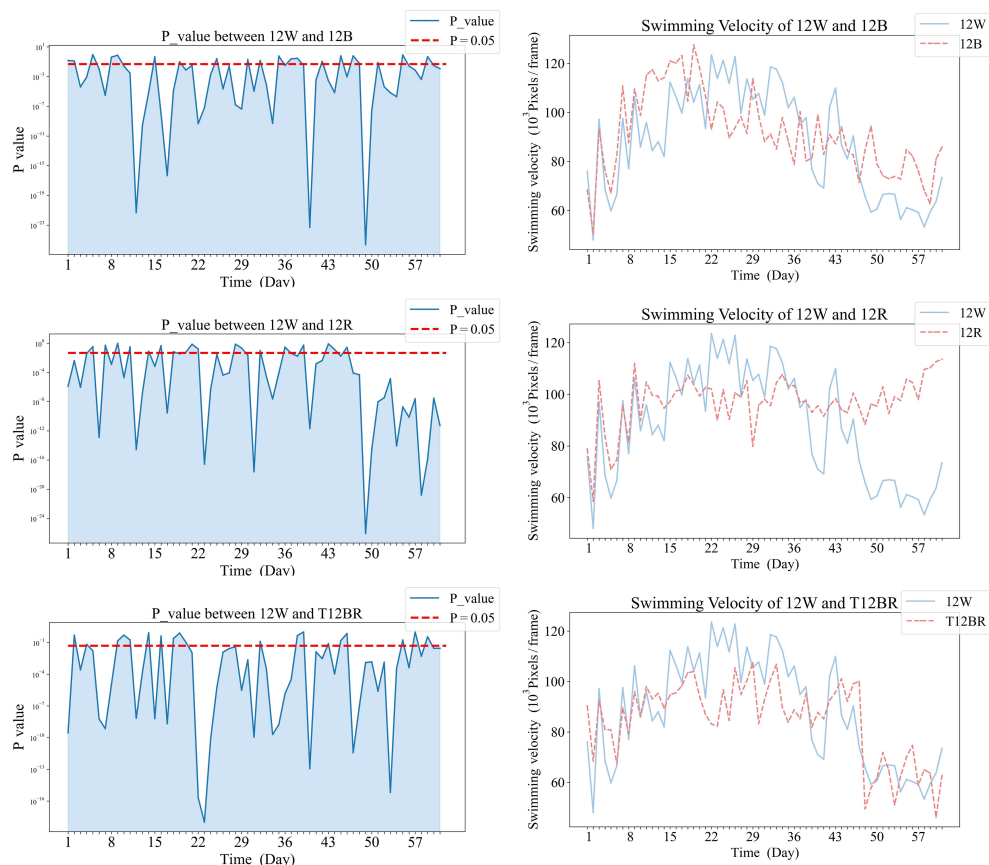


FIGURE 5

Comparative analysis of Swimming Velocity in 12W and pure color groups.

groups. Notably, the 3B9R and 6B6R groups displayed significant differences from the 12R group as early as the sixth week. In contrast, the 12W, 12B, and T12BR groups developed these differences more gradually, becoming apparent by the eighth week.

In Figure 5, panels (a1), (b1), and (c1) on the left display the daily variance in swimming velocity P-values between each pure color group (12B, 12R, T12BR) and the 12W group, over the initial eight-week period. Panels (a2), (b2), and (c2) on the right present

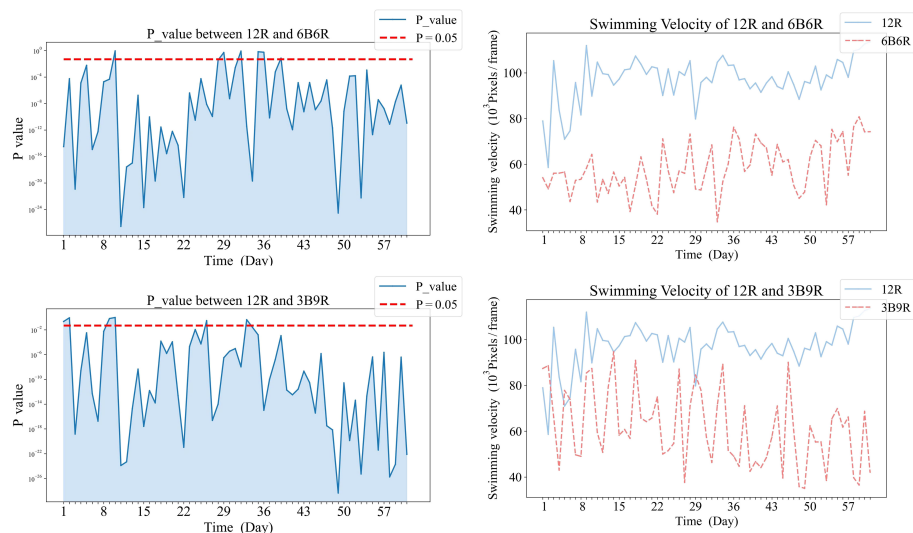


FIGURE 6

Comparative analysis of Swimming Velocity in 12R and mixed color groups.

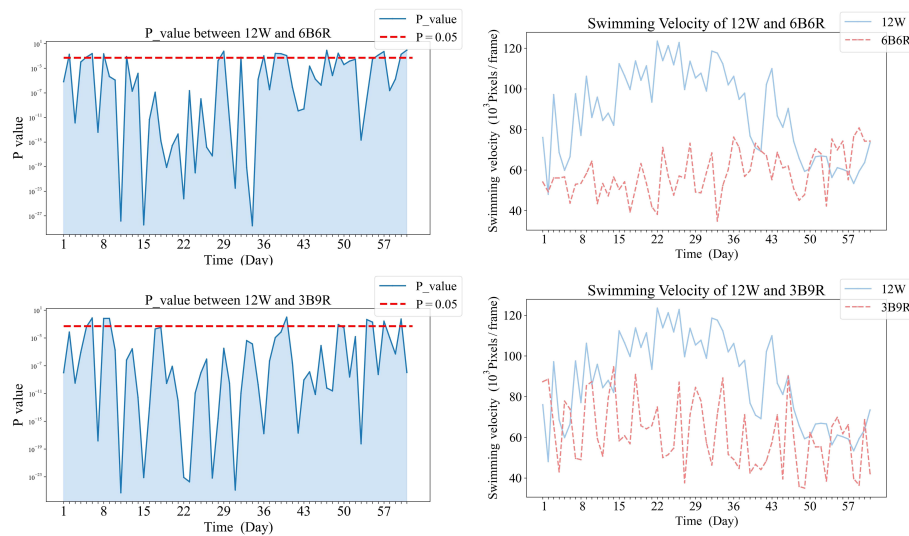


FIGURE 7  
Comparative analysis of Swimming Velocity in 12W and mixed color groups.

the corresponding swimming velocity values for these groups compared to the 12W group across the same duration. A P-value below 0.05 is indicative of a statistically significant difference. The naming convention for treatments is derived from the arrangement

of light colors within the 12-hour light period, denoted as xCyD: C color light for X hours, D color light for y hours. For instance, '3B9R' denotes blue light for 3 hours and 9 hours of red light, and '12B' denotes blue light for 12 hours.

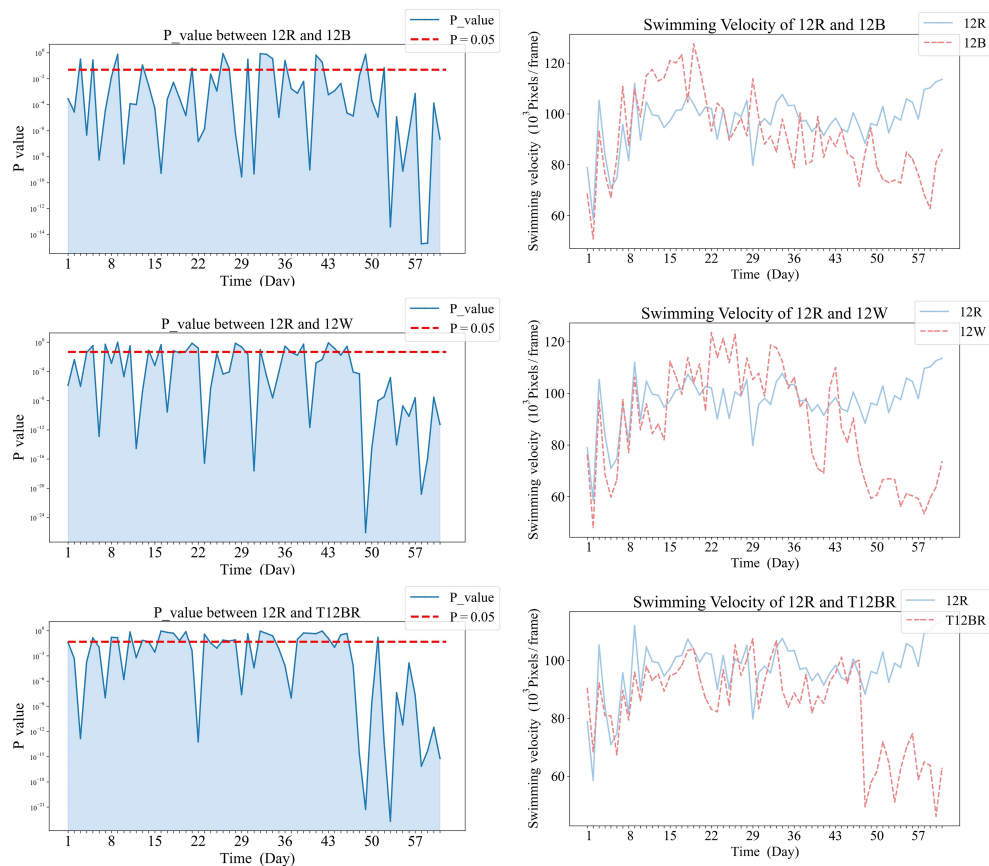


FIGURE 8  
Comparative analysis of Swimming Velocity in 12R and pure color groups.

In Figure 7, panels (a1), and (b1) on the left display the daily variance in swimming velocity P-values between each mixed color group (6B6R, 3B9R) and the 12W group, over the initial eight-week period. Panels (a2), and (b2) on the right present the corresponding swimming velocity values for these groups compared to the 12W group across the same duration. A P-value below 0.05 is indicative of a statistically significant difference. The naming convention for treatments is derived from the arrangement of light colors within the 12-hour light period, denoted as xCyD: C color light for X hours, D color light for y hours. For instance, '3B9R' denotes blue light for 3 hours and 9 hours of red light, and '12B' denotes blue light for 12 hours.

In Figure 8, panels (a1), (b1), and (c1) on the left display the daily variance in swimming velocity P-values between each pure color group (12B, 12W, T12BR) and the 12R group, over the initial eight-week period. Panels (a2), (b2), and (c2) on the right present the corresponding swimming velocity values for these groups compared to the 12R group across the same duration. A P-value below 0.05 is indicative of a statistically significant difference. The naming convention for treatments is derived from the arrangement of light colors within the 12-hour light period, denoted as xCyD: C color light for X hours, D color light for y hours. For instance, '3B9R' denotes blue light for 3 hours and 9 hours of red light, and '12B' denotes blue light for 12 hours.

In Figure 6, panels (a1), and (b1) on the left display the daily variance in swimming velocity P-values between each mixed color group (6B6R, 3B9R) and the 12R group, over the initial eight-week period. Panels (a2), and (b2) on the right present the corresponding swimming velocity values for these groups compared to the 12R group across the same duration. A P-value below 0.05 is indicative of a statistically significant difference. The naming convention for treatments is derived from the arrangement of light colors within the 12-hour light period, denoted as xCyD: C color light for X hours, D color light for y hours. For instance, '3B9R' denotes blue light for 3 hours and 9 hours of red light, and '12B' denotes blue light for 12 hours.

## 4 Discussion

In this study, the behavioral responses of steelhead trout under various light color combinations (12B, 12R, 12W, 3B9R, T12BR, 6B6R) were investigated to identify the most suitable light environment for steelhead trout cultivation. Light color is a significant factor in aquatic environments, with numerous studies investigating its effects on the physiology and behavior of cultured fish species. For example, A study on the impacts of blue, red, green, white, and natural sunlight on goldfish (*Carassius auratus*) found that blue light facilitated better growth and lower ventilation frequencies (Noureldin et al., 2021). It was found that steelhead trout are not recommended to be cultured under blue light, as it was associated with a reduction in liver total lipids and plasma glucose levels, along with an increase in serotonergic and dopaminergic activity in the brain (Karakatsouli et al., 2007). Meanwhile, another study found that steelhead trout exhibited good growth performance under yellow light Heydarnejad et al., 2013, indicating that different light colors have varying effects on steelhead trout. In contrast, this study found that the fish in the

12R group exhibited the highest swimming velocity and the lowest swimming angular velocity and GIOU, with significant differences ( $P < 0.05$ ) from the other groups. This indicates that fish under red light exhibit the highest amount of movement, preferring to swim over a large area rather than slowly wandering in a small radius. In the comparative analysis with the 12W group, it was found that only the 12R group exhibited significant differences with 12W in the end, indicating that red light has a greater behavioral impact on fish, as shown in Figures 5, 7.

Previous research has also discovered that steelhead trout exhibit an avoidance behavior towards red light (Luchiari and Pirhonen, 2008). The unique behavioral responses of steelhead trout under red light environment could potentially be explained by the physiological impacts of red light on fish. Several studies have found that under red light, steelhead trout exhibit hepatic stress, indicated by higher levels of alkaline phosphatase (ALP) ( $P < 0.05$ ), total bilirubin (T-BIL) ( $P < 0.05$ ), alanine aminotransferase (ALT), lysophosphatidic acid (LPA), and malondialdehyde (MDA) (Chen et al., 2022b); and hepatic oxidative stress, indicated by significantly decreased activities of catalase (CAT), glutathione peroxidase (GPx), glucose-6-phosphate dehydrogenase (G6PD), glutathione reductase (GR), glutathione S-transferase (GST), and acetylcholinesterase (AChE) in the liver (Guller et al., 2020).

The linkage between red light exposure and heightened fish activity may be a consequence of hepatic stress. As the liver plays a pivotal role in metabolic processes, its impaired function due to stress or oxidative damage may increase fish activity levels, likely as a stress response. Notably, this pattern of behavior under red light is not unique to steelhead trout. Similar hepatic stress responses have been observed in turbot (*Scophthalmus maximus*) (Wu et al., 2021), with juveniles exhibiting increased levels of Hsp70 mRNA and antioxidant enzymes when exposed to red and orange spectra. Moreover, goldfish (*Carassius auratus*) exposed to red light displayed reduced growth rates and heightened ventilation frequencies (Noureldin et al., 2021). Tiger puffer (*Takifugu rubripes*), too, exhibited avoidance behaviors to long-wavelength red light, with their movement rates correlating directly with light wavelength (Cai et al., 2023), echoing the increased activity levels noted in this study. These collective findings underscore the significant influence of red light on fish behavior and physiology. However, the specific mechanisms underpinning these effects, particularly how red light-induced hepatic stress and oxidative damage translate into altered behavior, warrant further investigation.

Beyond the impact of monochromatic light on fish, some studies have also investigated the effects of combinations of light colors on fish. For instance, a study on the influence of red, blue, green, white light, and three mixed light combinations on Nile tilapia (Elkadom et al., 2023), found that the growth performance under red/blue light was superior, whereas red light exhibited poorer growth performance, causing significant liver inflammation and damage, showing some similarities with this study. In contrast, this research investigates the effects of different light color combinations on the behavior of steelhead trout. It was discovered that the 3B9R group exhibited the lowest swimming velocity and highest GIOU,

indicating lower activity level and exercise expenditure than other groups. It has also been found (Chen et al., 2022b) that the 3B9R environment showed the best growth performance and no hepatic stress. This may be attributable to the spectral distribution under 3B9R conditions more closely to the natural light spectrum in trout habitats, thereby avoiding excessive stimulation and consequently resulting in lower activity levels. The spectral distribution of light in water is influenced by climatic conditions, incident angles, depth, and water quality (Liu et al., 2016; Hou et al., 2019). Light penetration in water varies with its wavelength; red light has poorer penetration. Due to longer light paths at dawn and dusk compared to noon (Wetzel, 1983; Ruchin, 2020), blue light dominates at the same depth in the morning and evening, while red light prevails at noon, which is similar to the light color distribution under the 3B9R condition. Thereby, 3B9R was considered a suitable light environment for trout cultivation.

In the comparative analysis based on the 12W and 12R groups, distinct behavioral patterns were observed in the mixed light groups (6B6R, 3B9R) as opposed to the pure light groups (12B, 12W, 12R, T12BR). Specifically, the results reveal that in the comparisons with the 12W group (showed in Figures 5, 7), mixed light groups (6B6R, 3B9R) briefly showed significant differences with 12W during the mid-term phase of the experiment, which subsequently disappeared, unlike the pure light groups (12B, 12R, T12BR). In the comparisons with the 12R group (showed in Figures 8, 6), the mixed color groups (6B6R, 3B9R) established long-term stable significant differences earlier, emerging by the sixth week, whereas the pure color groups (12B, 12W, T12BR) formed these differences at a later stage, becoming apparent by the eighth week, indicating a two-week advancement for the former. The key difference between mixed and pure light groups lies in the fact that 12W, 12B, 12R, and T12BR environments all provided light with constant spectral distribution, that is, a pure color light without spectral change (T12BR offers blue and red light, akin to violet light). In contrast, the 3B9R and 6B6R environments' light colors both have a blue-red-blue change, which might have led to the unique behavioral patterns observed in the mixed light groups. Previous research on the biochemical characteristics (Chen et al., 2022a) has found that fish in 3B9R and 6B6R environments showed myocardial stress with higher levels of lactate dehydrogenase (LDH). This physiological response may be correlated with the unique behavioral patterns of the mixed light groups. It suggests that a gradual light color transition phase is needed to facilitate fish adaptation and environments like 3B9R, and 6B6R should be cautious of abrupt light color changes.

Environmental light impacts fish behavior through the nervous system. Usually, light is detected by the eyes and the signal is sent to the brain for processing and interpretation. Some studies have delved into the effects of different environmental light colors on fish brain histology. For example, it was analyzed the brain histology of Nile tilapia under various light color environments and observed that red light exposure led to adverse histological changes in the brain (Elkadam et al., 2023), including significant neuronal degeneration, cerebral hemorrhaging, and pronounced

lesions. Conversely, fish exposed to red and blue light showed excellent cytoarchitecture. This may offer a potential explanation from a neurological and histological perspective for the negative results seen in this experiment under the 12R red light environment and the positive effects under the 6B6R and 3B9R red and blue light combinations. Further exploration is needed to deeply understand the mechanisms behind how light color influences fish behavior through the nervous system and brain histology.

This study currently still has several limitations, such as the accuracy of the MOT methods that could be further improved due to the lack of high-quality fish datasets. The behavioral metrics used primarily pertain to swimming behaviors of fish, with a dearth of metrics directly reflecting fish group interactions. Moreover, the internal mechanisms through which light affects histological, physiological, and biochemical features by impacting their neurological systems, leading to behavioral performance, remain unclear. In future research, more efficient and precise fish multiple object tracking methods could be employed, and additional fish behavior indicators could be proposed to comprehensively reflect both swimming behaviors and group interactions of fish. This would allow for a deeper investigation into the effects of light on the histological, physiological, and biochemical aspects of fish, and the underlying mechanisms of changes in fish behavior traits, thereby enhancing the understanding of fish behavior.

## 5 Conclusion

In the present study, the effects of various light color combinations on the behavior of juvenile steelhead trout were explored, integrating the analysis of how light color impacts on fish growth, physiology, and histology. Behavioral differences were discovered to emerge significantly earlier than physiological ones, manifesting as early as week eight. Notably, a distinct behavioral change was observed under the red light (12R) environment, characterized by increased swimming activity, which suggested the potential stress and discomfort experienced by trout. Consequently, the 12R environment appeared less favorable for long-term cultivation. Conversely, the mixed color light condition 3B9R, closely mimicked the natural light spectrum preferred by trout in their natural habitats, presenting a suitable light environment for steelhead trout cultivation where gradual light color transitions could help avoid stress. Existing brain histological analyses of fish also support these behavioral findings, particularly the adverse changes in brain histology observed under red light exposure contrasted sharply with more favorable outcomes under red and blue light conditions. Future research should continue to investigate the multifaceted effects of light on aquacultural species, employing more efficient and precise MOT methods for fish trajectory tracking. By utilizing more comprehensive behavioral metrics to quantify fish behavior, further studies can explore the potential relationships between fish behavior, physiology, and histology to identify more suitable aquacultural lighting environments, thereby enhancing fish welfare and productivity.



## Data availability statement

The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

## Ethics statement

All experimental methods of this study were performed under the Regulations of the Administration of Affairs Concerning Experimental Animals of China, as well as the Regulations of the Administration of Affairs Concerning Experimental Animals of Shandong Province. The study was conducted in accordance with the local legislation and institutional requirements.

## Author contributions

ZL: Data curation, Formal analysis, Investigation, Methodology, Software, Visualization, Writing – original draft. XC: Writing – review & editing, Investigation, Methodology. JH: Methodology, Software, Writing – review & editing. DA: Writing – review & editing. YZ: Writing – review & editing.

## References

- Baekelandt, S., Mandiki, S. N. M., Schmitz, M., and Kestemont, P. (2019). Influence of the light spectrum on the daily rhythms of stress and humoral innate immune markers in pikeperch *Sander lucioperca*. *Aquaculture* 499, 358–363. doi: 10.1016/j.aquaculture.2018.09.046
- Bewley, A., Ge, Z., Ott, L., Ramos, F., and Upcroft, B. (2016). "Simple online and realtime tracking," in *2016 IEEE International Conference on Image Processing (ICIP)*. AZ, USA: Phoenix. 3464–3468. doi: 10.1109/ICIP.2016.7533003
- Bugert, R. M., Bjornn, T. C., and Meehan, W. R. (1991). Summer habitat use by young salmonids and their responses to cover and predators in a small southeast alaska stream. *Trans. Am. Fisheries Soc.* 120, 474–485. doi: 10.1577/1548-8659(1991)120<0474:Shubys>2.3.CO;2
- Cai, H., Zhang, Y., Xiong, Y., Liu, Y., Sun, F., Zhou, Q., et al. (2023). Preference of juvenile tiger puffer for light spectrum and tank colours based on different body size and breeding background. *Animal* 17, 101021. doi: 10.1016/j.animal.2023.101021
- Chen, X., Zhou, Y., Huang, J., An, D., Li, L., Dong, Y., et al. (2022a). Blue and red light color combinations can enhance certain aspects of digestive and anabolic performance in juvenile steelhead trout *Oncorhynchus mykiss*. *Front. Mar. Sci.* 9. doi: 10.3389/fmars.2022.853327
- Chen, X., Zhou, Y., Huang, J., An, D., Li, L., Dong, Y., et al. (2022b). The effects of blue and red light color combinations on the growth and immune performance of juvenile steelhead trout, *Oncorhynchus mykiss*. *Aquaculture Rep.* 24, 101156. doi: 10.1016/j.aqrep.2022.101156
- Elkadom, E. M., Abd El-Kader, M. F., Bakr, B. A., Abozeid, A. M., and Mohamed, R. A. (2023). Impacts of various single and mixed colors of monochromatic LED light on growth, behavior, immune-physiological parameters, and liver and brain histology of Nile tilapia fingerlings. *Aquaculture* 577, 740007. doi: 10.1016/j.aquaculture.2023.740007
- Gong, L., Hu, Z., and Zhou, X. (2022). "A few samples underwater fish tracking method based on semi-supervised and attention mechanism," in *2022 6th international conference on robotics, control and automation (ICRCA)*. (Xiamen, China: IEEE), 18–22. doi: 10.1109/ICRCA55033.2022.9828911
- Guller, U., Onalan, S., Arabaci, M., Karatas, B., Yasar, M., and Kufrevioglu, O. I. (2020). Effects of different LED light spectra on rainbow trout (*Oncorhynchus mykiss*): in vivo evaluation of the antioxidant status. *Fish Physiol. Biochem.* 46, 2169–2180. doi: 10.1007/s10695-020-00865-x
- Hang, S., Zhao, J., Ji, B., Li, H., Zhang, Y., Peng, Z., et al. (2021). Impact of underwater noise on the growth, physiology and behavior of *Micropterus salmoides* in industrial recirculating aquaculture systems. *Environ. Pollut.* 291, 118152. doi: 10.1016/j.envpol.2021.118152
- Hartman, G. F. (1965). The Role of Behavior in the Ecology and Interaction of Underyearling Coho Salmon (*Oncorhynchus kisutch*) and Steelhead Trout (*Salmo gairdneri*). *J. Fisheries Res. Board Canada* 22, 1035–1081. doi: 10.1139/f65-095
- Head, A. B., and Malison, J. A. (2007). Effects of lighting spectrum and disturbance level on the growth and stress responses of yellow perch *perca flavescens*. *J. World Aquaculture Soc.* 31, 73–80. doi: 10.1111/j.1749-7345.2000.tb00700.x
- Heydarnejad, M. S., Fattollahi, M., and Khoshkam, M. (2017). Influence of light colours on growth and stress response of pearl gourami *Trichopodus leerii* under laboratory conditions. *J. Ichthyol.* 57, 908–912. doi: 10.1134/S0032945217060054
- Heydarnejad, M. S., Parto, M., and Pilevarian, A. A. (2013). Influence of light colours on growth and stress response of rainbow trout (*Oncorhynchus mykiss*) under laboratory conditions. *J. Anim. Physiol. Anim. Nutr. (Berl)* 97, 67–71. doi: 10.1111/j.1439-0396.2011.01243.x
- Hou, Z.-S., Wen, H.-S., Li, J.-F., He, F., Li, Y., Qi, X., et al. (2019). Effects of photoperiod and light Spectrum on growth performance, digestive enzymes, hepatic biochemistry and peripheral hormones in spotted sea bass (*Lateolabrax maculatus*). *Aquaculture* 507, 419–427. doi: 10.1016/j.aquaculture.2019.04.029
- Huang, J., Yu, X., Chen, X., An, D., Zhou, Y., and Wei, Y. (2022). Recognizing fish behavior in aquaculture with graph convolutional network. *Aquacul. Eng.* 98, 102246. doi: 10.1016/j.aquaeng.2022.102246
- Joher, G., Chaurasia, A., Stoken, A., Borovec, J., Kwon, Y., Michael, K., et al. (2022). ultralytics/yolov5: v6. 2-yolov5 classification models, apple m1, reproducibility, clearml and deci. ai integrations. *Zenodo*. doi: 10.5281/zenodo.7002879
- Karakatsouli, N., Papoutsoglou, S. E., Pizzonia, G., Tsatsos, G., Tsopekos, A., Chadio, S., et al. (2007). Effects of light spectrum on growth and physiological status of gilthead seabream *Sparus aurata* and rainbow trout *Oncorhynchus mykiss* reared under recirculating system conditions. *Aquacul. Eng.* 36, 302–309. doi: 10.1016/j.aquaeng.2007.01.005
- Li, J., Chen, W., Zhu, Y., Xuan, K., Li, H., and Zeng, N. (2023). Intelligent detection and behavior tracking under ammonia nitrogen stress. *Neurocomputing* 559, 126809. doi: 10.1016/j.neucom.2023.126809
- Li, X., Hao, Y., Zhang, P., Akhter, M., and Li, D. (2022). A novel automatic detection method for abnormal behavior of single fish using image fusion. *Comput. Electron. Agric.* 203, 107435. doi: 10.1016/j.compag.2022.107435
- Li, W., Li, F., and Li, Z. (2022). CMFTNet: Multiple fish tracking based on counterpoised JointNet. *Comput. Electron. Agric.* 198, 107018. doi: 10.1016/j.compag.2022.107018
- Li, D., Wang, G., Du, L., Zheng, Y., and Wang, Z. (2022). Recent advances in intelligent recognition methods for fish stress behavior. *Aquacul. Eng.* 96, 102222. doi: 10.1016/j.aquaeng.2021.102222

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## Conflict of interest

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- Li, D., Wang, Z., Wu, S., Miao, Z., Du, L., and Duan, Y. (2020). Automatic recognition methods of fish feeding behavior in aquaculture: A review. *Aquaculture* 528, 735508. doi: 10.1016/j.aquaculture.2020.735508
- Liu, Y., Li, B., Zhou, X., Li, D., and Duan, Q. (2024). FishTrack: Multi-object tracking method for fish using spatiotemporal information fusion. *Expert Syst. Appl.* 238, 122194. doi: 10.1016/j.eswa.2023.122194
- Liu, X., Zhang, Y., Shi, K., Lin, J., Zhou, Y., and Qin, B. (2016). Determining critical light and hydrologic conditions for macrophyte presence in a large shallow lake: The ratio of euphotic depth to water depth. *Ecol. Indic.* 71, 317–326. doi: 10.1016/j.ecolind.2016.07.012
- Luchiar, A. C., and Oliveira, J. J. (2014). The interaction of innate and imposed colour perception: a behavioural approach. *J. Ethol.* 32, 179–183. doi: 10.1007/s10164-014-0407-3
- Luchiar, A. C., and Pirhonen, J. (2008). Effects of ambient colour on colour preference and growth of juvenile rainbow trout *Oncorhynchus mykiss* (Walbaum). *J. Fish Biol.* 72, 1504–1514. doi: 10.1111/j.1095-8649.2008.01824.x
- Mandel, T., Jimenez, M., Risley, E., Nammoto, T., Williams, R., Panoff, M., et al. (2023). Detection confidence driven multi-object tracking to recover reliable tracks from unreliable detections. *Pattern Recognit.* 135, 109107. doi: 10.1016/j.patcog.2022.109107
- Mu, X., Zhen, W., Li, X., Cao, P., Gong, L., and Xu, F. (2019). A study of the impact of different flow velocities and light colors at the entrance of a fish collection system on the upstream swimming behavior of juvenile grass carp. *Water* 11, 322. doi: 10.3390/w11020322
- Norton, W. H., and Gutiérrez, H. C. (2019). The three-spined stickleback as a model for behavioural neuroscience. *PLoS One* 14, e0213320. doi: 10.1371/journal.pone.0213320
- Norton, W. H., Stumpfenhorst, K., Faus-Kessler, T., Folchert, A., Rohner, N., Harris, M. P., et al. (2011). Modulation of Fgfr1a signaling in zebrafish reveals a genetic basis for the aggression-boldness syndrome. *J. Neurosci.* 31, 13796–13807. doi: 10.1523/JNEUROSCI.2892-11.2011
- Noureddin, S. M., Diab, A. M., Salah, A. S., and Mohamed, R. A. (2021). Effect of different monochromatic LED light colors on growth performance, behavior, immune-physiological responses of gold fish, *Carassius auratus*. *Aquaculture* 538, 736532. doi: 10.1016/j.aquaculture.2021.736532
- Ruchin, A. B. (2020). Effect of illumination on fish and amphibian: development, growth, physiological and biochemical processes. *Rev. Aquaculture* 13, 567–600. doi: 10.1111/raq.12487
- Ruggerone, G. T., Quinn, T. P., McGregor, I. A., and Wilkinson, T. D. (1990). Horizontal and vertical movements of adult steelhead trout, *Oncorhynchus mykiss*, in the Dean and Fisher channels, British Columbia. *Can. J. Fisheries Aquat. Sci.* 47, 1963–1969. doi: 10.1139/f90-221
- Sanchez-Vazquez, F. J., Lopez-Olmeda, J. F., Vera, L. M., Migaud, H., Lopez-Patino, M. A., and Miguez, J. M. (2019). Environmental cycles, melatonin, and circadian control of stress response in fish. *Front. Endocrinol. (Lausanne)* 10. doi: 10.3389/fendo.2019.00279
- Takahashi, A., Kasagi, S., Murakami, N., Furufuji, S., Kikuchi, S., Mizusawa, K., et al. (2016). Chronic effects of light irradiated from LED on the growth performance and endocrine properties of barfin flounder *Verasper moseri*. *Gen. Comp. Endocrinol.* 232, 101–108. doi: 10.1016/j.ygcen.2016.01.008
- Takahashi, A., Kasagi, S., Murakami, N., Furufuji, S., Kikuchi, S., Mizusawa, K., et al. (2018). Effects of different green light intensities on the growth performance and endocrine properties of barfin flounder *Verasper moseri*. *Gen. Comp. Endocrinol.* 257, 203–210. doi: 10.1016/j.ygcen.2017.04.003
- Villamizar, N., Blanco-Vives, B., Migaud, H., Davie, A., Carboni, S., and Sánchez-Vázquez, F. J. (2011). Effects of light during early larval development of some aquacultured teleosts: A review. *Aquaculture* 315, 86–94. doi: 10.1016/j.aquaculture.2010.10.036
- Wang, Z., Xia, C., and Lee, J. (2021). Parallel fish school tracking based on multiple appearance feature detection. *Sensors (Basel)* 21, 3476. doi: 10.3390/s21103476
- Wang, H., Zhang, S., Zhao, S., Lu, J., Wang, Y., Li, D., et al. (2022a). Fast detection of cannibalism behavior of juvenile fish based on deep learning. *Comput. Electron. Agric.* 198, 107033. doi: 10.1016/j.compag.2022.107033
- Wang, H., Zhang, S., Zhao, S., Wang, Q., Li, D., and Zhao, R. (2022b). Real-time detection and tracking of fish abnormal behavior based on improved YOLOV5 and SiamRPN++. *Comput. Electron. Agric.* 192, 106512. doi: 10.1016/j.compag.2021.106512
- Wetzel, R. G. (1983). *WK Kellogg biological station* (Hickory Corners (USA: Michigan State Univ.). Limnology.
- Wu, L., Wang, Y., Li, J., Song, Z., Xu, S., Song, C., et al. (2021). Influence of light spectra on the performance of juvenile turbot (*Scophthalmus maximus*). *Aquaculture* 533, 736191. doi: 10.1016/j.aquaculture.2020.736191
- Xi, Y., Meng, X., Ying-Ping, H., Yi-Hong, Z., Johnson, D. M., and Zhi-Ying, T. (2019). Color-induced changes in oxygen consumption and swimming performance of juvenile bighead carp (*Aristichthys nobilis*). *Fish Physiol. Biochem.* 45, 1771–1777. doi: 10.1007/s10695-019-00671-0
- Xiao, G., Shao, T., Zhu, T., Li, Y., Mao, J., and Cheng, Z. (2016). "Attention region based approach for tracking individuals in a small school of fish for water quality monitoring," in *12th International Conference on Machine Learning and Data Mining (MLDM)*. (New York, NY: Springer), 756–760. doi: 10.1007/978-3-319-41920-6\_57
- Xu, J., Lin, C., Dai, H., Mao, J., Ke, S., Yin, R., et al. (2019). Influence of low-intensity light on phototactic behaviour of *Schizothorax oconnori* Lloyd. *River Res. Appl.* 36, 296–304. doi: 10.1002/rra.3580
- Yamanome, T., Mizusawa, K., Hasegawa, E., and Takahashi, A. (2009). Green light stimulates somatic growth in the barfin flounder *Verasper moseri*. *J. Exp. Zool. A Ecol. Genet. Physiol.* 311, 73–79. doi: 10.1002/jez.497
- Yang, L., Liu, Y., Yu, H., Fang, X., Song, L., Li, D., et al. (2020). Computer vision models in intelligent aquaculture with emphasis on fish detection and behavior analysis: A review. *Arch. Comput. Methods Eng.* 28, 2785–2816. doi: 10.1007/s11831-020-09486-2
- Zhang, X., Bian, Z., Yuan, X., Chen, X., and Lu, C. (2020). A review on the effects of light-emitting diode (LED) light on the nutrients of sprouts and microgreens. *Trends Food Sci. Technol.* 99, 203–216. doi: 10.1016/j.tifs.2020.02.031
- Zhang, H., Li, W., Qi, Y., Liu, H., and Li, Z. (2023). Dynamic fry counting based on multi-object tracking and one-stage detection. *Comput. Electron. Agric.* 209, 107871. doi: 10.1016/j.compag.2023.107871

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